

Universidade de São Paulo
Programa de Pós-Graduação em Ecologia

**Diversificação de espécies e da morfologia em serpentes da família Viperidae:
padrões e processos**

*Species and morphological diversification in snakes of the family Viperidae: patterns
and processes*

Laura Rodrigues Vieira de Alencar



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**Diversificação de espécies e da morfologia em serpentes da família Viperidae:
padrões e processos**

Laura Rodrigues Vieira de Alencar

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Orientador: Prof. Dr. Marcio R. C. Martins

Co-orientador: Prof. Dr. Tiago B. Quental

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Prof(a). Dr(a).

Prof(a). Dr(a).

Prof(a). Dr(a).

Prof(a). Dr(a).

Prof. Dr. Marcio Roberto Costa Martins
Orientador

*“There ought to be something very
special about the boundary
conditions of the universe - and
what can be more special than that
there is no boundary?”*

Stephen Hawking

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RESUMO

A diversidade de espécies e fenotípica pode variar consideravelmente entre grupos taxonômicos e ao longo do tempo em uma mesma linhagem. O estudo de tais variações tornou-se um dos principais objetivos da biologia evolutiva fornecendo informações importantes a respeito dos possíveis mecanismos que regulam a biodiversidade. Dessa forma, o objetivo geral da presente tese foi investigar os padrões da diversificação de espécies e da morfologia em um grupo cosmopolita de serpentes, a família Viperidae, e os potenciais processos subjacentes. Primeiramente, (1) reconstruímos as relações filogenéticas e estimamos os tempos de divergência entre as linhagens da família Viperidae utilizando uma abordagem Bayesiana. (2) Aplicando um método recentemente desenvolvido (BAMM), exploramos como as taxas de especiação e extinção variaram ao longo da radiação do grupo inferindo os possíveis processos reguladores. Por fim, (3) analisamos se a evolução do tamanho do corpo e as taxas de especiação variam nos diferentes habitats ocupados pelos viperídeos (terrestres vs arborícola). Nesta tese geramos a filogenia molecular de viperídeos mais completa até o momento utilizando sequências para 11 genes mitocondriais e nucleares abrangendo 79% das espécies viventes (264 terminais) e todos com exceção de um gênero. De maneira geral, foi possível obter relações filogenéticas robustas para o grupo com a maioria dos gêneros sendo monofilética. Os tempos de divergência obtidos indicam que os viperídeos começaram a diversificar em meados do Paleoceno tardio/meio do Eoceno inferindo idades um pouco mais tardias que o encontrado em estudos anteriores. Durante a radiação do grupo, um aumento nas taxas de especiação parece ter ocorrido durante a diversificação dos crotalíneos (pit vipers) em decorrência não só da evolução das fossetas loreais mas também como resultado de mudanças geológicas e climáticas na Ásia e da invasão do novo mundo. Após este rápido aumento inicial, as taxas de especiação desaceleraram em direção ao presente. Por fim, os resultados aqui apresentados indicam que apesar dos habitats arborícolas limitarem a evolução morfológica nos viperídeos, a evolução da arborealidade parece não afetar as taxas de especiação que permanecem similares entre linhagens arborícolas e terrestres. Isto sugere dois cenários: (1) a especiação acontece de forma independente das mudanças morfológicas nos viperídeos; ou (2) o isolamento geográfico seria um mecanismo importante na diversificação de linhagens arborícolas contrabalançando decréscimos nas oportunidades de especiação possivelmente relacionados às pressões seletivas impostas pelo ambiente arborícola. A presente tese contribui para entendermos mais sobre como evoluíram os viperídeos ao longo dos seus ~50 milhões de anos. Além de propor cenários e hipóteses a serem futuramente explorados com os viperídeos, elaboramos uma discussão ampla e conceitual a respeito dos possíveis mecanismos por trás da diversificação de espécies e da morfologia que poderiam também ser contemplados para outros grupos de organismos. Portanto, a

presente tese contribui não só para entendermos os mecanismos que geram e mantêm a diversidade de serpentes, mas também para enriquecer a discussão dos mecanismos que geram e mantêm a biodiversidade como um todo.

ABSTRACT

Species and phenotypic diversity may vary considerably between taxonomic groups and through time for a given lineage. The study of such variation became one of the main goals of evolutionary biology and provides important information related to the possible mechanisms regulating biodiversity. The general goal of the present thesis was to investigate the patterns of species and morphological diversification in a cosmopolitan group of snakes, the family Viperidae, and the potential underlying processes. First, (1) we estimated the phylogenetic relationships and divergence times between lineages of the family Viperidae using a Bayesian approach; then we (2) applied a recently developed method (BAMM) to explore how speciation and extinction rates varied during the radiation of the group suggesting possible underlying processes. Finally, (3) we analyzed if body size evolution and speciation rates showed distinct patterns among vipers occurring in different habitats (terrestrial vs arboreal). Herein we generated the most complete molecular phylogeny for vipers until this moment using sequences from 11 mitochondrial and nuclear genes comprising 79% of extant species (264 terminals) and all except one genus. In general, we were able to recover well supported phylogenetic relationships with most genera being monophyletic. Divergence time estimates suggested that vipers started to diversify around the late Paleocene/middle Eocene finding older ages than previous studies. During the group radiation, an increase in speciation rates seems to have occurred during the diversification of crotalines (pit vipers) not only due to the evolution of loreal pits but also as a result of climatic and geological changes in Asia and the invasion of the New World. After this rapid initial increase, speciation rates decelerated toward the present. Lastly, the results presented here suggest that although arboreal habitats constrain morphological evolution in vipers the evolution of arboreality does not seem to affect speciation rates, which remain similar among arboreal and terrestrial lineages. Our results suggest two distinct scenarios: (1) speciation could be independent of morphological evolution in vipers; or (2) geographic isolation would be an important mechanism underlying species diversification in arboreal lineages offsetting decreases in speciation opportunities potentially related to the selective pressures imposed by the arboreal environment. The present thesis contribute to increase our understanding about how vipers evolved during their ~50 million years. In addition to providing scenarios and hypotheses to be further explored with vipers, we elaborated a broad and conceptual discussion about the possible mechanisms underlying species and morphological diversification that might apply to other groups of organisms. Therefore, this thesis comprises a contribution that goes beyond the understanding of mechanisms generating and maintaining the diversity of snakes, but will hopefully enrich the discussion of mechanisms that generate and maintain biodiversity as a whole.

