

# Histopathology Reveals Environmental Stress in Dusky Flounder Syacium papillosum of the Yucatan Peninsula Continental Shelf

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# Abstract

The histological changes in the liver, kidney, spleen, and gills of *S. papillosum* from the continental shelf of the Yucatan Peninsula, Gulf of Mexico, and their statistical associations with environmental conditions and pollutants were assessed in 2010, 2011, and 2012. We evaluated the extension and severity of the lesions through a degree of tissue change (DTC) and, with the sum of the number of lesion types within each of their DTC stages, we determined the histological alteration index (HAI). The liver and kidney were the most affected organs, with HAI values > 100. Fish with the most severe damage were observed on the Campeche Bank and the Caribbean Sea, in contrast with those collected from the northern Yucatan continental shelf. The presence of foci cellular alteration and abundant melanomacrophage centers indicated that these flatfishes were chronically exposed to environmental stress factors. Redundancy analyses showed strong associations between HAI values and hydrocarbon and heavy metal concentrations in muscle. The integrated evaluation indicated that flatfish tissues respond actively to exposure to xenobiotics in the environment.

# Introduction

The Gulf of Mexico (GoM) is one of the world's large marine ecosystems (LMEs) (Sherman, 1994). Anthropogenic activities heavily influence this area through historic and chronic inputs of contaminants, such as heavy metals and hydrocarbons. The oil extraction and transportation processes in shallow and deep waters throughout the GoM often result in spills of a variety of magnitudes (Pulster et al., 2020). The release of chemical oil products due to intense maritime transportation represents an important pollution source that can have catastrophic effects on biota as well (Hook et al., 2014; Vidal-Martínez et al., 2019). Additionally, high concentrations of metals, hydrocarbons, and other chemical contaminants from the environment can be accumulated by marine organisms (Bazzi, 2014).

Fish, in contrast to invertebrates, are more sensitive to toxicants and have been considered potential indicators of environmental pollution (Authman et al., 2015). Therefore, the pathological histology of fish can be a helpful tool to evaluate and monitor aquatic ecosystems (Yancheva et al., 2016). Consequently, fish have been proposed as 'effect indicators' and 'accumulation indicators' because of the variety of ways in which they respond to pollutants such as oil spills, heavy metals, domestic sewage, and agricultural and industrial contamination (Botello et al., 2005; Hook et al., 2014; Authman et al., 2015).

Toxic compounds can be absorbed into the organism's organs, tissues, and cells, resulting in the development of different degrees of histological lesions, immune responses, and in many cases, harmful diseases.

Recently, Murawski et al. (2014) and Ali et al. (2014) reported skin lesions in red snapper (*Lutjanus campechanus*) and splenic responses in sea trout individuals exposed to different hydrocarbons, such as polycyclic aromatic hydrocarbons (PAHs), from the deep water horizon (DWH) oil spill. Despite

unexposed trout fish developing sporadic melanomacrophage centers (MMCs), those were significantly greater in number and size in the fish exposed in the GoM.

Histological lesions caused by environmental pollution have been well documented (Khoshnood, 2016; Yancheva et al., 2016). Several liver and gill lesions in marine fish have been associated with exposure to xenobiotic chemicals, such as neoplasms, non-neoplastic proliferative lesions, and specific degenerative/necrotic lesions, which are well established as histological bioindicators of response to pollutant exposure (Adams, 2003). Moreover, the presence of liver tumors in bottom-dwelling fish has been clearly associated with PAHs in sediment. However, it is likely that exposure to other chemicals can contribute either as an initiator or promoter (US EPA et al., 2012). As a response, fish health may reflect a low nutritional status and susceptibility to chemical exposure (Schmitt and Dethloff, 2000). Subsequently, physiological processes can be affected, causing fish death in more severe cases (Vázquez-Gómez et al., 2016). In this sense, the physiological responses in fish fitness (estimated as the condition factor (*K*)) may also be affected. Such responses have been observed as an inverse relationship between the condition factor K and histopathological changes of fishes in environmentally polluted conditions (Javed and Usmani, 2019).

In this context, one of the most significant threats in past years in the Gulf of Mexico (GoM) has been the DWH oil spill, which occurred in April 2010. The release of approximately 134 million gallons (3.19 million barrels) of oil at 1,600 m depth into the GoM required the addition of 1.84 million gallons of dispersant (Soto and Botello, 2013), the ecological impact of which is still under investigation. Although this oil spill was located within 555 km of the Yucatan shelf (YS) (CONAGUA, 2014), there was still concern that its effects could reach the YS. However, the organisms living in the YS could be affected by various other environmental problems not yet evaluated, such as the Cayo Arcas offshore crude oil loading terminal or the heavy maritime traffic of ships that transport petroleum coming from Aruba, Curacao, Maracaibo, and Trinidad and Tobago, all of which use the Yucatan Channel as a frequent route of transportation to North America (Ponce-Vélez and Botello, 2005). In addition, tourism activities from the Caribbean zone and the presence of natural hydrocarbon seeps could also contribute to YS pollution (Botello et al., 2005; Love et al., 2013).

The DWH oil spill established the importance of distinguishing local YS environmental damage sources from their external counterparts. Thus, an environmental baseline of the YS had to be established. This gave us an opportunity to start the GoM monitoring program. The National Institute of Ecology and Climatic Change Center (I.N.E.C.C.) together with other national marine and oceanological institutions implemented a monitoring plan to establish baseline environmental conditions in the Southern Gulf of Mexico after the DWH catastrophe. As a component of such baseline establishment, this paper has the primary purpose of recording flatfish (*Syacium papillosum*) histopathological injuries that may indicate the presence and effects of pollutants on the Yucatan shelf. We hypothesized that severe histological damage would be present in flatfishes of the sampling sites closest to the oil spill-impacted area and that histological alteration index scores (HAI) would be statistically associated with environmental and contamination variables in sediments and organisms.

We used histologic tools because they are relatively inexpensive and easy to measure compared to the use of expensive, complex, and sophisticated laboratory equipment. Furthermore, it is worth mentioning that fish histological biomarkers provide advantages over other biomarkers of environmental stress. Some advantages of histopathological analyses are the identification of specific target organs and tissues during exposure to potentially toxic pollutants. Moreover, histopathology can be used to establish the specific patterns of both acute and chronic deleterious effects of such pollutants on tissues (Au, 2004).

