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5	Utility of Satellite-Derived Burn Severity to Study Short- and Long-Term Effects of
6	Wildfire on Streamflow at the Basin Scale
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48 49

Abstract

50 We investigated the changes in hydrologic response in a forested catchment impacted by 51 wildfire in Colorado U.S.A. from the storm event to the inter-annual scales. We also 52 evaluated the utility of a remotely-sensed burn severity index to study post-fire shifts in 53 streamflow. At the storm-scale, we evaluated hydrologic shifts through changes in the 54 effective runoff (Q^*/P_{Tot}), peak streamflow (Q_{pk}) and response time (T_R/T_B) from 55 multiple hydrographs, while at seasonal and inter-annual-scales we quantified hydrologic 56 shifts through the runoff fraction (Q/P_{Tot}) and flow duration curves. Vegetation 57 anomalies were monitored through comparisons of the Normalized Burn Ratio (NBR) 58 between the burned and a hydrologically-similar, forested, neighboring, unburned 59 catchment. We found short-term acute and long-term chronic transient streamflow shifts 60 from the minute to the inter-annual scales. Flow duration curves indicate an order of 61 magnitude increase in maximum flows. Event-average Q*/P_{Tot} increased by two orders of 62 magnitude and Q_{pk} increased by one order of magnitude relative to multiple representative pre-fire events of similar precipitation intensities. Decreases in T_R/T_B 63 64 appear to be minimal. At the inter-annual scale, increases in the difference between 65 simultaneous unburned and burned NBR are associated with increases in Q/P_{Tot}. A 66 hydrologic recovery pathway is evident resembling a hysteresis effect driven by 67 vegetation re-growth. Results illustrate the non-steady physical processes that increase 68 flash-flooding risks post-fire in mountainous catchments and the utility of ΔNBR as a 69 hydrologic predictor in ungauged watersheds.

Keywords: Wildfire hydrology, remote sensing hydrology, Normalized Burn Ratio, Postfire streamflow shifts, forest hydrology, land cover change.

72 **1. Introduction**

73 Quantifying the magnitude and temporal extent of streamflow shifts in recently 74 burned mountain catchments is of primary interest due to the substantial increase in flash 75 flooding and debris flow risks downstream of the burned areas. This task is particularly 76 challenging in ungauged, topographically-complex catchments prone to intense, 77 convective precipitation. A significant amount of research has been devoted to this topic, 78 and several site-specific factors such as regional climate (Zhou et al., 2015), watershed 79 terrain characteristics (Wine and Cadol, 2016), burn area and severity (Moody et al. 80 2007; Benyon and Lane, 2013), and vegetation species (Kuczera, 1987; Heath et al., 81 2014) have been found to result in varying degrees of streamflow shifts in magnitude and 82 longevity (Kinoshita and Hogue, 2015). Despite previous work on using remote sensing 83 indices to investigate changes in runoff production, the utility of NBR to infer both short-84 and long-term hydrologic changes has not been fully evaluated. Linking vegetation burn 85 severity and hydrologic anomalies becomes a relevant topic for water and land managers 86 that often require site-specific information on the expected duration and magnitude of 87 fire-induced streamflow shifts.

Post-fire streamflow shifts have occurred globally across various climatic conditions and spatiotemporal scales but mostly in the U.S western states of California (Bart and Hope, 2010; Kinoshita and Hogue, 2015; Bart, 2016), New Mexico (Wine and Cadol, 2016), and Colorado (Larsen et al., 2009), and also in Portugal (Walsh et al., 1994), Spain (Cerdà and Lasanta, 2005), Israel (Inbar et al., 1998), and Australia (Zhou et al., 2015). Three streamflow metrics are usually investigated as indicators of hydrologic anomalies: (1) runoff volume (Helvey, 1980; Moody et al., 2008; Moody and Ebel, 2014; Ebel et al.,

95 2012); (2) peak streamflow (Moody and Martin, 2001b; Shakesby and Doerr, 2006); and 96 (3) time to peak streamflow (Neary et al., 2005). Post-fire runoff shifts have been 97 evaluated from the plot (Benavides-Solorio and MacDonald, 2001) to the catchment scale 98 (Moody et al., 2008). Such changes in runoff production generally occur because of the 99 formation or enhancement of water-repellant soils (DeBano, 2000) as well as significant 100 decreases in vegetation density and litter cover (Cerdà and Doerr, 2005). The post-fire 101 conditions enhance infiltration-excess runoff mechanisms (Shakesby and Doerr, 2006) 102 that result in positive peak streamflow shifts (Gottfried et al., 2003; MacDonald and 103 Huffman, 2004; Stoof et al., 2012; Mahat et al., 2016; Moreno et al., 2016) and shorter 104 response times (Baker et al., 2004; Neary et al., 2005; Cydzik and Hogue, 2009).

105 Fire severity effects on vegetation and soil properties and their subsequent recovery 106 play a significant role in the short- (acute) and long-term (chronic) hydrologic responses 107 of a burned region (Cerdà, 1998). In the case of long-term behavior, the post-fire 108 hydrologic restoration rates can vary from year-to-year depending on regional climatic 109 and vegetation conditions (Shin et al., 2013). Previous studies have noted the strong 110 coupling between soil, vegetation and hydrologic recovery pathways, usually finding this 111 process to occur within 2 to 7 years post-fire, in tandem with the re-establishment of 112 vegetation, stream-network connectivity and soil hydraulic properties (Moody and 113 Martin, 2001b; Cerdà and Doerr, 2005; Wittenberg et al., 2007; Mayor et al., 2007), 114 although in certain cases it may take longer. For instance, Kinoshita and Hogue (2015) 115 found elevated stream discharge during low flow seasons nearly ten years following a 116 Californian wildfire, and attributed the elevated runoff to reduced transpiration.

117 Remote sensing indices are used to investigate the severity and track recovery 118 pathways after wildfires. Among commonly used indices to study ecosystem damage 119 after fire are the Enhanced Vegetation Index (EVI), the Normalized Difference 120 Vegetation Index (NDVI) and, the more recently developed, Normalized Burn Ratio 121 (NBR). All are usually computed from Landsat or MODIS imagery. Both EVI and 122 NDVI were essentially developed to track canopy structural variations and health. 123 Thereby, they can also be used to infer vegetation damage after fires (Wittenberg et al., 124 2007; Casady et al., 2010; Kinoshita and Hogue, 2011; Wine and Cadol, 2016; Uyeda et 125 al., 2017). However, the application of EVI and NDVI to study post-fire vegetation 126 alterations involves conceptual limitations related to the spectral bands that do not 127 necessarily capture burn severity (Epting et al. 2005). Subsequently, the NBR was 128 defined to identify and quantify the effects of fire on vegetation by including the mid-129 infrared band (Cocke et al. 2005, Epting et al. 2005, Roy et al. 2006, Walz et al. 2007, 130 Loboda et al. 2007, Escuin et al., 2008, Weber et al. 2008). The difference between 131 consecutive pre- and post-fire scenes (i.e ΔNBR) is usually taken to represent burn extent 132 and severity (Miller et al. 2007, Moody et al, 2015). Despite recent advances in the 133 understanding of the relationship between post-fire soil and vegetation recovery and the 134 evolution of hydrologic shifts (Onda et al. 2008; Kinoshita and Hogue 2011), further 135 research remains necessary at the catchment scale (Woodsmith et al., 2004) to relate the 136 hydrologic anomalies to remotely-sensed imagery using paired-basin approaches. 137 Innovations are particularly germane for application to ungauged basins.

Post-fire hydrologic shifts beyond the plot and hillslope scale were studied by several
authors (Lane et al., 2006; Mayor et al., 2007; Heath et al., 2014; Wine and Cadol, 2016;

Bart and Hope, 2010; Bart, 2016). However, catchment scale studies are less frequently found than plot or hillslope scale investigations, and consequently streamflow dynamics at the watershed scale are less clearly understood under post-fire conditions (Moody and Martin, 2001a; Mayor et al., 2007). Furthermore, there is a paucity of post-fire studies with hydrological records available both pre- and post-fire (Woodsmith et al., 2004; Kinoshita and Hogue, 2015), making it difficult to quantify streamflow shifts relative to unburned conditions.

147 This study investigates catchment-scale hydrological shifts relative to unburned 148 conditions at a predominantly forested catchment in Colorado, USA. The broader 149 research questions are: (1) what are the potential links between basin-average NBR and 150 the observed post-fire streamflow anomalies? And (2) how could paired-basin ΔNBR be 151 used to quantify inter-annual runoff alterations and hydrologic recovery post-fire? The 152 focus of the study is on the use of remote-sensing data to detect and track a basin's 153 hydrologic response to wildfire impacts and to investigate the links with streamflow 154 shifts through process-based analyses. Pre- and post-fire catchment observations of 155 precipitation, streamflow and satellite-derived NBR are used. ANBR is computed from 156 simultaneous burned and unburned paired catchments, with the purpose of excluding the 157 effects of vegetation phenology, to infer changes due to the fire impacts that result in 158 runoff generation shifts.