O contraste no número de espécies entre os diferentes grupos de organismos constitui um dos padrões mais notáveis e intrigantes da diversidade biológica (Alfaro et al. 2009, Rabosky et al., 2013). Tais diferenças podem, dentre outros fatores, ser explicadas pelas diferentes taxas de diversificação (especiação menos extinção) associadas a cada grupo (Alfaro et al., 2009, Rabosky, 2014). Apesar de diferenças nas taxas médias de diversificação estarem por muitas vezes associadas a riqueza de espécies nos grupos atuais, tanto as taxas de especiação quanto as de extinção podem variar consideravelmente ao longo do tempo (e.g. Rabosky et al., 2014, Silvestro et al., 2015). Diversos são os mecanismos que poderiam explicar tais variações temporais (e.g. Glor, 2010, Losos and Mahler, 2010, Moen and Morlon, 2014). A invasão de novas áreas e o surgimento de inovações chave (Simpson, 1953), a competição por recursos (Silvestro et al., 2015) ou alterações bruscas no clima (Moen and Morlon, 2014), são citados como fatores que poderiam alterar as taxas de especiação e extinção de um dado grupo.

Assim como a diversidade de espécies, a diversidade fenotípica também pode variar consideravelmente entre grupos e ao longo do tempo. Entretanto, um grupo rico em espécies não necessariamente é também diverso fenotipicamente (e.g. *Plethodon*, Kozak et al., 2006) e vice-versa. De fato, a diversificação fenotípica e de espécies podem fornecer informações importantes a respeito dos possíveis mecanismos geradores de biodiversidade. Por exemplo, as famosas “radiações adaptativas” (see Glor, 2010, Losos and Mahler, 2010) estão frequentemente associadas a uma “explosão” tanto de espécies quanto de morfologia após a colonização de novas áreas ou zonas adaptativas (Simpson, 1953). Esta “explosão” é comumente relacionada à hipótese das oportunidades ecológicas (Schluter, 2000), onde as populações ocupam diferentes nichos culminando na formação de diversas espécies. Entretanto, existem situações em que a invasão de uma nova zona adaptativa (Simpson, 1953) limita ao invés de impulsionar a diversificação morfológica (e.g. Collar et al., 2009). Por fim, certos grupos podem ser considerados como radiações “não adaptativas” onde sugere-se que eventos de vicariância sucessivos podem ter culminado na formação de espécies com pouca variação morfológica (Kozak et al., 2006, Rundell and Price, 2009).

Para entendermos como evolui a biodiversidade é muito importante conhecer e entender a história filogenética das diferentes linhagens que a compõe. As árvores filogenéticas permitem que pesquisadores explorem mais profundamente a origem da diversidade biológica vivente e extinta, bem como os seus possíveis mecanismos reguladores. A partir da década de 80, o “pensamento filogenético” (“phylogenetic tree thinking”, Losos, 2011) tornou evidente a importância de incorporar as relações filogenéticas das espécies em inúmeras linhas de pesquisa. A filogenética

deixou de ser área exclusiva da sistemática passando também a ser interesse de outras áreas como a ecologia. Desde então, inúmeros foram os trabalhos descrevendo métodos visando incorporar uma perspectiva filogenética nas análises evolutivas, como o clássico trabalho de Felsenstein (1985). Atualmente é praticamente impossível publicar um trabalho comparativo (e.g. comparações entre espécies) sem utilizar uma abordagem filogenética (Losos, 2011). Essa crescente necessidade de se obter filogenias levou ao aumento tanto na disponibilidade de sequências moleculares como de métodos filogenéticos e comparativos. Hoje, é possível baixar sequências diretamente de bases de dados públicas (e.g. GenBank) e gerar árvores filogenéticas rapidamente para muitos grupos de estudos (e.g. Pearse and Purvis, 2013). Além disso, existem métodos que prometem inserir na filogenia táxons sem sequências moleculares usando apenas informações de hierarquia taxonômica (e.g. Jetz et al. 2012, mas ver crítica em Rabosky, 2015).

Apesar de toda a sua importância, as árvores filogenéticas muitas vezes são mal-interpretadas (Baum, 2014, Omland, 2014) e nos últimos anos diversos trabalhos foram publicados com o intuito de esclarecer alguns dos problemas conceituais (e.g. Baum et al., 2005, Baum e Smith, 2013, ver referências em Baum, 2014). Como frisado por Omland (2014) “Most biologists need not know how various phylogenetic algorithms work, but all biologists should know what trees represent and how trees inform our understanding of the process of evolution”. Apesar disso, dada a aparente rapidez com que filogenias podem ser geradas, muitas vezes estas deixam de ser o foco principal de um estudo tornando-se item da sua metodologia. No entanto, a utilização das filogenias apenas como uma ferramenta metodológica não pode esconder o seu principal objetivo e significado, além da real complexidade que existe na sua elaboração, nas decisões e nos cuidados que devem ser tomados por parte dos pesquisadores. Acima de tudo a filogenia é uma hipótese, e todo o trabalho que baseia-se nesta hipótese pode conter grandes deficiências se a árvore filogenética também as tiverem ou se não for bem interpretada. Assim, torna-se extremamente importante dedicar especial atenção à filogenia que está sendo usada no trabalho, cuidado esse por muitas vezes ignorado.

Entender os processos que moldam a biodiversidade depende também da habilidade em explorar as possíveis relações entre a evolução biológica e os padrões históricos (Sauquet et al., 2012). Para tal, integrar o uso de fósseis com filogenias moleculares incorporando dessa forma informações sobre a idade de um ou mais nós da filogenia, representa um passo importante nessa direção. Como resultado, datar uma filogenia significa converter seus comprimentos de ramo para o tempo geológico absoluto (em milhões de anos, e.g. Hedges and Kumar, 2009), sendo possível conhecer os tempos de divergência entre as linhagens. Trabalhos recentemente publicados elaboram estratégias para a escolha dos pontos de calibração (e.g. fósseis, eventos biogeográficos) como forma de diminuir o risco de incluir erros nas datações das filogenias (e.g. Parham et al., 2012,

Sauquet et al., 2012). Além das calibrações em si, a escolha dos *priors* ou idades mínimas e máximas das calibrações são fonte de erros potenciais e também devem ser investigadas com cuidado (Ho and Phillips 2009). Dessa forma, calibrar a filogenia em relação ao tempo é também passo crucial em trabalhos com perspectivas evolutivas (Sauquet et al., 2012).

