
Anvil Range Pit Lakes 2009 Evaluation of In-Situ Treatment

Report Prepared for
Yukon Government

On behalf of
Faro Mine Closure Planning Office

Report Prepared by



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SRK Project Number SRKNAC

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Executive Summary

Investigations undertaken in 2004 concluded that biological treatment should be continued in the Grum Pit and tested in the Faro Pit. The 2005 investigations in the Grum Pit concluded that biological treatment should be continued there. However, the 2005 fertilization program in the Faro Pit interfered with the performance of the Faro Mill water treatment system, and biological treatment was discontinued. Since then water quality and physical limnological monitoring were continued in the Faro and Vangorda pits, whilst fertilization and water quality monitoring were continued for the Grum Pit lake. In addition, since 2006 the rate of fertilization of the Grum Pit lake was reduced to ensure that nutrient concentrations are essentially depleted at the end of the season. This report provides presents and discusses the results from the 2008 biological treatment program in the Grum Pit and the monitoring results from the Faro and Vangorda Pit lakes.

The monitoring results indicated that nutrient concentrations of phosphate species have essentially been depleted from the surface water in the Faro Pit. Mass balance calculations indicate that, whilst there is some annual variability, on average the annual zinc loading to the pit lake is about 48 tonnes per year, of which about 24 % (11.6 tonnes) is sourced from the Zone 2 Pit water transfer. The balance of the loading represents the combined loadings from the wall rocks and the waste rock seepage that enters the pit lake. Monitoring results for seepage from the wall rocks suggest that the majority of this loading is likely from the waste rock dumps.

The results indicate that the Grum Pit again responded well to fertilization and that good algal growth was again achieved in 2009. Surface water mass balance calculations for 2004 through 2009 suggest that zinc loadings to the Grum Pit are on the order of 3000 to 4000 kg per annum. The loading appears to be associated with predominantly with wall rock runoff, and to a lesser extent with solute release from wall rock as the water level rises. Net zinc removal has matched the zinc loading to the pit lake during the 2009 summer period. Contrary to previous years, the concentrations in the pit lake did not increase over the 2008/09 winter season indicating a net removal of zinc occurred that may be a result of sulphate reduction occurring under the ice. Due to the removals, the volume weighted zinc concentration has decreased from in excess of 10 mg/L in 2005 to about 4.8 mg/L.

The leach extraction tests were used to verify the potential annual zinc loadings to the pit lake and then used to estimate steady state concentrations in outflow from the pit lake when it spills naturally at an elevation of about 1230 m ASL. The results suggest that the steady state zinc concentration in the steady state outflow (without any treatment) could still remain at about 1 mg/L. This is primarily due to the sulphidic wall rock remaining above the lake level. By raising the lake elevation by about 20 m, the zinc concentrations could potentially be reduced to about 0.1 mg/L in the long term.

Monitoring in the Vangorda Pit indicated that while the pH increased marginal in the near surface layer water, the layer acidified rapidly as a result of the acidity loads from the wall rocks and a in-pit waste rock dumps.

It is recommended that monitoring of both the Faro and Vangorda Pit be continued in 2010. Mass balance calculations should be carried forward to verify the current estimates of metal loadings to each pit. Furthermore, based on the zinc removal rates observed for the Grum Pit and the level of control established in 2008, it is recommended that the fertilization program be continued in 2010.

* * *

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- Appendix A In-situ Monitoring Results
- Appendix B Pit Lake Water Quality Monitoring Results

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1 Introduction

In-situ treatment of the Grum Pit lake has been ongoing since 2004, with the annual fertilization and monitoring carried out by Laberge Environmental Services. The findings from the previous programs (2004 to 2008) are reported elsewhere (SRK 2005, SRK 2006, SRK 2007, SRK 2008 and SRK 2009). The results to date indicate that the Grum Pit water responds rapidly to fertilization with excellent phytoplankton growth occurring within two weeks of the first fertilization program of each summer season. Zinc removal by phytoplankton has been demonstrated to the extent that total zinc concentrations in the near surface water annually decreases to or below 0.3 mg/L. However, annual turnover has resulted in the mixing of the clean surface layer with water of higher concentrations at depth which, together with ongoing sources, have annually replenished the zinc concentrations in the surface layer. Nevertheless, monitoring to date indicates that the zinc concentrations in the lake are decreasing annually indicating that the net removal of zinc is exceeding the rate of loading to the pit lake.

During 2009 the Grum Pit lake fertilisation program was continued. In conjunction with the annual fertilisation program for the Grum Pit lake, the Faro Pit and Vungorda Pit lakes have been monitored to assess potential changes in water quality over time. Supplemental to the 2009 monitoring Grum Pit lake program, samples of pit wall talus were obtained during 2009 to assess the potential for solute release from accumulated broken rock on the pit walls and benches of the Grum Pit to the rising water level. A bathymetry of the Grum Pit lake was also undertaken by Laberge during September 2009. As discussed later in this report, the survey indicated that the actual volume of water contained in the pit lake is significantly lower than had been inferred from the capacity curve used for load and volume calculations. As result, all of the historical calculations have been re-run to accommodate this change and are presented herein.

This report summarises and presents the outcomes from the 2009 monitoring programs together with the updated performance estimates based on the revised Grum Pit lake volume.

2 Fertilization and Monitoring Program

2.1 Grum Pit

Laberge Environmental Services again completed the pit lake fertilization and monitoring programs commencing 16 June 2009, as summarised in Table 2.1. As in previous years, fertilization of the Grum Pit entailed distribution by boat of about 1.5 drums (320 L) of the fertilizer mix to the surface of the pit lake during each fertilization event. The fertilizer addition was repeated until 5 August 2009.

Samples from the water column were collected at six depths from surface (1, 3, 5, 15, 30 and 50m – note samples were collected at 40 m in previous years) on 14 July, 11 August and 8 September 2009. Water samples were submitted to Maxxam Analytics Inc. for detailed chemical analysis.

Concurrent with each water sampling event, CTD casts using a Seabird were made of the water column and in-situ measurements were completed at each water sample depth. Field parameters included measurement of pH, temperature, TDS conductivity and dissolved oxygen.

The fertilization and monitoring programs were consistent with those of 2006 to 2008.

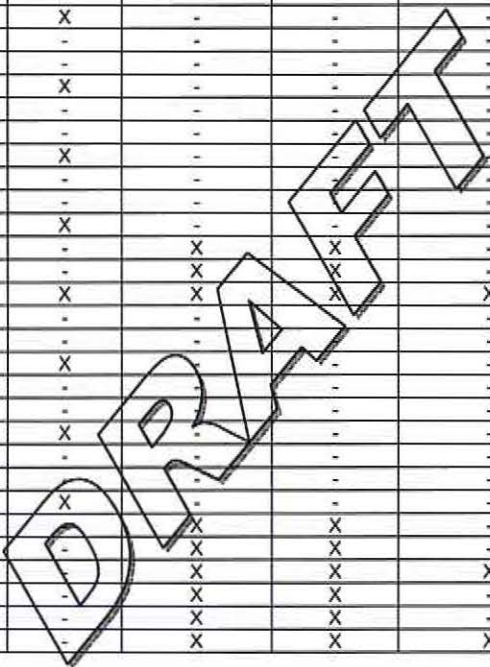
2.2 Vangorda and Faro Pits

Whilst neither the Vangorda nor the Faro Pit lake was fertilized during the 2009 season, water samples were collected from six depths in Vangorda Pit (1, 3, 5, 15, 30 and 40m) and seven depths in Faro pit (1, 3, 5, 15, 30, 60 and 80m) and submitted for detailed analysis as per the schedule shown in Table 2.1. As with the Grum Pit lake monitoring program, CTD casts were made concurrently with the in-situ measurements conducted at each of the sample depths. Field parameters included measurement of pH, temperature, TDS conductivity and dissolved oxygen.

All water samples were submitted to Maxxam Analytics Inc. for analysis. Complete field parameter results are provided in Appendix A and analytical results for the water quality samples are provided in Appendix B.

Table 2.1 Summary of Fertilization and Monitoring Program for 2008

Date	Pit Lake	Fertilize	Field Parameters	Water Quality	Chlorophyll 'a'
30 Apr 09 (under ice)	Vangorda	-	X	-	-
	Faro	-	X	-	-
	Grum	-	X	-	-
16-18 Jun 09	Vangorda	-	X	-	-
	Faro	-	X	-	-
	Grum	X	X	-	X
16 Jun 09	Vangorda	-	-	-	-
	Faro	-	-	-	-
	Grum	X	-	-	-
24 Jun 09	Vangorda	-	-	-	-
	Faro	-	-	-	-
	Grum	X	-	-	-
1 Jul 09	Vangorda	-	-	-	-
	Faro	-	-	-	-
	Grum	X	-	-	-
8 Jul 09	Vangorda	-	-	-	-
	Faro	-	-	-	-
	Grum	X	-	-	-
14-16 Jul 09	Vangorda	-	X	X	-
	Faro	-	X	X	-
	Grum	X	X	X	X
22 Jul 09	Vangorda	-	-	-	-
	Faro	-	-	-	-
	Grum	X	-	-	-
29 Jul 09	Vangorda	-	-	-	-
	Faro	-	-	-	-
	Grum	X	-	-	-
5 Aug 09	Vangorda	-	-	-	-
	Faro	-	-	-	-
	Grum	X	-	-	-
11-13 Aug 09	Vangorda	-	X	X	-
	Faro	-	X	X	-
	Grum	-	X	X	X
8-10 Sep 09	Vangorda	-	X	X	-
	Faro	-	X	X	-
	Grum	-	X	X	X



3 Results and Discussion

3.1 Faro Pit

3.1.1 In Situ Monitoring

As in past years, the in-situ monitoring results (see Figure 3.1) confirmed the annual development of a thermocline in the Faro Pit at a depth of about 5 to 6 m below the water surface. The results further suggest that a partial mixing event had occurred at about mid June, with the temperatures for 17 June 2009 indicating a middle layer at about 7°C, whilst the surface water at about 11 °C. Thereafter a 'normal' temperature profile was observed, with a maximum recorded temperature of about 14.5 °C measured in August. (Note that this may not necessarily represent the maximum temperature reached during the season since monitoring was only on a monthly basis.)

The EC results indicated that a chemocline continues to exist at a depth of between 15 and 30 m below surface. It is important to note that since the samples are taken at a fixed depth from the surface of the pit lake, the change in the pit lake surface elevation could account for some of the changes observed due to the fluctuation in surface elevation caused by pumping, treating and discharging annually. The absolute elevations of the samples therefore fluctuate proportionally with the fluctuation in pit lake elevation, whereas the absolute elevation of the chemocline would remain constant over time. Nevertheless, the EC results indicate relatively fresh inflows occur during the early season as indicated by the lower EC values for June 2009 compared to those observed for April 2009 (note that ice melt would have contributed to this). Later in the season the upper layer salinity increases again as the EC values approaches those for April 2009.

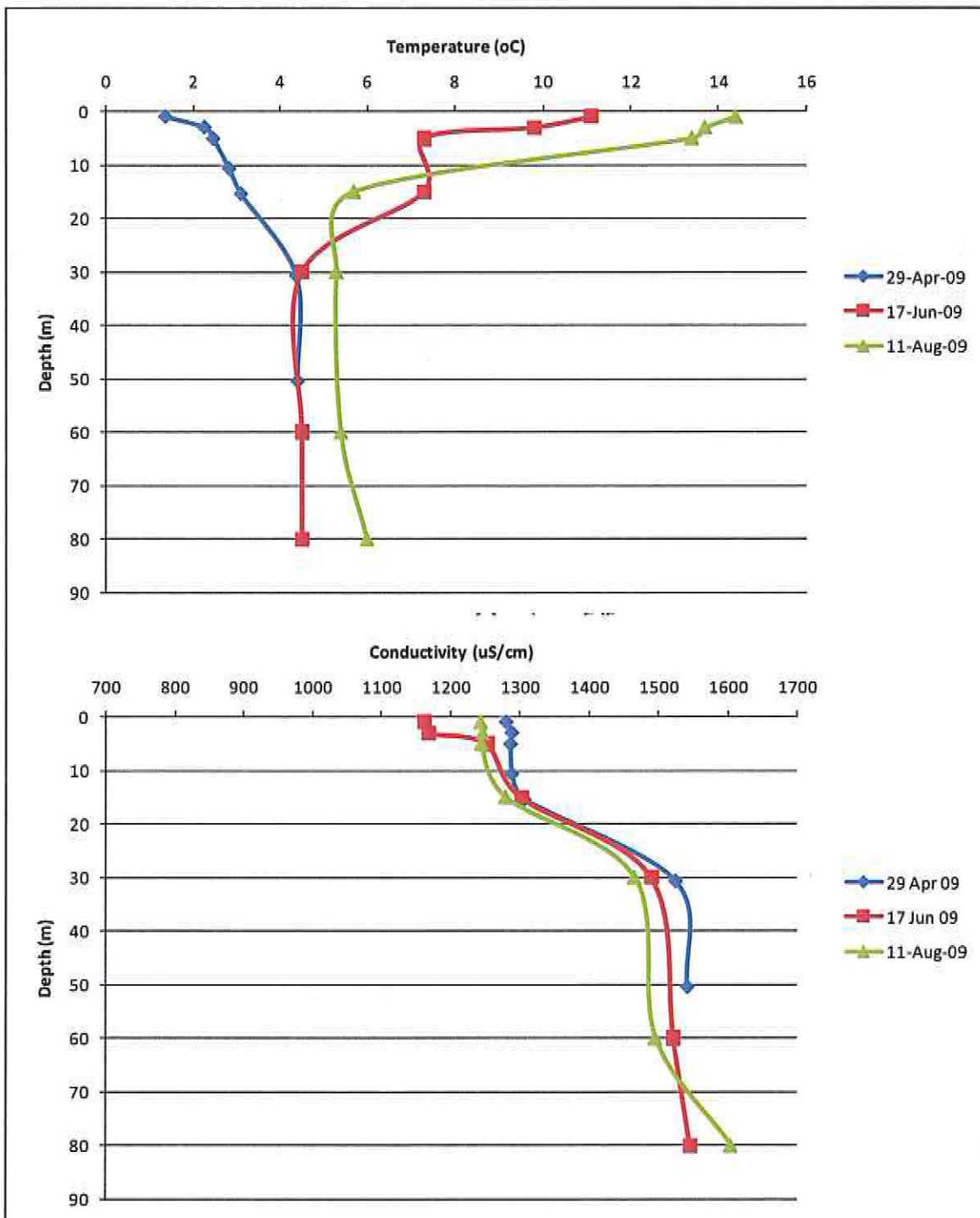


Figure 3.1 In-situ Parameters Measured in the Faro Pit

3.1.2 Nutrients

Analytical results for total phosphorus (P) concentrations for 2005 to 2009 at various depths in the Faro Pit lake are shown in Figure 3.2. Note that the August 9, 2006 result for a depth of 30 m and the June 12, 2007 result at a depth of 80 m appear to be anomalous.

The increase in the phosphorus concentration that occurred during 2005 was as a result of fertilization. In 2006, concentrations in general decreased in the near surface water (to a depth of about 15 m below surface) ranging from about 0.002 to 0.004 mg/L, and a further decrease was observed in 2007, with concentrations trending below the detection limit of 0.002 mg/L by the end of summer. The concentration at depth (30 m and below) appears to have decreased as well, and has an average concentration of slightly below 0.004 mg/L. The decreasing trend was observed to continue through 2008 with concentrations at all depths at or below about 0.002 mg/L. Whilst the results for 2009 appear to be of a similar spread and average concentration as for 2008, it should be noted that the detection limit for 2009 was variable, with the majority of the results reported as below the detection limit of 0.005 mg/L. The plot shows these results as at the detection limit, hence the similarity to the 2008 results.

Analytical results for ammonia (as N) are shown in Figure 3.3. Ammonia-N was present in the pit lake prior to the fertilization event undertaken in 2005. The results reported for April 2009 were all below a detection limit of 0.005 and are considered anomalous. The balance of the results for 2009 indicated a net decrease in the surface layer concentrations, whereas the concentrations at depth (30m and deeper) remained approximately the same as for 2008. Overall however concentrations are clearly decreasing within the pit lake.

Note that ammonia in waste rock seepage could be contributing to ammonia concentrations in the pit lake on an ongoing basis; based on the surface water concentrations the contribution however appears to be decreasing over time.

3.1.3 Algal Growth

Whilst the Faro Pit was fertilized three times in 2005, and a rapid increase in chlorophyll 'a' occurred during June and July of that year, the chlorophyll 'a' concentration has in subsequent years remained at or near the detection limit with no significant algal growth indicated during 2006 or 2007. During 2008, a single low value (0.005 ug/L) was recorded in early August for the sample taken at a depth of 1 m. No Chlorophyll measurements were completed for the Faro Pit lake water during 2009.

3.1.4 Zinc Concentrations

Total zinc concentrations at various monitoring depths over time are shown in Figure 3.4. The zinc concentrations confirm the presence of a chemocline in the Faro Pit at a depth between 15 and 30 m below surface, as discussed previously.

As noted in previous assessments, the monitoring results for 2005 suggested a net removal of zinc from the near surface water layers, as was evident in the 1, 3 and 5 m results for that year. The results for 2006 showed a net increase from 2005, with no decrease in the near surface zinc concentrations over the summer period, and the results for 2007 indicated zinc concentrations in the near surface water were similar to that observed in 2006. However, 2008 and again in 2009, zinc concentrations at all sampling depths down to 15 m showed a significant increase. These results suggest that the loading of zinc to the pit lake may be accelerating.

At depth, zinc concentrations appear to be marginally increasing over time with average zinc concentrations of 2.6 mg/L (2006), 2.7 mg/L (2007), 3.0 mg/L (2008) and 3.5 mg/L (2009), a for all samples in the 30 to 80 m depth range.

Iron concentrations also appear to be increasing at depth, as shown in Figure 3.5. This could suggest that iron precipitates (e.g. iron oxy-hydroxides) may be dissolving and that the increase in zinc is a result of desorption from these dissolving phases. Within the surface layer iron clearly remains at low concentrations. However it is possible that iron is entering the pit lake within acidic surface water runoff. Since the pit lake is slightly alkaline in pH, the iron forms iron oxy-hydroxides and the precipitates then settle over time from the surface layer, passing through the chemocline and accumulating at depth. Since the conditions at depth are reducing (as indicated by measured redox values of < -10 mV), the iron precipitates are slow reduced and dissolves within the deeper layer.

When the iron precipitates formed in the surface layer where zinc concentrations are elevated, zinc would have been sorped and transported to the bottom of the pit lake. Dissolution of the iron precipitates would result in the release of the sorped zinc and hence a concurrent increase in both iron and zinc concentrations occurs.

Overall these results indicate that the net zinc loading to the pit lake is increasing, in particular to the water layer above the chemocline. Some of the zinc is removed from the surface layer but is again released at depth.

An approximate water and load balance has been prepared to illustrate the distribution of loadings. The results are shown in Table 3.1. As shown, the average zinc concentration in the upper layer has increased over time. The load estimate accounts for the changes brought about by pumping and treating and the net loading from the Zone II pit (which is assumed to remain approximately constant, 100,000 m³/year at 115 mg/L Zn, 3120 mg/L SO₄). The results indicate that there are significant annual variations, which may be due to differences in annual rainfall (low rainfall resulting in lower loads and visa versa) and possible earthworks within and around the pit. During 2008 the Yukon experienced a very wet summer period and as shown the estimated loading to the pit lake was almost double the average loading. The average total zinc loading for the period 2005 to 2009 has been about 47 tonnes per year. About one quarter of the load is coming from the Zone II pit, and the remaining 75 % would be from the pit wall rocks and seepage from waste rock dumps within the pit lake catchment. Note that the negative change for 2006 was assumed to have been the result of fertilization.