To account for the effects of severe weather within this mountain watershed, analyses are conducted under warm-season precipitation as measured by rain gauges situated within and nearby the catchment. Six years pre- and post-fire are analyzed as a representative period for precipitation and streamflow in regards to vegetation recovery,

while also ensuring a representative number of storm events to synthesize hydrologicpatterns.

165 The objectives of this study are to: (1) evaluate the utility of NBR in estimating 166 transient shifts in runoff fraction, peak streamflow and time to peak streamflow, and (2) 167 examine the coupled ecosystem-hydrologic response post-fire from the event to the inter-168 annual scale by combining ground stations and remote-sensing image analysis. The focus 169 will be on the Camp Creek catchment located within Waldo Canyon in Colorado. Table 1 170 summarizes previous research and reports that have investigated the hydrological and 171 societal impacts of the 2012 Waldo Canyon fire in Colorado, a unique and widely studied 172 fire which also serves as a case study in the current paper.

173 **Table 1:** Summary of previous research concerning the 2012 Waldo Canyon fire.

Study/Report	Study area	Summary of findings		
Verdin et al., 2012	22 basins affected by the Waldo Canyon burn scar	Model projected debris flow probabilities for Camp Creek were 24%, 45% and 55% for 2-year, 10-year, and 25-year storms, respectively.		
Young and Rust, 2012	5 watersheds and 26 pour points within Waldo Canyon burn scar	Moderate and high severity burns occur on steep slopes, producing projected erosion rates of up to 31 t/ha for a 2-year runoff event, 61 t/ha for a 5-year event and 90 t/ha for a 10-year event.		
Rosgen et al., 2013	Four major watersheds affected by the Waldo Canyon fire (including Camp Creek)	Projected annual change in water yield of 66 mm in Camp Creek due to reduction in forest cover. Comparisons of pre- and projected post-fire water yield indicates an increase in water yield by 1.95*10 ⁶ m ³ in Camp Creek. Project slow vegetative recovery rate due to coarse textured soils and low precipitation magnitudes relative to potential evapotranspiration.		
Staley et al., 2015	12 drainage basins within Waldo Canyon burn scar	Total rainfall for debris-flow producing flash floods ranged from 7.2-34.9 mm with storm durations spanning from 10-125 minutes. Flash floods occurred following short bursts of high intensity rainfall events.		
Kinoshita et al., 2016	Colorado Springs, CO	Hazard mitigation techniques often confound interrelated post-fire processes, thereby hindering predictions of post-fire natural responses.		
Chin et al., 2016	Waldo Canyon burn scar	Barriers installed to block post-fire debris flow facilitated downstream erosion and channel degradation thereby enhancing the hazard and requiring further alterations to the landscape.		

174 **2.** Study area, hydrology and remote sensing data

175 2.1 Wildfire coverage and study area

Between June 23rd and July 10th of 2012, the Waldo Canyon fire burned over 74 km² 176 of land near Colorado Springs, CO. The wildfire extent, general area topography and 177 178 vegetation distribution are shown in Figure 1 and Appendix A.1. The proximity of the 179 fire to a large city (Colorado Springs, Figure 1) enhanced the human impact of the Waldo 180 Canyon fire, damaging 346 homes and killing two people (City of Colorado Springs, 181 2013). The U.S. Forest Service's Inter-Agency Burn Area Emergency Response (BAER) 182 team classified 28% of the burn area as very low severity, 23% as low, 29% as moderate 183 and 20% high severity (see Figure A.1). Camp Creek is a 24.4 km² catchment with 78% 184 of its drainage area affected by this fire (see Figure 1). Table 2 summarizes wildfire 185 spatiotemporal characteristics within Camp Creek, indicating burned area and wildfire 186 period.

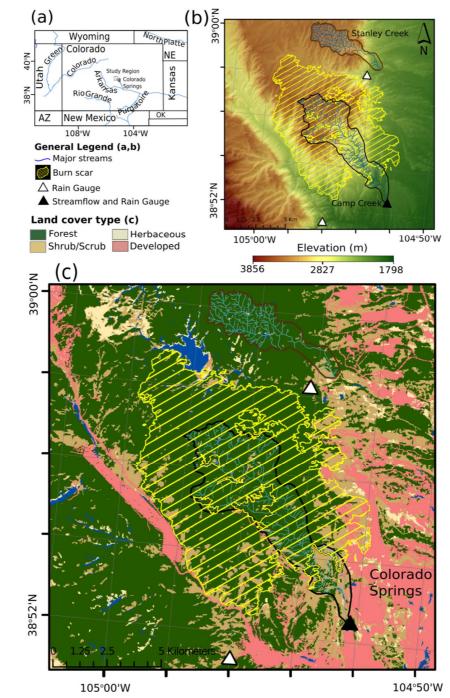
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 Table 2: Wildfire characteristics in Camp Creek and hydrologic evaluation periods.

Value
19
78
5.4
Jun 23 – Jul 10, 2012
Apr 2006 – June 2018
Apr 2006 – June 2012
July 2012 –June 2018

188

According to the BAER team, Camp Creek had 36% of the burn area as low severity, 37% moderate severity and 5% high severity (Young and Rust, 2012; Rosgen et al., 2013). A neighboring unburned control catchment with similar topographic, channel network and land cover characteristics is Stanley Creek. This is a 9 km² watershed 193 located within 8.7 km distance from Camp Creek that was used as a control catchment for



194 comparing unburned to burned (i.e. Camp Creek) NBR values (see Figure 1).

195

Figure 1: (a) Regional location of the study area including the Waldo Canyon fire near the Colorado
Springs city limits. (b) General relief and channel network distribution at Camp Creek (burned) and Stanley
Creek (unburned) catchments. (c) Land cover type at Camp Creek (burned) and Stanley Creek (unburned)
catchments. On panels (b) and (c), burned areas are hatched in yellow. Also shown within the main legend
are the used USGS and NOAA rain and streamflow gauge stations.

201	Table 3 documents geomorphic and land cover characteristics of the two (burned and
202	unburned) catchments. Table 3 and Figure 1 indicate that both Camp and Stanley Creeks
203	hold similar relief, channel network distribution, mean slope, general aspect and
204	vegetation cover. Additionally, given the short distance that separates them, they share
205	similar climates. Such similarities result in similar spectral signatures of the NBR (see
206	section 4.1). The seasonality of precipitation in Colorado's central mountains is complex,
207	with intense precipitation possible all year. However, during the warm season (April to
208	October) the precipitation is likely to be convective and, even under unburned conditions,
209	it can trigger instantaneous flood risk that is enhanced by the steep slopes of this region
210	(Mahoney et al., 2015; Moreno et al. 2012, 2013).
211	

212 213 214 215 **Table 3:** Topographic, land cover and hydrologic characteristics for Camp Creek (burned) and Stanley Creek (unburned) catchments. Seasonal and annual averages span from 2006 – 2018. Stanley Creek does not have a streamflow gauge or in-catchment rain gauge, and therefore only topographic and vegetation

14	not have a streamnow gauge or in-catchment rain gauge, a	nd therefore only topographic and vegetation
15	characteristics are reported.	

Feature	Camp Creek (burned) catchment	Stanley Creek (unburned) catchment
Outlet coordinates	104.872° W, 38.875° N	104.890° W, 38.975° N
Total area [km ²]	24.4	9.0
Length of main channel [km]	13.8	8.3
Slope of main channel [mkm ⁻¹]	127.4	95.1
Seasonal total precipitation [mm]	3960	-
Wettest season during study period	2015	-
Driest season during study period	2012	-
Annual average streamflow [m ³ /s]	0.030	-
Seasonal average streamflow [m ³ /s]	0.052	-
Mean elevation [m]	2539	2670
Minimum/maximum elevations [m]	1917/2936	2112/2857
Std. elevation [m]	287.0	206.4
Mean slope [%]	20.6	14.01
Std. slope [%]	10.5	8.8
Slope aspect (%)	SE 18.38 E 17.13	E 22.91 SE 15.27
	S 15.31 NE 13.16	NE 14.92 S 12.43
Major vegetation class 1 (% area)	Evergreen Forest (77.5)	Evergreen Forest (84.5)
Major vegetation class 2	Shrub/Scrub (11.8)	Deciduous Forest (7.3)
Major vegetation class 3	Deciduous Forest (4.5)	Shrub/Scrub (6.2)
Kirpich (1985) Concentration time [min]	66.3	50.4

217 2.2 Observed precipitation and streamflow data

This study uses multi-year data from three precipitation gauges and one streamflow gauge spanning the warm season (April through October, without snow events) in the analysis of pre- and post-fire hydrologic patterns to emphasize rainfall-driven responses. Figure 1 shows the regional location of the study gauge stations while Table 4 illustrates station codes, type, measurement frequency and missing data, and Table 2 specifies the evaluation periods according to the data availability.