Thus, the study aims were (i) to establish quantitative values through a histological index, (ii) to determine whether the histological changes in *S. papillosum* exhibit differences among subregions from east to west of the YS, and (iii) to determine whether the occurrence of histological damage of *S. papillosum* is associated with natural physicochemical environmental variables, heavy metals, nutrients, and/or hydrocarbons at the seascape level.

# **Materials and Methods**

## Study sites and fish sampling

The study area comprised 87 sampling sites on the continental shelf of the Yucatan Peninsula from Isla Arena, Campeche in the west to Playa del Carmen, Quintana Roo in the east (Fig. 1). The sampling sites covered the northern shelf of the entire peninsula between 15 and 200 m depth.

Three oceanographic cruises (OC) were carried out during consecutive years: Gomex-1 (10–21 September 2010) and Gomex-2 (September 23 to October 3, 2011), during the rainy season, and Gomex-3 (from November 28 to December 2012) during the early north winds season. Once at the sampling sites, the fish were collected by net trawling and identified by the staff of the Ichthyology Laboratory at CINVESTAV Mérida (see Vidal-Martínez et al., 2019 for details).

In each OC, a total of 62 environmental variables from water and sediments were measured (hydrocarbons and heavy metals), including bile PAH metabolites in organisms (Supplementary material Table S1). In addition, nutrients and physicochemical water parameters (e.g., oxygen (mg/l), salinity (psu), and temperature (°C) were measured (Supplementary material Table S1). The physicochemical characteristics and concentrations of hydrocarbons and heavy metals in sediments and organisms were determined at the Marine Geochemistry Laboratory (CINVESTAV-IPN-Mérida) by applying standardized methods (Buchanan, 1984; Brassard, 1997; Farrington et al., 1983). PAH metabolite concentrations in organisms were provided by the Ecotoxicology Laboratory.

## Figure 1

Immediately after being caught, fish morphometric data were taken. The fish were weighed (g) and measured for total length (cm) and then dissected. The length-weight relationship (LWR) of *S. papillosum* was estimated using the equation  $W = \alpha SL^{\beta}$  (Ricker, 1975), where W = total weight (g), SL = standard

length (cm),  $\alpha$  is the ordinate to the origin and  $\beta$  is the intercept, also known as the allometry coefficient. Based on the estimations of LWR, we used the condition factor (K) proposed by Bagenal and Tesch (1978) (K = 100 \* W/SL<sup> $\beta$ </sup>) to assess the physiological condition of the fish, as well as the effect of pollutant exposure on this physiological biomarker (Collier et al., 2013; Snyder et al., 2019).

## Histopathological analyses

Each fish was euthanized by brain puncture to avoid interference of anesthetics that at some point could cause some mechanical damage to tissues, mainly in gills (Boijink et al., 2017). Sections of gills, kidney, liver, and spleen tissues were fixed immediately in a 10% buffered formalin solution to preserve biological tissues and avoid autolysis or putrefaction.

The tissues were processed following the methodology by Humason (1962) and Luna (1968), dehydrated in an automatic processor (Histokinette), and embedded in paraffin (melting point 56°C). Afterward, sections were cut to 5 µm thickness using a microtome and mounted in Entellan® for microscopy. Finally, the slides were stained with hematoxylin and eosin (H&E) for morphological examination under an Olympus BX50 compound microscope. Using a Q imagine digital camera of 5-megapixel resolution, photographs of normal and histopathological tissues were taken for the histological baseline. All the material corresponding to this investigation is available for consultation in the Aquatic Pathology Laboratory Histology Section of CINVESTAV-IPN Mérida.

Histological alterations in tissues and organs were evaluated semi-quantitatively based on two criteria according to Schwaiger et al. (1997) and Simonato et al. (2008): (a) extent of the damage (EOD), where the alterations were classified into three degrees of incidence, and (b) degree of tissue change (DTC) based on the severity of the lesions and the possibility of recovery of tissue and organs (Table 1).

EOD	The extent of the Damage	DTC	Histopathological alterations
Grade 1	There are no focal alterations	Stage I	Changes that do not damage the organ to such an extent that it cannot repair itself if conditions improve.
Grade 2	Focal changes	Stage II	Changes that are more severe and affect the associated organ or tissue function.
Grade 3	Extensive damage	Stage III	Changes that prevent the organ from repairing itself even if conditions improve.

Table 1 Evaluation of the histopathological alterations in relation to extent of the damage (EOD) and the degree

## Table 1

The histological alteration index (HAI) was calculated with the equation used by Silva and Martinez (2007), Simonato et al. (2008) and proposed by Poleksic and Mitrovic-Tutundzic (1994).

$$HAI = 1 \cdot \sum I + 10 \cdot \sum II + 100 \cdot \sum III$$

In this equation, I, II, and III are the lesion stages (DTCs) of the histological alterations. The HAI was calculated for each fish and tissue. The average HAI was ranked into five categories: 0-10, normal organ; 11-20, slightly to the moderately damaged organ; 21-50, moderately to the heavily damaged organ; 51-100, severely damaged organ, and > 100, irreversibly damaged organ. The histopathological alterations were compared among the fish of sampling sites of the same cruise and afterward compared among the fish HAI data on different oceanographic cruises.

#### Data analyses

Normality of both EOD mean values and HAI scores were obtained for each organ for each oceanographic cruise and then assessed with a Shapiro–Wilks normality test. If normality was not reached, a nonparametric ANOVA test (Kruskal–Wallis) with a level of significance of *P*<0.05 was used to determine potential differences in HAI values among subregions (see below).