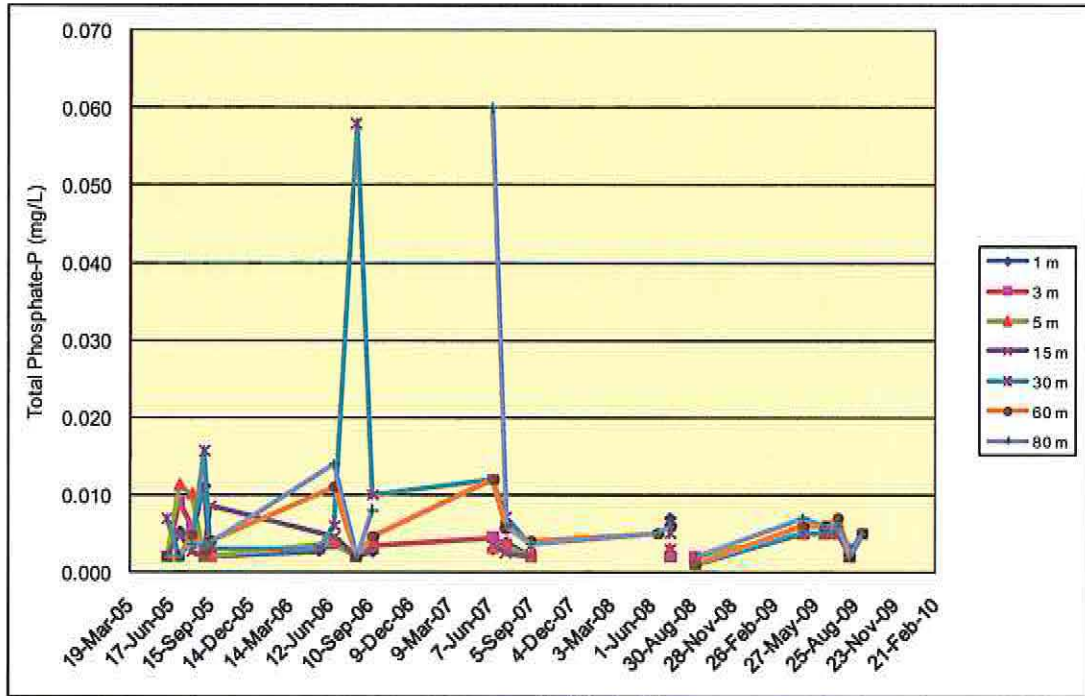


Figure 3.2 Total Phosphate-P Concentrations Measured in the Faro Pit

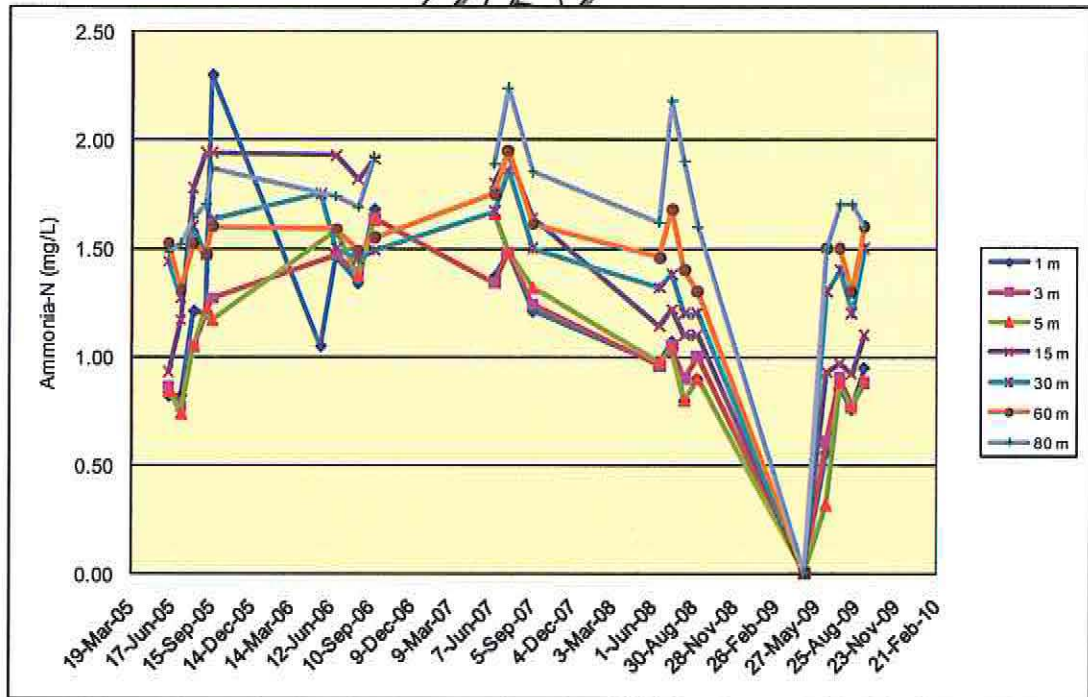


Figure 3.3 Ammonia-N Concentrations Measured in the Faro Pit

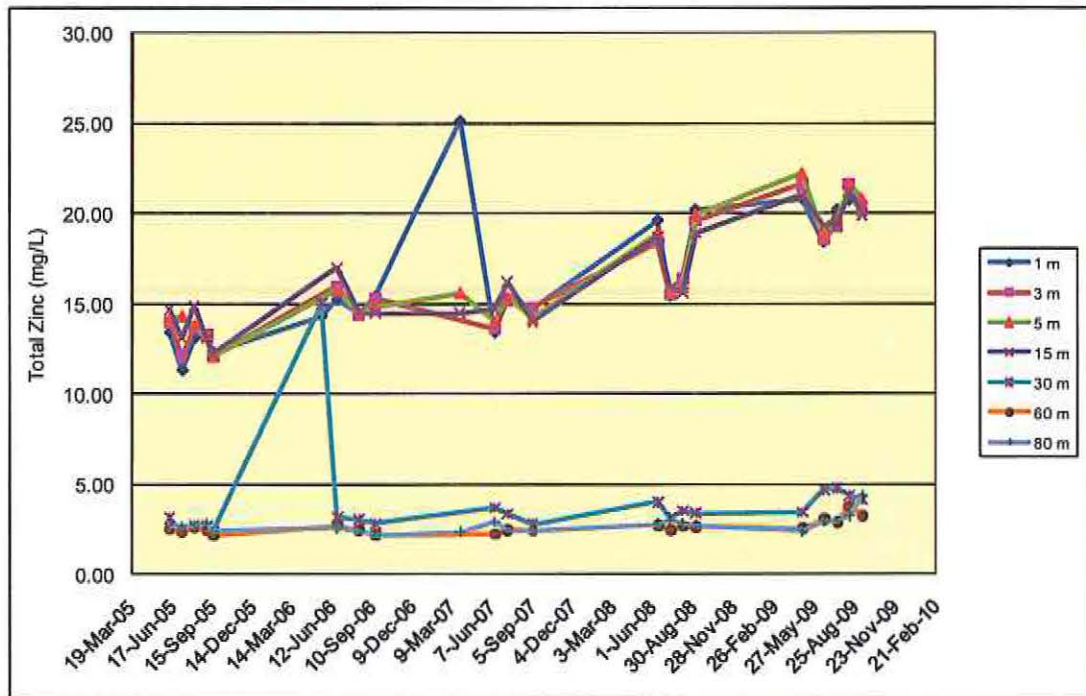


Figure 3.4 Total Zinc Concentrations Measured in the Faro Pit

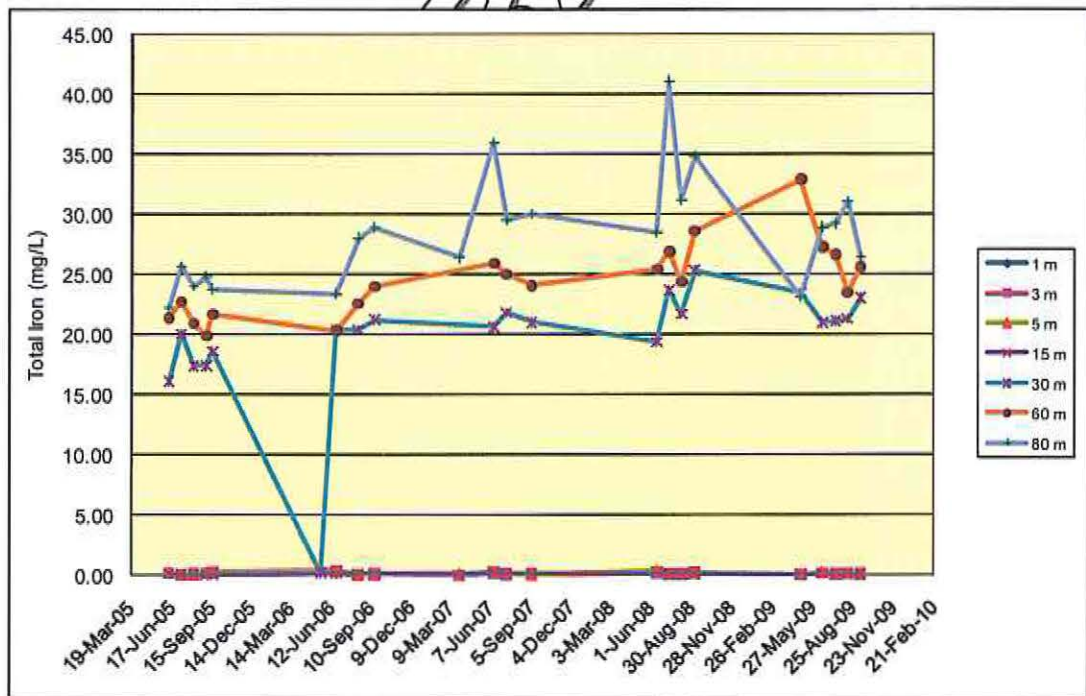


Figure 3.5 Total Iron Concentrations Measured in the Faro Pit .

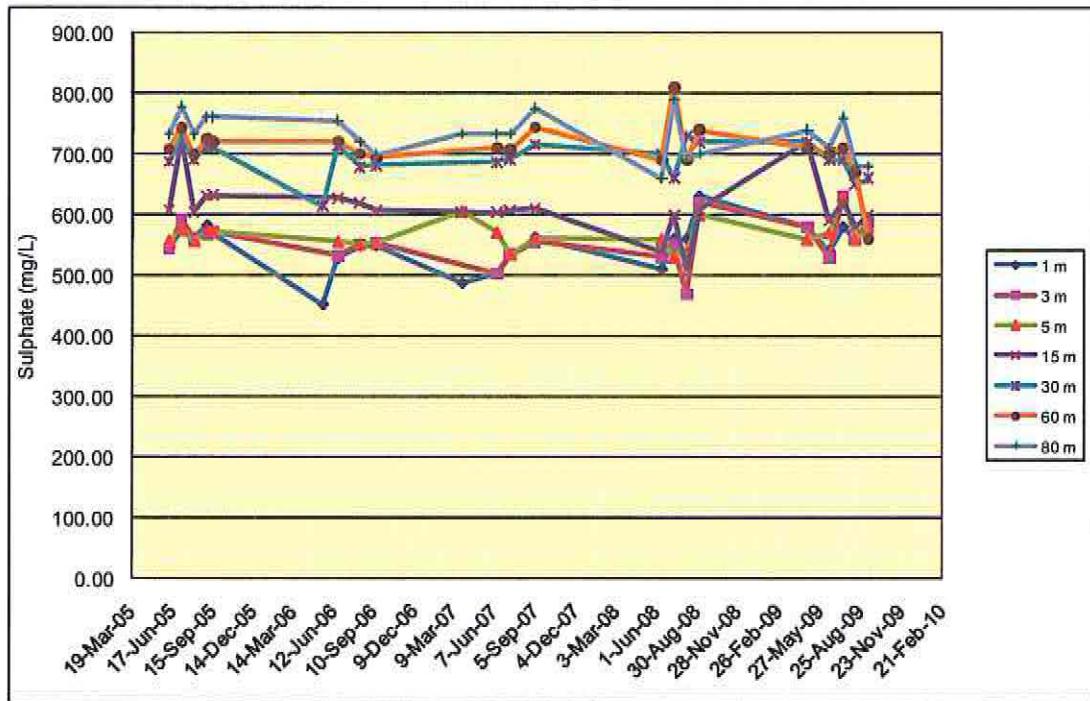


Figure 3.6 Sulphate Concentrations Measured in the Faro Pit .

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Table 3.1 Estimated Net Loadings to the Faro Pit

Date	Elevation m	Total Volume (m3)	Change (m3)	Volume to	Zn	Change (tonnes/yr)	Pumped (tonnes/yr)	NET LOAD (tonnes/yr)	Load Z2 pit (tonnes/yr)	Other sources (tonnes/yr)
				15 m (m3)	(mg/L)					
18-May-05	1176.5	5.12E+07	1.60E+06	1.12E+07	14.1	-21.8	20.9	42.7	11.6	31.1
19-Sep-05	1174.5	4.96E+07			12.1					
31-May-06	1176.1	5.08E+07	1.20E+06	1.12E+07	15.4	32.4	18.2	50.6	11.6	39.0
8-Sep-06	1174.5	4.96E+07			15.0					
22-May-07	1176.5	5.11E+07	1.40E+06	1.12E+07	13.9	-6.4	19.8	13.4	11.6	1.8
11-Sep-07	1174.6	4.97E+07			14.4					
10-Jun-08	1176.4	5.11E+07	1.50E+06	1.12E+07	18.9	58.3	28.9	87.2	11.6	75.6
2-Sep-08	1174.5	4.96E+07			19.7					
29-Apr-09	1176.4	5.11E+07	1.50E+06	1.12E+07	21.4	7.8	31.3	39.1	11.6	27.5
9-Sep-09	1174.5	4.96E+07			20.4					
Date	Elevation m	Total Volume (m3)	Change (m3)	Volume to 15 m (m3)	SO4 (mg/L)	Change (tonnes/yr)	Removed (pumped) (tonnes/yr)	NET LOAD (tonnes/yr)	Load Z2 pit (tonnes/yr)	Other sources (tonnes/yr)
18-May-05	1176.5	5.12E+07	1.60E+06	1.12E+07	564	212	918	918	315	603
19-Sep-05	1174.5	4.96E+07			583					
31-May-06	1176.1	5.08E+07	1.20E+06	1.12E+07	560	-192	675	483	315	168
8-Sep-06	1174.5	4.96E+07			566					
22-May-07	1176.5	5.11E+07	1.40E+06	1.12E+07	545	73	782	855	315	540
11-Sep-07	1174.6	4.97E+07			572					
10-Jun-08	1176.4	5.11E+07	1.50E+06	1.12E+07	535	477	863	1339	315	1024
2-Sep-08	1174.5	4.96E+07			615					
29-Apr-09	1176.4	5.11E+07	1.50E+06	1.12E+07	610	-307	898	591	315	276
9-Sep-09	1174.5	4.96E+07			588					

It is not possible to distinguish between loadings from the wall rocks and the waste rock seepage because the flows and concentrations from the Faro Valley dumps, for example, vary considerably over the season (e.g. 20 to 121 mg/L zinc, and flows ranging from 2 to >1000 L/min) and only spot measurements of flows and concentrations are available. Zinc concentrations in wall rock seeps have been noted to range from less than the detection limit to 875 mg/L; however, flows other than those associated with the Faro Valley Dumps are generally low. This suggests that the majority of the loading is likely associated with seepage from the waste rock dumps.

Similar calculations were completed for sulphate loadings and included in **Error! Reference source not found.** These calculations indicate an annual average loading to the pit lake of about 840 tonnes of sulphate, comprising about 315 tonnes from the Zone 2 Pit and the balance originating from other sources (wall rocks and waste rock dumps within the catchment).

3.1.5 Summary

In summary, subsequent to fertilization in 2005, orthophosphate concentrations have remained below detection, and total phosphorus is being depleted from the surface layer to a depth of about 15 m. At depth, while there appeared to be a marginal increase in total phosphorus, a net decrease in the concentrations is occurring.

Whilst the chemocline continues to exist and is isolating the water at depth from the water near surface, there is a detectable increase in the zinc concentration at depth. This increase is occurring concurrently with an increase in iron concentration which suggests that the zinc increase is likely due to its desorption from dissolving iron precipitates transported by gravity from the surface layer of the pit lake.

Mass balance calculations indicate that while there is some annual variability in the total loadings, on average the annual zinc loading to the pit lake is about 47 tonnes per year, of which about 25 % (11.6 tonnes) is estimated to be sourced from the Zone 2 Pit water transfer. The balance of the loading represents the combined loadings from the wall rocks and the waste rock seepage that enters the pit lake. Monitoring results for seepage from the wall rocks suggest that the majority of this loading is likely present in seepage from the waste rock dumps.

3.2 Grum Pit

3.2.1 Pit Lake Bathymetry

A bathymetric survey of the Grum Pit lake was completed by Laberge during 2009 and is provided in Appendix C. The survey was used to calculate the approximate Grum Pit lake capacity curve below the water and the compared to the volume capacity curve that was used in the past for water balance calculations. The comparison of the capacity curves between the original and that based on the bathymetric results, as shown in Figure 3.7 indicate that the initial volume of the Grum Pit lake was vastly overestimated. The vertical line indicates the lake elevation as at the time of the bathymetric survey. Because of this large error, the capacity curve for the water and load balance calculations has been revised. This change is reflected in the following sections of the report pertaining to the Grumm Pit lake.

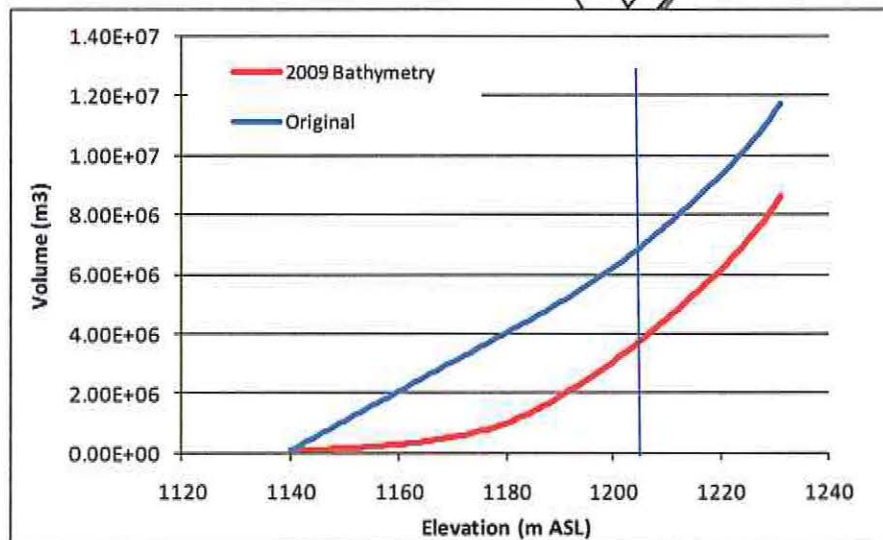


Figure 3.7 Comparison between Assumed and Actual Grum Pit Lake Capacity Curve

3.2.2 Water Level

The Grum Pit lake elevation over time is illustrated in Figure 3.8. As shown, the rate of rise increased significantly during 2008 due, in part, to the fact that this part of the Yukon experienced significantly more rain in 2008 than normal, i.e. about 2 times normal during the summer months. However, Laberge noted an increase in the rate and extent of till sloughing in some areas of the east pit wall during that time as well.

The effect of the would be expected to be twofold; first the till sloughing into the pit lake would cause an increase in the pit lake level and second, the shift of the wall could cause a 'narrowing' or decrease of the planar surface area of the pit at lake level (i.e. the lake would be narrower than if the

wall had remained in place). As a result, the rate of rise from the same inflow would be higher. (The flipside of this is that at some future time it may be possible that the pit lake area becomes wider than it would otherwise have been and the rate of rise could decrease.)

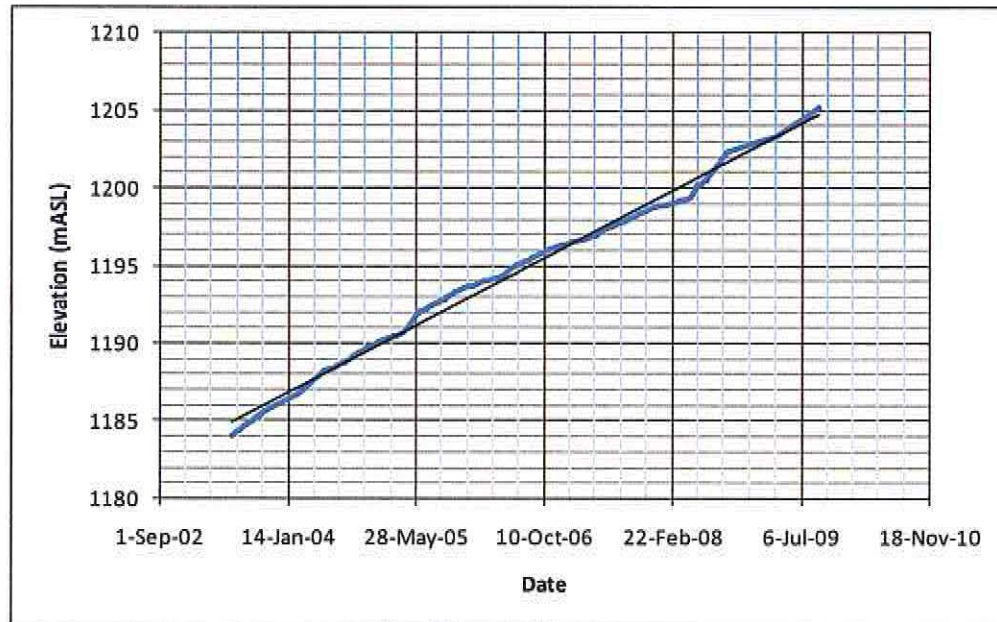


Figure 3.8 Grum Pit Lake Elevation over Time

The sloughing of the pit wall adds a complexity to the pit lake inventory calculations since the pit capacity above the water level is based on the topography of the pit before the sloughing had occurred. Although pit sloughing may have contributed to the difference between the original assumed capacity curve and that based on the bathymetric survey, the sloughing does not account for the entire volume change since the pit lake elevation increased only by about 4 m during 2008 which, at most, equates to about 500 000 m³ (based on the updated capacity curve).

As shown, by September 2009 the lake elevation had risen to about 1205 m ASL.

3.2.3 In-situ Monitoring

The in-situ monitoring results, as shown in Figure 3.9 indicate that as in previous years a thermocline developed in the Grum Pit to a depth of about 5 m below the water surface. A maximum surface water temperature of about 14°C was recorded in August. The electrical conductivity (EC) profiles indicate a freshwater surface layer after ice-melt, which remained for most of the open water season. The results for June indicated a marginal increase in salinity at the surface. At depth however, the EC values were inconsistent, ranging from about 1050 uS/cm in April to 1200 uS/cm in August. This variation is likely due to instrument error as such large variations are not common or readily explainable and are inconsistent with results from previous years. Whilst showing similar trends the

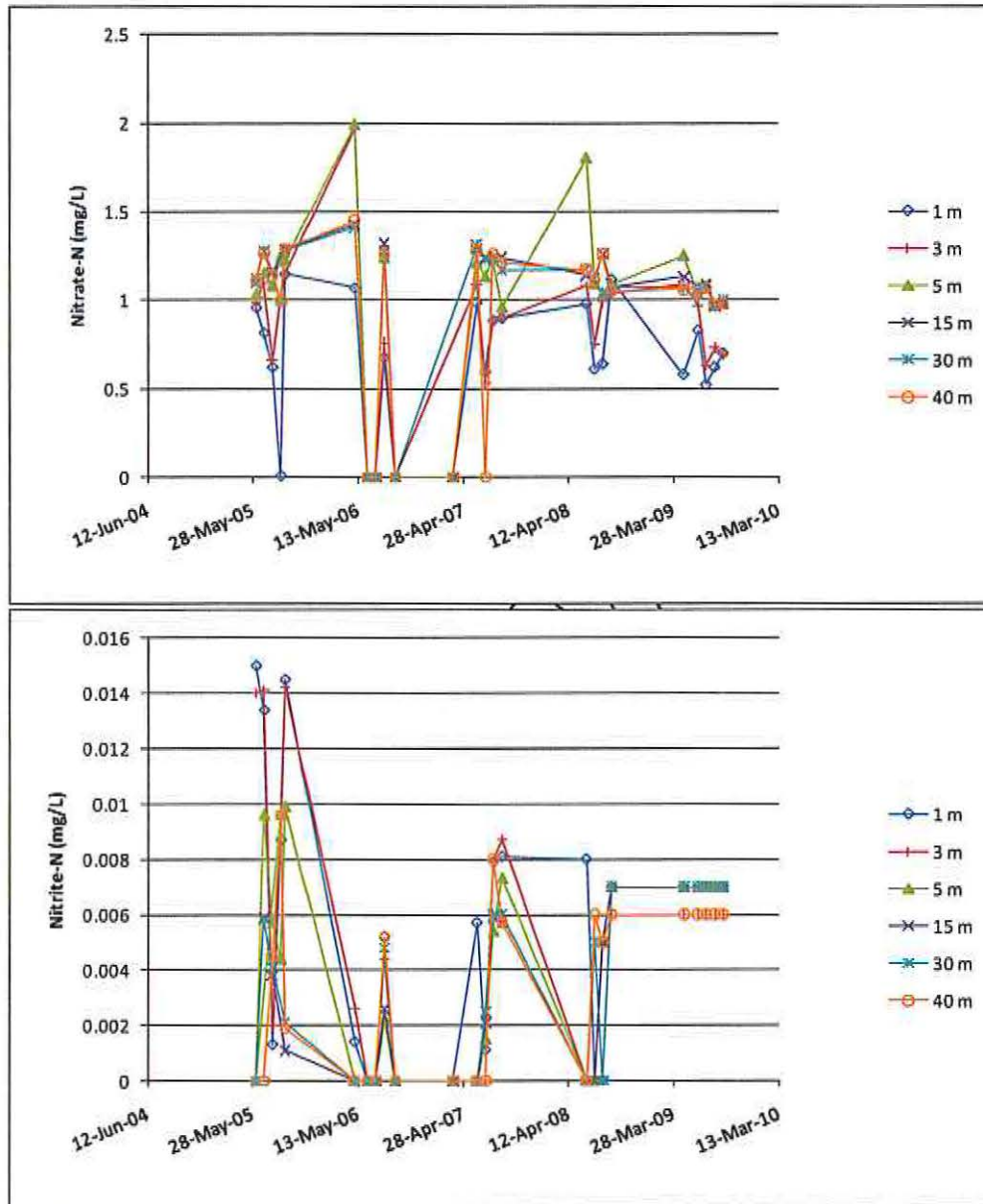


Figure 3.13 Nitrate-N and Nitrite-N Concentrations at Various Depths in the Grum Pit

3.2.5 Algal Growth

Based on visual observations made by field staff, rapid growth with high algal densities was again achieved in the Grum Pit. Whilst only limited chlorophyll 'a' analyses were completed for 2009, the results supported the visual observations. In general, based on the available results as shown in Figure 3.14, the chlorophyll 'a' concentrations were lower than recorded in 2005, but has been

consistently peaking at about 10 ug/L since the fertilization program was curbed to limit the nutrient availability at the end of the season.