224

 Table 4: Precipitation and streamflow gauges used in this study.

Code	Туре	Time Step	% Missing data
USGS 07103703	Streamflow	15 min	0.52
USGS 07103703	Precipitation	5 min	2.6
USGS 07103800	Precipitation	5 min	5.7
NOAA COOP055352	Precipitation	1 hour	1.3

225

Camp Creek's main stream is ephemeral, with zero flow values outside of active
precipitation events. It has 15-minute streamflow and 5-minute precipitation data
available through the United States Geological Survey (USGS) National Water
Information System and National Oceanic and Atmospheric Administration (NOAA)
Cooperative Observer Program (COOP).

231

232 2.3 Remote sensing data

This study uses MOD13Q1 MODIS imagery to calculate the NBR and quantify fireinduced changes within the burned (Camp Creek) relative to the unburned (Stanley Creek) catchment. While EVI and NDVI are based on near infrared and visible light to quantify vegetation's density and health, NBR enhances the distinction between burned and unburned surfaces with the addition of the mid-infrared spectral band. The MOD13Q1 MODIS provides imagery for calculating 16-day composites of NBR data at 239 250-m resolution and is available from February 2000 – present (NASA LP DAAC,
240 2000).

241

242 **3. Methods**

243 3.1 Evolution of catchment burn severity and streamflow pre- and post-fire

244 NBR values were calculated from the near-infrared (NIR) and mid-infrared (MIR) 245 MODIS bands (Roy et al., 2006) for both Camp and Stanley Creeks. The phenologic 246 cycles in the spectral signal of vegetation prevent the use of a single- reference, pre-fire 247 NBR image for Camp Creek to estimate NBR changes with respect to standard 248 conditions. Instead, comparing simultaneous NBR composites at paired Camp Creek and 249 Stanley Creek catchments helps in isolating the effects of fire on vegetation from natural 250 variability. Therefore, a NBR time series analysis was conducted to determine if both (i.e. 251 burned and unburned) catchment's spectral signal (mean and variability) preserved a high 252 correlation coefficient during pre-fire conditions (Lhermitte et al., 2010; Veraverbeke et 253 al., 2010; Diaz-Delgado and Pons, 2001; Veraverbeke et al., 2010). Further, regression 254 analysis was applied (e.g. logarithmic, exponential, linear, power law, and polynomial) to 255 best fit the first two distributional moments during the 2006 through 2012 period. The 256 best predictor was used to adjust NBR values at Stanley to Camp Creek pre-fire and to 257 forecast post-fire NBR at Camp Creek as if it was unburned. Catchment average and 258 standard deviation NBR values associated with the burned and unburned conditions were 259 computed throughout the April – October season over the evaluation period excluding 260 any periods with observed snow on the ground. In a typical year, and for clear-sky 261 conditions, a maximum of 14 average NBR composites (2 per month over the 7-month season) were available. ΔNBR was then computed as the difference between adjusted
unburned and burned conditions at Camp Creek as illustrated by equation (1) (DiazDelgado and Pons, 2001):

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- 266

$$\Delta NBR(t) = NBR_{unburned}(t) - NBR_{burned}(t)$$
(1)

267

268 Where t is the simultaneous period for which the difference is calculated. All terms 269 represent catchment-average values. NBRunburned and NBRburned correspond to the adjusted 270 NBR from Stanley Creek and the simultaneous values at Camp Creek, respectively. 271 Although this difference (i.e. ΔNBR) only makes sense post-fire, the calculation pre-fire 272 should yield values close to zero, as a way to ensure that both adjusted and actual NBR 273 values were very similar pre-fire. Theoretically, ΔNBR ranges from -2 (negative two) 274 indicating healthier vegetation and 2 (positive two) for high severity burn with 0 (zero) 275 meaning conditions similar to pre-fire. This methodology facilitated the tracking of post-276 fire vegetation recovery relative to the unburned conditions under similar climatic 277 conditions.

Flow duration curves (FDCs) are widely used to study exceedance probabilities of mean and extreme hydrologic values. In some cases, they have also been used to evaluate the effects of catchment disturbances on flow distribution, providing statistical information on streamflow variability (Lane et al., 2005; Lane et al., 2006; Kinoshita and Hogue, 2015). FDCs were created for the pre- and post-fire study periods in Camp Creek using daily average streamflow data to identify fire-related shifts in streamflow frequency distribution.

285 *3.2 Cross-scale temporal streamflow shifts*

286 The subsequent event-scale analyses incorporated ΔNBR by using the average values 287 associated with the closest-in-time 16-day composite of each month, depending on the 288 date of the streamflow event. Similarly, the seasonal and inter-annual-scale analyses took 289 the average of each year's 14 NBR composites to compute Δ NBR between the burned 290 and unburned vegetation. Regarding the streamflow metrics, event, daily and seasonal 291 scale values were computed to assess runoff changes post-fire relative to pre-fire 292 conditions. Event-scale calculations were conducted for a total of 76 hydrologic events, 293 38 associated with pre-fire and another 38 with post-fire conditions. Table 2 summarizes 294 the pre-fire and post-fire time spans. All periods with snowfall or presence of snow on the 295 ground were removed from the analyses. To ensure that initial soil moisture conditions 296 were low, events with minimal precipitation ($P \le 3 \text{ mm}$) five days prior were selected, so 297 that the hydrologic effects due to land cover change, rather than high antecedent soil 298 moisture, could be isolated. Figure 2 illustrates a typical response hydrograph where Q 299 represents the total streamflow through the basin outlet as a result of both base flow, Q_B, 300 and event flow, Q*. Q* represents the total water volume per unit catchment area as a 301 result of the precipitation depth, P_{Tot}. A graphical base flow separation method was used 302 to divide response hydrographs between Q_B and Q^{*} (Figure 2; Dingman, 2015). At the 303 monthly and seasonal scales, Q represents the total streamflow volume leaving the 304 catchment through its outlet as a result of a monthly or seasonal precipitation input, P_{Tot}. 305 Calculations for the monthly and seasonal Q and P_{Tot} are made for the months and 306 seasons associated with each of the 78 maximum hydrologic events in Camp Creek.

Based on those streamflow metrics, four comparative streamflow response metrics were used: (a) event effective runoff per unit precipitation, Q^*/P_{Tot} (mm/mm), (b) event peak streamflow, Q_{pk} (m³/s), (c) event time to peak fraction of hydrograph base time, T_R/T_B , (min/min) and (d) monthly and seasonal runoff coefficient, Q/P_{Tot} (mm/mm).

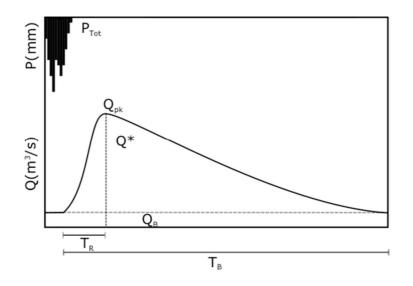


Figure 2: Typical watershed response hydrograph indicating how event flow (Q^*) , total event precipitation (P_{Tot}), time to peak (T_R), time base (T_B) and maximum observed flow (Q_{pk}) are extracted. Q_B represents base flow.

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High values of Q^*/P_{Tot} and Q_{pk} represent high runoff yields and maximum flows. Small values of T_R/T_B indicate a faster-paced hydrograph rise whereas larger values are associated with a slower rise. This cross-scale analysis facilitated the assessment of acute and chronic post-fire anomalies in runoff production per unit drainage area and precipitation depth so that runoff changes could be mostly attributed to vegetation alterations produced within the burn scars.