Nonmetric multidimensional scaling (NMDS) was performed with all the HAI values of the four organs examined for the fish collected at every sampling site, considering the depth and zone as grouping factors. For these analyses, we used Primer 6 software, based on the Bray–Curtis dissimilarity index, to determine the patterns that are formed concerning HAI with respect to depth (10-49 m, 50-100 m, 101-200 m) and zone (Campeche bank, Central Yucatan, Caribbean) factors within the YS and for three Gomex cruises from 2010–2012. Due to their geographical proximity, all of the individual fish obtained were grouped a priori into subregions: Campeche bank in the western YS (sampling sites A2, A3, A4, A5, and B6, B7, B8, B9, B10), Central Yucatan (sampling sites F28, F29, G33, G34, G35, H39, H40, I42, I43, I45, and J47, J48), and the Caribbean (sampling sites L58, L59, M63, O71, O72, O73, O74, O75, P77, P78, P79, and P80). Redundancy analyses (RDA) were performed between dependent variables (HAI values), independent variables (62 variables: physicochemical, nutrient, and contaminant data for sediments and organisms), and K as a covariable (Supplementary Material Table S1). Monte Carlo permutations were used to determine the significance of the canonical axes based on 4,999 permutations. The independent variables were selected by forwarding stepwise analyses, keeping only the variables that explained the greatest amount of variance (10%) and that did not present a variance inflation factor (VIF) higher than four. The significance of the statistics was determined with p < 0.05. RDAs were performed with CANOCO software for Windows 5.0 (ter Braak and Smilauer, 2012).

## **Results**

## Water quality

Physicochemical parameters of water and pollutants in sediments and organisms are presented in the Supplementary Material (Table S1). Analyses results showed that except for salinity, the physicochemical parameters of water (i.e., oxygen, temperature) were variable with respect to the three OCs (Fig. 2).

## Fish sampled

A total of 153 adult flatfish (*Syacium papillosum*) individuals were captured from 15 sites in 2010, 19 sites in 2011, and 17 sites in 2012 out of 87 sampled sites per year. Fish were captured between 15 and 200 meters in depth (Fig. 1(B) and Supplementary Material Table S1). Most fish collected during Gomex-1 came from sites at 100 m depth at the eastern and western boundaries of the Yucatan Shelf (Supplementary Material Table S1: transects A and P, respectively). For Gomex-2 and Gomex-3, it was also possible to collect fish from the F-M transects at 50 to 200 meters depth (Fig. 1). The size and weight of collected fishes varied between 20 and 28 cm of total length and from 105 to 270 g of weight. The total length of fishes in Gomex-2 and Gomex-3 presented significant differences compared to Gomex-1, and the weight of the fishes from Gomex-1 and Gomex-2 had significant differences with respect to Gomex-3 (H = 6.97; df = 2; p = 0.0307 and H = 27.92; df = 2; p = 0.0001 to length and weight, respectively) (Fig. 3; Supplementary Material Table S1).

### Figure 3

## Histopathological findings

The external fish examination did not show clinical signs of diseases, injuries, or malformations. Some of the lesions observed were caused by mechanical damage during trawling. Between eight and 14 types of histopathological alterations were identified in the selected organs (Supplementary material Table S2). The spleen was the organ with the fewest histological lesions, and the liver and kidney were the tissues that showed the most significant histopathologies by number and severity. Prevalence values of the lesions found varied between 3% (gills hypertrophy) and 100% of melanomacrophage centers (MMCs) in the spleen, kidney, and liver. The prevalence values of each lesion were variable among cruises. There was a general reduction of lesion prevalence for the second year with respect to the first, increasing again for the third year (Supplementary material Table S2), although the histopathologies were focal in most cases.

The spleen presented eight lesions; however, the most prevalent alteration in the three cruises was the presence of MMCs (Supplementary material Table S2; Fig. 4). This alteration reached a prevalence higher than 80%, which was maintained for all three OCs.

### Figure 4

Thirteen histological damages were identified in the gills (Supplementary material Table S2). The prevalence of hypertrophy was 90% in samples from Gomex-1, while in Gomex 2 and Gomex-3, the values decreased to 20% and 3%, respectively. The prevalence of hyperplasia for the three cruises presented values > 89%. The prevalence of inflammation reflected in mucus cells and lamellar fusion tended to decrease through each of the OCs from 90–40%, while telangiectasia increased its prevalence through every OC (Supplementary material Table S2). The lamellae presented edema, metaplasia, and parasites

(mainly protozoa), with a prevalence of 35 to 61%. The pathological alterations caused by the presence of parasites were usually limited to slight inflammation around the parasite cyst (Fig. 5).

## Figure 5

In the kidneys, MMCs had a high prevalence in each subsequent oceanographic cruise, with 85, 95, and 100%, respectively. However, cellular atrophy and necrosis were the most severe lesions, and they tended to increase through every OC (Supplementary material Table S2; Fig. 6). The rest of the pathologies observed in this organ varied in percentage on all three cruises.

### Figure 6

In the hepatic tissue, we identified 13 lesions where cellular atrophy, fatty degeneration, MMCs, and necrosis were the most prevalent (>80%) (Supplementary material Table S2, Fig. 7). Additionally, there was an increase in the prevalence of these lesions through the OCs. Injuries such as inflammation, hemorrhage, granulomas, dilation of sinusoids, adenomas, and parasites showed lower prevalence (< 50%). Again, the injuries caused by the presence of parasites were focal (slight inflammation).

### Figure 7

The histological damage like focal inflammation in the kidney, liver, and spleen and focal telangiectasia in the gills was slight on most sampling sites. However, in two sampling sites of Gomex-2 (G34 and H40) and Gomex-3 (G35 and H40) (Fig. 1 (B)), the fishes showed severe histological damage such as fatty degeneration in the liver accompanied by cellular atrophy, necrosis, and hypertrophy in the renal epithelial cells, as well as abundant MMCs in the kidney and spleen.

### Quantification of histological damage in S. papillosum

Table 1 shows the classification given for the extent of the damage (EOD) and histological alteration index (HAI) in organs and tissues for the different lesions observed in the liver, spleen, kidney, and gills of flatfish collected from the Yucatan shelf during the Gomex cruises. The extent of the damage (EOD) of the pathological alterations in the evaluated organs was mainly focal (Grade 2) and generally classified in DTC stage II, except for the neoplastic lesion, severe necrosis, and structural disintegration, which were classified as DTC stage III (Supplementary material Table S2).

The mean values of the EOD and HAI values in the four organs evaluated are shown in Table 2. The highest values of HAI in the liver and kidney are from the coastal zone of the Yucatan shelf, mainly off the coast of Progreso and in front of the upwelling zone, close to the Mexican-Caribbean coast.