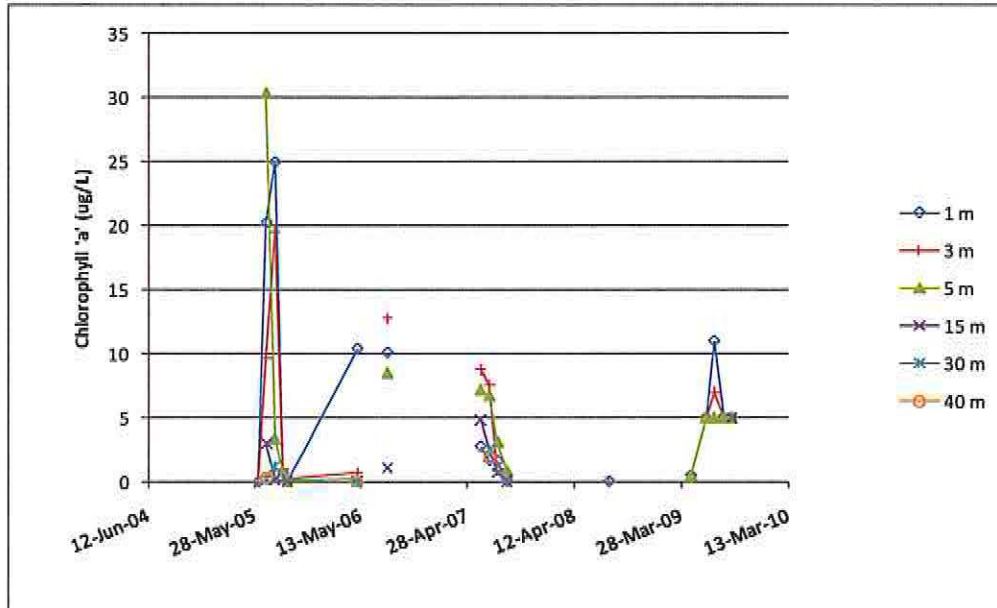


Figure 3.14 Chlorophyll 'a' Concentrations at Various Depths in the Grum Pit

3.2.6 Sulphate Concentrations and Loadings

The sulphate concentrations at various depths over time in the Grum Pit lake are shown in Figure 3.15. As shown, the sulphate concentrations remain approximately constant, slightly above about 400 mg/L. Since water is accumulating in the pit lake, and since the sulphate concentrations are remaining approximately constant at the end of each open water season, it is clear that there is an ongoing source of sulphate that is flowing into the pit lake (i.e. the sulphate concentration is not being diluted by fresh water inflows and direct precipitation even though the level of the lake is rising). Over the term of the monitoring period, were there no additional sources of sulphate, the sulphate concentration would have been lowered by more than 40 % (i.e. base don dilution the sulphate concentration would have decreased from about 425 to about 250 mg/L). The actual decrease in sulphate concentrations is about 10 %.

The total sulphate inventory in the water body over time is shown in Figure 3.16. If the end-of-season inventories are compared, it appears that there is net sulphate loading on the order of about 110 to 120 tonnes entering the pit lake each year.

3.2.8 Zinc Loadings

As discussed in a preceding section, the sulphate inventory in the pit lake has steadily increased over time indicating a net loading. Since zinc is an oxidation product similar to sulphate, it is reasonable to expect that there is a net loading of zinc to the pit lake as well. However, the total loading is difficult to determine because, concurrent with the loading, biological treatment has led to the removal of zinc. The assessment of zinc removal achieved by biological treatment is tentative at best since it relies on the net balance of zinc before and after each season. This is because the total loading cannot be determined within the available monitoring data. Nevertheless, in contrast to sulphate, the annual monitoring results have shown that the inventory of zinc has fluctuated within an approximately constant range since 2005, as shown in Figure 3.19.



Figure 3.19 Grum Pit Zinc Inventory

The results in the above figure further shows that up until 2008, each year the inventory increased over the winter months by about 4 to 5 tonnes of zinc. This trend did not occur during the 2008/09 winter period. The results further indicate that the inventory decreased substantially from the 2008 end-of-season to the 2009 end-of-season, suggesting that: i) winter removal occurred likely due to sulphate reduction (which is confirmed by the reduction in concentrations at depth) and, ii) the annual removal achieved in 2009 was approximately the same as the loading to the pit lake. Since the removal appears to have equalled the loading, the removal can be quantified only if the loading is known.

To provide a basis for estimating the annual loading to the pit lake, a series samples were obtained from the pit wall rocks. The sample locations were distributed around the different exposed lithological units as shown in Figure 3.20. The samples were subjected to leach extraction tests conducted at a solid:water ratio of 1:3 to estimate solute release rates. The results are summarized in Table 3.3. As shown, the results indicate very high concentrations of zinc may be released from the sulphide unit. The sulphide unit is also a source of copper, iron, lead and nickel. All of the units release sulphate to various degrees.

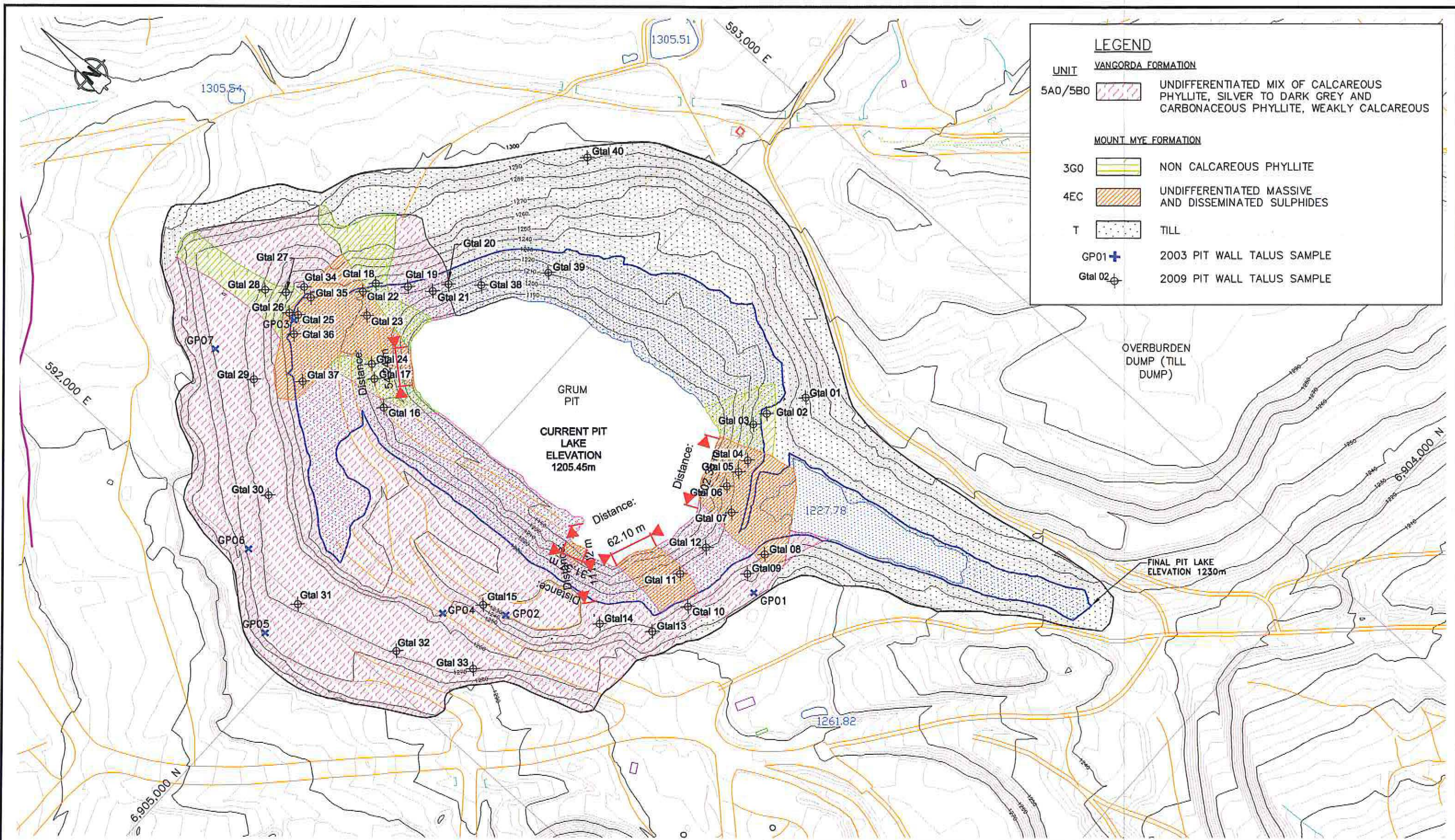
Table 3.3 Total and Dissolved Zinc Concentrations in Surface Layer (< 1 m depth)

Lithology		pH	Sulphate mg/L	Copper mg/L	Iron mg/L	Lead mg/L	Nickel mg/L	Zinc mg/L
3G0 Non Calcareous Phyllite	Average	-	832	0.00074	0.0062	0.00016	0.0066	0.0070
	Minimum	7.07	72	0.00005	0.001	0.000005	0.00002	0.0001
	Maximum	7.89	1858	0.0013	0.014	0.00048	0.019	0.018
4EC Sulphides (massive, disseminated)	Average	-	1201	4.32	1.15	1.14	0.562	160
	Minimum	2.52	266	0.00023	0.005	0.00225	0.00541	0.115
	Maximum	8.1	2912	37.5	596	247	4.1	800
5A0/5B0 Mix of Calcareous and Carbonaceous Phyllite	Average	-	238	0.00078	0.008	0.00064	0.0019	0.045
	Minimum	7.47	10	0.00025	0.002	0.000035	0.00038	0.0005
	Maximum	8.08	943	0.0026	0.057	0.0064	0.0094	0.61
Till	Average	-	184	0.0022	0.0535	0.00024	0.0012	0.0049
	Minimum	7.65	7	0.00053	0.005	0.00004	0.0007	0.0008
	Maximum	8.19	385	0.0051	0.185	0.00048	0.0018	0.010
Uncertain	Average	-	390	0.0086	0.06	0.030	0.040	6.63
	Minimum	7.08	2	0.0003	0.002	0.000018	0.00015	0.0005
	Maximum	8.06	1337	0.022	0.153	0.115	0.127	25.3

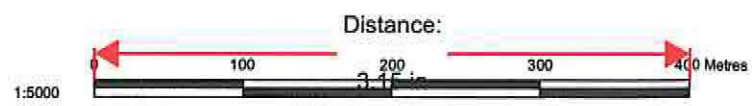
Based on the average zinc concentration about 480 mg Zn may on average be released from the sulphide rich units. Assuming the talus depth on the benches of about 0.5 m and an equivalent depth of fracturing/oxidation on the sulphide wall rocks, every square meter of pit wall face may release about 0.43 kg of zinc (up to 2.2 kg at the maximum concentration) to the water column when inundated. The samples of uncertain classification (7.5 % of the samples) indicate that there may be other materials present that could release zinc; however the surface area they represent is unknown.

Based on an average annual water level rate of rise of about 3 m, and an estimated pit lake perimeter exposure length for the sulphide units (see Figure 3.20) of about 300 m, and assuming a vertical wall, about 9000 m² of rock was inundated during 2009. This equates to a potential net annual loading of about 388 kg, not including any loading associated with wall rock runoff. The balance of the wall rock inundated would contribute less than 5 kg of zinc. (This assumes that 7.5 % of the samples were of the 'uncertain' category with a higher soluble zinc content.)

The corresponding sulphate release from the sulphides would have been about 2.91 tonnes, with the remainder of the wall rock contributing about 1.99 tonnes. As noted before, the annual sulphate loading is in the order of 110 to 120 tonnes per year. The balance of the loading (110 - 5 = 105 tonnes) is present in the pitwall runoff and any inflows to the lake.



LEGEND	
VANGORDA FORMATION	
5A0/5B0	UNDIFFERENTIATED MIX OF CALCAREOUS PHYLLITE, SILVER TO DARK GREY AND CARBONACEOUS PHYLLITE, WEAKLY CALCAREOUS
MOUNT MYE FORMATION	
3G0	NON CALCAREOUS PHYLLITE
4EC	UNDIFFERENTIATED MASSIVE AND DISSEMINATED SULPHIDES
T	TILL
GP01	2003 PIT WALL TALUS SAMPLE
Gtal 02	2009 PIT WALL TALUS SAMPLE



SRK Consulting
Engineers and Scientists
Vancouver B.C.

SRK JOB NO.: 1CY001.035
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Assuming that the pit wall runoff contains sulphate concentrations similar to those obtained in the leach extraction tests, then the annual pit wall rock runoff together with the wall surface areas may be used to estimate the loadings that enter the lake by this mechanism. The estimated exposed surface areas above the water level for each lithology are summarised in Table 3.4. The table also shows the average concentration achieved in the leach extraction test and the calculated average concentration for the entire pit wall weighted according to the surface area. This would represent the average concentration of sulphate and zinc in the runoff from direct precipitation. Based on the total wall rock surface area and the annual rainfall, the estimated annual runoff (assuming 100 % runoff coefficient) is about 275 000 m³.

Table 3.4 Summary of Wall Rock Exposure and Average Leach Extraction Concentrations

Unit	Description	Area (m ²)	Distribution	SO ₄ (mg/L)	Zn (mg/L)
4EC	Sulphides	43455	7.5%	1201	160
3GO	Non calcareous Phyllite	29026	5.0%	832	0.007
5A0/5B0	Undifferentiated Phyllites	268956	46.3%	285*	1.51*
-	Till	239289	41.2%	184	0.0049
Total		588725	100.0%	-	-
Surface Area Weighted Average Concentration (mg/L)				339	12.7
Loading in Runoff (kg)				93 300	3485

Notes: * includes 'uncertain' samples

As shown in Table 3.4, the estimated sulphate loading from the wall rock is estimated to be about 93.3 tonnes. Together with the estimated loading from the inundation of the wall rock, the total loading is in the order of about 98 tonnes. Since there is likely to be some sulphate in the runoff that enters the open pit as natural runoff, the total estimate accounts for most of the estimated annual loading of between 110 and 120 tonnes (see Section 3.2.6) and is considered to be sufficiently accurate for the purposes of this assessment. The approach also appears to be reasonable for estimating wall rock loadings. As well, it is considered reasonable to use the same method for estimating the loading for zinc. Accordingly the estimated total zinc loading (inundation + runoff) is about 3.88 tonnes.

As shown in Figure 3.19, during the summer season the zinc inventory initially decreased and then increased slight to a level equivalent to the start of the season, i.e. showing no net change. In the context of the preceding load calculations, the treatment effectively removed about 3.88 tonnes of zinc from the water column. Similar or higher loadings would have existing for preceding years because the surface areas exposed would have been greater, as would the rise in the water elevation. Therefore, zinc removals on average would appear to have ranged from about 3.5 to 4 tonnes per year, which were offset by similar or slightly higher annual loadings. Furthermore, as noted previously, in contrast to other winter seasons, zinc was also removed from the water column during

the 2008/09 winter season, with approximately 4 tonnes removed during this period, likely due to sulphate reduction occurring beneath the ice.

3.2.9 Estimated Post Spill Loadings and Concentrations

Water quality predictions previously completed for the Grum Pit lake were based on the estimated loadings presented in Table 3.5 (extracted from SRK, 2005). As discussed above, the actual zinc loadings from the wall rocks and releases from talus during inundation is about ten times higher than previously estimated (3000 to 4000 compared to 350 kg per annum). The sulphate loadings have also been underestimated (65 tonnes per annum compared to actual loadings of about 110 tonnes per annum – see Table 3.5).

Table 3.5 Initial Estimates of Net Loadings to the Grum Pit (SRK 2005)

Parameter	Estimated Annual Loading (kg/year)
Cl	83
SO ₄	65 000
Ca	14 000
Mg	12 000
K	350
Na	420
Al	7
Cd	0.6
Co	3
Cu	0.5
Fe	180
Pb	2
Mn	45
Ni	16
Zn	350

As the pit lake elevation rises, more and more of the reactive sulphide wall rock will be inundated so that the loadings from this source will decrease over time. The estimated spill elevation at the time of spill (at lowest elevation in the slot cut, spill elevation of about 1230 mASL) is shown in Figure 3.20. At that elevation most of the sulphidic wall rock will be inundated. The surface areas for the remaining areas above the water level have been estimated and used to calculate the loadings from the pit wall rocks at that time. The water balance was used to then calculate the steady state concentrations in water flowing through the lake (i.e. for the discharge) without any treatment. The results are shown in Table 3.6. As the results in the table suggest, zinc concentrations in the outflow are projected to remain at about 1 mg/L, without any treatment. Note however that the calculations do not consider the potential reduction in oxidation rates as the sulphide content of the surface materials are depleted over time. A few other elements may also remain slightly elevated. These concentrations are generally caused by the exposed sulphidic wall rocks that continue to release metals at elevated concentrations.

Table 3.6 Summary of Estimated Concentrations at Spill Elevation 1230 m ASL

Unit	4EC	3G0	5A0/5B0	Till	Weighted Wall rock Runoff Concentrations mg/L	Annual Loading t/year	Concentrations in Outflow mg/L
Description	Sulphides	Non calcareous Phyllite	Undifferentiated Phyllite	Till			
Area (m2) Distribution	7673 2%	15701 4%	209493 49%	191835 45%			
Parameter	Average Leach Extraction Concentrations						
	mg/L	mg/L	mg/L	mg/L			
Sulphate	1202	832	285	185	277	42.280	118
Hardness CaCO3	634	808	313	222	296	45.238	127
Aluminium Al	9.46	0.005	0.014	0.077	0.212	0.032	0.091
Antimony Sb	0.008	0.0002	0.001	0.002	0.001	0.00018	0.000
Arsenic As	0.258	0.0002	0.001	0.002	0.006	0.001	0.002
Barium Ba	0.019	0.021	0.031	0.043	0.036	0.005	0.015
Beryllium Be	0.006	0.00001	0.00001	0.00001	0.00002	0.00002	0.00005
Bismuth Bi	<0.00005	0.00001	0.00001	0.00001	0.00005	0.0000008	0.000002
Boron B	<0.05	0.050	<0.05	<0.05	0.002	0.00028	0.0008
Cadmium Cd	0.224	0.0002	0.0030	0.00004	0.0056	0.0009	0.0024
Calcium Ca	175	181	67	59	70	10.665	30
Chromium Cr	0.062	0.00010	0.00018	0.00045	0.00141	0.00022	0.00060
Cobalt Co	0.218	0.0005	0.0025	0.0001	0.0053	0.0008	0.0023
Copper Cu	4.327	0.001	0.001	0.002	0.080	0.012	0.034
Iron Fe	116.654	0.006	0.017	0.054	2.140	0.327	0.917
Lead Pb	1.146	0.0002	0.0072	0.0002	0.0244	0.004	0.010
Lithium Li	0.020	0.010	0.009	0.007	0.008	0.0013	0.0035
Magnesium Mg	47.8	86.7	35.1	18.2	29.6	4.523	12.7
Manganese Mn	6.315	0.014	0.030	0.030	0.143	0.022	0.061
Mercury Hg	0.800	0.010	0.060	0.010	0.049	0.0075	0.021
Molybdenum Mo	0.009	0.002	0.007	0.007	0.007	0.0010	0.0029
Nickel Ni	0.563	0.007	0.011	0.001	0.016	0.0025	0.0070
Phosphorus P	0.549	0.002	0.002	0.018	0.019	0.0029	0.0082
Potassium K	1.19	2.55	3.49	3.22	2.80	0.428	1.20
Selenium Se	0.006	0.004	0.002	0.007	0.004	0.0007	0.0019
Silicon Si	2.189	0.197	0.450	1.110	0.770	0.118	0.330
Silver Ag	0.0004	0.00002	0.00002	0.00001	0.00002	0.000003	0.00001
Sodium Na	0.446	0.974	0.727	4.320	2.354	0.360	1.01
Strontium Sr	0.208	0.362	0.202	0.327	0.280	0.043	0.120
Sulphur (S)	406	274	96	78	100	15.277	43
Thallium Tl	0.024	0.001	0.0004	0.00004	0.0007	0.00010	0.0003
Tin Sn	0.001	0.00003	0.00005	0.00002	0.00005	0.000007	0.00002
Titanium Ti	0.220	0.001	0.001	0.004	0.006	0.0010	0.003
Uranium U	0.057	0.001	0.001	0.004	0.003	0.00053	0.001
Vanadium V	0.049	0.0002	0.0002	0.001	0.0015	0.00023	0.0006
Zinc Zn	160.5	0.007	0.045	0.005	2.924	0.447	1.252
Zirconium Zr	0.015	0.0001	0.0002	0.0004	0.0006	0.00008	0.0002

A reduction in the exposed sulphidic wall rock can be achieved by increasing the spill elevation of the pit lake. For example, raising the lake elevation by 20 m by installing a dam in the slot cut would inundate essentially all of the sulphidic wall rocks. A revised water quality prediction was completed for a spill elevation of 1250 m ASL. The results are shown in Table 3.7. As shown the steady state zinc concentration in the outflow would decrease to about 0.1 mg/L without treatment.