322

323

325 *3.3 Streamflow shifts and burn severity links*

In order to explore the co-dependence of Q^*/P_{Tot} , Q_{pk} and T_R/T_B with event 326 327 precipitation properties such as precipitation total (P_{Tot}), average and maximum intensity (Iave and Imax) and land cover change due to burn severity (i.e. ΔNBR), a variance 328 329 contribution approach was applied including the 76 study events. First, scatterplots and 330 correlation coefficients between pairs of Imax, PTot and Iavg were calculated to assess their 331 level of independence prior to the application of the variance contribution analysis. Then 332 Q^*/P_{Tot} , Q_{pk} and T_R/T_B were plotted against ΔNBR with regards to their P_{Tot} , I_{max} and I_{avg} 333 to investigate the potential links among groups of three variables (e.g. Q*/P_{Tot}, ΔNBR 334 and P_{Tot}). Empirical density distributions were plotted and Kolmogorov-Smirnov tests 335 were conducted in regards to the pre- and post-fire distributions. Subsequently, a variance 336 contribution analysis was conducted for variable arrays containing each of the event 337 streamflow properties $(Q^*/P_{Tot}, Q_{pk} \text{ and } T_R/T_B)$ and all other independent variables 338 including precipitation characteristics and ΔNBR (I_{max}, P_{Tot}, I_{avg} and ΔNBR). The method 339 includes all measured predictors into multiple linear combinations to find a subset that 340 explains the highest percent of the variance in the predictands. Best subsets (Heinze et al. 341 2018; Olejnik et al. 2010) is a technique that relies on exhaustive searches for the best 342 groups of the variables using an efficient branch-and-bound algorithm. The procedure fits 343 2^{P} (i.e. 16) models, where P=4 is the number of predictors in the dataset. The Akaike 344 information criterion (AIC; Akaike, 1973) and the coefficient of determination (R^2) 345 facilitate model selection by providing an estimator of the relative quality of statistical 346 models for a given set of data. Results are presented in pie charts, model regression plots 347 and tables for pre- and post-fire events. Finally, inter-annual Q/PTot were analyzed relative to simultaneous ΔNBR values to identify recovery patterns of vegetation
properties and their potential controls on runoff production at Camp Creek.

350

351 4. Results

352 *4.1 Evolution of catchment burn severity and streamflow pre- and post-fire*

353 A sequence of NBR field values from May 2012 (pre-fire month) and August 2012 354 (post-fire month) and their corresponding NBR_{May}-NBR_{Aug} difference illustrates the 355 effects that fire had on vegetation in the Waldo Canyon region (See Figure 3). In Figures 356 3(a) and 3(b), NBR values range between -1 and 1 pre- and post-fire with an evident shift 357 to negative values post-fire indicating removal of vegetation by fire. The NBR_{May}-358 NBR_{Aug} difference illustrates changes of up to 2 NBR units in the most severely burned 359 areas of Camp Creek. The pre-fire average basin values of NBR (i.e. NBR_{May}) for Camp 360 Creek and Stanley Creek are 0.24 and 0.29, respectively. For post-fire conditions, the 361 average NBR_{Aug} for Camp Creek is -0.03 while the unburned value for Stanley is 0.36.

362 NBR regression analysis found that either linear, logarithmic or polynomial models 363 provided the same coefficient of determination ($R^2=0.80$) between simultaneous 364 catchment-average values at Camp and Stanley Creeks during the 2006 through 2012 365 (pre-fire) period (see Table B.1). A linear model was then selected to express NBR at 366 Camp Creek as a function of the values at Stanley Creek as shown by Figures B.1 and 367 B.2. This results in synthetic time series of NBR at Camp Creek as if it was unburned 368 thereby facilitating the comparison of burned and unburned conditions for post-fire 369 events within Camp creek.

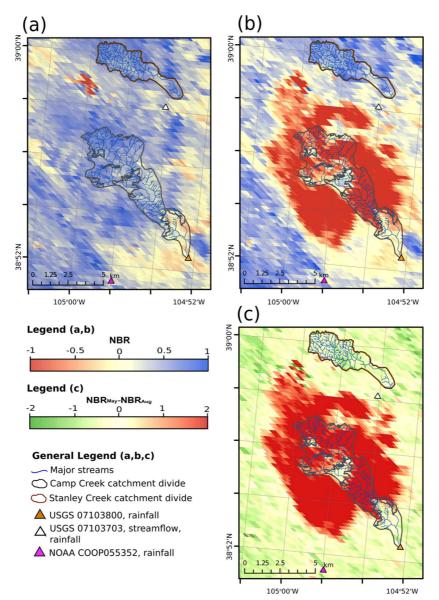




Figure 3: Spatial distribution of NBR during (a) pre-fire conditions in May 2012 and (b) post-fire conditions in August 2012. (c) Regional NBR_{May}-NBR_{Aug} difference for the Waldo Canyon including Camp and Stanley Creeks; negative values (green pixels) in the NBR_{May}-NBR_{Aug} difference map indicate healthier vegetation but positive (red pixels) indicate burn severity. Note that ΔNBR spans between -2 and 2 since NBR ranges from -1 to 1.

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Time series of daily total precipitation P, bi-weekly mean and standard deviation
NBR, and daily average streamflow Q, over Camp Creek catchment are shown in Figure
d during the 12-year evaluation period. For reference, Stanley Creek's adjusted mean
NBR time series have also been added to Figure 4. Seasonally, NBR peaks in late July
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and early August of each year. The Waldo Canyon wildfire abruptly decreases the postfire Camp Creek's catchment-average NBR and increases its spatial variability as reflected by the larger standard deviation bars. However, after fire, NBR follows a recovery pathway gradually increasing each year. Further NBR's spatial variability appears to decrease faster returning to pre-fire values in about three years after the wildfire event.

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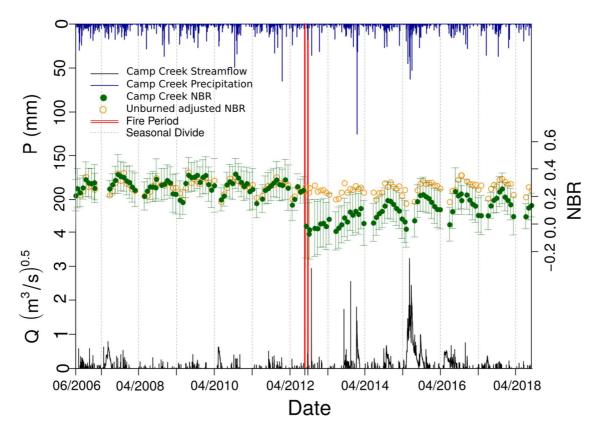




Figure 4: Time series of daily total precipitation (P, blue lines), daily average streamflow (Q, black lines) and monthly basin average and standard deviation of Normalized Burn Ratio (NBR, green dots and bars) for Camp Creek. Red lines represent the start and end of the Waldo Canyon wildfire. For reference, simultaneous catchment-average NBR values for Camp Creek predicted from the Stanley Creek time series have been added (hollow, orange circles). Q values are elevated to a 0.5 power with the purpose to better visualize the lowest flow rates.

397 In Camp Creek, the pre-fire (i.e. 2006-2012) and post-fire (i.e. 2012-2018) mean 398 seasonal precipitation values were 357 and 416 mm, respectively. This difference can be 399 explained by the exceptionally dry year of 2012 (year of the wildfire) with a precipitation 400 total of 256 mm and the exceptionally wet year of 2015 with 663 mm of precipitation. In general, Camp Creek showed lower streamflow responses to precipitation events pre-fire 401 402 (Figure 4). Changes in discharge are evident post-fire. The magnitude of streamflow 403 events increased significantly post-fire, with the largest pre-fire Q_{pk} (over the 2006 – 404 2012 period) being 0.26 m³/s while the largest post-fire Q_{pk} of 10.42 m³/s occurred in 405 May of 2015 (almost three years after the fire), equivalent to, at least, an order of 406 magnitude increase with respect to the maximum Qpk which occurred during the pre-fire 407 period. Similar peak streamflow events occurred the same year and month after the 408 wildfire in July of 2012 (8.64 m³/s) and one year later in August of 2013 (6.54 m³/s).

409 Warm-season flow duration curves (FDCs) are created from daily average streamflow 410 data for the pre- and post-fire study periods and indicate changes in Camp Creek's flow 411 regime post-fire (Figure 5). FDCs over a 20-year pre-fire time period (1992-2012) are 412 used as the baseline to quantify fire impacts. Pre- and post-fire FDCs are not distributed 413 over the entire range of exceedance probabilities because Camp Creek is an ephemeral 414 stream. The rightward shift of the post-fire FDC indicates that all portions of the flow 415 regime are affected by burn scars, and the streamflow at each probability value is, on 416 average, an order of magnitude larger across the range of probabilities. High flows (flows 417 at 1% exceedance probability; Smakhtin, 2001) increase from 0.74 m³/s pre-fire to 2.5 m³/s post-fire. Low flows (flows 99% exceedance probability; Brown et al., 2005) do not 418 419 increase in magnitude, remaining at 0 m³/s post-fire. However, the shift in streamflow distribution post-fire indicates that the number of zero flow days slightly decreased.
Throughout the 20 years prior to the wildfire, zero flow days occurred 88.9% of the time
and over the 6-year post-fire time span, this value decreased to 85.7% for a difference of
3.2% reduction in zero flow days.