#### Table 2

Mean values of the extent of damage (EOD) and histological alteration index (HAI) in the liver, spleen, kidney and gills of the flatfish *Syacium papillosum* from the continental shelf of the Yucatan Peninsula. These data were obtained during oceanographic cruises <u>Gomex 1, 2, and 3 and for comparison, the most recent data from Gomex 6 were included.</u>

		GOMEX 1	GOMEX 2	GOMEX 3		
LIVER	EOD ±SD	1.34 ± 0.27 a	1.29 ± 0.29 a	1.37 ± 0.12 a		
	HAI ±SD	131 ± 93 c	108 ± 59 c	35 ± 40 b		
	Largest value	313	301	142		
	Smallest value	1	1	2		
SPLEEN	EOD ±SD	1.21 ± 0.36 b	1.13 ± 0.37 a	1.28 ± 0.29 b		
	HAI ±SD	2 ± 3 b	3±4a	5±7b		
	Largest value	12	12	32		
	Smallest value	1	1	1		
KIDNEY	EOD ±SD	1.28 ± 0.46 b	1.25 ± 0.28 a	1.41 ± 0.18 b		
	HAI ±SD	97±70 c	58 ± 60 b	19 ± 18 b		
	Largest value	232	224	134		
	Smallest value	1	1	2		
GILLS	EOD ±SD	1.56 ± 0.49 c	1.34 ± 0.22 b	1.27 ± 0.23 a		
	HAI ±SD	30±19 c	7±6b	4±3 a		
	Largest value	72	28	16		
	Smallest value	5	2	1		
*different letters denote significant differences between years of liver, kidney, spleen, and gill samplings ( $P < 0.05$ ).						

#### Table 2

The NMDS analyses showed that, although the HAI values of some fish collected in different zones of the Yucatan shelf were high (> 100), they were not associated with depth or sampling zone (Fig. 8 (A) and Fig. 8 (B)). In contrast, Fig. 8C shows differences among oceanographic cruises (Fig. 8C).

### Figure 8

RDA between HAI and abiotic variables: contaminants, physicochemical parameters, and nutrients

The redundancy analyses (RDA) showed significant and positive associations between the HAI values of the liver, kidney, spleen, and gills (dependent variables) and the independent variables such as polycyclic aromatic hydrocarbons (PAHs) and heavy metals concentrations. Although other variables, such as nutrients, water physicochemical variables, and K as a covariable, were considered, these were not positively associated with the histological index. For Gomex-1, the RDA accounted for 72.6% of the total variance and was highly significant for the first and all canonical axes (F = 6.3473; P value = 0.0226; 4,999 permutations). The HAI of the liver and kidney (L\_HAI and K\_HAI) were positively associated with the total hydrocarbons in muscle (TH (M)) and with aliphatic hydrocarbons in sediments (ALIPH (S)), while the spleen HAI values (S\_HAI) were positively associated with total polycyclic aromatic hydrocarbons in sediment (TPAHs (S)) and with vanadium in muscle (V (M)). For gill HAI (G\_HAI), there was a close association with hydroxypyrene bile metabolites (HPY (BM)) (Fig. 9 (A)). For Gomex-2, the RDA accounted for 72.6% of the total variance and was highly significant for the first and all four canonical axes (F = 9.928; P value = 0.0004; 4,999 permutations). The HAI for kidney and liver (L\_HAI and K\_HAI) were positively associated with PAHs of low molecular weight in sediment (LMWPAHs (S)), lead and barium in muscle (Pb (M) and Ba (M)), while the spleen and gills HAI (S\_HAI and G\_HAI) were positively associated with cadmium in muscle (Cd (M)) (Fig. 9 (B)). For Gomex-3, the RDA accounted for 93.9% of the total variance and was highly significant for all four canonical axes (F = 2.307; P value = 0.0120; 4,999 permutations). The HAI of the liver and spleen (L\_HAI and S\_HAI) were positively associated with PAHs of high molecular weight (HMWPAHs (L)) and total PAHs in sediment (TPAHs (S)). The kidney HAI (K\_HAI) was positively associated with the UCM and nickel in muscle (UCM (M) and Ni (M)), while the gill HAI (G\_HAI) was closely associated with low molecular weight PAHs (PAHs LMW (M)), phenanthrene (PHE (BM)) and iron in muscle (Fe (M)) (Fig. 9 (C)).

Figure 9

# Discussion

Our original hypothesis that severe histological lesions would be present in *S. papillosum* at the closest sites to the DWH oil spill, and therefore at the boundary of the Yucatan shelf, was not fulfilled. The distribution of lesions in *S. papillosum* from the Yucatan shelf showed that these occurred at all sampling sites, being more prevalent and severe in the eastern and western margins of the YS and in the coastal zone in front of Progreso harbor. Regarding the severity of the lesions, the DTC of the collected fish from all Gomex oceanographic cruises showed severe alterations in the liver and kidney (stage III), which can compromise the normal function of the organs. The observed histological damage could be the consequence of chronic exposure to hydrocarbons and heavy metals present in the sediments (Fig. 9). Since there are no oil extraction activities in the YS, a possible explanation for the presence of pollutants causing *S. papillosum* injuries is the proximity of the sampling sites to the oil extraction operations of the Campeche Bank in the western YS. Peters et al. (2021) mention that the eddies formed by the loop current create strong currents that redistribute the water in the surface layers, transporting the pollutants mainly on continental shelves and along the coastal zones of GoM. In addition, according to

Love et al. (2013), the study area matches spatially with the presence of hydrocarbon seeps in the Yucatán Peninsula, which could be related to the most severe histological injuries and highest HAI scores in *S. papillosum*.

With respect to the severity of pathological changes (DTC), most alterations did not hurt the organ structure and function of liver. Fatty degeneration, MMCs, and necrosis increased their prevalence in fishes of Gomex-2 and Gomex-3. The prevalence of lesions such as necrosis in liver tissue is focal and does not represent an irreversible change in the liver tissues as a whole. These lesions have been associated with biochemical disturbances such as degenerative alterations (granular, vacuolar, hydropic, and fatty degeneration) (Monteiro et al., 2005; Rajeshkumar and Munuswamy, 2011). In addition, we occasionally found foci of cellular alteration (FCA), structural disintegration, and coagulative necrosis, which are pathologies that remain and can extend despite the disappearance of the causal factor. These lesions have been reported previously and linked to PAH exposure (Myers et al., 2003). The extension of these lesions can often affect liver functions, such as the detoxification process and fat metabolism, carbohydrates, and proteins (Straif et al., 2005).