Table 3.7 Summary of Estimated Concentrations at Spill Elevation 1250 m ASL

Unit	4EC	3G0	5A0/5B0	Till	Weighted Wall rock Runoff Concentrations	Annual Loading t/year	Concentrations in Outflow mg/L
Description	Sulphides	Non calcareous Phyllite	Undifferentiated Phyllite	Till			
Area (m2) Distribution	603 0.20%	12396 4.1%	142000 47%	149056 49%			
Parameter	Average Leach Extraction Concentrations				mg/L		
	mg/L	mg/L	mg/L	mg/L			
Sulphate	1202	832	285	185	260	28.460	80
Hardness CaCO3	634	808	313	222	289	31.646	89
Aluminium Al	9.46	0.005	0.014	0.077	0.063	0.007	0.019
Antimony Sb	0.008	0.0002	0.001	0.002	0.001	0.00012	0.000
Arsenic As	0.256	0.0002	0.001	0.002	0.002	0.000	0.001
Barium Ba	0.019	0.021	0.031	0.043	0.037	0.004	0.011
Beryllium Be	0.006	0.00001	0.00001	0.00001	0.00002	0.00000	0.00001
Bismuth Bi	<0.00005	0.00001	0.00001	0.00001	0.000005	0.0000006	0.000002
Boron B	<0.05	0.050	<0.05	<0.05	0.002	0.00022	0.0006
Cadmium Cd	0.224	0.0002	0.0030	0.00004	0.0019	0.0002	0.0006
Calcium Ca	175	181	67	59	68	7.458	21
Chromium Cr	0.062	0.00010	0.00018	0.00045	0.00043	0.00005	0.00013
Cobalt Co	0.218	0.0005	0.0025	0.0001	0.0017	0.0002	0.0005
Copper Cu	4.327	0.001	0.001	0.002	0.010	0.001	0.003
Iron Fe	116.654	0.006	0.017	0.054	0.266	0.029	0.081
Lead Pb	1.146	0.0002	0.0072	0.0002	0.0058	0.001	0.002
Lithium Li	0.020	0.010	0.009	0.007	0.008	0.0009	0.0024
Magnesium Mg	47.8	86.7	35.1	18.2	28.9	3.166	8.9
Manganese Mn	6.315	0.014	0.030	0.030	0.042	0.005	0.013
Mercury Hg	0.800	0.010	0.060	0.010	0.035	0.0038	0.011
Molybdenum Mo	0.009	0.002	0.007	0.007	0.007	0.0007	0.0020
Nickel Ni	0.563	0.007	0.011	0.001	0.007	0.0008	0.0021
Phosphorus P	0.549	0.002	0.002	0.018	0.011	0.0012	0.0034
Potassium K	1.19	2.55	3.49	3.22	2.85	0.312	0.87
Selenium Se	0.006	0.004	0.002	0.007	0.005	0.0005	0.0014
Silicon Si	2.189	0.197	0.450	1.110	0.767	0.084	0.235
Silver Ag	0.0004	0.00002	0.00002	0.00001	0.00001	0.000001	0.00000
Sodium Na	0.446	0.974	0.727	4.320	2.498	0.273	0.77
Strontium Sr	0.208	0.362	0.242	0.327	0.284	0.031	0.087
Sulphur (S)	406	274	96	78	95	10.388	29
Thallium Tl	0.024	0.001	0.0004	0.00004	0.0003	0.00003	0.0001
Tin Sn	0.001	0.00003	0.00005	0.00002	0.00003	0.000004	0.00001
Titanium Ti	0.220	0.001	0.001	0.004	0.003	0.0003	0.001
Uranium U	0.057	0.001	0.001	0.004	0.003	0.00029	0.001
Vanadium V	0.049	0.0002	0.0002	0.001	0.0007	0.00008	0.0002
Zinc Zn	180.5	0.007	0.045	0.005	0.342	0.037	0.105
Zirconium Zr	0.015	0.0001	0.0002	0.0004	0.0003	0.00004	0.0001

3.2.10 Summary

The results indicate that the Grum Pit again responded well to fertilization and that good algal growth was again achieved in 2009. Consistent with previous years, elimination of late summer fertilization has again been shown to effectively control nutrient levels by the end of the season.

Surface water mass balance calculations for 2004 through 2009 suggest that zinc loadings to the Grum Pit are on the order of 3000 to 4000 kg per annum. The loading appears to be associated with predominantly with wall rock runoff, and to a lesser extent with solute release from wall rock as the water level rises. Net zinc removal has matched the zinc loading to the pit lake during the 2009 summer period. Contrary to previous years, the concentrations in the pit lake did not increase over

the 2008/09 winter season. Rather, a net removal of zinc occurred. The mechanism appears to be a result of sulphate reduction, with concurrent metal precipitation as a sulphide mineral.

Due to the removals depth weighted zinc concentration has decreased from in excess of 10 mg/L in 2005 to about 4.8 mg/L.

Mass balance calculations further showed that the net loading of sulphate to the pit lake is about 110 to 120 tonnes per year.

The leach extraction tests indicate that most of the lithological units release sulphate whereas most of the zinc release is associated with the sulphidic wall rocks. A potential correlation was established between the sulphate loadings from the wall rocks and the leach extraction test results. This correlation was extended to verify the potential annual zinc loadings to the pit lake and then used to estimate steady state concentrations in outflow from the pit lake when it spills naturally at an elevation of about 1230 m ASL. The results suggest that the steady state zinc concentration in the steady state outflow (without any treatment) could still remain at about 1 mg/L. This is primarily due to the sulphidic wall rock remaining above the lake level. By raising the lake elevation by about 20 m, the zinc concentrations could potentially be reduced to about 0.1 mg/L in the long term.

3.3 Vangorda Pit

3.3.1 Concentrations

The pH trends over time are shown in Figure 3.21 for each of the monitoring depths. The results show that the surface water acidified in early 2006 but returned to circum neutral conditions by the end of that year. During 2007 surface water again acidified however the pH did not increase again to circum neutral conditions and remained acidic through to early 2008. During 2009, the first sampling event showed a slight increase in the pH in the surface water layer due to ice melt and spring freshet inflows. By the second monitoring event, the entire water column was at the same pH level.

The results for sulphate, zinc, cobalt and nickel concentrations indicated that complete mixing had occurred at the end of 2008, as shown in Figure 3.22 and Figure 3.23. As with the pH profiles, while there was a distinct difference between the surface and deeper water for the first sampling event, this distinction became less significant during the subsequent sampling events unlike the results for 2008. These results suggest that the loadings to the surface layer are sufficiently high to nullify the effects of dilution from the freshet flows.

Cadmium and copper concentrations as shown in Figure 3.24 appear to decrease slightly at depth and did not show significant increases even though the entire water column is acidic.

3.3.2 Estimated Loadings

As shown in Figure 3.25, conductivity measurements suggest that even though it acidified during this period, the water in the surface layer freshened rapidly in the spring of 2009, as in preceding years. Overall, however there is a net increase in the EC indicating a net accumulation of solutes in the pit lake. Since the water transfer rates for the pit lake were not available it was not possible to complete overall water and load balances for the Vangorda Pit for 2009. The estimated average annual loadings were determined to the end of the 2008 monitoring period, by obtaining the difference between the estimated inventories as at August 2005 and June 2008, and then normalising the net loading to the pit to an annual rate, are shown in the last column of Table 3.8.

The average annual loadings are well above previous estimates. For example the current estimate of the zinc loading is almost twice the estimate derived in 2007 the reason for this appears to be the acidification of the water column. Acidification negates the effects of solubility controls and would cause accumulated precipitates on the pit lake floor to dissolve. This is likely to greatly enhance dissolved and total metal concentrations. Consequently, solute loadings entering the pit lake with surface runoff are likely to be below the current estimates.

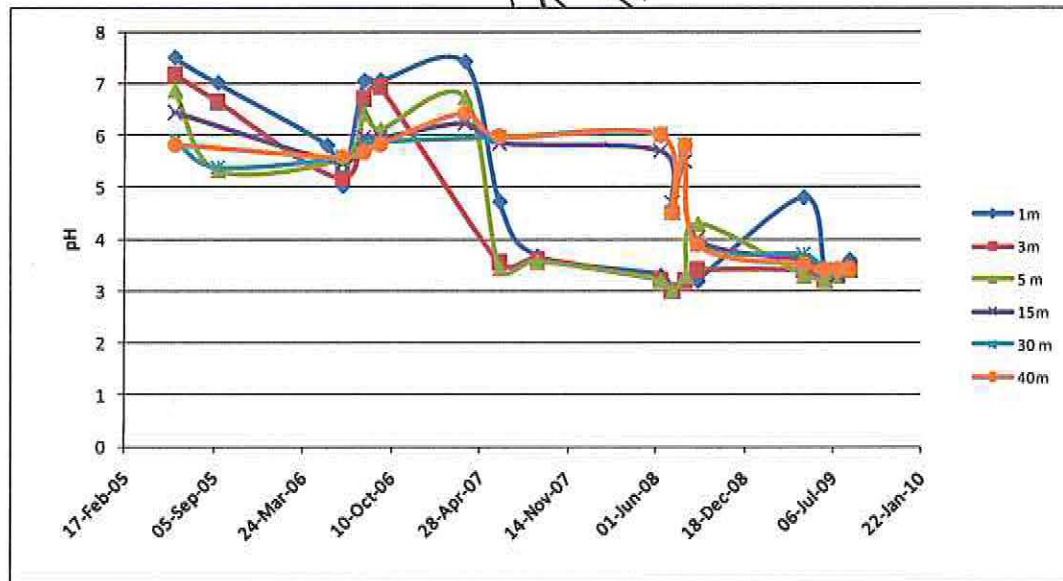


Figure 3.21 Vangorda Pit pH Trends

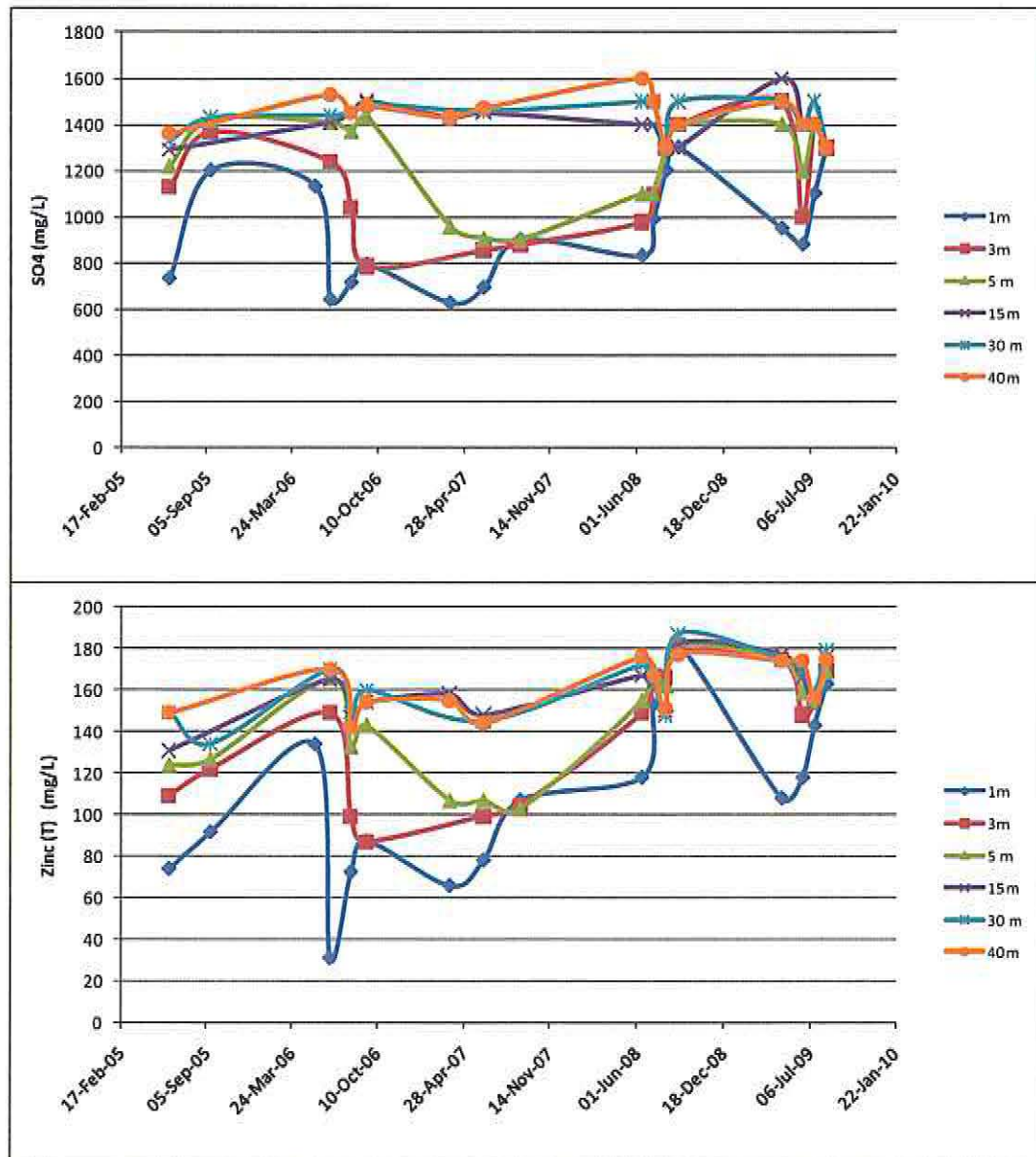


Figure 3.22 Vangorda Pit Lake Sulphate and Zinc Concentration Trends

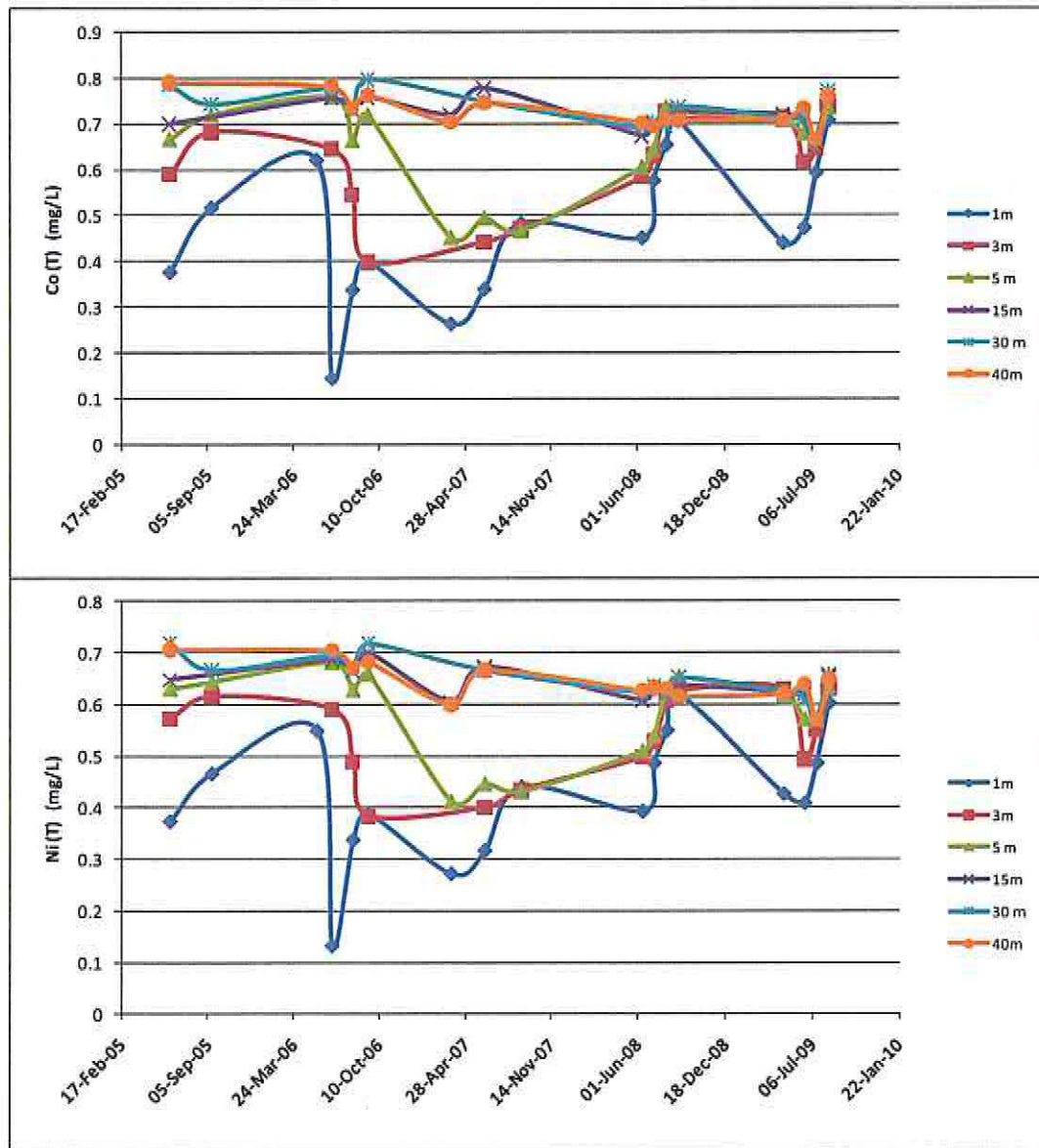


Figure 3.23 Vangorda Pit Lake Cobalt and Nickel Concentration Trends

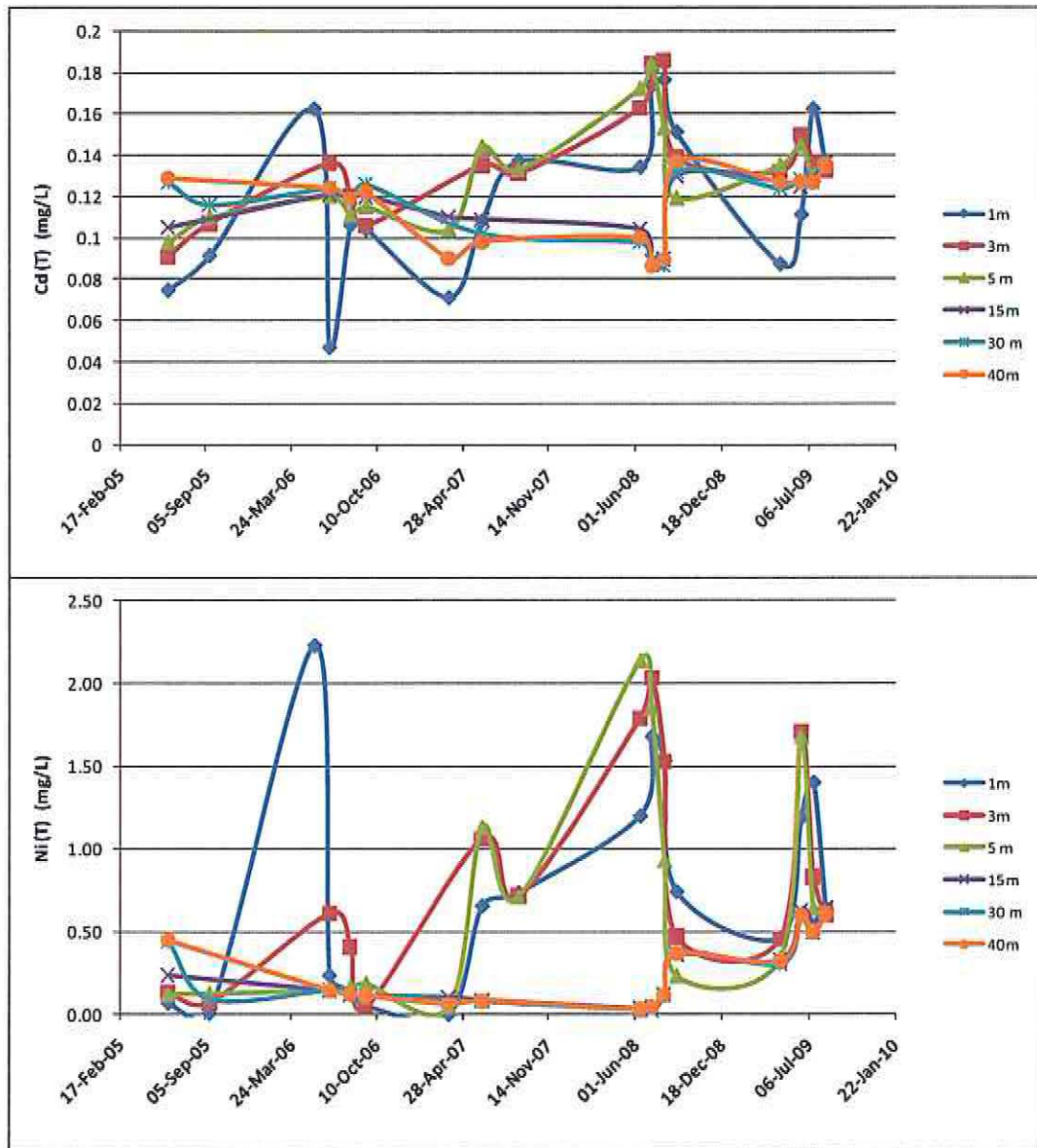


Figure 3.24 Vangorda Pit Lake Cadmium and Copper Concentration Trends

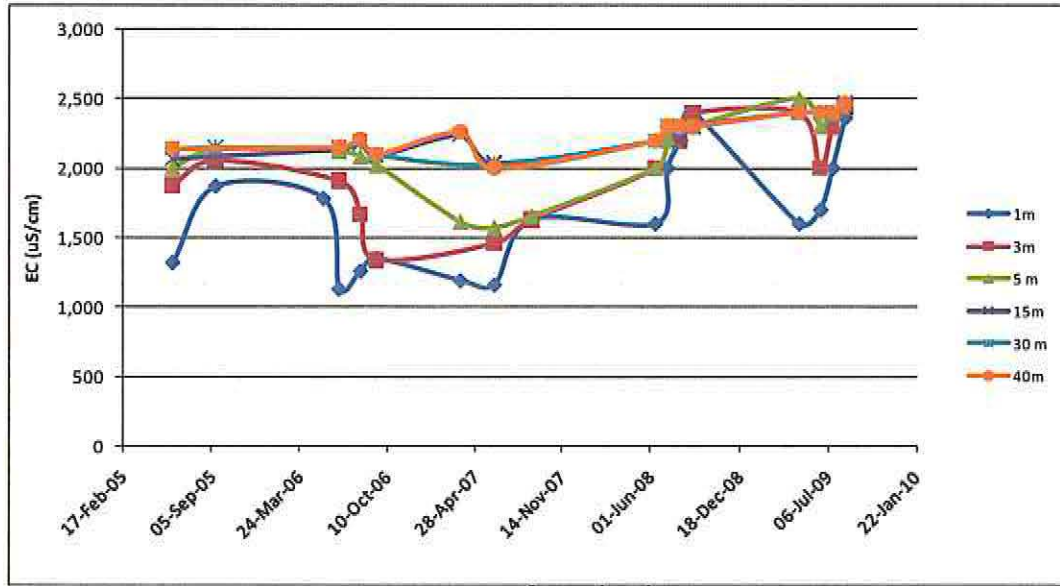


Figure 3.25 Electrical Conductivity Trends in the Vangorda Pit Lake

Table 3.8 Summary of Estimated Annual Loading to the Vangorda Pit Lake

Parameter	Inventory (tonnes)		Average Loading (tonne/year)
	Aug-05	Jun-08	
Volume (m ³)	725,816.93	1,787,986	386,147
Sulphate SO ₄	980	2,208	447
Cadmium	0.078	0.236	0.055
Cobalt	0.489	1.103	0.223
Copper	0.051	1.557	0.548
Iron	22	74	19
Manganese	29	73	16
Nickel	0.442	0.969	0.192
Zinc	88	279	70

4 Conclusions and Recommendations

4.1 Conclusions

The conclusions from the 2009 biological treatment program for Grum Pit lake, and the monitoring of the Faro and Vangorda Pit lakes can be summarised as follows.