Sendo o objetivo geral desta tese explorar os padrões por trás da diversificação de espécies e da morfologia em serpentes da família Viperidae e os possíveis processos que geram estes padrões, dedicamos o primeiro capítulo a reconstruir relações filogenéticas robustas entre as linhagens do grupo. Para tal, usamos um método filogenético Bayesiano integrando sequências moleculares de bases públicas e inéditas para 79% das espécies que compõe a família. Cabe ressaltar que uma curadoria cuidadosa foi realizada após a obtenção dos dados moleculares. Esta curadoria se mostrou importante devido à erros inerentes a grandes bases de dados. Além disso, utilizando fósseis, datamos tais relações filogenéticas o que permitiu também investigarmos as divergências entre as linhagens do grupo ao longo do tempo geológico (Capítulo I). De posse da filogenia datada, exploramos a dinâmica de diversificação que caracterizou a radiação evolutiva dos viperídeos (Capítulo II). Especificamente, o segundo capítulo foca em como e por que as taxas de especiação e extinção variaram ao longo do tempo nesta família de serpentes. Por fim, dado que o fenótipo pode sugerir possíveis paisagens adaptativas nas quais as linhagens diversificaram, analisamos se a evolução do fenótipo e as taxas de especiação dos viperídeos apresentam padrões similares em linhagens ocupando um mesmo habitat porém distintos entre eles (Capítulo III).

Modelo de estudo

A família Viperidae representa um dos grupos mais notáveis de serpentes do mundo, por sua alta diversidade de espécies (329 espécies), sua ampla distribuição (Uetz e Hosek, 2014), seu mecanismo complexo de inoculação de veneno (Kardong e Lavin-Murcio, 1993) e por sua grande importância médica (Wüster et al., 2008). Fazem parte dessa família serpentes conhecidas pelo público em geral como as cascavéis e jararacas. Acredita-se que o grupo tenha se originado na Ásia ou na África e atualmente ocorre tanto no Velho quanto no Novo Mundo (Greene, 1997, Wüster et al., 2008). Os viperídeos apresentam grande variação no número de espécies que compõe suas diferentes linhagens. A subfamília Viperinae, restrita ao Velho Mundo, possui em torno de 98 espécies, enquanto Crotalinae, amplamente distribuída no Velho e Novo Mundo, possui em torno de 232 espécies. Mais dramática ainda é a diferença na riqueza de espécies entre crotalíneos e sua subfamília irmã, Azemiopinae, a qual compreende apenas duas espécies restritas ao sudeste asiático. A alta riqueza da subfamília Crotalinae é frequentemente associada a eventos que potencialmente levaram a um aumento da formação de novas espécies como o surgimento da

fosseta loreal, característica exclusiva deste grupo, e a invasão do Novo Mundo (Greene, 1997, Burbrink et al., 2012).

Além da grande variação na riqueza de espécies, os viperídeos também impressionam pela diversidade em sua ecologia. Espécies deste grupo ocorrem nas maiores latitudes e altitudes já registradas para qualquer outra serpente: acima de 65° N em *Vipera berus* e 47° S em *Bothrops ammodytoides*, e até 4800 m de altitude em *Gloydius himalayanus* no Nepal e 4570 m em *Crotalus triseriatus* no México; Maritz et al., *in review*). Abrangem espécies tanto ovíparas quanto vivíparas (Fenwick et al., 2012) e possuem uma dieta extremamente ampla incluindo, por exemplo, mamíferos, répteis, anfíbios, aves e invertebrados, sendo algumas espécies mais generalistas que outras (Greene, 1997, Martins et al., 2002). A maior parte das espécies é principalmente terrestre, mas linhagens associadas ao ambiente arborícola surgiram em pelo menos oito gêneros (ver Capítulo III). Por ser um grupo excepcionalmente heterogêneo tanto no número de espécies que compõe cada linhagem quanto em relação à ecologia, a família Viperidae é um ótimo modelo para explorar os padrões e processos envolvidos na diversificação de espécies e da morfologia em serpentes.

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MOLECULAR PHYLOGENY AND DIVERGENCE TIME ESTIMATES OF VIPERS (VIPERIDAE): A BAYESIAN APPROACH

Alencar, L.R.V., Quental, T., Grazziotin, F.G., Alfaro, M.L., Martins, M., Venzon, M., and Zaher, H.

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Abstract. Snakes of the cosmopolitan family Viperidae comprise around 329 venomous species distributed in three subfamilies, Azemiopinae, Viperinae, and Crotalinae. Phylogenetics of vipers have greatly improved in the past decades due to the access to new DNA sequences. Here, we present the most complete molecular phylogeny for the Viperidae to date comprising sequences from nuclear and mitochondrial genes representing 79% of all living vipers and all except one genus. We also investigated the time of divergence between lineages using six fossils to time-calibrate the tree. Our phylogenetic analyses retrieved high support values for the monophyly of the family and subfamilies, and 22 out of 27 genera for which we could test for monophyly, as well as well-supported intergeneric relationships throughout the family. In contrast to other studies, we were able to recover a strongly supported sister clade to the New World crotalines that comprises *Gloydius*, *Ovophis*, *Protobothrops* and *Trimeresurus gracilis*. Time of divergence estimates suggested that crown vipers started to radiate around the late Paleocene to middle Eocene with subfamilies crown group ages most likely dating back to the Eocene. The invasion of the New World might have taken place sometime close to the Oligocene/Miocene boundary. Although some phylogenetic aspects still require further investigation (e.g. relationships among New World vipers), the present study represents an important addition to the knowledge of the phylogenetic relationships and tempo of diversification in vipers.

INTRODUCTION

Vipers form a monophyletic lineage of venomous snakes comprising about 329 species distributed worldwide. Because vipers are considered a medically important group, different aspects of their biology have been widely studied (e.g. Fenwick et al., 2012; Greene, 1997; Martins et al., 2001) but some aspects of their phylogenetic relationships are still poorly understood. Species are

currently arranged in 34 genera belonging to three subfamilies: Viperinae, Azemiopinae, and Crotalinae (Uetz and Hosek, 2014). Viperines, or the "true vipers", comprise 98 species whereas Azemiopinae comprises only two and both subfamilies are restricted to the Old World. Crotalinae, or the "pit vipers", is the most diverse and widely distributed lineage of vipers, comprising about 229 species (~70% of all viperid species) (Campbell and Lamar, 2004; Phelps, 2010; Uetz and Hosek, 2014) occurring both in the Old and New World.