The severity of the histological damages in sampling sites near the coastal area of Yucatan (B6) may be due to strong freshwater discharge from the continent during the rainy season, which contains different compounds, including nutrients such as HCO3, SO4, Cl, Ca, Mg, Na, K, NO2, NO3, and NH4 (Delgado et al., 2010; Herrera-Silveira and Morales-Ojeda, 2010). Additionally, there is evidence of PAHs of low and high molecular weight in sediments all along the coastal lagoons of Yucatán, presumably carried from the continent to the coastal region by subterranean currents (Gold-Bouchot et al., 2014). Consequently, fish are exposed to all these compounds in the coastal zone, which in turn could produce the histopathologies found in *S. papillosum*. On the other hand, histopathological damages observed in more deepest sampling stations (G34, G35, and H40) could be related to additional factors such as pollutants dispersion and frequent maritime traffic (including oil transportation), which often results in frequent small scale spill, that is one of the main stressors in approximately 50% of the YS (Ocaña et al., 2019).

Furthermore, the RDA analyses suggested associations or causal relationships between the concentrations of hydrocarbons, heavy metals, and the HAI scores of *S. papillosum*. Despite the positive association between histopathological lesions and the presence of pollutants shown by our analyses, we should not jump to conclusions regarding the source of such pollutants being from the oil industry only. As Vidal-Martínez et al. (2019) point out, the concomitant effect of other environmental variables and other water sources of the discharge into the YS should be considered. To develop a histological lesion, high pollutant concentrations are not always necessary, and low doses are usually sufficient to trigger their development (Améndola-Pimenta et al., 2020), which indicates localized environmental conditions could lead to similar fish histopathologies elsewhere on the GoM.

Although no solid evidence was found of DWH spill-originate pollutants reaching the YS, the decreasing tendency in prevalence and severity of lesions found in *S. papillosum* throughout the Gomex cruises is similar to the pattern observed by Murawski et al. (2014). These authors found a similar pattern while

sampling considerably nearer to the DWH oil spill. The incidence of skin lesions reported in their 2011 samplings had decreased by 53% when they sampled again by 2012, with the severity of lesions also declining. A likely reason for the observed decreasing prevalence and severity of lesions in Murawski's research and our own could be the death of the more severely affected fish. It is not yet reasonable to conclude that the histopathologies reported in this paper have been necessarily caused by pollutants originated from the DWH or that the observed decrease in lesion prevalence observed by Murawski and ourselves can be indeed attributed to the death of the most severely affected fish. More precise conclusions on these matters will require additional long-term research.

# Conclusions

The prevalence and high HAI values that we found in the examined organs of flatfish that inhabit the Yucatan platform are evidence that there are specific areas with poor environmental quality. The histopathologies described in this work also indicate that these fish are chronically exposed to stress sources.

These findings thereby constitute a relevant set of baseline data to be considered for monitoring the biological effects of pollutants in the YS. The Yucatan shelf has several anthropogenic and natural stress factors, thusly requiring continuous biological and environmental monitoring is necessary to better understand their effect on the YS communities. Equally important will be to experimentally test the potential relationship between contaminants such as hydrocarbons and heavy metals with histological damage in *S. papillosum*.

# **Abbreviations**

L\_HAI, Liver-histological alteration index; S\_HAI, spleen-histological alteration index; K\_HAI, kidneyhistological alteration index; G\_HAI, gill-histological alteration index; TH (M), total hydrocarbons (muscle); ALIPH (S), aliphatic hydrocarbons (sediment); TPAHs (S), total polycyclic aromatic hydrocarbonssediment; BaP(BM), benzo a pyrene (bile metabolites); HPY (BM), hidroxypyrene (bile metabolites); NAPH (BM), naphthalene (bile metabolites); HMWPAHs (M), Polycyclic aromatic hydrocarbons-high molecular weight (muscle); PHE(BM), phenanthrene (bile metabolites); Ni (M), nickel (muscle); Ba(M), barium (muscle); Cd (M), cadmium (muscle); Pb (M), lead (muscle); V (M), vanadium (muscle); Fe (M), iron (muscle); Pyrene(M), pyrene (muscle); Perylene (L), Perylene (liver); LMWPAHs (S), Polycyclic aromatic hydrocarbons-low molecular weight (sediment); NH4, ammonia; UCM (M), unresolved complex mixture (muscle); LMWPAHs (M), Polycyclic aromatic hydrocarbons-low molecular weight (muscle); HMWPAHs (L), Polycyclic aromatic hydrocarbons-high molecular weight (liver)

# Declarations

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### **Author Contributions**

Gold-Bouchot and Agurirre-Macedo designed the sampling network; E.D. Couoh-Puga, M.L. Aguirre-Macedo, M.C. Chávez-Sánchez: Conceptualization of the histological study; E.D. Couoh-Puga, M.C. Chávez-Sánchez, O.A. Centeno-Chalé: Data curation; E.D. Couoh-Puga, O.A. Centeno-Chalé: Formal analysis; V.M. Vidal-Martínez, M.L. Agurirre-Macedo: Funding acquisition; E.D. Couoh-Puga, V. Vidal-Martínez, M.L. Agurirre-Macedo, M.C. Chávez-Sánchez: Methodology; V. Vidal-Martínez, M.L. Agurirre-Macedo: Project administration; V.M. Vidal-Martínez, M.L. Agurirre-Macedo: Resources; E.D. Couoh-Puga, O.A. Centeno-Chalé: Software; E.D. Couoh-Puga: Roles/Writing - original draft; E.D. Couoh-Puga, M.L. Agurirre-Macedo, M.C. Chávez Sánchez: Writing - review & editing: V.M. Vidal-Martínez, M.L. Agurirre-Macedo, M.C. Chávez Sánchez: Writing - review & editing: V.M. Vidal-Martínez, M.L. Agurirre-Macedo, M.C. Chávez Sánchez: Writing - review & editing: V.M. Vidal-Martínez, M.L. Agurirre-

### Conflict of interest declaration

The authors report no potential conflicts of interest.