4.1.1 Faro Pit

In summary, subsequent to fertilization in 2005, orthophosphate concentrations have remained below detection, and total phosphorus is being depleted from the surface layer to a depth of about 15 m. At depth, while there appeared to be a marginal increase in total phosphorus, a net decrease in the concentrations is occurring.

Whilst the chemocline continues to exist and is isolating the water at depth from the water near surface, there is a detectable increase in the zinc concentration at depth. This increase is occurring concurrently with an increase in iron concentration which suggests that the zinc increase is likely due to its desorption from dissolving iron precipitates transported by gravity from the surface layer of the pit lake.

Mass balance calculations indicate that while there is some annual variability in the total loadings, on average the annual zinc loading to the pit lake is about 47 tonnes per year, of which about 25 % (11.6 tonnes) is estimated to be sourced from the Zone 2 Pit water transfer. The balance of the loading represents the combined loadings from the wall rocks and the waste rock seepage that enters the pit lake. Monitoring results for seepage from the wall rocks suggest that the majority of this loading is likely present in seepage from the waste rock dumps.

4.1.2 Grum Pit

The results indicate that the Grum Pit again responded well to fertilization and that good algal growth was again achieved in 2009. Consistent with previous years, elimination of late summer fertilization has again been shown to effectively control nutrient levels by the end of the season. Near-shore sampling at shallow depths indicated that the biological removal of zinc occurs across the entire surface and is not limited to the central areas alone. The results further suggest that loadings from wall rocks and other inflows appear to be rapidly treated such that plumes do not appear to be formed within close proximity of the shores.

Surface water mass balance calculations for 2004 through 2009 suggest that zinc loadings to the Grum Pit are on the order of 3000 to 4000 kg per annum. The loading appears to be associated with predominantly with wall rock runoff, and to a lesser extent with solute release from wall rock as the water level rises. Net zinc removal has matched the zinc loading to the pit lake during the 2009 summer period. Contrary to previous years, the concentrations in the pit lake did not increase over

the 2008/09 winter season. Rather, a net removal of zinc occurred. The mechanism appears to be a result of sulphate reduction, with concurrent metal precipitation as a sulphide mineral.

Due to the removals depth weighted zinc concentration has decreased from in excess of 10 mg/L in 2005 to about 4.8 mg/L.

Mass balance calculations further showed that the net loading of sulphate to the pit lake is about 110 to 120 tonnes per year.

The leach extraction tests indicate that most of the lithological units release sulphate whereas most of the zinc release is associated with the sulphidic wall rocks. A potential correlation was established between the sulphate loadings from the wall rocks and the leach extraction test results. This correlation was extended to verify the potential annual zinc loadings to the pit lake and then used to estimate steady state concentrations in outflow from the pit lake when it spills naturally at an elevation of about 1230 m ASL. The results suggest that the steady state zinc concentration in the steady state outflow (without any treatment) could still remain at about 1 mg/L. This is primarily due to the sulphidic wall rock remaining above the lake level. By raising the lake elevation by about 20 m, the zinc concentrations could potentially be reduced to about 0.1 mg/L in the long term.

4.1.3 Vangorda Pit

The monitoring results for 2009 indicated a slight increase in the pH of the surface water due to spring freshet inflows. However, acidity loadings nullified this increase and the pH of the surface again decreased to values equivalent to the deeper water. The results further indicated that the entire water column mixed at the end of 2008, unlike in previous years no clear chemocline developed that separated the surface water from the deeper water. Since no significant increase in zinc concentrations occurred for the water at depth it is concluded that all of the treatment sludges that were deposited in the pit have now dissolved and any further increases in solute concentrations would be a result of loadings associated with surface inflows from wall rocks and in-pit waste rock dumps.

4.2 Recommendations

4.2.1 Faro Pit

It is recommended that profiling of the pit lake continue to be undertaken at regular intervals during 2010 and mass balance calculations be completed to verify estimates of overall current loadings and support future water quality predictions. It is further recommended that improved monitoring of the flows from the Faro Valley dump be undertaken to enable a better estimation of the total loading from this source so that the pit lake loadings can further be broken down to distinguish between waste rock and wall rock loadings. Improved monitoring of the Faro Valley dump flows would require that a regular monitoring station be established to allow weekly or fortnightly monitoring of flow and water quality.

4.2.2 Grum Pit

Based on the observed performance of biological treatment in the Grum Pit, it is recommended that the fertilization program be continued in 2010. Fertilization should again be terminated by mid-August to prevent the build-up of residual nutrients in the water column.

The higher than predicted loadings to the pit lake are due to the inundation of talus on the benches as the water level rises and loadings associated with the surface runoff from the wall rocks. These loadings however are decreasing. The removal of zinc appears to further be facilitated by sulphate reduction under the ice. It is therefore recommended that an under-ice monitoring event be undertaken to determine if hydrogen sulphide is being generated.

4.2.3 Vangorda Pit

At this time, it is not foreseen that the Vangorda Pit will be utilized as an ancillary treatment system. Fertilization of this pit lake is therefore not recommended. However, it is recommended that the water column water quality profile be monitored to verify the load calculations. Monitoring should be undertaken on a monthly basis, commencing at open water conditions, and should be continued to September 2010. We further recommend that in future, should additional transfers occur, the volumes and quality of water transferred should be determined to enable evaluation of the impacts on the pit lake water quality by this operational strategy.

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5 References

SRK Consulting, 2004, *Anvil Range Pit Lakes, Assessment of Post Closure Conditions*. Prepared for Deloitte and Touche Inc., January 2004.

SRK Consulting, 2005, *Anvil Range Pit Lakes, Evaluation of In Situ Treatment*. Prepared for Deloitte and Touche Inc., January 2005.

SRK Consulting, 2006, *Anvil Range Pit Lakes, Evaluation of In Situ Treatment*. Prepared for Deloitte and Touche Inc., July 2006.

SRK Consulting, 2007, *Anvil Range Pit Lakes, Evaluation of In Situ Treatment*. Prepared for Deloitte and Touche Inc., July 2007.

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Appendix A
In Situ Monitoring Results

Grum Pit Insitu Results

Date	time	Station	pH	Temp °C	Cond. μ S/cm	D.O. mg/L	Secchi (m)	Comments
17-Jun-09	14:00	GL-1	6.91	11.0	1027	6.2	2.0	Grum Lake
17-Jun-09	14:15	GL-2	7.60	7.5	999	6.1	2.9	Former "Fish" limno corral, now control
17-Jun-09	14:25	GL-5	7.60	5.0	1054	5.6		
17-Jun-09	14:40	GL-15	7.65	4.5	1050	5.5		
17-Jun-09	14:50	GL-30	7.67	4.5	1054	5.5		
17-Jun-09	15:00	GL-50	7.68	4.2	1053	5.4		
11-Aug-09	18:40	GL-1	7.45	13.3	1000		0.9	Grum Lake
11-Aug-09	19:05	GL-3	7.94	13.5	1100		4.5	Former "Fish" limno corral, now control
11-Aug-09	19:18	GL-5	7.91	13.5	1200			Total depth: 62 m
11-Aug-09	19:34	GL-15	7.77	7.1	1200			Chlorophyll samples collected at 1m, 3m and 5m
11-Aug-09	19:45	GL-30	7.76	5.6	1200			
11-Aug-09	19:57	GL-50	7.79	5.0	1200			CTD cast ON 18:17, OFF 18:24
12-Aug-09	10:20	GLS-1	8.53	13.8	852	7.4		
12-Aug-09	10:35	GLS-2	6.81	13.8	860	10.4		Flow at GLS-2: 100 mL/sec
12-Aug-09	10:53	GLS-3	8.83	13.8	859	7.5		
12-Aug-09	11:20	GLS-4	8.86	13.9	859	7.0		
12-Aug-09	11:03	GLS-5	8.79	13.6	859	6.9		
12-Aug-09	11:13	GLS-6	8.86	13.6	857	7.3		
12-Aug-09	11:30	GLS-7	8.88	13.7	866	8.8		

Faro Pit Insitu Results

Date	time	Station	pH	Temp °C	Cond. u S/c	D.O. mg/L	Secchi (m)	comments
16-Jun-09	17:40:00	FL-1	7.32	11.1	1163	8.6	4.5	Total depth 85 m at sampling site
16-Jun-09	17:55:00	FL-3	7.38	9.8	1168	6.9		Chlorophyll a sampled at 1, 3 and 5 m
16-Jun-09	18:20:00	FL-5	7.39	7.3	1252	6.0		CTD cast ON: 17:20, OFF: 17:32
16-Jun-09	18:35:00	FL-15	7.30	7.3	1302	6.0		
16-Jun-09	18:50:00	FL-30	7.15	4.5	1490	1.6		
16-Jun-09	19:05:00	FL-60	6.97	4.5	1522	1.3		
16-Jun-09	19:20:00	FL-80	6.86	4.5	1546	0.7		
12-Aug-09	13:49	FL-1	6.87	14.4	1243	5.2	7.8	Total depth 84 m at sampling site
12-Aug-09	13:52	FL-3	7.13	13.7	1246	5.3		Chlorophyll a sampled at 1, 3 and 5 m
12-Aug-09	14:00	FL-5	7.32	13.4	1244	5.4		CTD cast ON: 13:22, OFF: 13:32
12-Aug-09	14:12	FL-15	7.39	5.7	1279	5.6		
12-Aug-09	14:22	FL-30	7.32	5.3	1465	1.2		
12-Aug-09	14:36	FL-60	7.04	5.4	1495	1.1		
12-Aug-09	14:51	FL-80	6.95	6.0	1604	1.0		

Vangorda Pit Insitu Results

Date	time	Station	pH	Temp °C	Cond. u S/c	D.O. mg/L	Secchi (m)	comment
18-Jun-09	10:00	VL-1	3.54	12.5	1675	4.7	1.5	CTD cast ON: 09:18, OFF: 09:24
18-Jun-09	10:15	VL-3	3.44	10.4	1951	2.6		
18-Jun-09	10:30	VL-5	3.61	5.0	2200	2.3		
18-Jun-09	10:45	VL-15	3.85	5.0	2300	2.3		
18-Jun-09	11:00	VL-30	3.99	4.0	2340	1.0		
18-Jun-09	11:15	VL-40	4.02	3.8	2330	0.9		
12-Aug-09	16:35	VL-1	3.78	12.3	2304	4.4	1.2	CTD cast ON: 16:20 , OFF: 16:26
12-Aug-09	16:41	VL-3	4.00	6.9	2297	1.1		
12-Aug-09	16:56	VL-5	3.99	5.9	2329	1.1		
12-Aug-09	17:05	VL-15	3.95	5.8	2393	1.1		
12-Aug-09	17:12	VL-30	4.00	6.1	2416	1.0		Chlorophyll samples at 1m and 3m
12-Aug-09	17:22	VL-40	3.99	5.7	2431	1.1		

Appendix B
Pit lake Water Quality Monitoring Results

Grum Pit - 2005-2008 Monitoring Results

RESULTS OF ANALYSIS

Table with columns for Depth (m) and rows for Physical Tests, Dissolved Anions, Nutrients, Total Metals, and Dissolved Metals. Each row lists a parameter and its values across various depths and sampling dates.

Grum Pit - 2005 -2008 Monitoring Results
RESULTS OF ANALYSIS

Table with columns for Depth (m), Sample ID, Date Sampled, ALS Sample ID, Nature, Physical Tests (Hardness, Conductivity, ORP, pH, Total Suspended Solids), Dissolved Anions (Alkalinity, Sulphate, Chloride, Fluoride, Bromide), Nutrients (Ammonia Nitrogen, Nitrate Nitrogen, Nitrite Nitrogen, Dissolved ortho-Phosphate, Total Phosphate), Total Metals (Aluminum, Antimony, Arsenic, Barium, Beryllium, Bismuth, Boron, Cadmium, Calcium, Chromium, Cobalt, Copper, Iron, Lead, Lithium, Magnesium, Manganese, Molybdenum, Nickel, Phosphorus, Potassium, Selenium, Silicon, Silver, Sodium, Strontium, Thallium, Tin, Titanium, Uranium, Vanadium, Zinc), Dissolved Metals (Aluminum, Antimony, Arsenic, Barium, Beryllium, Bismuth, Boron, Cadmium, Calcium, Chromium, Cobalt, Copper, Iron, Lead, Lithium, Magnesium, Manganese, Molybdenum, Nickel, Phosphorus, Potassium, Selenium, Silicon, Silver, Sodium, Strontium, Thallium, Tin, Titanium, Uranium, Vanadium, Zinc), and Organic Parameters (Chlorophyll a (a,b)).

Grum Pit - 2005-2008 Monitoring Results
RESULTS OF ANALYSIS

Table with 24 columns (1-40) and 30 rows of data. Rows include: Depth (m), Sample ID, Date Sampled, ALS Sample ID, Nature, Physical Tests (Hardness, Conductivity, ORP, pH, Total Suspended Solids), Dissolved Anions (Alkalinity, Sulphate, Chloride, Fluoride, Bromide), Nutrients (Ammonia Nitrogen, Nitrate Nitrogen, Nitrite Nitrogen, Nitrite/Nitrate Nitrogen, Dissolved ortho-Phosphate, Total Phosphate), Total Metals (Aluminum, Antimony, Arsenic, Barium, Beryllium, Bismuth, Boron, Cadmium, Calcium, Chromium, Cobalt, Copper, Iron, Lead, Lithium, Magnesium, Manganese, Molybdenum, Nickel, Phosphorus, Potassium, Selenium, Silicon, Silver, Sodium, Strontium, Thallium, Tin, Titanium, Uranium, Vanadium, Zinc), Dissolved Metals (Aluminum, Antimony, Arsenic, Barium, Beryllium, Bismuth, Boron, Cadmium, Calcium, Chromium, Cobalt, Copper, Iron, Lead, Lithium, Magnesium, Manganese, Molybdenum, Nickel, Phosphorus, Potassium, Selenium, Silicon, Silver, Sodium, Strontium, Thallium, Tin, Titanium, Uranium, Vanadium, Zinc), and Organic Parameters (Chlorophyll a (a,b)).

Faro Pit - 2005 to 2009 Monitoring Results
RESULTS OF ANALYSIS

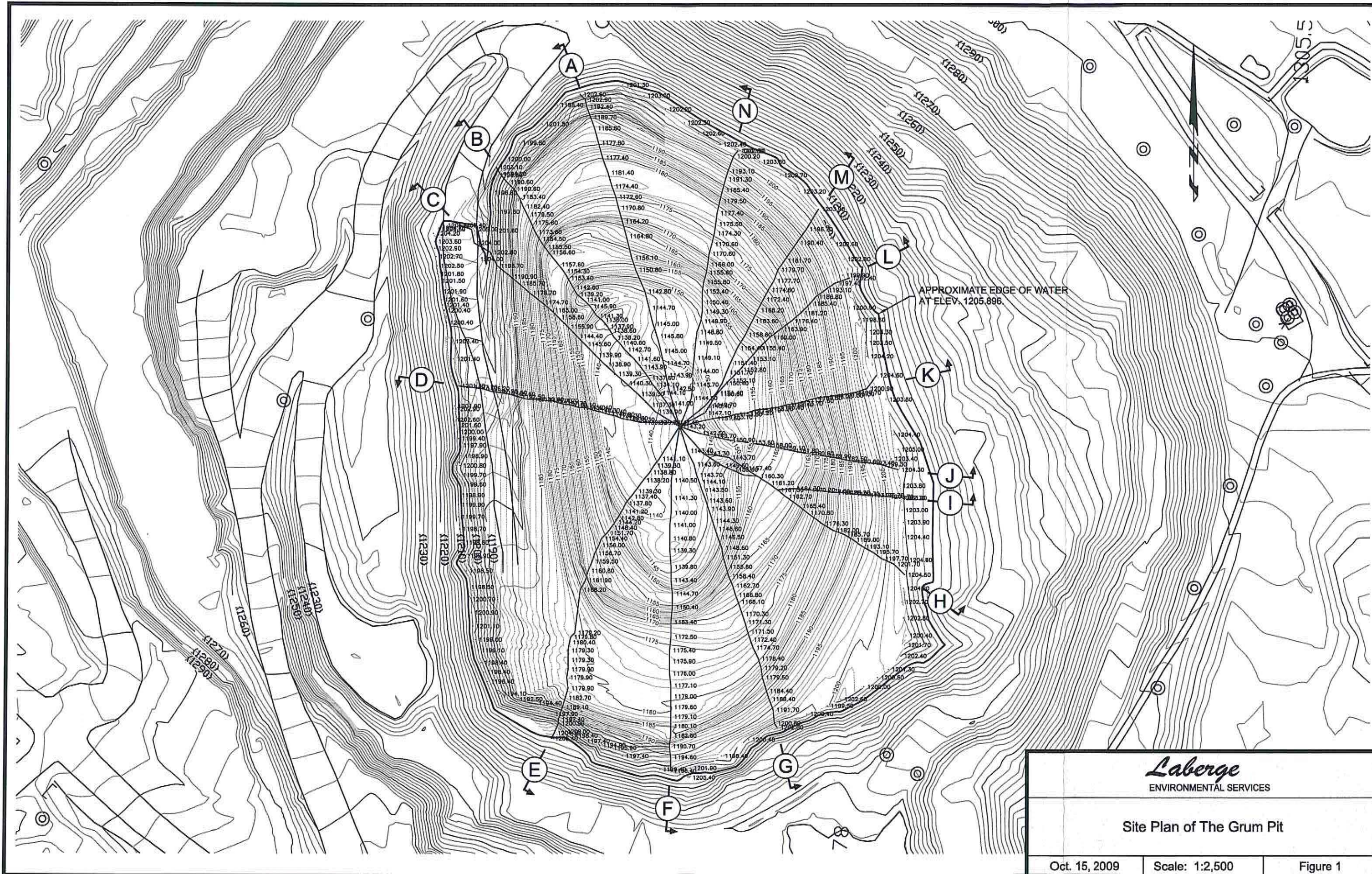
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Faro PII - 2005 to 2009 Monitoring Re
RESULTS OF ANALYSIS

Table with columns for Depth (m) and sample locations (1, 3, 5, 15, 30, 60, 80) for various parameters including Physical Tests, Dissolved Anions, Nutrients, Total Metals, and Dissolved Metals.