A robust understanding of the phylogenetic affinities of vipers is important not only for systematics itself but also for ecology and macroevolution, and even for other fields not directly related to biology such as medicine, pharmacology, and toxicology (Castoe and Parkinson, 2006). The systematics of vipers has greatly improved in the past decades, especially regarding the phylogenetic position and monophyly of the three subfamilies. Although previous studies based on both morphological and molecular evidence suggest that *Causus* should comprise a separated subfamily (e.g. Groombridge, 1986, Herrmann and Joger, 1995), more recent studies now agree in the placement of *Causus* inside Viperines (although its position within the subfamily remains under debate, see Herrmann et al., 1999 and Wüster et al., 2008, Pyron et al., 2013). The genus *Azemiops* has been considered as part of viperines but is now undoubtedly considered as a distinct subfamily closely related to the crotalines (Liem et al., 1971, Orlov et al., 2013, Pyron et al., 2013). However, crotalines relationships closer to the root have been a controversial issue, in particular with respect to whether *Hypnale*, *Garthius*, *Deinagkistrodon*, *Calloselasma* and *Tropidolaemus* form a monophyletic clade (Malhortra and Thorpe, 2004, Malhortra et al., 2010) or a paraphyletic configuration (Parkinson et al., 2002, Castoe and Parkinson, 2006). Morphological studies suggested a polyphyletic origin of NW pit vipers (e.g. Brattstrom, 1964, Burger, 1971) but molecular works strongly support the monophyly of the clade suggesting a single invasion of the New World via the Beringian Land Bridge, which seems to be the predominant view (Parkinson et al., 2002, Castoe and Parkinson, 2006, Wüster et al., 2008). However, neither morphological nor molecular studies were able to confidently recover the intergeneric relationships among most NW genera (e.g. Gutberlet and Harvey, 2002, Castoe and Parkinson, 2006).

In the past years, the access to new Viperidae DNA sequences has greatly improved the limited phylogenetic inference done solely based on morphological data (see Castoe and Parkinson, 2006) but few studies (Fenwick et al., 2012; Pyron et al., 2013; Wüster et al., 2008) that have used molecular data have investigated the phylogenetic relationships of vipers in a broader phylogenetic context. The pioneering study by Wüster et al. (2008) included all except one Viperidae genera in their molecular analysis but was limited to 85 species and therefore explored the phylogenetic relationships among higher taxa. Although recent works by Fenwick et al. (2012) and Pyron et al. (2013) included 218 and 209 Viperidae terminals respectively, the phylogenetic relationships of

more inclusive lineages and the tempo of diversification underlying the divergence among those lineages are still unclear and debatable.

The choice of proper fossils and/or biogeographic events (Benton et al., 2009; Ho and Phillips, 2009; Sauquet et al. 2012) is of central importance in dating analyses because it is not possible to estimate absolute ages from molecular data alone (Ho and Phillips, 2009). Wüster et al. (2008) used four fossils and two biogeographic events to calibrate their genus-level tree and Fenwick et al. (2012) two fossils and one biogeographic event. These authors found different diversification times for some lineages (e.g. Viperidae crown group), which could be related to differences in sampling effort or choice of calibration points (Parham et al., 2012; Sauquet et al., 2012). A survey of the snake fossil record suggests it is possible to use additional fossils for conducting dating analysis of vipers (see Material and Methods) avoiding biogeographic events, which have been suggested to be problematic on dating studies (Sauquet et al. 2012). A calibration setting comprising only fossils and a wider inclusion of current species should therefore greatly improve our understanding of the tempo of vipers' diversification.

Herein we assembled the most complete time-calibrated molecular dataset of the family Viperidae to investigate the phylogenetic relationships and time of divergence among vipers using a Bayesian approach.

MATERIAL AND METHODS

Taxon sampling and data acquisition

The Reptile Database (Uetz and Hosek 2014) currently recognizes 329 species in the family Viperidae. From those, we included 261 species as terminal taxa in the ingroup of our phylogenetic tree. As in Pyron et al. (2013), we also included *Bothrops colombiensis*, *B. isabelae* and *Crotalus tortugensis* as distinct species although some researchers (e. g., Campbell and Lamar, 2004; Castoe and Parkinson, 2006; Uetz and Hosek 2014) consider the first two as synonyms of *B. atrox* and the third as synonym of *C. atrox* (but see Grismer, 1999). Thus, our sampling for Viperidae encompasses 264 taxa corresponding to 79.5% of the diversity presently described for the family. This taxon sampling comprises all the three known subfamilies with the following sampling schemes (sampled species/number of described species): Azemiopinae (1/2), Viperinae (72/98), and Crotalinae (191/232). Additionally, we included as outgroup 97 species from different families: Boidae, Elapidae, Colubridae, Dipsadidae, Homalopsidae, Natricidae, Atractaspididae, Lamprophiidae, Psamphiidae, and Xenodermatidae. The species *Indotyphlops braminus* (Scoleophidia: Typhlopidae) was used to root our phylogenetic tree (Pyron et al., 2013).

Our molecular matrix is composed of sequences from 11 genes, six mitochondrial (12S, 16S, cytb, cox1, nd2, nd4) and five nuclear (bndf, c-mos, jun, nt3, rag1). We used sequences for all species of Viperidae available in GenBank up to April 2014 (Table S1). We also included sequences available in the Barcode of Life Database (BOLDSYSTEMS, <http://www.boldsystems.org/>) to complement information for the Cox1 gene (Table S1). We provided new DNA sequences for 27 viper species (Table S1) for eight genes, including a species not previously included in GenBank (*Causus lichtensteini*), totaling 165 new sequences for vipers. All new sequences were obtained following standard PCR and sequencing protocols as described in Graziotin et al. (2012). Both strands of the PCR products were sequenced, and the trimming and assembling procedures were performed using the default parameters in the program GENEIOUS v.5 (Biomatters, available at <http://www.geneious.com>).

Our dataset represents a significant improvement in sampling compared to the most recently published viper phylogenies. We include 46 more species than Fenwick et al. (2012) and 55 more than Pyron et al. (2013). Moreover, our molecular dataset comprises 11 genes and 1192 sequences whereas Fenwick et al. (2012) and Pyron et al. (2013) comprise 785 and 854 sequences for 4 and 10 genes respectively. Therefore, the study presented herein comprises the most complete phylogenetic analysis for the family to the present.