## Statement of data availability

The authors declare that the data supporting the findings of this study are available within the paper and its Supplementary Information files. Histological slides data for the tissues changes assessment reported in this article have been deposited at Aquatic Pathology Laboratory of CINVESTAV-IPN Mérida unit, with the names of GOMEX 2010, GOMEX 2011 and GOMEX 2012. All other relevant data generated and analyzed during this study, which include micro photography of histological sections and computational data, are included in this article and its supplementary information. Source data are provided with this paper. Should any raw data files be needed in another format they are available from the corresponding author upon reasonable request. Source data are provided with this paper.

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## References

- Adams, S.M. (2003). Establishing Causality between Environmental Stressors and Effects on Aquatic Ecosystems. Human and ecological risk assessment: HERA 9:1, 17–35. https://doi.org/10.1080/713609850
- 2. Ali, A.O., Hohn, C., Allen, P.J., Ford, L., Dail, M.B., Pruett, S., Petrie-Hanson, L. (2014). The effects of oil exposure on peripheral blood leukocytes and splenic melano-macrophage centers of Gulf of Mexico fishes. Marine pollution bulletin 79: 87–93. https://doi.org/10.1016/j.marpolbul.2013.12.036
- Améndola-Pimenta, M., Cerqueda-García, D., Zamora- Briseño, J., Couoh-Puga, D., Montero-Muñoz, J., et al. (2020). Toxicity evaluation and microbiota response of the lined sole *Achirus lineatus* (Chordata: Achiridae) exposed to the light petroleum water accommodated fraction (WAF). Journal of toxicology and environmental health. Part A. DOI:10.1080/15287394.2020.1758861
- Authman M.M.N., Zaki, M.S., Khallaf, E.A., Abbas, H.H. (2015). Use of Fish as Bioindicator of the Effects of Heavy Metals Pollution. Journal of aquaculture research & development, 06(04). doi:10.4172/2155-9546.1000328
- 5. Au, D.W.T. (2004). The application of histo-cytopathological biomarkers in marine pollution monitoring: a review. Marine Pollution Bulletin, 48(9–10), 817–834. doi:10.1016/j.marpolbul.2004.02.032
- 6. Bagenal, T., and Tesch, F.W. (1978). Age and growth. In: Bagenal T (Ed) Methods for assessment of fish production in freshwater waters. Blackwell Scientific Publication, pp. 101–136.
- Bazzi, A.O. (2014). Heavy metals in seawater, sediments and marine organisms in the Gulf of Chabahar, Oman Sea. Journal of Oceanography and Marine Science, 5(3), 20–29. doi:10.5897/joms2014.0110
- Boijink, C.L., Maciel, P.O., Tavares-Dias, M., Iwashita, M.K.P., Morais, M.S., Hide, D.M.V., et al. (2017). Anesthesia by sprinkling method in the gills of tambaqui *Colossoma macropomum* does not influence intensity and morphology of monogeneans. Brazilian journal of biology = Revista brasleira de biologia., 77: 1–5. https://doi.org/10.1590/1519-6984.15915
- 9. Botello, A.V., Rendón von Osten, J., Gold-Bouchot, G., Agraz-Hernández, C.M. (2005). Golfo de México, Contaminación e Impacto Ambiental: Diagnóstico y Tendencias. 2da Edición. Univ Autón de Campeche, Univ. Nal. Autón. de México, Instituto Nacional de Ecología, 696 p
- Brassard, P. (1997). Measurement of Eh and pH in aquatic sediments, in: Mudroch, Azcue A, Mudroch PJM (eds) Mannual of Physico-chemical Analysis of Aquatic sediment, CRC Press, Inc. Lewis Publishers, pp 47–69
- 11. Buchanan, J.B. (1984). Sediment analysis. In: Holme NA, McIntyre AD (ed) Methods for the Study of Marine Benthos, Blackwell Scientific, London, pp. 41–65
- Collier, T.K., Anulacion, B.F., Arkoosh, M.R., Dietrich, J.P., Incardona, J.P., Johnson, L.L., Myers, M.S. (2013) Effects on Fish of Polycyclic Aromatic Hydrocarbons (PAHs) and Naphthenic Acid Exposures. Fish Physiol - Organic Chemical Toxicology of Fishes-, 33, 195–255. doi:10.1016/b978-0-12-398254-4.00004-2

- Comisión Nacional del Agua. (2014). Monitoreo ambiental y evaluación de la calidad del agua en las costas del Golfo de México y mar caribe. Ciudad de México. (CGACC\_2014\_Monitoreo\_Ambiental\_CONAGUA), pp 1–83
- Delgado, C., Pacheco, J., Cabrera, A., Batllori, E., Orellana, R., Bautista, F. (2010). Quality of groundwater for irrigation in tropical karst environment: the case of Yucatán, Mexico. Agricultural Water Manage. 97 (10), 1423–1433. https://doi.org/10.1016/j. agwat.2010.04.006.
- Farrington, J.W., Goldberg, E.D., Risebrough, R.W., Martin, J.H., Bowen, V.T. (1983). U.S. 'Mussel Watch' 1976–1978: An overview of the trace-metal, DDE, PCB, hydrocarbon, and artificial radionuclide data. Environmental science & technology, *17*:8 490–496
- 16. Gold-Bouchot, G., Ceja-Moreno, V., Chan-Cocom, E., Zapata-Perez, O. (2014). Petroleum hydrocarbons, fluorescent aromatic compounds in fish bile and organochlorine pesticides from areas surrounding the spill of the Kab121 well, in the Southern Gulf of Mexico: a case study. Journal of environmental biology, 35:1 147–156
- Herrera-Silveira, J.A., Morales-Ojeda, S.M. (2010). Subtropical Karstic Coastal Lagoon Assessment, Southeast Mexico, the Yucatan Peninsula Case, in: Kennish, M.J., Paerl, H.W. (Eds.), Coastal Lagoons: Critical Habitats of Environmental Change. CRC Marine Science Series, CRC press, Boca Raton, FL, pp. 307–333.
- Hook, S.E., Gallagher, E.P., Batley, G.E. (2014). The role of biomarkers in the assessment of aquatic ecosystem health. Integrated environmental assessment and management, 10:3 327–341. https://doi.org/10.1002/ieam.1530
- 19. Humason, G.L. (1962). Animal tissue techniques. Freeman, San Francisco, p 468. https://doi.org/10.5962/bhl.title.5890
- Javed, M., and Usmani, N. (2019). An Overview of the Adverse Effects of Heavy Metal Contamination on Fish Health. Proceedings of the National Academy of Sciences, India - Section B: Biological Sciences 89((2)):389–403. https://doi.org/10.1007/s40011-017-0875-7
- 21. Khoshnood, Z. (2016). Using Biomarkers in Ecotoxicology: What and Why? Focus on Sciences, 2, 2, 1–2.
- 22. Love, M., Baldera, A., Yeung, C., Robbins, C. (2013). The Gulf of Mexico Ecosystem: A Coastal and Marine Atlas. New Orleans, LA: Ocean Conservancy, Gulf Restoration Center, 162 p
- 23. Luna, L.G. (1968). Manual of histologic staining methods of the Armed Forces Institute of Pathology, 3rd edition. McGraw-Hill, New York, 258 p.
- 24. Monteiro, S.M., Mancera, J.M., Fontaínhas-Fernandes, A., Sousa, M. (2005). Copper induced alterations of biochemical parameters in the gill and plasma of *Oreochromis niloticus*. Comparative biochemistry and physiology. Toxicology & pharmacology: CBP, 141:375–383
- 25. Murawski, S.A., Hogarth, W.T., Peebles, E.B., Barbeiri, L. (2014). Prevalence of External Skin Lesions and Polycyclic Aromatic Hydrocarbon Concentrations in Gulf of Mexico Fishes, PostDeepwater Horizon. Transactions of the American Fisheries Society, 143:4 1084–1097. https://doi.org/10.1080/00028487.2014.911205