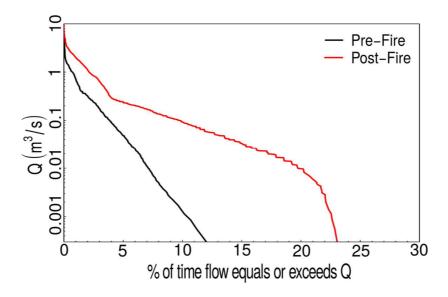


Figure 5: Pre- (black) and post- (red) fire flow duration curves for Camp Creek watershed computed from daily average streamflow data. Pre-fire flow duration curves were created using data from April 1992 – June 2012 and post-fire from July 2012 –July 2018.

428

424

429 *4.2 Cross-scale temporal streamflow shifts*

This section synthetizes the observed wildfire-induced streamflow shifts from the event to seasonal time scales. Figure 6 illustrates the pre- and post-fire Q^*/P_{Tot} for 76 hydrologic events (38 pre- and 38 post-fire) and their associated daily and monthly average Q/P_{Tot} . Symbols are colored by their corresponding total precipitation P_{Tot} (mm). A summary of the mean absolute and relative changes between pre- and post-fire values, at each temporal scale, is shown in Table 5. Figure 6 and Table 5 indicate increases in effective runoff per unit precipitation from

the event to the monthly time scales. Nonetheless, the largest difference in runoffbetween pre- and post-fire is observed at the event scale, with two orders of magnitude

439 (i.e 10²) higher values than pre-fire (see Table 5). Increases in streamflow magnitudes
440 occur across a wide range of precipitation totals.

441

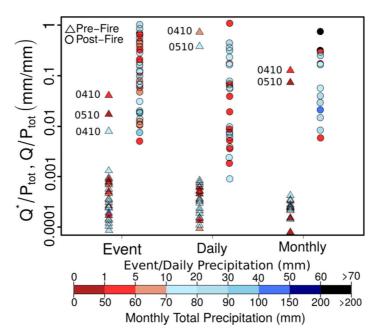


Figure 6: Q*/P_{Tot} calculated at the storm and Q/P_{Tot} at the daily and monthly time scales for Camp Creek
watershed. Triangles represent pre-fire and circles, post-fire hydrologic conditions. Labeling indicates the
month (two-digit) and year (two-digit) when some large runoff pre-fire values occurred.

446

442

447 **Table 5:** Cross-scale average, absolute and relative mean differences of event Q^*/P_{Tot} and daily and 448 monthly Q/P_{Tot} values for pre- and post-fire conditions.

Catchment	Temporal Scale	Pre-fire Average (mm/mm)	Post-fire Average (mm/mm)	Mean Absolute Change (mm/mm)	Relative increase (Post-fire/ Pre-fire)
	Event	0.0025	0.2131	+0.2168	1.1×10^{2}
Camp Creek	Daily	0.0349	0.1175	+0.0826	3.4×10^{0}
	Monthly	0.0135	0.1413	+0.1278	1.1×10^{1}

449

450 *4.3 Streamflow shift links with burn severity*

This section illustrates the differing hydrologic response patterns between pre- and post-fire conditions for individual precipitation events at Camp Creek and the role of burn severity on the observed acute (event) and chronic (inter-annual) streamflow shifts. Figure 7 provides an example case of the corresponding response hydrographs for two such events that are similar in P_{Tot} and I_{avg} . Figure 7(b) illustrates a post-fire storm event with a P_{Tot} of 27.6 mm and I_{avg} of 6.12 mm/hr that produced a Q_{pk} of 8.64 m³/s and a Q^*/P_{Tot} of 0.85. These values (i.e. Q_{pk} and Q^*/P_{Tot}) are, at least, one and two orders of magnitude, respectively, larger than the pre-fire values shown in Figure 7(a).

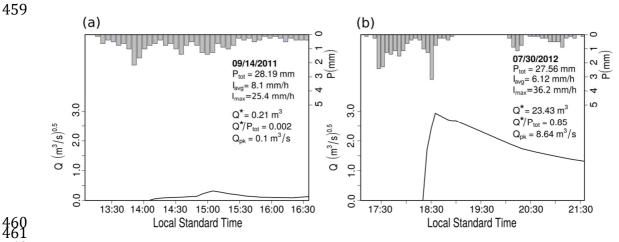
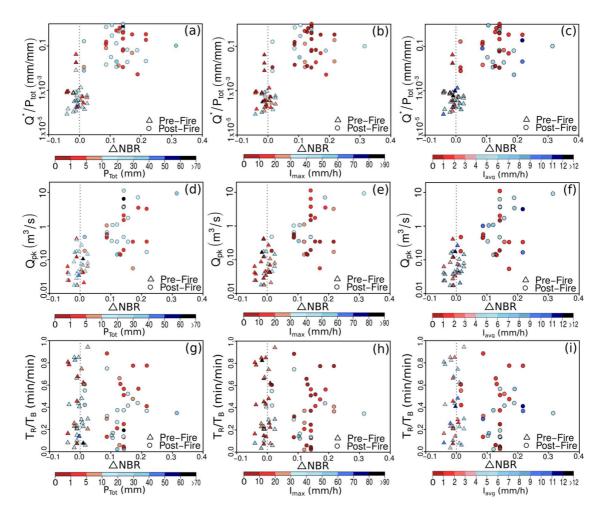


Figure 7: Example case hyetographs (5-min) and corresponding response hydrographs (15-min) for similar (i.e. I_{avg} and P_{Tot}) precipitation events before (a) and after (b) the Waldo Canyon fire at Camp Creek. Occurrence dates, total event precipitation (P_{Total}), mean precipitation intensity (I_{avg}), maximum 5-min precipitation intensity (I_{max}), event flow volume (Q^*), effective runoff fraction (Q^*/P_{Total}) and peak flow (Q_{pk}) values are indicated within each panel.

467

468 The total precipitation five days prior to each event is 1.2 mm pre-fire, and 3.1 mm 469 post-fire, indicating that the antecedent soil moisture between the two events is similar. 470 The nature and properties of precipitation play a definitive role on the magnitude and 471 pace of the response hydrograph and this is the reason why the following results are 472 illustrated in terms of I_{max} , I_{avg} and P_{Tot} .

To expand on the analysis of Q^*/P_{Tot} , Q_{pk} and T_R/T_B for the pre- and post-fire study events, Figure 8 illustrates their relationships with ΔNBR and classifies according to precipitation properties (i.e. P_{Tot} , I_{max} , I_{avg}) for the 76 pre- and post-fire storm events. To help understand average pre- and post-fire behaviors, Table 6 further synthetizes event477 based mean, absolute and relative hydrograph shifts. From Figure 8 and Table 6, it can 478 be inferred that, at the event scale, and for a wide range of P_{Tot} , I_{max} and I_{avg} values, 479 increases in two orders of magnitude (i.e. 10^2) on Q*/P_{Tot} are observed post-fire relative 480 to pre-fire conditions. Similarly, increases of one order of magnitude (i.e. 10^1) on Q_{pk} 481 occur, on average, across a large spectrum of P_{Tot}, I_{max} and I_{avg} . Additionally, T_R/T_B only 482 shows a reduction of 10% relative to pre-fire conditions, illustrating a slight decrease on 483 the average time to peak relative to pre-fire conditions.



485

Figure 8: Event scale scatterplots of Q^*/P_{Tot} , Q_{pk} and T_R/T_B vs. closest-in-time Δ NBR values for Camp Creek during multiple isolated warm-season storm events. Hydrologic events are classified by P_{Tot} (mm; left-hand side column), I_{max} (mm/h; center column) and I_{avg} (mm/h; right-hand side column).

490	Besides ΔNBR , P_{Tot} and I_{avg} appear to contribute to the variability of all three
491	streamflow metrics (i.e. Q^*/P_{Tot} , Q_{pk} and T_R/T_B) often regulating (i.e. enhancing or
492	reducing) the effect of ΔNBR . For instance, for a moderate precipitation event (e.g.
493	moderate P_{Tot} or I_{avg}) under a high ΔNBR , high Q^*/P_{Tot} , Q_{pk} and low T_R/T_B tend to occur.
494	However, a similar situation will occur under moderate ΔNBR in the presence of high
495	P_{Tot} and I_{avg} . Both scenarios result in an increase of streamflow values.

- 496
- 497

Table 6: Average Q*/P_{Tot}, Q_{pk} and T_R/T_B at Camp Creek along with mean absolute and relative changes for 498 all events shown in Figure 8 for pre- and post-fire conditions.