2.2 Phylogenetic analyses and divergence time estimation

Sequences were aligned with MAFFT (Katoh et al., 2002) using the E-INS-i algorithm for 12S, 16S, nd4, nd2, cox1 and rag1, and FFT-NS-i for bndf, cmos, cytb, jun and nt3. For both algorithms all the parameters were set as default. Resulting alignments were visually inspected in Mesquite v 2.75 (Maddison and Maddison, 2009). Sequences were concatenated with SequenceMatrix v 1.7.8 (Vaidya et al., 2011) in two different ways: (1) all sequences for all the 11 genes (361 terminals), and (2) all sequences from 16S, 12S, cytb and nd4 genes (genes with sequences representing the majority of the species included in the phylogenetic analysis, 357 terminals). Our main dataset (dataset 1) contained alignments with 10712 base pairs (bp) and dataset 2 contained 4419 bp. Although recent studies found no evidence that missing data would lead to inaccurate estimates in Bayesian phylogenetic analyses and time divergence estimates (Filipski et al., 2014; Wiens and Morrill, 2011; but see Lemmon et al., 2009), we used dataset 2 to assess any potential influence of missing data in our phylogenetic estimates by comparing its results to our main dataset 1. The comparison between datasets also allowed us to evaluate any improvements in the phylogenetic relationships of vipers when adding nuclear genes information (main dataset) given that dataset 2 comprises only mitochondrial genes. We used PartitionFinder

1.0.1 (Lanfear et al., 2012) with the greedy algorithm and linked branch lengths to select the best partition scheme and the best models of nucleotide substitution for our molecular matrix partitioned by gene and codon position for each dataset above. The Bayesian Information Criterion (BIC) was used as the optimality criterion to select the best partition schemes and best models of nucleotide substitution for each dataset (Table S2).

We estimated the phylogenetic relationships and time of divergence between species using a Bayesian framework implemented in the program BEAST v1.8.1 (Drummond and Rambaut, 2007) for each dataset. Substitution rates were estimated under a relaxed uncorrelated lognormal clock (Drummond et al., 2006) that allows different branches to have independent rates. Extinction has played a major role in shaping diversification patterns in the tree of life (Raup 1986; Quental and Marshall 2010), so using a Yule model, which assumes extinction rates equal to zero, as our tree prior would be inappropriate. Thus, we used a Birth-Death speciation model accommodating incomplete taxon sampling as our tree prior (Stadler, 2009).

Calibrating a phylogenetic tree is a critical step in molecular dating analyses and caution is required in choosing the points to be used (Sanders et al., 2010; Sauquet et al., 2012). To time calibrate our phylogenetic tree, we followed recent published guidelines and protocols that aim to help researchers in choosing fossils to be used as calibration points (Benton et al., 2009; Parham et al., 2012; Sauquet et al., 2012). We chose to include in our analyses six fossil records from the literature, two of them described as Viperidae and the other four positioned throughout the outgroup. Fossils details and justifications on the choice of calibration schemes are described below.

1. Stem-Alethinophidia

Haasiophis terrasanctus Tchernov, Rieppel, Zaher, Polcyn, Jacobs 2000. Set as the most recent common ancestor of *Indotyphlops braminus* and Alethinophidia. A complete, articulated specimen recovered from the Ein Yabrud quarries, near Ramallah, West Bank Palestinian Territories (Hebrew University of Jerusalem, Paleontological Collections, HUIJ-Pal. EJ 695). The position of the Ein Yabrud quarries within the Cenomanian is unclear, and may fall either in the Early Cenomanian Bet-Meir Formation (Chalifa and Tchernov, 1982) or the Late Cenomanian Amminadav Formation (Chalifa, 1985). Root minimum age constrained to 93.9 Ma. We follow Head (2015) in using an age range that reflects the overall uncertainty of the position of Ein Yabrud, with the minimum age set for the Cenomanian (Ogg and Hinnov, 2012). Hsiang et al. (2015) concluded that *Haasiophis terrasanctus* is positioned as a stem-Alethinophidia instead of a crown-Alethinophidia in their combined, molecular and morphological, unconstrained analysis (their figure 3). Hsiang et al.'s

unconstrained topology represents a more inclusive hypothesis for the position of *Haasiophis terrasanctus*. Because Hsiang et al.'s analysis incorporates the known phylogenetic uncertainties related to the position of this fossil, and it is based on the largest set of unconstrained evidence, we prefer to use *Haasiophis* as a stem-althinophidian instead of a crown-taxa as traditionally used.

2. Stem-Colubroidea

Procerophis sahnii Rage, Folie, Rana, Singh, Rose, Smith 2008. Set as the most recent common ancestor of Colubroidea and the remaining snakes. Minimum age constrained to 54 Ma. *Procerophis sahnii* is known from the early Eocene (early Ypresian) Cambay Formation of Vastan Lignite Mine, Gujarat, India (Sahni et al., 2006; Rana et al., 2008; Rage et al., 2008). A single well-preserved posterior trunk vertebra (Rana Collection from Vastan, VAS 1014). The shape of the vertebra, prezygapophyseal processes and paracotylar foramina suggest the fossil belongs to the Colubroidea (Rage et al., 2008).

3. Stem-Boinae

Titanoboa cerrejonensis Head, Bloch, Hastings, Bourque, Cadena, Herrera, Polly, Jaramillo 2009. Set as the most recent common ancestor of Boinae and the remaining snakes. Minimum age constrained to 58 Ma. (Jaramillo et al., 2007; Head, 2015). *Titanoboa cerrejonensis* is known from the middle-late Paleocene (Palynological zone Cu-02) Cerrejón Formation of La Puente Pit, Cerrejón Coal Mine, Guajira Peninsula, Colombia (Jaramillo et al., 2005, 2007). A single preloacal vertebra (Florida Museum of Natural History, UF/IGM 1). Paracotylar foramina morphology and a convex anterior zygosphenic margin suggests this fossil belongs to the Boinae (Head et al., 2009, 2015; Albino and Brizuela, 2014). Additionally, a preliminary analysis of cranial elements by Head et al. (2013) places it in the stem lineage of Boinae.

4. Stem-Elapidae

Elapidae gen. & sp. indet. Kuch, Müller, Mödden, Mebs 2006. Set as the most common recent ancestor of Elapidae and the remaining snakes. Minimum age constrained to 19.5 Ma. (set as the minimum age estimate for the MN2 unit of the European Land Mammal Age) (Kuch et al., 2006; Hilgen et al., 2012). Several tubular fangs deposited in the collection of the Institut Für Geowissenschaften der Universität Mainz, Germany, and referred to an early elapid species from the Early Miocene (Aquitania) of the Oppenheim/Nierstein quarry, Mainz basin, Rhineland-

Palatinate, Germany. Kuch et al. (2006) considered the material to belong to a member of the family Elapidae based on confident similarities in the curvature and suture present between the fossil material and present-day elapids.