- 26. Myers, M.S., Johnson, L.L., Collier, T.K. (2003). Establishing the Causal Relationship between Polycyclic Aromatic Hydrocarbon (PAH) Exposure and Hepatic Neoplasms and Neoplasia-Related Liver Lesions in English Sole (*Pleuronectes vetulus*). Human and ecological risk assessment: HERA, 9:1,67–94. http://dx.doi.org/10.1080/713609853
- Ocaña, F.A., Pech, D., Simões, N., Hernández-Ávila, I. (2019). Spatial assessment of the vulnerability of benthic communities to multiple stressors in the Yucatan Continental Shelf, Gulf of Mexico. Ocean & Coastal Management, https://doi.org/10.1016/j.ocecoaman. 2019.104900
- 28. Peters, E. M., Zaragoza Álvarez, R. A., Herzka, Sharon Z., y Herguera García, J. C. (Eds.) (2021). Introducción. En S. Z. Herzka, R. A. Zaragoza Álvarez, E. M. Peters y G. Hernández Cárdenas. (Coord. Gral.). Atlas de línea base ambiental del golfo de México, México: Consorcio de Investigación del Golfo de México."
- 29. Poleksic, V., and Mitrovic-Tutundzic, V. (1994). Fish gills as a monitor of sublethal and chronic effects of pollution, in: Müller R, Lloyd R (eds) Sublethal and chronic effects of pollutants on freshwater fish. Cambridge Univ. Press, pp 339–352
- 30. Ponce-Vélez, G., Botello, A.V. (2005). Niveles de hidrocarburos en el Golfo de México, in: Botello AV, Rendón-von Osten J, Gold-Bouchot G, Agraz-Hernández C (eds) Golfo de México Contaminación e Impacto Ambiental: Diagnóstico y Tendencias, 2da Edición. Univ Autón de Campeche, Univ Nal Autón de México, Instituto Nacional de Ecología, pp 269–298.
- 31. Pulster, E.L., Gracia, A., Snyder, S.M., Deak, K., Fogleson, S., Murawski, S.A. (2020). Chronic sublethal effects observed in wild caught fish following two major oil spills in the Gulf of Mexico: *Deepwater Horizon* and Ixtoc 1 (Chap. 24). In: Murawski SA, Ainsworth C, Gilbert S, Hollander D,
- 32. Rajeshkumar, S., and Munuswamy, N. (2011). Impact of metals on histopathology and expression of HSP 70 in different tissues of Milk fish (*Chanos chanos*) of Kaattuppalli Island, South East Coast, India. Chemosphere, 83:4 415–421. 10.1016/j.chemosphere.2010.12.086
- 33. Ricker, W.E. (1975). Computation and interpretation of biological statistics of fish populations.
- 34. Bulletin of the Fisheries Research Board of Canada, 191, 1–382.
- 35. Schmitt, C.J., and Dethloff, G.M. (2000), Biomonitoring of Environmental Status and Trends (BEST) Program: selected methods for monitoring chemical contaminants and their effects in aquatic ecosystems. U.S. Geological Survey, Biological Resources Division, Columbia, (MO): Information and Technology Report USGS/BRD-2000–0005. 81 pp.
- 36. Schwaiger, J., Wanke, R., Adam, S., Pawert, M., Hönnen, W., Triebskorn, R. (1997). The use of histopathological indicators to evaluate contaminant-related stress in fish. Journal of Aquatic Ecosystem Stress and Recovery, 6: 75–86. https://doi.org/10.1023/A:1008212000208
- 37. Sherman, K. (1994), Sustainability, biomass yields, and health of coastal ecosystems: An ecology perspective. Marine ecology progress series, 112: 277–301.
- 38. Simonato, J.D., Guedes, C.L.B., Martinez, C.B.R. (2008). Biochemical, physiological, and histological changes in the neotropical fish *Prochilodus lineatus* exposed to diesel oil. Ecotoxicology and environmental safety, 69:1 112–120. https://doi.org/10.1016/j.ecoenv.2007.01.012