Metric	Pre-fire Event Average	Post-fire Event Average	Mean Absolute Change	Relative Change (Post-fire/ Pre-fire)
Q*/P _{Tot} (mm/mm)	0.002	0.231	0.229	1.1×10^{2}
$Q_{pk}(m^3/s)$	0.090	1.678	1.588	1.9×10^{1}
T_R/T_B (min/min)	0.387	0.357	-0.03	0.9

499

500 A further look to the probability density functions (Figure 9) and Kolmogorov-501 Smirnov (K-S) tests (Table 7) illustrate that precipitation properties (e.g. P_{Tot}, I_{max}, I_{avg}) 502 do not appear to change significantly from pre- to post-fire, except for a possible 503 reduction in average intensity (I_{avg} ; Fig 9 and Table 7) post-fire. Contrastingly, ΔNBR 504 distributions and K-S test illustrate an evident right-ward (increasing) shift post-fire (Fig. 505 9(d) and Table 7). The density functions of post-fire Q*/P_{Tot} and Q_{pk} are also right-shifted 506 relative to pre-fire (Fig.9 (e), (f) and Table 7). The shifts in the distribution of T_R/T_B 507 cannot be clearly identified (Fig 9(g); Table 7).

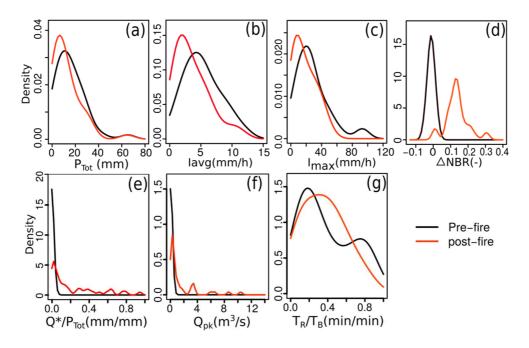




Figure 9: Density distributions comparison between pre (black) and post-fire (red) event-based precipitation (a) P_{Tot} , (b) I_{avg} , (c) I_{max} , (d) ΔNBR , and streamflow hydrograph (e) Q^*/P_{Tot} , (f) Q_{pk} and (g) T_R/T_B , properties from samples of 38 events pre- and 38 post-fire. Please refer to Table 7 with Kolmogorov-Smirnov tests for pre- and post-fire distributions comparison.

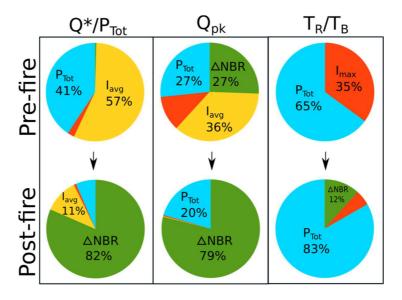
Table 7: Kolmogorov-Smirnov test results for the distributions shown in Figure 9. The seven study variables (shown in Figure 9) are replicated in column 1 (i.e. Metric). The K-S maximum distance statistic (D) is presented in column 2. The null hypothesis H₀ is that the two series are drawn from the same distribution. The K-S theoretical threshold value for α =0.005 and n=38 is D_{n, α}= 0.276.

		,
Metric	D	\mathbf{H}_{0}
P _{Tot} (mm)	0.2368	Accept
$I_{avg}(mm)$	0.3421	Reject
$I_{max}(mm)$	0.2632	Accept
ΔNBR	0.9211	Reject
Q*/P _{Tot} (mm/mm)	0.8947	Reject
$Q_{pk}(m^3/s)$	0.7632	Reject
T_R/T_B (min/min)	0.1579	Accept

519

Precipitation characteristics (i.e. P_{Tot} , I_{max} , I_{avg}) do not seem to be strongly correlated as shown by Figure C.1. The variance contribution results shown in Figure 10 and further detailed in Figure D.1 and Table D.1 help quantify the relationships between burn severity, precipitation properties and hydrologic response both pre- and post-fire. For the pre-fire events, precipitation-related predictors contribute in 98% (P_{Tot} and I_{avg}), 73% (P_{Tot} and I_{avg}) and 100% (P_{Tot} and I_{max}) to the total variability that is possible to explain of each predictand (Q^*/P_{Tot} , Q_{pk} and T_R/T_B ; see Table D.1) when the four predictors are included within the best possible linear model combinations. Contrastingly, for the postfire condition, and except for T_R/T_B , ΔNBR displaces all precipitation properties as the main driver of hydrologic response contributing 82% and 79% to the explained variability of Q^*/P_{Tot} and Q_{pk} respectively. Consistent to previous results, T_R/T_B remains mostly controlled by P_{Tot} .





533

539

Figure 10: Results of the variance contribution analysis to identify the marginal contribution of each of the four predictors (i.e. I_{avg} , P_{Tot} , I_{max} and ΔNBR) to the explained variability of the three predictands (i.e. Q^*/P_{Tot} , Q_{pk} and T_R/T_B) pre- and post-fire. Please refer to Figure D.1 and Table D.1 to see the full results including all linear model combinations and the maximum percent of explained variance of each model.

At the inter-annual scale, burn scars undergo a recovery process with vegetation re-growth eventually leading to canopy densities similar to pre-fire conditions (in the absence of species replacement or succession), and as this process occurs a hysteresis effect on the values of Q/P_{Tot} is observed. Figure 11 illustrates this process through 544 seasonal average Q/P_{Tot} plotted against the corresponding seasonal average ΔNBR 545 throughout the inter-annual catchment's evaluation period.

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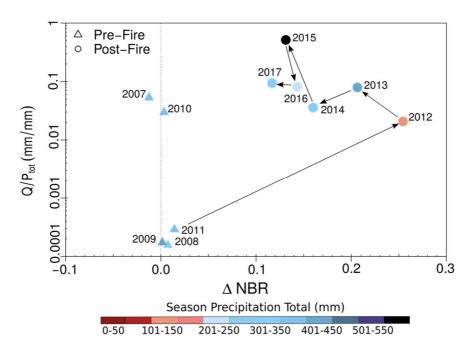


Figure 11: Seasonal Q/P_{Tot} vs. average Δ NBR for Camp Creek during six years before and six after the wildfire occurrence. Each season is classified by its total precipitation. Arrows connect the sequence of years post-fire to identify vegetation re-growth and runoff coefficient trends.

552

547

Figure 11 shows that the pre-fire years (2007 through 2011) present low values with Δ NBR ranging between -0.015 and 0.015 and relatively low Q/P_{Tot} values, except by 2007 and 2010. Between 2011 and the 2012 fire year, an increase in Δ NBR from 0.015 in 2011 to 0.26 in 2012 is associated with an increase in Q/P_{Tot} from 0.0003 to 0.02 in 2012 with this period being one of the driest in terms of low seasonal precipitation (i.e. 122 mm). The 2013 – 2017 seasonal sequence indicates a trend toward decreasing Δ NBR values, although runoff efficiencies remain higher than many of the pre-fire years shown.

561 **5. Discussion**

562 5.1 Evaluating post-fire hydrologic response shifts through NBR

563 Previous studies have described post-fire anomalies and their relation to remotely-564 sensed indices (Moody et al. 2008; Kinoshita and Hogue, 2011, Benyon et al. 2013, 565 Moody et al. 2015). For example, Moody et al. (2008 and 2015) linked pre- and post-fire 566 Δ NBR to short-term soil hydraulic alterations and anomalies in peak streamflows post-567 fire. Kinoshita and Hogue (2011) evaluated the annual evolution in runoff production of a 568 burned watershed in California but using time series of the EVI. In an effort to quantify 569 vegetation recovery, Jin et al. (2012) used multi-year times series of NBR to evaluate 570 changes in forest albedo. The results presented in this manuscript represent the first effort 571 to systematically evaluate the utility of NBR to reveal transient hydrologic catchment 572 responses observed at multiple temporal scales (i.e. from event to inter-annual) for up to 573 six years of vegetation regrowth.