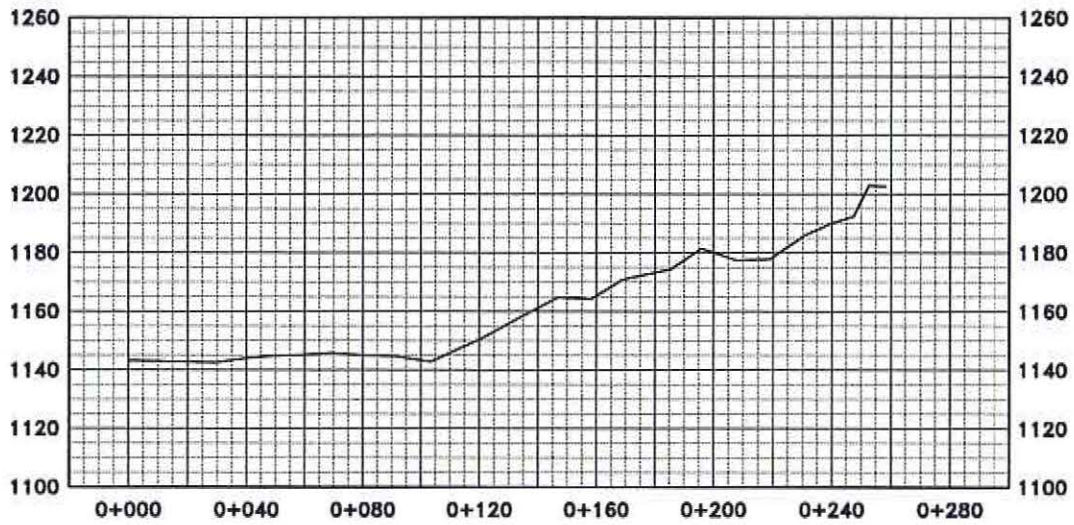
Faro Pit - 2005 to 2009 Monitoring Re
RESULTS OF ANALYSIS

Depth (m)	1							3							5							15							30							60							80						
	FL-1	FL-3	FL-5	FL-15	FL-30	FL-60	FL-80	FL-1	FL-3	FL-5	FL-15	FL-30	FL-60	FL-80	FL-1	FL-3	FL-5	FL-15	FL-30	FL-60	FL-80	FL-1	FL-3	FL-5	FL-15	FL-30	FL-60	FL-80	FL-1	FL-3	FL-5	FL-15	FL-30	FL-60	FL-80	FL-1	FL-3	FL-5	FL-15	FL-30	FL-60	FL-80							
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Titanium T-Ti	...																																																
Uranium T-U	...																																																
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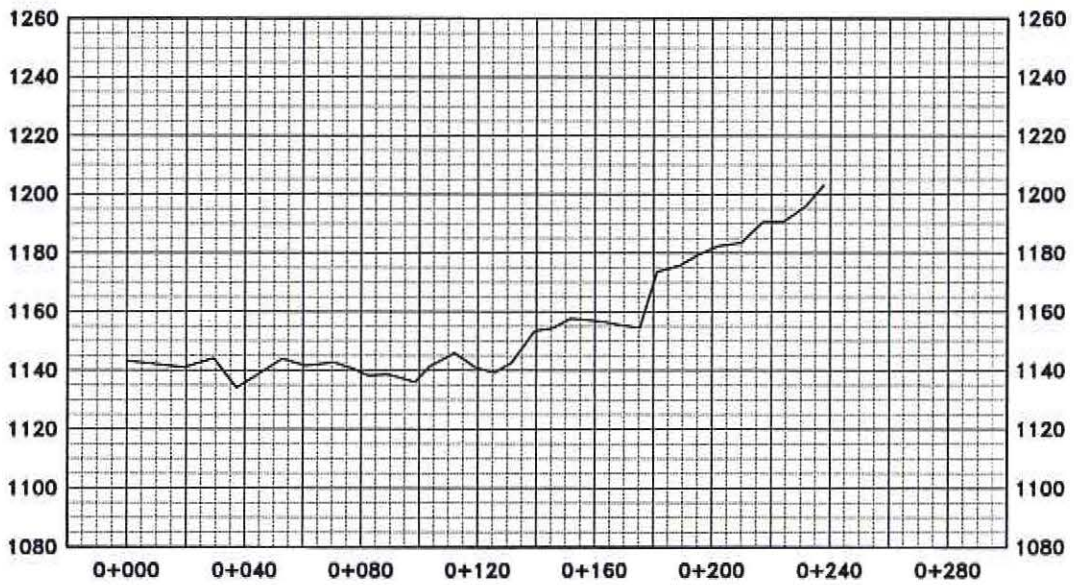


Site Plan of The Grum Pit		
Oct. 15, 2009	Scale: 1:2,500	Figure 1

Section A



Section B



Laberge
ENVIRONMENTAL SERVICES

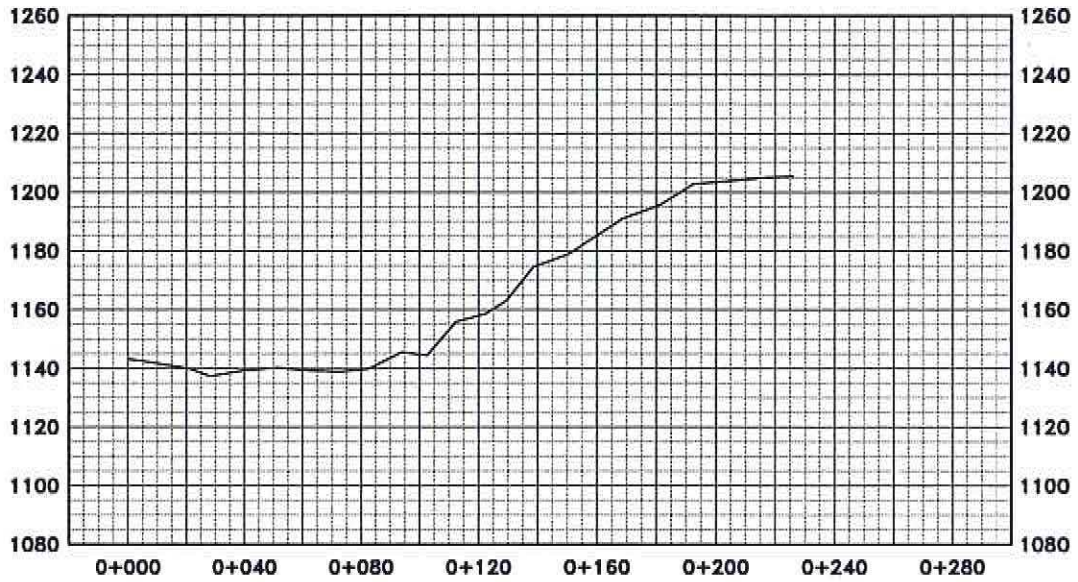
Grum Pit - Sections A & B

Oct. 15, 2009

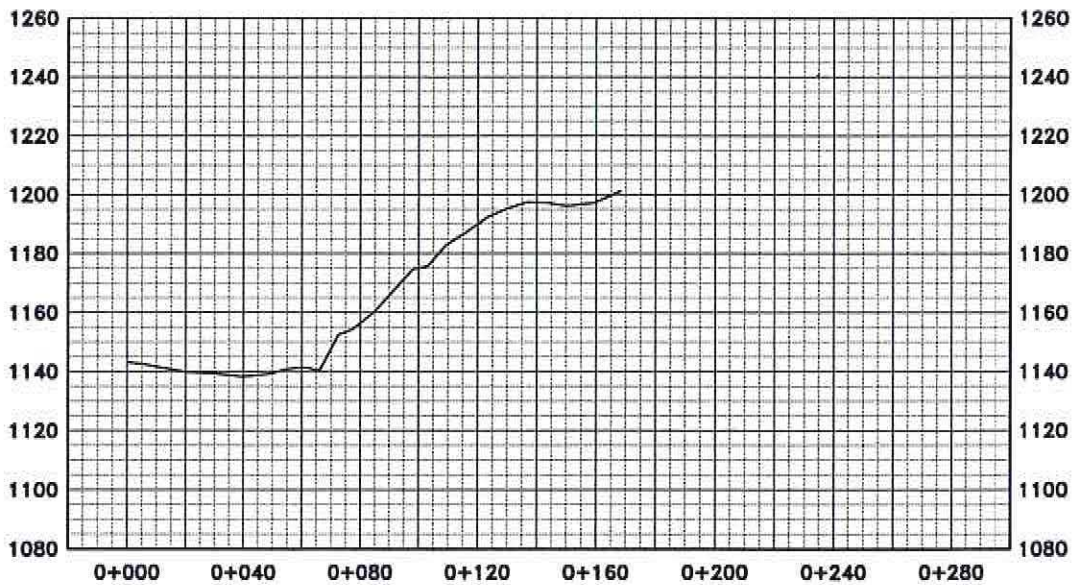
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Figure 2

Section C



Section D



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ENVIRONMENTAL SERVICES

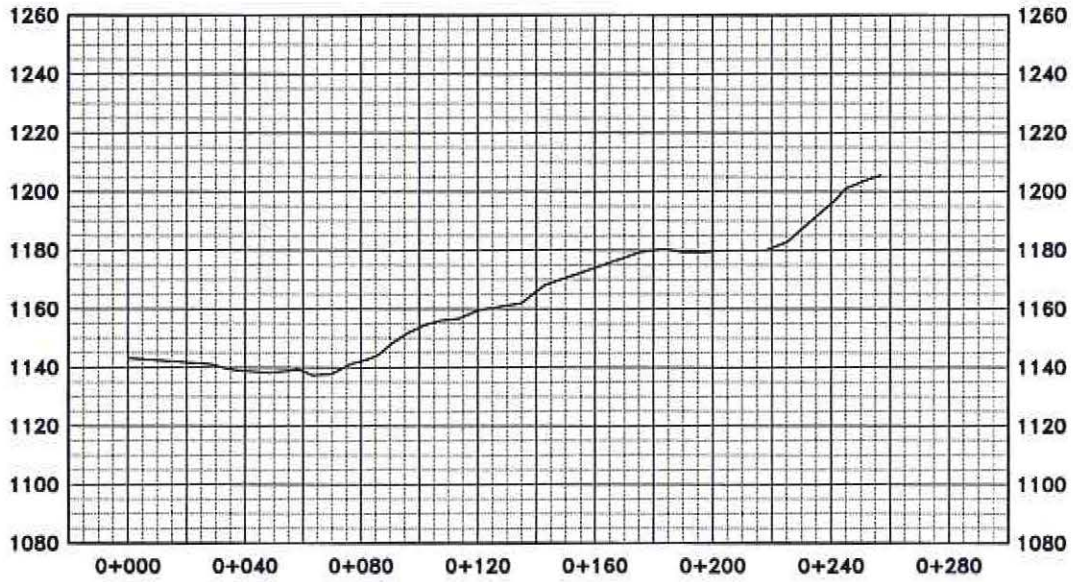
Grum Pit - Sections C & D

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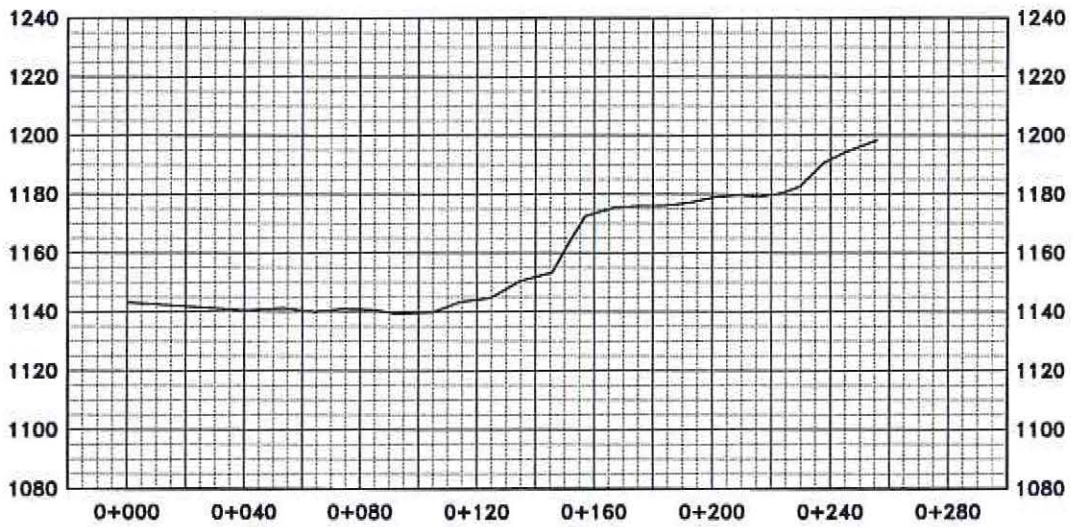
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Figure 3

Section E



Section F



Laberge
ENVIRONMENTAL SERVICES

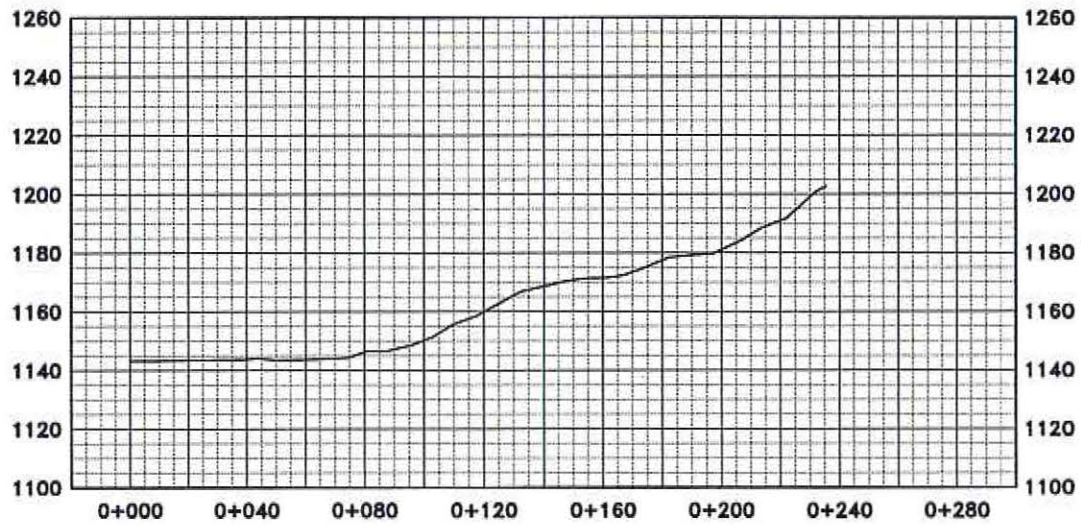
Grum Pit - Sections E & F

Oct. 15, 2009

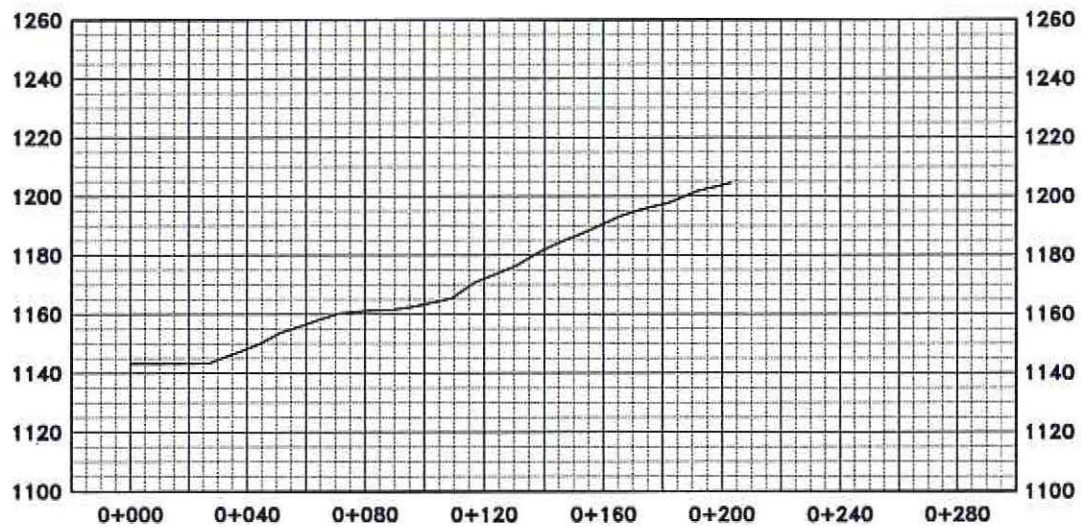
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Figure 4