5. Stem-Viperidae

Viperidae. gen. & sp. indet. Kuch, Müller, Mödden, Mebs 2006. Set as the most common recent ancestor of Viperidae and the remaining snakes. Minimum age constrained to 19.5 Ma. (set as the minimum age estimate for the MN2 unit of the European Land Mammal Age; Kuch et al., 2006; Hilgen et al., 2012). Several fangs deposited in the collection of the Institut Für Geowissenschaften der Universität Mainz, Germany, and referred to a viperid species from the Early Miocene (Aquitanian) of the Oppenheim/Nierstein quarry, Mainz basin, Rhineland-Palatinate, Germany. Kuch et al. (2006) referred the fangs to a member of the family Viperidae based on details of the curvature, length, presence of prominent anterior blade ridges, and shape of the venom discharge orifice.

6. Stem-*Sistrurus*

Sistrurus sp. indet. Parmley and Holman 2007. Set as the most common recent ancestor of *Sistrurus catenatus* and the remaining snakes. Minimum age constrained to 10.3 Ma. (set as the minimum age estimate for the Clarendonian NALMA; Hilgen et al., 2012). A well preserved trunk vertebra (University of Nebraska State Museum, UNSM 100251) from the late Miocene of the Ash Hollow formation (Clarendonian), recovered from fluvial channel fill of the Pratt Slide site, located near Johnstown, Brown County, Nebraska, USA. The fossil is referred to the extant genus *Sistrurus* Parmley and Holman (2007) based on the presence of zygosphenal spine considered a derived trait unique to the genus.

We used lognormal prior distributions to incorporate uncertainty surrounding fossil calibration. Lognormal is often the most appropriate distribution for calibrations using paleontological data assigning the highest point probability for the calibration point to be somewhat older than the oldest fossil (Ho and Phillips, 2009). In all cases, hard minimum bounds were set accordingly to the minimum age of the geological interval that the fossil belongs. The establishment of maximum bounds to calibrate prior distributions is regarded as critical (Ho and Phillips, 2009) especially when dealing with groups with a relatively poor fossil record like vipers. We set the maximum 95% in three different manners: around 61.5 mya, when the calibration point located at

the divergence between the Viperidae and their sister group, 47.4 mya, when calibration point was located in the ingroup, and around 145 mya in the other cases. 61.5 mya is the age of divergence between Viperidae and the outgroup estimated by Wüster et al. (2008), 47.4 mya is the age of the most recent common ancestor (MRCA) of Viperidae estimated by Wüster et al. (2008) and 145 mya refers to the age of the oldest snake fossil known (Colston et al., 2013; Hofstetter, 1960).

For each concatenated dataset we ran four replicates of our phylogenetic analyses with different starting trees or random seed saving the parameters and trees every 1000 interactions. In total, we ran eight phylogenetic analyses. Chain length was set to 300 million generations but we stopped the analyses as soon as they reached convergence. For each concatenated dataset we kept the replicate(s) that achieved the higher likelihood values. We used the program Tracer v 1.6 (Rambaut and Drummond, 2007) to check for convergence, the effective sample size (ESS) of parameters and to verify burn-in. When appropriate we used logCombiner 1.8 to combine posteriors of replicates of the same concatenated dataset that converged to the same higher likelihood values. We used TreeAnnotator v 1.8.1 (Drummond and Rambaut, 2007) to generate maximum credibility trees.

RESULTS

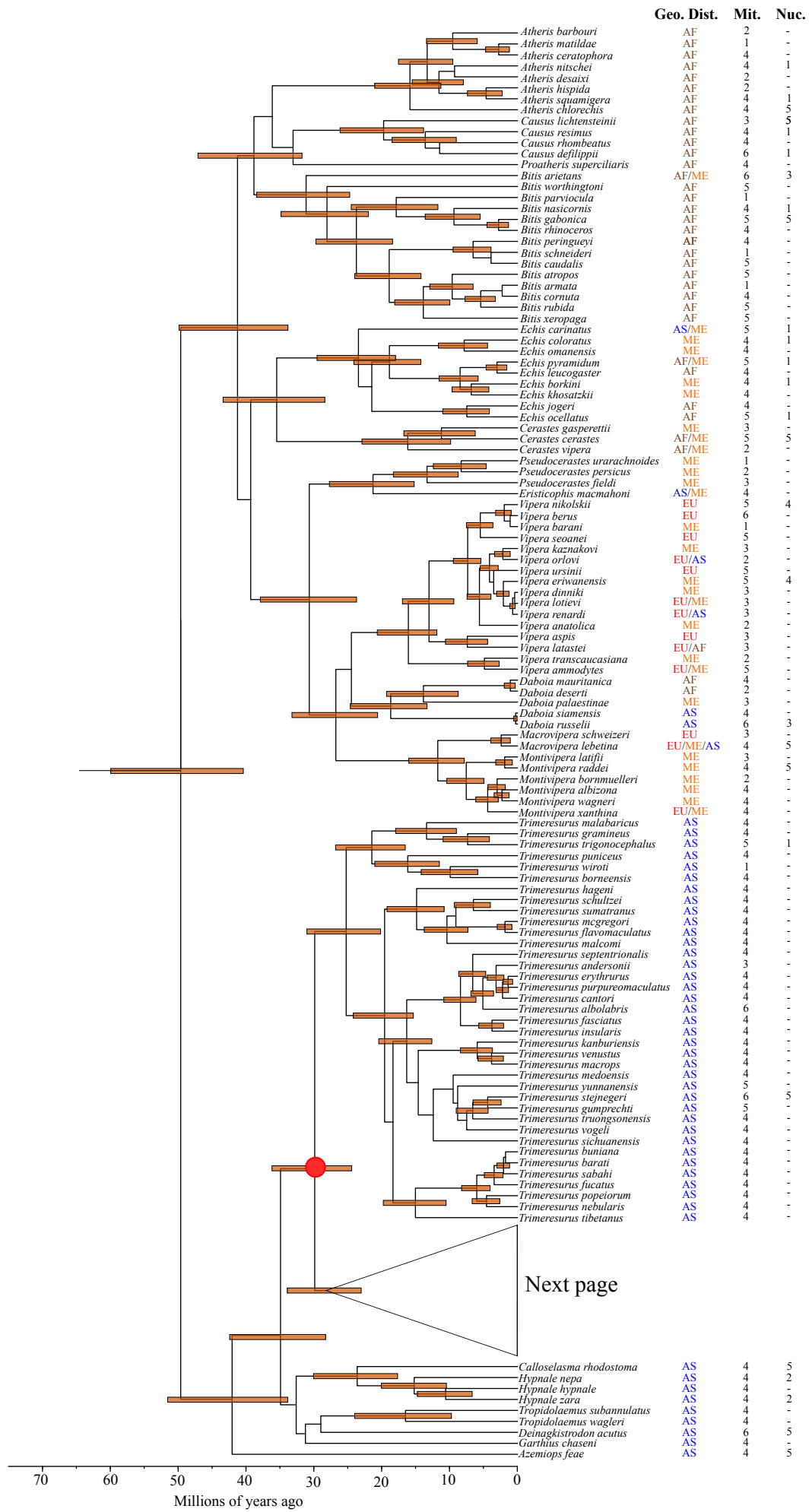
We generated maximum credibility trees after discarding burn-in with a posterior probability limit of 95% (Figure 1, Figure S1, Figure S2). Both datasets yielded similar topologies and divergent times estimates for the family Viperidae (Figures 1 and S2, Table 1). Thus, results will focus on our main dataset (dataset 1) comparing with dataset 2 when relevant for the discussion. We also note that the small differences in dating analyses are not significant given that estimates for both datasets show considerable overlap in their 95% high posterior density interval (Table 1, Figure 1 and Figure S2). The complete maximum credibility tree for dataset 1 including the outgroup and calibration points can be visualized through Figure S1.