574 In this study, catchment-scale hydrologic response to fire was evaluated for a 575 large sample (i.e. 76) of precipitation events and seasonal totals that were similar to pre-576 fire values in an ephemeral, mountainous, forested catchment. A six-year, pre-fire time 577 period provided a comparison timeframe to evaluate the shifts in hydrologic response at 578 the watershed scale. Regarding the first research question, results showed that a drop in 579 the catchment-average satellite-derived normalized burned ratio (NBR) is accompanied 580 by an increase in the amount of effective runoff, contributing to larger magnitude 581 streamflow responses in volume and peak flow rate suggesting a degree of coupling 582 between burn severity and hydrologic response. Additionally, flow frequencies showed a 583 consistent shift toward a larger range of exceedance probabilities post-fire, also indicating

584 a slight decrease in the number of zero flow days with respect to the 20 years of record 585 previous to the wildfire date in 2012. Previous studies had observed this effect from 586 single events or a smaller sample of events (Stoof et al., 2012; Verdin et al., 2012; Young 587 and Rust, 2012; Rosgen et al., 2013). The influence of burn severity on vegetation 588 propagates from the minute to the inter-annual time scales indicating that the mechanisms 589 of runoff generation endure for a significant period of time post-fire, regardless of the 590 precipitation characteristics. The associated increases in volume and flow rates are 591 suggestive of enhanced infiltration excess overland flow (Mayor et al., 2007). Three 592 particular events in April and May of 2010 show large runoff efficiencies, similar to post-593 fire events. We hypothesize these could be due to a prolonged snow-melt period 594 extending from late-April to early May. Out of the three evaluation metrics (i.e. Q*/P_{Tot}, 595 Q_{pk} and T_R/T_B), Q*/P_{Tot} was the most significantly affected for all precipitation events 596 post-fire. This means that the efficiency of runoff volume production is significantly 597 enhanced regardless of the precipitation total from the storm to the seasonal scales. This 598 result is similar to previous studies investigating fire impacts on stream hydrology. 599 Kinoshita and Hogue (2015) also noted elevated streamflow during low flow 600 seasons for up to 10 years post-fire, producing a newly perennial system. However, 601 the increase in peak streamflow shown in Camp Creek's post-fire was not observed 602 in Lane et al. (2006) or Kinoshita and Hogue (2015). The reasons both studies 603 provide for this is the rapid breakdown or altogether absence of fire-induced soil 604 hydrophobicity.

605 On the other hand, although Cydzic and Hogue (2009) found significant 606 reductions in basin lag time for six large to moderate post-fire runoff events, attributing

607 the change to reduced infiltration associated with loss of vegetation and increased 608 imperviousness, the present study did not find significant changes in the time to peak 609 flow relative to the total time base of the hydrographs. One reason for this could be the 610 ephemeral condition of Camp Creek added to the relative small size of the catchment as 611 compared to larger watersheds studied by Cydzic and Hogue.

612 In answering the second science question related to the utility of ΔNBR to quantify inter-613 annual runoff alterations and hydrologic recovery post-fire, the development of a 614 variance contribution analysis provided further element tools to link satellite-derived 615 burned severity to the observed shifts in hydrologic response. For the case of this single 616 wildfire event, no direct mathematical relations were found to exactly predict the 617 magnitude of expected change for particular combinations of ΔNBR , P_{Tot}, I_{max} or I_{avg}. 618 Nonetheless, the cross-scale analyses conducted outlined that observed event-scale shifts 619 in watershed stream response characteristics are found to shape longer-term runoff ratios 620 post-wildfire. Further, post-fire vegetation recovery is a key determinant in the evolution 621 of runoff as noted by Wittenberg et al. (2007), Casady et al. (2009) and Kinoshita and 622 Hogue (2011) that conducted their analyses through time series of Δ EVI. We selected 623 paired-basin Δ NBR as it removes the effects of phenology and enhances the distinction 624 between burned and unburned surfaces with the addition of the mid-infrared spectral 625 band. As Δ NBR decreased several months after the wildfire occurred, seasonal average 626 runoff began recovering toward pre-fire values following a hysteresis pattern (see Figure 627 11) with the burn severity but also precipitation depth as exemplified by the exceptionally wet year of 2015. Very strong positive ENSO phases, like the one in 2015, enhance the 628 629 effect of fire severity through significant increases in Q/P_{Tot} of up to, at least, one order of 630 magnitude relative to runoff fractions during post-fire years. The Waldo Canyon fire was 631 a rather severe wildfire judging by the long recovery time relative to normal-year 632 hydrologic (e.g. non-ENSO; 2008, 2009, 2010) pre-fire conditions. We argue that this 633 hysteresis effect rules post-fire hydrologic responses as measured by Q*/P_{Tot} and possibly 634 also Q_{pk} . The time and magnitude of such an effect depends on the severity of the burn 635 and vegetation species pre- and post-fire. While a clear mathematical or statistical 636 relationship between the Δ NBR, Q*/P_{Tot} and Q_{pk} was not found, the balance of evidence 637 shows correlation between hydrologic response properties and ΔNBR . The lack of 638 identifying a clear relation can be attributed to the highly complex watershed response 639 processes and their feedbacks that rule post-fire runoff responses. Additional factors 640 include antecedent soil moisture conditions and snow processes, but also differences in 641 soil hydraulic properties determined by the short-term presence of an ash layer (Woods 642 and Balfour, 2010; Ebel et al., 2012). As noted by previous authors, post-fire event runoff 643 generating processes and magnitudes are dynamic (Shakesby and Doerr, 2006; Larsen et 644 al., 2009) and therefore cannot likely be captured by a single vegetation metric.

645 Although the results presented herein are based on a single basin's response, the 646 study underlines the potential use of paired-basin ANBR for process understanding and 647 predicting acute and chronic anomalies in vegetation and runoff and their linked recovery 648 after a wildfire event. Along with precipitation characteristics and antecedent soil 649 moisture conditions, NBR is useful in determining mean expected shifts in effective 650 runoff and peak streamflow from the event to the inter-annual time scales. NBR can also be used to reveal the vegetation recovery pathway that restores hydrologic conditions (i.e. 651 652 hysteresis curve), demonstrating its usability to track the evolution of the annual runoff 653 anomalies relative to pre-fire conditions. Despite no straightforward relations being found 654 between precipitation characteristics, antecedent soil moisture and burn severity, the 655 findings of this study are of utility to hyper-resolution, process-based modelers (including 656 the newly developed National Water Model (Gochis et al. 2018)) that account for time-657 evolving vegetation and land cover status. Linking satellite-derived indices to such 658 hydrologic response metrics could also result in improvements of the rainfall-runoff 659 hydrograph calculation methods (e.g. U.S. Natural Resources Conservation Service 660 (SCS) method and/or unit and synthetic hydrographs) after a wildfire event. In an 661 operational context, land managers and decision makers could rely on the usability of 662 NBR and \triangle NBR to better monitor, understand and estimate both event and long-term 663 responses of burned watersheds, particularly those that are ungauged.

664

665 5.2 Study scope and limitations

666 Results provide consistent evidence of the exacerbated post-fire runoff responses 667 and their occurrence from the hourly to the inter-annual time scales. However inherent 668 uncertainties arise when taking a lumped approach to vegetation dynamics. For example, 669 within-burn heterogeneity might play a role in enhancing or inhibiting water flow 670 connectivity after fire. A distributed approach could help shed light on the post-fire 671 vegetation dynamics of each sub-catchment, which might not be well captured by the 672 catchment-average Δ NBR. Additionally, uncertainties are likely to influence results when 673 three point-based rain gauges are applied to analyze catchment-scale streamflow 674 dynamics. Spatially distributed, bias-corrected radar-based precipitation data is preferred 675 when quantifying catchment scale streamflow dynamics because it provides an estimation of rainfall spatial variability with respect to the burned areas. Further, the present study
did not explicitly consider the effects of the fire severity on the soil hydraulic properties
as directly measured from the terrain. These results underscore the importance of in situ
soil measurements and field validated data for storm-scale hydrologic projections.

680

681 6. Conclusions

Acute (short-term) and chronic (long-term, 6 years post-fire) transient streamflow shifts caused by wildfires are evaluated for a forested, mountain catchment under warm season precipitation. Along with the quantified shifts, the utility of the remotely sensed Normalized Burn Ratio (NBR) is explored and quantified from the event to inter-annual scales to study the transient post-fire changes in streamflow. Results can be summarized as:

Observed reductions in NBR due to wildfire are concurrent with increases in runoff
 production (i.e. Q*/P_{Tot}) under similar event precipitation depths, intensities and
 antecedent soil moisture conditions.

691 2. Daily average flow duration curves show increases of an order of magnitude in the692 maximum streamflow values for the same exceedance probability events.

693 3. Increases in effective runoff (Q^*/P_{Tot}) and runoff fraction (Q/P_{Tot}) are observed at the 694 event, daily, monthly, and seasonal scales for similar precipitation totals pre- and 695 post-fire. The largest absolute increase in runoff production is observed at the event 696 scale (Q^*/P_{Tot}) increases in two orders of magnitude in average) indicating a 697 significant shift in flash flood probability post-fire. 4. Of the three response hydrograph metrics, Q^*/P_{Tot} and Q_{pk} were the ones that illustrated the largest positive changes. This indicates that regardless of the precipitation type or intensity, enhanced runoff generation mechanisms were able to transform available precipitation into quick flow resulting in taller, wider response hydrographs. The response time was not necessarily identified as a critical shift variable in this catchment, meaning that travel and residence times may not necessarily be affected by wildfires on ephemeral catchments of this type.