Section G



Section H



Laberge

ENVIRONMENTAL SERVICES

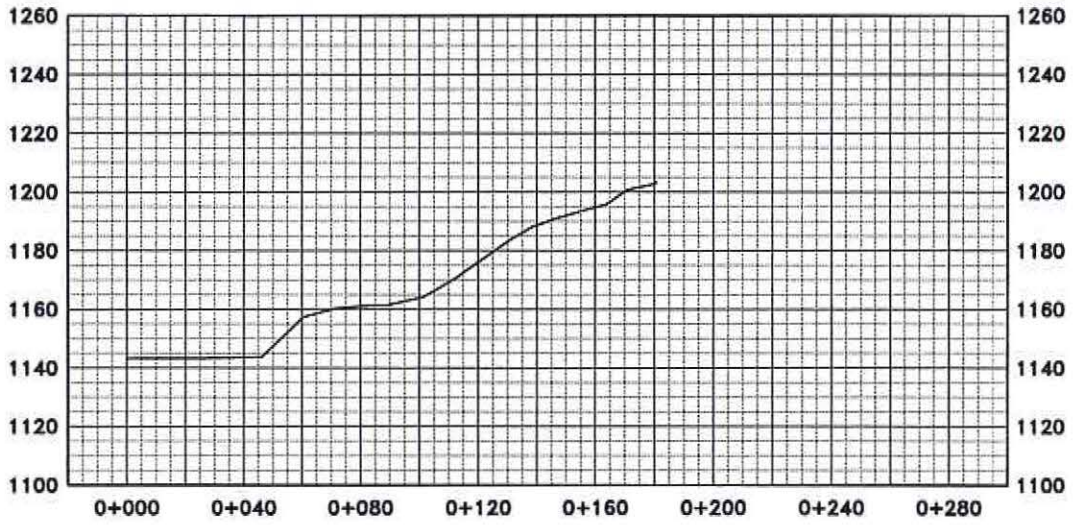
Grum Pit - Sections G & H

Oct. 15, 2009

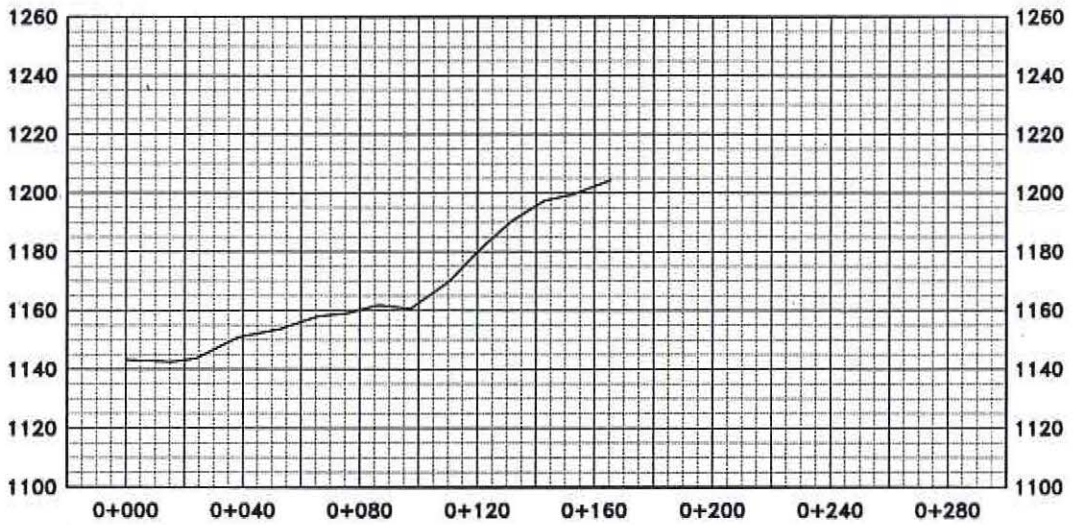
Scale: H,V 1:2,500

Figure 5

Section I



Section J



Laberge
ENVIRONMENTAL SERVICES

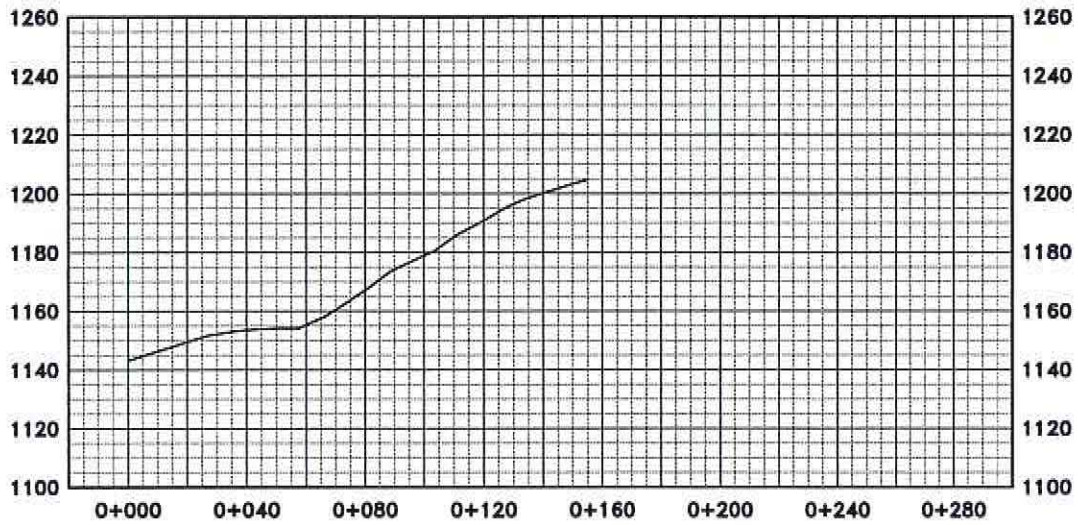
Grum Pit - Sections I & J

Oct. 15, 2009

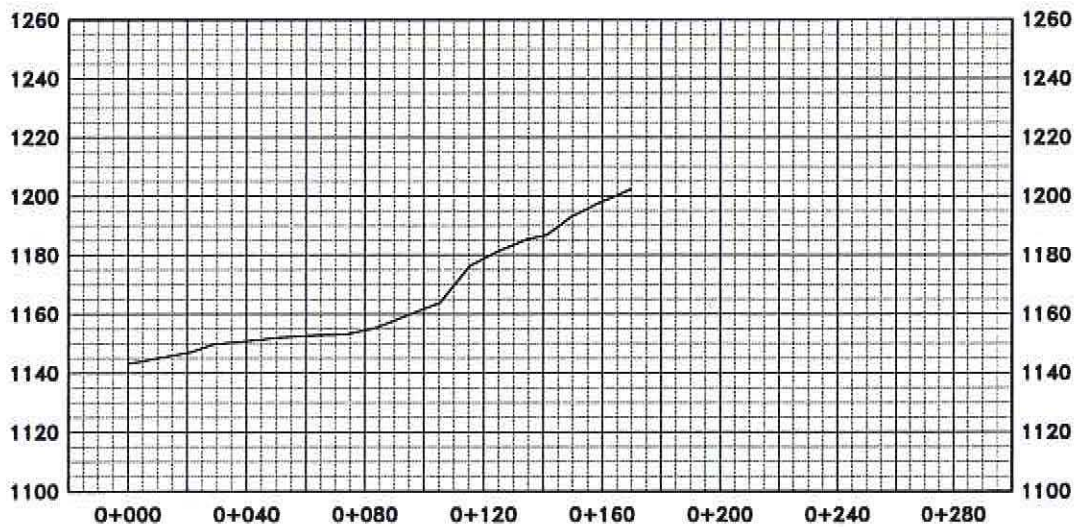
Scale: H,V 1:2,500

Figure 6

Section K



Section L



Laberge
ENVIRONMENTAL SERVICES

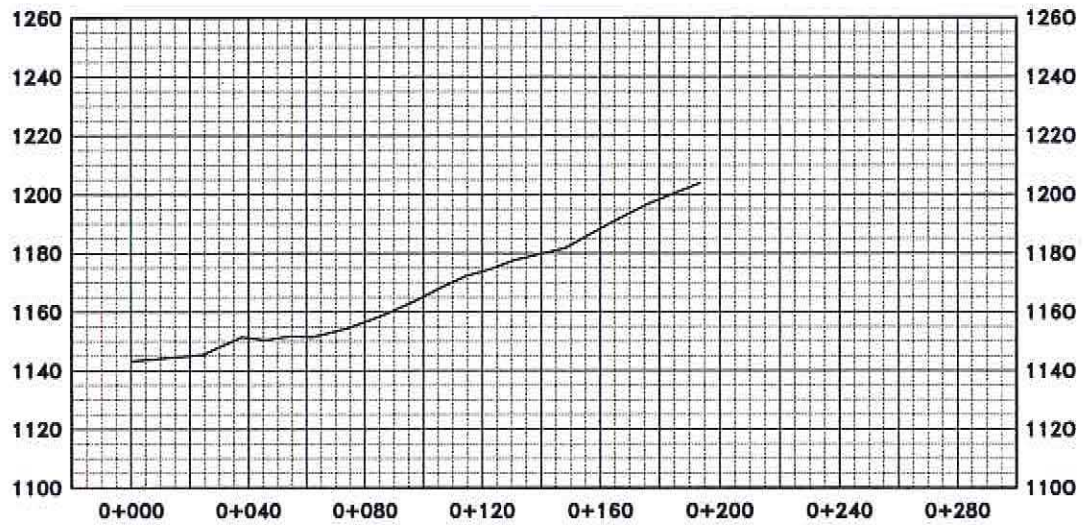
Grum Pit - Sections K & L

Oct. 15, 2009

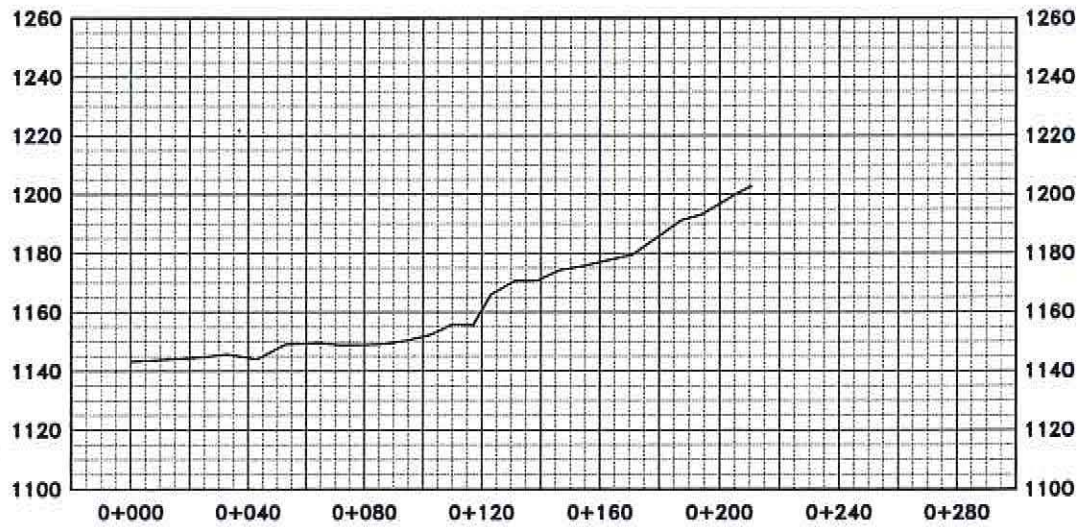
Scale: H,V 1:2,500

Figure 7

Section M



Section N



Laberge
ENVIRONMENTAL SERVICES

Grum Pit - Sections M & N

Oct. 15, 2009

Scale: H,V 1:2,500

Figure 8

Leachate Analysis

			LOD	3G0	3G0	3G0
Lithology						
Sample ID				GTTAL 03	GTTAL 27	GTTAL 28
Parameter	Method	Units	LOD			
Volume Nanopure water		mL	1	750	750	750
Sample Weight		g	0.01	250	250	250
pH	Meter		0.01	7.89	7.07	7.55
Redox	Meter	mV	1	325	339	332
Conductivity	Meter	uS/cm	1	241	2521	2378
Acidity (to pH 4.5)	Titration	mg CaCO3/L	0.1	#N/A	#N/A	#N/A
Total Acidity (to pH 8.3)	Titration	mg CaCO3/L	0.1	3.7	6.9	6.6
Alkalinity	Titration	mg CaCO3/L	0.1	64.9	15.5	24.3
Sulphate	Turbidimetry	mg/L	1	72	1858	1735
Ion Balance						
Major Anions	Calculation	meq/L	0.01	2.80	39.02	36.63
Major Cations	Calculation	meq/L	0.01	2.40	35.76	32.77
Difference	Calculation	meq/L	0.01	0.39	3.26	3.87
Balance (%)	Calculation	%	0.01	7.6%	4.4%	5.6%
Dissolved Metals						
Hardness CaCO3	Calculation	mg/L	0.5	116	1780	1630
Aluminum Al	ICP-MS	mg/L	0.0002	0.011	0.004	0.004
Antimony Sb	ICP-MS	mg/L	0.00002	0.00007	0.0001	0.0003
Arsenic As	ICP-MS	mg/L	0.00002	0.00009	0.0002	0.0002
Barium Ba	ICP-MS	mg/L	0.00002	0.0466	0.0201	0.0189
Beryllium Be	ICP-MS	mg/L	0.00001	<0.00001	<0.00005	<0.00005
Bismuth Bi	ICP-MS	mg/L	0.000005	<0.000005	<0.00003	<0.00003
Boron B	ICP-MS	mg/L	0.05	<0.05	<0.3	<0.3
Cadmium Cd	ICP-MS	mg/L	0.000005	0.000015	0.00087	0.00007
Calcium Ca	ICP-MS	mg/L	0.05	17.4	440	333
Chromium Cr	ICP-MS	mg/L	0.0001	<0.0001	<0.0005	<0.0005
Cobalt Co	ICP-MS	mg/L	0.000005	0.000235	0.00013	0.00142
Copper Cu	ICP-MS	mg/L	0.00005	0.00131	0.0006	0.001
Iron Fe	ICP-MS	mg/L	0.001	0.014	0.007	0.006
Lead Pb	ICP-MS	mg/L	0.000005	0.000482	0.00012	0.00009
Lithium Li	ICP-MS	mg/L	0.0005	0.0036	0.018	0.008
Magnesium Mg	ICP-MS	mg/L	0.05	17.7	166	195
Manganese Mn	ICP-MS	mg/L	0.00005	0.00916	0.0051	0.0446
Mercury Hg	ICP-MS	ug/L	0.01	<0.01	<0.05	<0.05
Molybdenum Mo	ICP-MS	mg/L	0.00005	0.00019	0.003	0.0044
Nickel Ni	ICP-MS	mg/L	0.00002	0.00058	0.0118	0.0194
Phosphorus P	ICP-MS	mg/L	0.002	<0.002	<0.01	<0.01
Potassium K	ICP-MS	mg/L	0.05	1.53	2.35	2.28
Selenium Se	ICP-MS	mg/L	0.00004	0.00008	0.0105	0.0109
Silicon Si	ICP-MS	mg/L	0.1	0.203	<0.5	<0.5
Silver Ag	ICP-MS	mg/L	0.000005	0.00003	<0.00003	<0.00003
Sodium Na	ICP-MS	mg/L	0.05	0.89	1.54	0.64
Strontium Sr	ICP-MS	mg/L	0.00005	0.0954	0.708	0.59
Sulphur (S)	ICP-MS	mg/L	3	24	614	559
Thallium Tl	ICP-MS	mg/L	0.000002	0.000234	0.00124	0.00076
Tin Sn	ICP-MS	mg/L	0.00001	<0.00001	<0.00005	<0.00005
Titanium Ti	ICP-MS	mg/L	0.0005	<0.0005	<0.003	<0.003
Uranium U	ICP-MS	mg/L	0.000002	0.000241	0.00087	0.00126
Vanadium V	ICP-MS	mg/L	0.0002	<0.0002	<0.001	<0.001
Zinc Zn	ICP-MS	mg/L	0.0001	0.0062	0.0179	0.0034
Zirconium Zr	ICP-MS	mg/L	0.0001	<0.0001	<0.0005	<0.0005

Leachate Analysis

Lithology		3G0	4EC	4EC	4EC
Sample ID		GTTAL 17	GTTAL 04	GTTAL 05	GTTAL 06
Parameter	Method				
Volume Nanopure water		750	750	750	750
Sample Weight		250	250	250	250
pH	Meter	7.72	2.52	2.90	5.52
Redox	Meter	319	508	416	259
Conductivity	Meter	930	3270	1924	528
Acidity (to pH 4.5)	Titration	#N/A	1225	192.3	#N/A
Total Acidity (to pH 8.3)	Titration	5.6	1850	975.0	174.9
Alkalinity	Titration	28.0	#N/A	#N/A	3.9
Sulphate	Turbidimetry	495	2298	1189	266
Ion Balance					
Major Anions	Calculation	10.87	47.88	24.77	5.62
Major Cations	Calculation	10.47	43.33	24.85	5.61
Difference	Calculation	0.41	4.55	-0.08	0.01
Balance (%)	Calculation	1.9%	5.0%	-0.2%	0.0%
Dissolved Metals					
Hardness CaCO3	Calculation	512	181	19.3	118
Aluminum Al	ICP-MS	0.0033	37.9	26.7	<0.02
Antimony Sb	ICP-MS	0.00044	0.0346	0.032	0.002
Arsenic As	ICP-MS	0.00052	1.89	0.53	0.003
Barium Ba	ICP-MS	0.0213	0.005	0.01	0.016
Beryllium Be	ICP-MS	<0.00001	0.0018	0.002	<0.001
Bismuth Bi	ICP-MS	<0.000005	<0.0001	<0.0005	<0.0005
Boron B	ICP-MS	<0.05	<1	<5	<5
Cadmium Cd	ICP-MS	0.000069	0.0057	0.484	0.115
Calcium Ca	ICP-MS	115	35	2	23
Chromium Cr	ICP-MS	<0.0001	0.12	0.075	<0.01
Cobalt Co	ICP-MS	0.000465	0.74	0.0942	0.0631
Copper Cu	ICP-MS	0.00075	0.907	8.26	0.033
Iron Fe	ICP-MS	0.003	596	144	0.132
Lead Pb	ICP-MS	0.00011	2.45	0.152	3.47
Lithium Li	ICP-MS	0.0221	0.06	<0.05	<0.05
Magnesium Mg	ICP-MS	54.5	22.7	3	15
Manganese Mn	ICP-MS	0.0118	1.34	4.03	2.04
Mercury Hg	ICP-MS	<0.01	0.8	<1	<1
Molybdenum Mo	ICP-MS	0.0012	0.013	<0.005	<0.005
Nickel Ni	ICP-MS	0.00139	0.303	0.133	0.298
Phosphorus P	ICP-MS	<0.002	0.993	<0.2	<0.2
Potassium K	ICP-MS	6.53	0.8	<1	<1
Selenium Se	ICP-MS	0.00048	0.0092	<0.004	<0.004
Silicon Si	ICP-MS	0.288	<2	<10	<10
Silver Ag	ICP-MS	<0.000005	0.0013	0.0007	<0.0005
Sodium Na	ICP-MS	1.75	0.5	<1	<1
Strontium Sr	ICP-MS	0.516	0.06	0.017	0.045
Sulphur (S)	ICP-MS	169	681	410	157
Thallium Tl	ICP-MS	0.000296	0.0281	0.0008	0.159
Tin Sn	ICP-MS	0.00004	0.0008	<0.001	<0.001
Titanium Ti	ICP-MS	<0.0005	0.22	<0.05	<0.05
Uranium U	ICP-MS	0.000856	0.0098	0.0299	<0.0002
Vanadium V	ICP-MS	<0.0002	0.049	<0.02	<0.02
Zinc Zn	ICP-MS	0.0075	10.5	396	102
Zirconium Zr	ICP-MS	<0.0001	0.015	<0.01	<0.01

Leachate Analysis

		4EC	4EC	4EC	4EC
Lithology		GTTAL 07	GTTAL 08	GTTAL 22	GTTAL 23
Sample ID	Method				
Parameter	Method				
Volume Nanopure water		750	750	750	750
Sample Weight		250	250	250	250
pH	Meter	7.27	3.51	2.74	6.77
Redox	Meter	247	413	509	356
Conductivity	Meter	630	1752	4100	1257
Acidity (to pH 4.5)	Titration	#N/A	33.3	275.0	#N/A
Total Acidity (to pH 8.3)	Titration	82.7	348.8	1150	28.5
Alkalinity	Titration	13.8	#N/A	#N/A	18.6
Sulphate	Turbidimetry	322	1042	2912	796
Ion Balance					
Major Anions	Calculation	6.98	21.71	60.67	16.96
Major Cations	Calculation	6.73	21.05	52.80	14.43
Difference	Calculation	0.25	0.66	7.87	2.53
Balance (%)	Calculation	1.8%	1.5%	6.9%	8.1%
Dissolved Metals					
Hardness CaCO3	Calculation	263	691	1760	700
Aluminum Al	ICP-MS	<0.01	4.59	5.66	0.002
Antimony Sb	ICP-MS	0.001	0.0006	0.002	0.0025
Arsenic As	ICP-MS	<0.001	0.0163	0.0692	0.0001
Barium Ba	ICP-MS	0.028	0.0158	0.0177	0.0287
Beryllium Be	ICP-MS	<0.0005	0.0033	0.0197	<0.00005
Bismuth Bi	ICP-MS	<0.0003	<0.0001	<0.0001	<0.00003
Boron B	ICP-MS	<3	<1	<1	<0.3
Cadmium Cd	ICP-MS	0.317	0.191	0.125	0.0497
Calcium Ca	ICP-MS	90.5	166	445	217
Chromium Cr	ICP-MS	<0.005	0.003	0.048	<0.0005
Cobalt Co	ICP-MS	0.0299	0.461	0.575	0.00803
Copper Cu	ICP-MS	0.011	37.5	2.9	0.0007
Iron Fe	ICP-MS	<0.05	19.4	172	0.006
Lead Pb	ICP-MS	1.32	0.25	0.657	0.188
Lithium Li	ICP-MS	<0.03	0.014	<0.01	0.005
Magnesium Mg	ICP-MS	8.9	66.9	158	38.5
Manganese Mn	ICP-MS	0.952	40	17.5	0.16
Mercury Hg	ICP-MS	<0.5	<0.2	<0.2	<0.05
Molybdenum Mo	ICP-MS	<0.003	<0.001	0.001	0.0006
Nickel Ni	ICP-MS	0.042	0.322	4.1	0.0152
Phosphorus P	ICP-MS	<0.1	<0.04	0.105	<0.01
Potassium K	ICP-MS	<0.5	<0.2	<0.2	1.23
Selenium Se	ICP-MS	<0.002	0.0028	0.0062	0.0006
Silicon Si	ICP-MS	<5	4.03	<2	<0.5
Silver Ag	ICP-MS	<0.0003	0.0003	0.0004	0.00003
Sodium Na	ICP-MS	<0.5	<0.2	<0.2	0.49
Strontium Sr	ICP-MS	0.203	0.349	0.623	0.209
Sulphur (S)	ICP-MS	110	366	939	254
Thallium Tl	ICP-MS	0.0563	0.00339	0.00141	0.00924
Tin Sn	ICP-MS	<0.0005	<0.0002	<0.0002	<0.00005
Titanium Ti	ICP-MS	<0.03	<0.01	<0.01	<0.003
Uranium U	ICP-MS	<0.0001	0.0577	0.383	0.00003
Vanadium V	ICP-MS	<0.01	<0.004	<0.004	<0.001
Zinc Zn	ICP-MS	46.9	90.1	164	12.1
Zirconium Zr	ICP-MS	<0.005	<0.002	<0.002	<0.0005

Leachate Analysis

Lithology		4EC	4EC	4EC	4EC
Sample ID		GTTAL 25	GTTAL 26	GTTAL 35	GTTAL 36
Parameter	Method				
Volume Nanopure water		750	750	750	750
Sample Weight		250	250	250	250
pH	Meter	3.92	8.1	6.45	7.35
Redox	Meter	459	246	373	339
Conductivity	Meter	3410	1354	1734	1013
Acidity (to pH 4.5)	Titration	17.2	#N/A	#N/A	#N/A
Total Acidity (to pH 8.3)	Titration	1350	7.9	167.3	78.6
Alkalinity	Titration	#N/A	62.4	7.8	22.1
Sulphate	Turbidimetry	2605	845	1214	576
Ion Balance					
Major Anions	Calculation	54.27	18.85	25.45	12.44
Major Cations	Calculation	53.36	16.95	22.35	11.79
Difference	Calculation	0.91	1.90	3.10	0.65
Balance (%)	Calculation	0.8%	5.3%	6.5%	2.7%
Dissolved Metals					
Hardness CaCO3	Calculation	1420	843	953	521
Aluminum Al	ICP-MS	0.839	0.0022	0.009	<0.002
Antimony Sb	ICP-MS	0.012	0.00041	<0.0004	0.0031
Arsenic As	ICP-MS	0.293	0.00025	0.0066	0.001
Barium Ba	ICP-MS	0.002	0.0235	0.0219	0.0344
Beryllium Be	ICP-MS	0.003	<0.00001	<0.0002	<0.0001
Bismuth Bi	ICP-MS	<0.0005	<0.000005	<0.0001	<0.00005
Boron B	ICP-MS	<5	<0.05	<1	<0.5
Cadmium Cd	ICP-MS	1.04	0.000739	0.158	0.0911
Calcium Ca	ICP-MS	488	219	224	152
Chromium Cr	ICP-MS	<0.01	<0.0001	<0.002	<0.001
Cobalt Co	ICP-MS	0.35	0.0015	0.191	0.0303
Copper Cu	ICP-MS	2.29	0.00023	0.005	0.0026
Iron Fe	ICP-MS	1.69	0.005	<0.02	<0.01
Lead Pb	ICP-MS	0.249	0.00225	1.72	0.991
Lithium Li	ICP-MS	<0.05	0.0082	0.011	<0.005
Magnesium Mg	ICP-MS	49	72.1	95.4	34.2
Manganese Mn	ICP-MS	1.75	0.101	6.82	0.475
Mercury Hg	ICP-MS	<1	<0.01	<0.2	<0.1
Molybdenum Mo	ICP-MS	<0.005	0.0214	<0.001	<0.0005
Nickel Ni	ICP-MS	0.512	0.00541	0.491	0.0781
Phosphorus P	ICP-MS	<0.2	<0.002	<0.04	<0.02
Potassium K	ICP-MS	<1	1.76	1.1	1.57
Selenium Se	ICP-MS	<0.004	0.00443	0.0153	0.0007
Silicon Si	ICP-MS	<10	0.347	<2	<1
Silver Ag	ICP-MS	0.0007	<0.000005	0.0001	0.00011
Sodium Na	ICP-MS	<1	0.66	0.4	0.18
Strontium Sr	ICP-MS	0.196	0.424	0.102	0.113
Sulphur (S)	ICP-MS	975	295	378	186
Thallium Tl	ICP-MS	0.0042	0.00174	0.00323	0.00247
Tin Sn	ICP-MS	<0.001	<0.00001	<0.0002	<0.0001
Titanium Ti	ICP-MS	<0.05	<0.0005	<0.01	<0.005
Uranium U	ICP-MS	0.0314	0.00516	0.00008	<0.00002
Vanadium V	ICP-MS	<0.02	<0.0002	<0.004	<0.002
Zinc Zn	ICP-MS	800	0.115	98.2	43.1
Zirconium Zr	ICP-MS	<0.01	<0.0001	<0.002	<0.001