Phylogenetic analyses resulted in a relatively well-resolved tree with most nodes having high posterior probabilities ($pp > 0.95$, represented by nodes with bars of age estimates in Figure 1) confirming the monophyly of the family and subfamilies (Viperinae, Azemiopinae and Crotalinae). Phylogenetic relationships among subfamilies were also recovered with high support with Viperinae being sister to a clade formed by Azemiopinae and Crotalinae. Of the 27 genera for which we could test for monophyly, 22 presented high posterior probabilities in favor of their monophyly (see Discussion for details). Among viperines, posterior probability showed high support for a clade comprising the genera *Atheris*, *Causus*, *Proatheris* and *Bitis*. Eurasian vipers (*Pseudocerastes*, *Eristicophis*, *Vipera*, *Macrovipera* and *Montivipera*) also formed a well-supported

clade (Figure 1). Among the remaining lineages of the subfamily Viperinae, *Echis* and *Cerastes* also showed a robust sister relationship (Figure 1). Regarding the affinities in the subfamily Crotalinae, *Hypnale*, *Garthius*, *Deinagkistrodon*, *Calloselasma* and *Tropidolaemus* formed a weak supported monophyletic clade sister to the remaining crotalines but *Hypnale* and *Calloselasma* formed a well-supported clade (Figure 1). We recovered with great confidence a clade comprising the genera *Protobothrops*, *Gloydus*, *Ovophis*, and *T. gracilis* as the sister clade to the New World pit vipers, with almost all nodes showing strongly supported relationships (Figure 1). Among the NW pit vipers, we recovered high posterior probabilities for clades comprising the Middle American genera *Atropoides*, *Cerrophidion*, and *Porthidium*, another comprising the Bothropoides (*Bothrops* and *Bothrocophias*), as well as high supported clades formed by *Sistrurus* + *Crotalus*, and *Ophryacus* + *Mixcoatlus* (Figure 1). However, more inclusive relationships and some higher-level relationships were recovered with low support among New World pit vipers (see Figure 1).

Table 1. Comparison of age divergence estimates (in millions of years ago; mean and 95% high posterior density interval) between datasets analyzed in the present study, Fenwick et al. (2012), and Wüster et al. (2008). tmrca = time of maximum recent common ancestor, *approximate age.

	Dataset 1 (this study)		Dataset 2 (this study)		Wüster et al. 2008		Fenwick et al. (2012)*	
Root (<i>Indotyphlops braminus</i> vs others)	111.73	95.3 - 134.03	108.52	94.87 - 129.27	-	-	-	-
tmrca Viperidae	49.67	40.44 - 59.98	52.41	43.17 - 62.58	47.4	38.1 - 57.4	25.5	-
tmrca Viperinae	41.3	33.86 - 49.91	44.86	36.87 - 53.75	39.7	31.9 - 48.1	22.5	-
tmrca Crotalinae	34.96	28.31 - 42.45	37.65	30.71 - 45.27	31.2	25.4 - 37.8	17.5	-
tmrca <i>Trimeresurus</i>	25.27	20.21 - 31.07	27.37	22.16 - 32.97	-	-	-	-
tmrca NW Crotalinae	25.68	20.99 - 31.05	27.86	23.18 - 33.53	22.1	17.9 - 26.9	12.5	-
tmrca <i>Lachesis</i>	8.44	5.37 - 11.96	9.26	5.93 - 12.95	6.5	3.5 - 9.8	-	-
tmrca <i>Sistrurus</i>	14.1	10.07 - 18.21	14.29	10.16 - 18.45	-	-	-	-