5. Density distributions and K-S tests showed significant positive shifts for Δ NBR, Q*/P_{Tot} and Q_{pk} post-fire, despite P_{Tot}, I_{avg}, I_{max} showed average decreases from a set of storm events analyzed post-fire.

6. A variance contribution analysis further supports the strong dependence of hydrologic
responses on precipitation properties pre-fire, but ΔNBR was a primary contributor on
post-fire hydrologic response.

711 7. A hysteresis effect was found at the inter-annual scale between the seasonal runoff 712 fraction (i.e. Q/P_{Tot}), ΔNBR and P_{Tot} that illustrates the strong controls of soil and 713 vegetation conditions on the corresponding runoff production at multiple temporal 714 scales post-fire. For this particular case, 6 years of vegetation re-growth and soil 715 hydraulic recovery post-fire are not enough time to return to pre-fire hydrologic 716 conditions as the runoff mechanisms enhanced by the wildfire are still producing 717 larger streamflow values comparable to pre-fire years with exceptional rainfall or 718 snow seasons.

720 In summary, paired-basin ΔNBR was found to be of high utility for hydrologic 721 analyses as it indicated vegetation status through its temporal correlation with the 722 efficiency of runoff generation. This finding supports the hypothesis that the remotely-723 sensed ΔNBR could potentially be used to gain process understanding and improve 724 predictions in burned basins as they are susceptible to yielding much greater runoff ratios 725 than what is expected during pre-fire conditions. $\triangle NBR$ can also be used to monitor the 726 basin's vegetation and hydrologic response recovery and improve transient modeling of 727 the effects of vegetation status on hillslope and channel runoff. Future work could focus 728 on conducting physically-based, distributed hydrologic observations and modeling to 729 disaggregate the spatio-temporal variability of runoff-generating mechanisms that are 730 responsible for the shifts observed across the temporal scales illustrated in this study.

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737	and suggestions significantly improved the quality of this manuscript.
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739	
740	

743 Appendix A: Waldo Canyon Burn Scar

Areal extent of the Waldo Canyon wildfire obtained through the United States Forest Service (USFS) Remote Sensing Applications Center (Figure A.1). This classification indicates that Camp Creek watershed was primarily affected by moderate and high burn severities.

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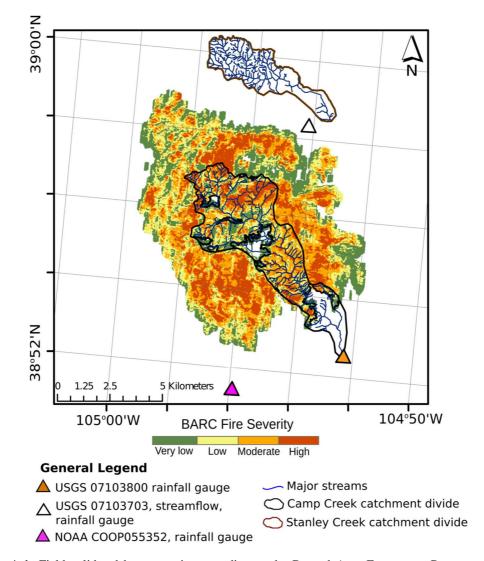


Figure A.1: Field validated burn severity according to the Burned Area Emergency Response (BAER)
team for the Waldo Canyon Fire and accessed through the Remote Sensing Applications Center (RSAC)
user interface.

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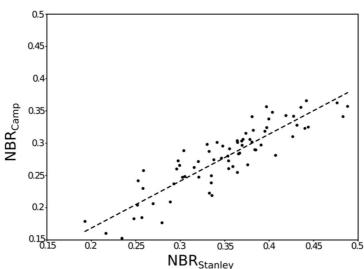
Appendix B: Unburned, adjusted NBR values from control catchment 754

755 For the post-fire period, unburned values at Camp Creek are obtained from a 756 linear regression with Stanley Creek, a neighboring, unburned catchment with similar 757 vegetation cover and type and therefore spectral signature.

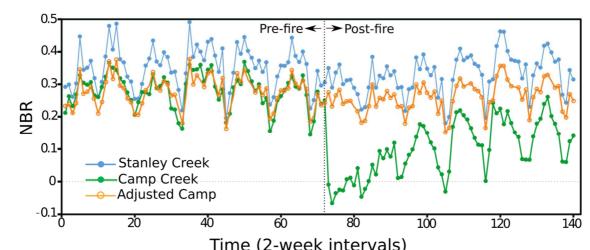
758

759 Table B.1: Regression models for simultaneous catchment average NBR values at Camp and Stanley 760 Creek basins

Model	Expression	\mathbf{R}^2
Linear	NBR _{camp} =0.0228 + 0.7281NBR _{Stanley}	0.80
Exponential	NBR _{camp} = $0.1 + e^{2.85941NBRStanley}$	0.76
Logarithmic	NBR _{camp} =0.2408 Ln(NBR _{Stanley})+0.5349	0.80
Polynomial	$NBR_{camp} = -0.9476 (NBR_{Stanley})^2 + 1.3786 NBR_{Stanley} - 0.0848$	0.80
Power	NBR _{Camp} =0.7577(NBR _{Stanley}) ^{0.959}	0.79

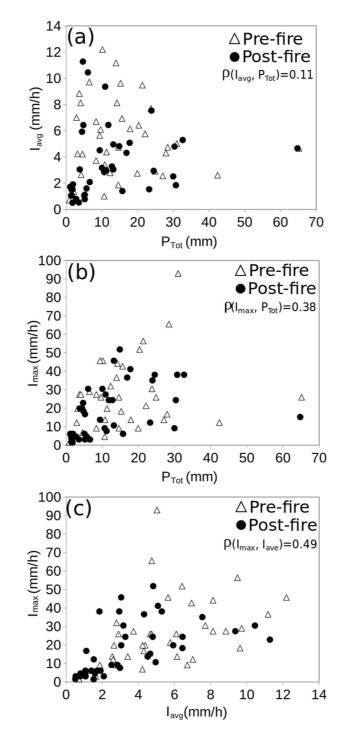


NBR_{Stanley} Figure B.1: Scatterplot of simultaneous catchment average NBR values at Camp (Y-axis) and Stanley (X-762 763 764 765 axis) Creek watersheds, and linear fit during pre-fire (2006-2012) conditions.



768 769 770 771
 Time (2-week intervals)

 Figure B.2: Time series of biweekly, simultaneous catchment average NBR values at Stanley (unburned)
 and Camp (burned) creeks, along with the linearly adjusted time series at Camp Creek from the values at Stanley Creek. Values are only for the warm season of each year between 2006 and 2018.

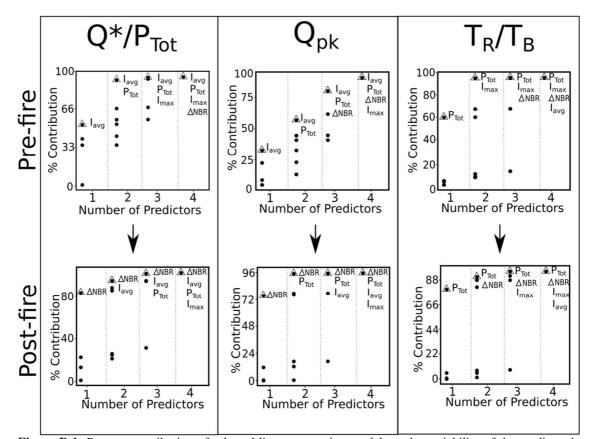


801 Appendix C: Event precipitation characteristics pre- and post-fire in Camp Creek

802

Figure C.1: Event-based scatterplots between (a) I_{avg} and P_{Tot} , (b) I_{max} and P_{Tot} , (c) I_{max} and I_{avg} , for 38 isolated precipitation events pre- (triangles) and 38 post- (circles) fire. Pearson correlation coefficients (ρ) are shown within each plot.

809



811 812 Figure D.1: Percent contribution of selected linear regression models to the variability of the predictands (i.e. Q^*/P_{Tot} , Qpk and T_R/T_B columns) when 1, 2, 3 and 4 predictors (i.e. I_{avg} , P_{Tot} , I_{max} and ΔNBR) are included in all possible linear combinations for pre- and post-fire (rows) events. Triangles indicate the best models for each number of predictors and their percent contribution to each of the predictands.

Table D.1: Coefficient of determination (R²) as indicative of the fraction of explained variance between the best combination of the four predictors (see Figure D.1.) and each of the predictands for pre- and post-fire conditions.

	Q*/P _{Tot}	Qpk	T_R/T_B
Pre-fire	0.58	0.66	0.55
Post-fire	0.57	0.70	0.74

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