- 39. Silva AG, Martinez BR (2007) Morphological changes in the kidney of a fish living in an urban stream. Environmental toxicology and pharmacology, 23:2 185–192. https://doi.org/10.1016/j.etap.2006.08.009
- 40. Snyder, S.M., Pulster, E.L., Murawski, S.A. (2019). Associations between chronic exposure to polycyclic aromatic hydrocarbons and health indices in Gulf of Mexico Tilefish (*Lopholatilus chamaeleonticeps*) post Deepwater Horizon. Environmental Toxicology and Chemistry, 1–13 https://doi.org/10.1002/etc.4583.
- 41. Soto, L.A., and Botello, A.V. (2013). Legal Issues and Scientific Constraints in the Environmental Assessment of the Deepwater Horizon Oil Spill in Mexico Exclusive Economic Zone (EEZ) in the Gulf of Mexico. International Journal of Geosciences, 4:5 39–45. http://dx.doi.org/10.4236/ijg.2013.45B007
- 42. Straif, K., Baan, R., Grosse, Y., Secretan, B., El Ghissassi, F., Cogliano, V. (2005). Carcinogenicity of polycyclic aromatic hydrocarbons. The Lancet Oncology, 6: 931
- 43. Ter Braak, C.J.F., Šmilauer, P. (2012). Canoco Reference Manual and User's Guide:Software for Ordination, Version 5 0. Microcomputer Power, Ithaca, USA.
- 44. US Environmental Protection Agency., US Geological Survey., US Fish and Wildlife Service. (2012). Toxic Contaminants in the Chesapeake Bay and its Watershed: Extent and Severity of Occurrence and Potential Biological Effects, USEPA Chesapeake Bay Program Office, Annapolis, MD, 175 p
- 45. Vázquez-Gómez, G., Rubio-Lightbourn, J., Espinosa-Aguirre, J.J. (2016). Mecanismos de acción del receptor de hidrocarburos de arilos en el metabolismo del benzo [a] pireno y el desarrollo de tumores. Revista especializada en ciencias químico-biológicas, TIP19 (1):54–67. https://dx.doi.org/10.1016/j.recqb.2016.02.006
- 46. Vidal-Martínez, V.M., Velázquez-Abunader, I., Centeno-Chalé, O.A., May-Tec, A.L., Soler–Jiménez, L.C., Pech, D., et al. (2019). Metazoan parasite infracommunities of the dusky flounder *Syacium papillosum* as bioindicators of environmental conditions in the continental shelf of the Yucatan Peninsula, Mexico. Parasites & Vectors, 12: 277. https://doi.org/10.1186/s13071-019-3524-6
- 47. Yancheva, V., Velcheva, I., Stoyanova, S., Georgieva, E. (2016). Histological biomarkers in fish as a tool in ecological risk assessment and monitoring programs: A review. Applied ecology and environmental research, 14:1 47–75. http://dx.doi.org/10.15666/aeer/1401\_047075





Study area. A) Sampling transect network for the Gomex oceanographic cruises (GMX) over the Yucatan continental shelf during 2010–2012, A-E transects of the region near the Campeche bank (Campeche bank zone); F-K transects of the central zone of the Yucatán shelf (Yucatán zone), L-Q transects of the Mexican Caribbean (Yucatan-Quintana Roo zone). B) Sampling sites where fish were collected during the Gomex oceanographic cruises 2010–2012



## Figure 2

Statistical differences in the oxygen and temperature during three Gomex (GMX) OCs. A) mean oxygen  $\pm$  ED., B) mean temperature  $\pm$  E.D., and C) mean salinity  $\pm$  E.D. Different letters denote significant differences (p < 0.05). We used the statistical test Kruskal–Wallis nonparametric one-way ANOVA.



Statistical differences in the length and weight of all flatfish for three Gomex (GMX) oceanographic cruises. A) Mean total length  $\pm$  E.D., B) Total weight average  $\pm$  E.D. Different letters denote significant differences (p< 0.05). We used the Kruskal–Wallis nonparametric one-way ANOVA test.



### Figure 4

Histological damage in the spleen during the Gomex 1, 2, and 3 cruises, a) normal tissue with focal melanomacrophage centers ( $\uparrow$ ). H & E staining at 10x. Images b) and c) splenic tissue with abundant melanomacrophage centers ( $\uparrow$ ). H & E staining at 10x and 4x



Histological damage in gills of flatfish *Syacium papillosum* during oceanographic cruises Gomex 1, 2, and 3 in a); healthy gill tissue, H & E stain 4x; b); increase in mucus at the branchial apex (metaplasia) ( $\uparrow$ ), H & E stain 10x, and c); destruction of the epithelial lamella and epithelial lifting in gills with severe lamellar edema ( $\uparrow$ ), hypertrophy (\*), blood congestion ( $\Delta$ ), and pyknotic cells (o). H & E staining 40x



### Figure 6

Histological damage observed in the kidney during the Gomex 1, 2, and 3 cruises. a) Renal tissue with severe dilation in the glomerular vessels ( $\uparrow$ ), 40x H & E staining; b) renal tissue with abundant MMCs ( $\nearrow$ ), glomerular atrophy ( $\Delta$ ) and dilation of glomerular blood vessels ( $\uparrow$ ),40x H & E staining, and c) renal tissue with abundant MMCs, H & E staining 10x



Histological damage observed in the liver during the Gomex 1, 2, and 3 cruises, a) normal liver tissue, with slight fatty degeneration ( $\Box$ ), H & E staining 4x; b) congestion of hepatic sinusoids ( $\Box$ ), MMCs (o), severe fatty degeneration (\*), H & E staining 40x, and c) liver tissue with abundant MMCs ( $\uparrow$ ), H & E staining 40 x



Nonmetric multidimensional scaling (NMDS) of the histological alteration index (HAI) of the dusky flounder *Syacium papillosum*. A) NMDS analysis between HAI values with the depth classification of the Gomex 1, 2, and 3 cruises. B) NMDS analysis between HAI values with zone classification of the Gomex 1, 2, and 3 cruises. C) NMDS analysis between HAI values and years of Gomex cruises within the Yucatan platform.



## Figure 9

A-C Redundancy analyses (RDA) of the environmental variables and histological alteration index HAI of *Syacium papillosum* of the three Gomex cruises (1-2010, 2-2011, and 3-2012). (A) For Gomex-1, the RDA accounted for 72.6% of the total variance and was highly significant for the first and all canonical axes (F= 6.3473; *P* value= 0.0226; 4,999 permutations). (B) For Gomex-2, the RDA accounted for 72.6% of the total variance and was highly significant for the first and all four canonical axes (F= 9.928; *P* value= 0.0004; 4,999 permutations). (C) For Gomex-3, the RDA accounted for 93.9% of the total variance and was highly significant for all four canonical axes (F= 2.307; *P* value= 0.0120; 4,999 permutations).

# **Supplementary Files**

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