Leachate Analysis

Lithology		4EC	5A0/5B0	5A0/5B0	5A0/5B0
Sample ID		GTTAL 37	GTTAL 12	GTTAL 09	GTTAL 10
Parameter	Method				
Volume Nanopure water		750	750	750	750
Sample Weight		250	250	250	250
pH	Meter	6.89	7.86	8.08	7.94
Redox	Meter	360	275	228	259
Conductivity	Meter	669	232	196	265
Acidity (to pH 4.5)	Titration	#N/A	#N/A	#N/A	#N/A
Total Acidity (to pH 8.3)	Titration	234.5	4.6	5.0	4.9
Alkalinity	Titration	8.7	51.5	79.0	78.6
Sulphate	Turbidimetry	355	72	37	71
Ion Balance					
Major Anions	Calculation	7.57	2.53	2.35	3.05
Major Cations	Calculation	7.85	2.21	1.94	2.69
Difference	Calculation	-0.28	0.32	0.41	0.36
Balance (%)	Calculation	-1.8%	6.8%	9.6%	6.3%
Dissolved Metals					
Hardness CaCO3	Calculation	139	108	94.7	130
Aluminum Al	ICP-MS	<0.004	0.0113	0.0159	0.0089
Antimony Sb	ICP-MS	0.0024	0.00045	0.00013	0.00053
Arsenic As	ICP-MS	0.0039	0.00057	0.00022	0.00025
Barium Ba	ICP-MS	0.0269	0.0324	0.0515	0.0473
Beryllium Be	ICP-MS	<0.0002	<0.00001	<0.00001	<0.00001
Bismuth Bi	ICP-MS	<0.0001	<0.000005	<0.000005	<0.000005
Boron B	ICP-MS	<1	<0.05	<0.05	<0.05
Cadmium Cd	ICP-MS	0.112	0.000069	0.000009	0.000011
Calcium Ca	ICP-MS	39.3	11.6	14	22.4
Chromium Cr	ICP-MS	<0.002	<0.0001	0.0002	0.0001
Cobalt Co	ICP-MS	0.0774	0.000435	0.000087	0.000198
Copper Cu	ICP-MS	0.013	0.00057	0.00059	0.00025
Iron Fe	ICP-MS	<0.02	0.008	0.006	0.003
Lead Pb	ICP-MS	2.3	0.00154	0.000098	0.000065
Lithium Li	ICP-MS	<0.01	0.0086	0.0051	0.0119
Magnesium Mg	ICP-MS	9.9	19.2	14.5	18
Manganese Mn	ICP-MS	0.612	0.01	0.00463	0.00428
Mercury Hg	ICP-MS	<0.2	0.02	<0.01	<0.01
Molybdenum Mo	ICP-MS	<0.001	0.0376	0.00872	0.00161
Nickel Ni	ICP-MS	0.454	0.0025	0.00038	0.00057
Phosphorus P	ICP-MS	<0.04	<0.002	<0.002	<0.002
Potassium K	ICP-MS	0.7	1.2	1.29	2.34
Selenium Se	ICP-MS	<0.0008	0.00372	0.00049	0.00042
Silicon Si	ICP-MS	<2	0.304	0.272	0.251
Silver Ag	ICP-MS	0.0001	<0.000005	0.000011	<0.000005
Sodium Na	ICP-MS	<0.2	0.46	0.34	0.66
Strontium Sr	ICP-MS	0.152	0.0598	0.0728	0.107
Sulphur (S)	ICP-MS	120	24	12	24
Thallium Tl	ICP-MS	0.0152	0.00063	0.000046	0.000041
Tin Sn	ICP-MS	<0.0002	0.00007	<0.00001	<0.00001
Titanium Ti	ICP-MS	<0.01	<0.0005	<0.0005	<0.0005
Uranium U	ICP-MS	0.00008	0.000241	0.000375	0.000726
Vanadium V	ICP-MS	<0.004	<0.0002	<0.0002	<0.0002
Zinc Zn	ICP-MS	163	0.0101	0.0014	0.0013
Zirconium Zr	ICP-MS	<0.002	<0.0001	0.0002	<0.0001

Leachate Analysis

Lithology		5A0/5B0	5A0/5B0	5A0/5B0	5A0/5B0
Sample ID		GTTAL 13	GTTAL 14	GTTAL 15	GTTAL 16
Parameter	Method				
Volume Nanopure water		750	750	750	750
Sample Weight		250	250	250	250
pH	Meter	8.00	7.98	7.99	7.47
Redox	Meter	283	287	297	320
Conductivity	Meter	144	155	250	1161
Acidity (to pH 4.5)	Titration	#N/A	#N/A	#N/A	#N/A
Total Acidity (to pH 8.3)	Titration	4.7	4.1	4.0	6.2
Alkalinity	Titration	74.7	70.6	65.8	30.2
Sulphate	Turbidimetry	10	23	70	695
Ion Balance					
Major Anions	Calculation	1.70	1.89	2.77	15.08
Major Cations	Calculation	1.46	1.60	2.50	13.67
Difference	Calculation	0.24	0.30	0.27	1.41
Balance (%)	Calculation	7.5%	8.5%	5.1%	4.9%
Dissolved Metals					
Hardness CaCO3	Calculation	70.5	78	118	674
Aluminum Al	ICP-MS	0.0524	0.0203	0.0133	0.0033
Antimony Sb	ICP-MS	0.0003	0.00024	0.00057	0.00248
Arsenic As	ICP-MS	0.00039	0.00029	0.00104	0.00262
Barium Ba	ICP-MS	0.0426	0.0332	0.0422	0.02
Beryllium Be	ICP-MS	<0.00001	<0.00001	<0.00001	<0.00001
Bismuth Bi	ICP-MS	<0.000005	<0.000005	<0.000005	<0.000005
Boron B	ICP-MS	<0.05	<0.05	<0.05	<0.05
Cadmium Cd	ICP-MS	0.000016	0.000008	0.000009	0.00341
Calcium Ca	ICP-MS	21.2	10.7	18.8	125
Chromium Cr	ICP-MS	0.0002	0.0002	0.0002	<0.0001
Cobalt Co	ICP-MS	0.000035	0.000112	0.000125	0.00286
Copper Cu	ICP-MS	0.00054	0.00025	0.00042	0.00088
Iron Fe	ICP-MS	0.057	0.01	0.004	0.005
Lead Pb	ICP-MS	0.000182	0.000545	0.000142	0.00642
Lithium Li	ICP-MS	0.0011	0.0065	0.0058	0.0194
Magnesium Mg	ICP-MS	4.27	12.4	17.2	87.8
Manganese Mn	ICP-MS	0.00094	0.00437	0.00314	0.0241
Mercury Hg	ICP-MS	<0.01	<0.01	<0.01	<0.01
Molybdenum Mo	ICP-MS	0.00224	0.00051	0.0228	0.0125
Nickel Ni	ICP-MS	0.00054	0.00059	0.00127	0.00328
Phosphorus P	ICP-MS	0.002	<0.002	<0.002	<0.002
Potassium K	ICP-MS	1.4	1.07	5.02	3.37
Selenium Se	ICP-MS	0.00013	0.00015	0.00087	0.00186
Silicon Si	ICP-MS	1.49	0.373	0.283	0.402
Silver Ag	ICP-MS	<0.000005	<0.000005	<0.000005	0.00002
Sodium Na	ICP-MS	0.45	0.31	0.46	2.04
Strontium Sr	ICP-MS	0.0631	0.0392	0.0536	0.617
Sulphur (S)	ICP-MS	3	7	23	222
Thallium Tl	ICP-MS	0.000049	0.000078	0.000136	0.000521
Tin Sn	ICP-MS	0.00007	<0.00001	0.00004	<0.00001
Titanium Ti	ICP-MS	0.0011	<0.0005	<0.0005	<0.0005
Uranium U	ICP-MS	0.000926	0.000214	0.00124	0.000495
Vanadium V	ICP-MS	<0.0002	<0.0002	<0.0002	<0.0002
Zinc Zn	ICP-MS	0.0017	0.0068	0.0011	0.611
Zirconium Zr	ICP-MS	<0.0001	<0.0001	<0.0001	<0.0001

Leachate Analysis

Lithology		5A0/5B0	5A0/5B0	5A0/5B0	5A0/5B0
Sample ID		GTTAL 19	GTTAL 20	GTTAL 21	GTTAL 29
Parameter	Method				
Volume Nanopure water		750	750	750	750
Sample Weight		250	250	250	250
pH	Meter	7.65	7.47	7.78	7.69
Redox	Meter	318	336	324	328
Conductivity	Meter	242	1470	542	1359
Acidity (to pH 4.5)	Titration	#N/A	#N/A	#N/A	#N/A
Total Acidity (to pH 8.3)	Titration	5.1	6.5	5.7	6.4
Alkalinity	Titration	64.8	27.7	45.5	40.1
Sulphate	Turbidimetry	63	943	254	845
Ion Balance					
Major Anions	Calculation	2.61	20.20	6.20	18.41
Major Cations	Calculation	2.36	18.54	5.47	16.77
Difference	Calculation	0.24	1.66	0.73	1.64
Balance (%)	Calculation	4.9%	4.3%	6.2%	4.6%
Dissolved Metals					
Hardness CaCO3	Calculation	114	919	262	832
Aluminum Al	ICP-MS	0.0143	0.0035	0.0055	0.008
Antimony Sb	ICP-MS	0.00073	0.00009	0.00035	<0.0001
Arsenic As	ICP-MS	0.00319	0.00007	0.00025	0.0005
Barium Ba	ICP-MS	0.0164	0.0238	0.0198	0.0193
Beryllium Be	ICP-MS	<0.00001	<0.00001	<0.00001	<0.00005
Bismuth Bi	ICP-MS	<0.000005	<0.000005	<0.000005	<0.00003
Boron B	ICP-MS	<0.05	<0.05	<0.05	<0.3
Cadmium Cd	ICP-MS	0.000011	0.000339	0.000059	0.00007
Calcium Ca	ICP-MS	21.5	245	49.1	217
Chromium Cr	ICP-MS	<0.0001	<0.0001	<0.0001	<0.0005
Cobalt Co	ICP-MS	0.00034	0.000346	0.000028	0.00108
Copper Cu	ICP-MS	0.00045	0.00034	0.00032	0.0025
Iron Fe	ICP-MS	0.003	0.006	0.003	<0.005
Lead Pb	ICP-MS	0.000073	0.000041	0.000053	0.00014
Lithium Li	ICP-MS	0.0121	0.0202	0.0121	0.005
Magnesium Mg	ICP-MS	14.7	74.7	34	70.5
Manganese Mn	ICP-MS	0.00508	0.0275	0.00087	0.0251
Mercury Hg	ICP-MS	<0.01	<0.01	<0.01	<0.05
Molybdenum Mo	ICP-MS	0.00119	0.00503	0.0135	0.0061
Nickel Ni	ICP-MS	0.00108	0.00281	0.00117	0.0094
Phosphorus P	ICP-MS	<0.002	<0.002	<0.002	<0.01
Potassium K	ICP-MS	2	3.51	7.5	2.35
Selenium Se	ICP-MS	0.00034	0.00641	0.00043	0.0115
Silicon Si	ICP-MS	0.509	0.285	0.372	<0.5
Silver Ag	ICP-MS	0.000007	0.000006	0.000008	<0.00003
Sodium Na	ICP-MS	0.67	1.18	0.66	1.5
Strontium Sr	ICP-MS	0.16	0.922	0.13	0.662
Sulphur (S)	ICP-MS	23	335	85	282
Thallium Tl	ICP-MS	0.000089	0.000064	0.000177	0.00006
Tin Sn	ICP-MS	<0.00001	0.00001	<0.00001	<0.00005
Titanium Ti	ICP-MS	<0.0005	<0.0005	<0.0005	<0.003
Uranium U	ICP-MS	0.000607	0.000734	0.000831	0.00346
Vanadium V	ICP-MS	<0.0002	<0.0002	<0.0002	<0.001
Zinc Zn	ICP-MS	0.0011	0.0151	0.0043	0.0165
Zirconium Zr	ICP-MS	<0.0001	<0.0001	<0.0001	<0.0005

Leachate Analysis

Lithology		5A0/5B0	5A0/5B0	5A0/5B0	5A0/5B0
Sample ID		GTTAL 30	GTTAL 31	GTTAL 32	GTTAL 33
Parameter	Method				
Volume Nanopure water		750	750	750	750
Sample Weight		250	250	250	250
pH	Meter	7.97	7.78	8.00	8.07
Redox	Meter	311	319	307	302
Conductivity	Meter	216	573	309	265
Acidity (to pH 4.5)	Titration	#N/A	#N/A	#N/A	#N/A
Total Acidity (to pH 8.3)	Titration	4.6	5.4	4.7	4.2
Alkalinity	Titration	65.1	53.6	73.9	96.7
Sulphate	Turbidimetry	51	279	99	61
Ion Balance					
Major Anions	Calculation	2.36	6.88	3.54	3.20
Major Cations	Calculation	2.21	5.92	3.30	2.71
Difference	Calculation	0.15	0.97	0.25	0.50
Balance (%)	Calculation	3.3%	7.6%	3.6%	8.4%
Dissolved Metals					
Hardness CaCO3	Calculation	107	290	160	133
Aluminum Al	ICP-MS	0.0165	0.0078	0.0092	0.0095
Antimony Sb	ICP-MS	0.00062	0.00021	0.00041	0.00013
Arsenic As	ICP-MS	0.00292	0.00013	0.00072	0.00037
Barium Ba	ICP-MS	0.029	0.0167	0.0246	0.0259
Beryllium Be	ICP-MS	<0.00001	<0.00001	<0.00001	<0.00001
Bismuth Bi	ICP-MS	<0.000005	<0.000005	<0.000005	<0.000005
Boron B	ICP-MS	<0.05	<0.05	<0.05	<0.05
Cadmium Cd	ICP-MS	0.000008	0.00001	0.000005	<0.000005
Calcium Ca	ICP-MS	31.9	60.4	28.6	18.5
Chromium Cr	ICP-MS	<0.0001	<0.0001	<0.0001	<0.0001
Cobalt Co	ICP-MS	0.000541	0.000494	0.00041	0.000074
Copper Cu	ICP-MS	0.00268	0.00068	0.00044	0.00081
Iron Fe	ICP-MS	0.005	0.002	0.002	0.002
Lead Pb	ICP-MS	0.000075	0.000035	0.000064	0.00015
Lithium Li	ICP-MS	0.0046	0.0064	0.0087	0.0042
Magnesium Mg	ICP-MS	6.56	33.9	21.6	21
Manganese Mn	ICP-MS	0.00703	0.0154	0.0041	0.00089
Mercury Hg	ICP-MS	<0.01	0.17	0.04	0.01
Molybdenum Mo	ICP-MS	0.0016	0.0015	0.00126	0.00189
Nickel Ni	ICP-MS	0.00122	0.00183	0.00143	0.0006
Phosphorus P	ICP-MS	<0.002	<0.002	<0.002	<0.002
Potassium K	ICP-MS	1.73	2.93	2.51	1.52
Selenium Se	ICP-MS	0.00018	0.00044	0.00034	0.00082
Silicon Si	ICP-MS	0.643	0.466	0.302	0.409
Silver Ag	ICP-MS	<0.000005	0.00005	0.00001	<0.000005
Sodium Na	ICP-MS	0.72	0.76	0.61	0.36
Strontium Sr	ICP-MS	0.235	0.203	0.0998	0.0771
Sulphur (S)	ICP-MS	19	87	32	20
Thallium Tl	ICP-MS	0.000015	0.000049	0.000045	0.000201
Tin Sn	ICP-MS	<0.00001	<0.00001	<0.00001	<0.00001
Titanium Ti	ICP-MS	0.0011	<0.0005	<0.0005	<0.0005
Uranium U	ICP-MS	0.000834	0.00192	0.00102	0.000617
Vanadium V	ICP-MS	<0.0002	<0.0002	<0.0002	<0.0002
Zinc Zn	ICP-MS	0.0019	0.002	0.0014	0.0005
Zirconium Zr	ICP-MS	<0.0001	<0.0001	<0.0001	<0.0001

Leachate Analysis

Lithology		Till	Till	Till	Till
Sample ID		GTTAL 39	GTTAL 40	GTTAL 02	GTTAL 38
Parameter	Method				
Volume Nanopure water		750	750	750	750
Sample Weight		250	250	250	250
pH	Meter	7.93	7.99	7.65	8.19
Redox	Meter	317	313	347	304
Conductivity	Meter	423	142	500	721
Acidity (to pH 4.5)	Titration	#N/A	#N/A	#N/A	#N/A
Total Acidity (to pH 8.3)	Titration	4.7	4.2	5.0	6.2
Alkalinity	Titration	68.1	76.9	57.7	48.0
Sulphate	Turbidimetry	173	7	174	385
Ion Balance					
Major Anions	Calculation	4.97	1.68	4.78	8.98
Major Cations	Calculation	4.48	1.45	4.84	8.11
Difference	Calculation	0.49	0.23	-0.06	0.87
Balance (%)	Calculation	5.2%	7.3%	-0.6%	5.1%
Dissolved Metals					
Hardness CaCO3	Calculation	209	66	214	400
Aluminum Al	ICP-MS	0.016	0.263	0.0203	0.007
Antimony Sb	ICP-MS	0.00065	0.0038	0.00121	0.00056
Arsenic As	ICP-MS	0.00077	0.00189	0.00315	0.00037
Barium Ba	ICP-MS	0.0313	0.0704	0.0497	0.0217
Beryllium Be	ICP-MS	<0.00001	0.00001	<0.00001	<0.00001
Bismuth Bi	ICP-MS	<0.000005	0.000006	<0.000005	<0.000005
Boron B	ICP-MS	<0.05	<0.05	<0.05	<0.05
Cadmium Cd	ICP-MS	0.000021	0.000038	0.000044	0.000069
Calcium Ca	ICP-MS	46	20.4	67.7	102
Chromium Cr	ICP-MS	<0.0001	0.0008	0.0001	<0.0001
Cobalt Co	ICP-MS	0.00006	0.00011	0.000053	0.000205
Copper Cu	ICP-MS	0.00168	0.00514	0.00175	0.00053
Iron Fe	ICP-MS	0.01	0.185	0.014	0.005
Lead Pb	ICP-MS	0.00004	0.000485	0.000335	0.000119
Lithium Li	ICP-MS	0.0052	0.0023	0.0054	0.0153
Magnesium Mg	ICP-MS	23	3.66	10.9	35.1
Manganese Mn	ICP-MS	0.00788	0.00361	0.0981	0.00965
Mercury Hg	ICP-MS	<0.01	<0.01	<0.01	<0.01
Molybdenum Mo	ICP-MS	0.00352	0.0179	0.00191	0.0031
Nickel Ni	ICP-MS	0.0007	0.00165	0.00081	0.00181
Phosphorus P	ICP-MS	0.006	0.02	0.028	<0.002
Potassium K	ICP-MS	3.33	2.15	3.8	3.58
Selenium Se	ICP-MS	0.00069	0.00453	0.0229	0.00052
Silicon Si	ICP-MS	1.16	1.48	1.4	0.4
Silver Ag	ICP-MS	<0.000005	0.000011	0.000006	<0.000005
Sodium Na	ICP-MS	4.52	1.53	10.7	0.53
Strontium Sr	ICP-MS	0.22	0.107	0.299	0.681
Sulphur (S)	ICP-MS	54	<3	53	126
Thallium Tl	ICP-MS	0.00001	0.000011	0.000022	0.000102
Tin Sn	ICP-MS	<0.00001	0.00002	<0.00001	<0.00001
Titanium Ti	ICP-MS	0.0007	0.0074	<0.0005	<0.0005
Uranium U	ICP-MS	0.00567	0.00649	0.00409	0.00111
Vanadium V	ICP-MS	0.0003	0.0021	0.0009	<0.0002
Zinc Zn	ICP-MS	0.0059	0.0028	0.0008	0.0104
Zirconium Zr	ICP-MS	<0.0001	0.0004	<0.0001	<0.0001

Leachate Analysis

		Uncertain- 3G0? 4EC?	Uncertain- 5A0/5B0	Uncertain- 5A0/5B0? 3G0? 4EC?	
Lithology					
Sample ID		GTTAL 24	GTTAL 11	GTTAL 34	Blank
Parameter	Method				
Volume Nanopure water		750	750	750	750
Sample Weight		250	250	250	-
pH	Meter	7.18	8.06	7.15	7.08
Redox	Meter	351	263	332	300
Conductivity	Meter	2057	221	419	<1
Acidity (to pH 4.5)	Titration	#N/A	#N/A	#N/A	#N/A
Total Acidity (to pH 8.3)	Titration	58.5	4.0	6.4	3.6
Alkalinity	Titration	21.9	102.9	25.4	3.0
Sulphate	Turbidimetry	1337	27	194	2
Ion Balance					
Major Anions	Calculation	28.29	2.62	4.55	#N/A
Major Cations	Calculation	25.32	2.40	4.08	#N/A
Difference	Calculation	2.97	0.22	0.47	#N/A
Balance (%)	Calculation	5.5%	4.4%	5.4%	#N/A
Dissolved Metals					
Hardness CaCO3	Calculation	1220	117	199	<0.5
Aluminum Al	ICP-MS	0.002	0.042	0.0034	0.0002
Antimony Sb	ICP-MS	0.0011	0.00049	0.00147	<0.00002
Arsenic As	ICP-MS	0.0005	0.00038	0.00064	<0.00002
Barium Ba	ICP-MS	0.0287	0.0559	0.0272	0.00003
Beryllium Be	ICP-MS	<0.00005	<0.00001	<0.00001	<0.00001
Bismuth Bi	ICP-MS	<0.00003	<0.000005	<0.000005	<0.000005
Boron B	ICP-MS	<0.3	<0.05	<0.05	<0.05
Cadmium Cd	ICP-MS	0.0456	0.000009	0.00203	<0.000005
Calcium Ca	ICP-MS	254	20.2	42.9	<0.05
Chromium Cr	ICP-MS	<0.0005	<0.0001	<0.0001	<0.0001
Cobalt Co	ICP-MS	0.0349	0.000084	0.00369	<0.000005
Copper Cu	ICP-MS	0.0022	0.0003	0.00057	0.0004
Iron Fe	ICP-MS	0.153	0.025	0.002	<0.001
Lead Pb	ICP-MS	0.115	0.000127	0.00496	0.000018
Lithium Li	ICP-MS	0.01	0.0102	0.0035	<0.0005
Magnesium Mg	ICP-MS	143	16.1	22.3	<0.05
Manganese Mn	ICP-MS	0.333	0.00215	0.0685	0.00005
Mercury Hg	ICP-MS	<0.05	<0.01	<0.01	<0.01
Molybdenum Mo	ICP-MS	0.0011	0.0015	0.00577	<0.00005
Nickel Ni	ICP-MS	0.127	0.00049	0.0357	0.00015
Phosphorus P	ICP-MS	<0.01	<0.002	<0.002	<0.002
Potassium K	ICP-MS	1.79	2.06	1.26	<0.05
Selenium Se	ICP-MS	0.0018	0.00044	0.00061	<0.00004
Silicon Si	ICP-MS	<0.5	0.527	0.309	<0.1
Silver Ag	ICP-MS	0.00003	<0.000005	0.000012	<0.000005
Sodium Na	ICP-MS	0.78	0.35	0.78	<0.05
Strontium Sr	ICP-MS	0.4	0.109	0.165	<0.00005
Sulphur (S)	ICP-MS	459	10	62	<3
Thallium Tl	ICP-MS	0.00316	0.000061	0.00195	<0.000002
Tin Sn	ICP-MS	<0.00005	<0.00001	<0.00001	0.00009
Titanium Ti	ICP-MS	<0.003	0.0006	<0.0005	<0.0005
Uranium U	ICP-MS	0.00004	0.000712	0.000132	<0.000002
Vanadium V	ICP-MS	<0.001	<0.0002	<0.0002	<0.0002
Zinc Zn	ICP-MS	25.3	0.0016	1.23	0.0005
Zirconium Zr	ICP-MS	<0.0005	<0.0001	<0.0001	<0.0001