Bitis



Echis

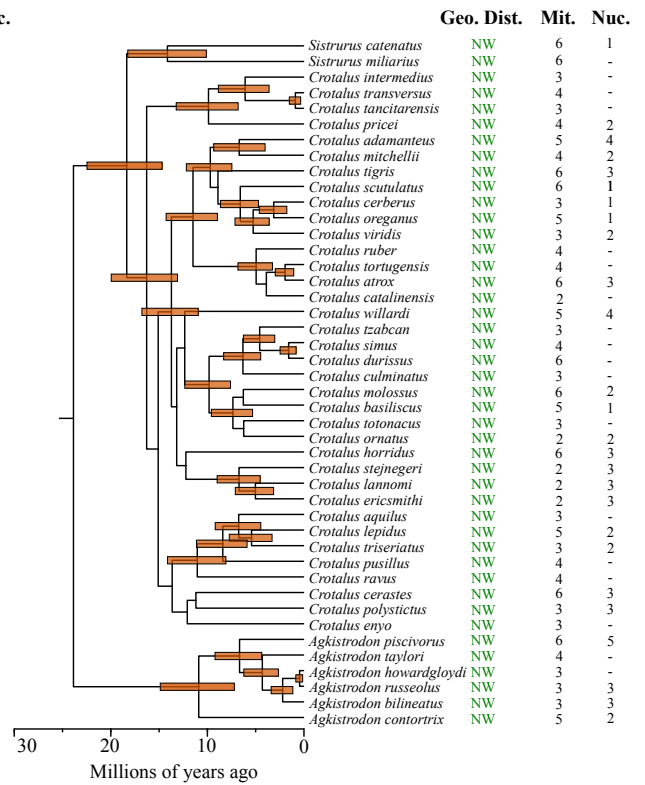
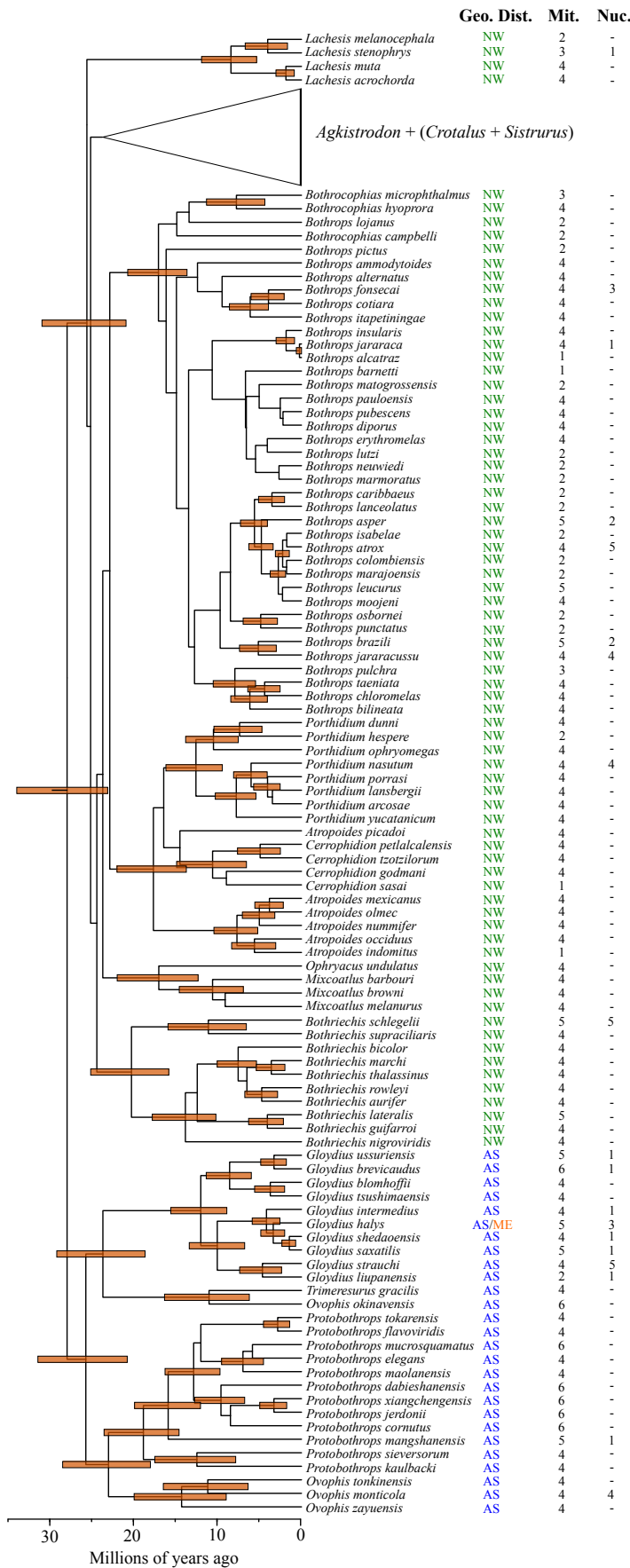


Vipera



Trimeresurus

Next page



Bothrops



Crotalus

Figure 1. Maximum credibility tree generated for the family Viperidae comprising all genes and species (dataset 1). Bars showing the 95% high posterior density interval of age estimates were added only for nodes with posterior probability values equal to or higher than 0.95. Geographic distribution (Geo. Dist.) and the number of mitochondrial (Mit.) and nuclear (Nuc.) gene sequences for each taxon are also shown. AF (brown) = Africa, ME (orange) = Middle East, EU (red) = Europe, AS (blue) = Asia, NW (green) = New World. Red circle indicates the position of the speciation rate shift (see Chapter II). Images of vipers by Thor Hakonsen (<http://thorhakonsen.com/>).

Divergence time estimates suggest that the ancestral of vipers diverged from its sister group in the Paleocene (around 64.5 Mya) and that crown vipers started to diversify around the late Paleocene to middle Eocene (59.9 - 40.4 Mya) (95% high posterior density interval of the main dataset). During the Eocene, around 49.91 - 33.86 Mya, the Viperinae subfamily started to diversify. The crown group of crotalines is placed around 42.45 - 28.31 Mya in the middle Eocene to late Oligocene. New World pit vipers started to radiate around the middle/late Oligocene to the early Miocene (31.05 - 20.99 Mya). Information regarding divergence time estimates of vipers is summarized in Figure 1 and Table 1.

DISCUSSION

Phylogenetic inference

Phylogenetic relationships among vipers recovered in the present study agree in many aspects with those from previously published works (e.g. Pyron et al., 2013; Wüster et al., 2008). Although all subfamilies and the majority of genera were considered monophyletic we could not recover monophyly for the following genera: *Bothrops* and *Bothrocophias*, *Atropoides*, *Trimeresurus* and *Ovophis*.

The Andean *Bothrops lojanus* have been considered as *incertae sedis* and although morphological analyses recovered the species within *Bothrops* (Carrasco et al., 2012), in the present study *B. lojanus* appears within the genus *Bothrocophias* with weak support. The distribution of *B. lojanus* in terrestrial forested habitats in Ecuador and the occurrence of some *Bothrocophias* species in this same region and habitat (Campbell and Lamar, 2004; Harvey et al., 2005) suggest this might be a possible relationship. However, the few gene sequences available for *B. lojanus* (only two mitochondrial gene sequences) may be preventing us to find a reliable phylogenetic position for the species. Irrespective to the position of *B. lojanus*, the monophyly of Bothropoides (*Bothrops* +

Bothrocophias) have been recovered with great confidence in both morphological and molecular analyses (Carrasco et al., 2012; Pyron et al., 2013; present study).

Another genus not recovered as monophyletic was *Atropoides*. The monophyly of *Atropoides* has been controversial because of the uncertainty in the phylogenetic position of *A. picadoi* (Castoe and Parkinson, 2006; Castoe et al., 2009; Jadin et al., 2010). While using morphological characters Jadin et al. (2010) found strong support for the monophyly of *Atropoides*, molecular studies failed to find strong support for the monophyly of the genus (Castoe and Parkinson, 2006; Castoe et al., 2009). In the present study, we could not recover the phylogenetic position of *A. picadoi* with confidence. Both *Atropoides picadoi* and the remaining species of the genus are represented by sequences of four mitochondrial genes (except for *A. indomitus* that had a sequence of nd4 gene only). That is, molecular data available for *Atropoides* species seems to be insufficient to elucidate the relationship between *A. picadoi* and the remaining *Atropoides*.

Like previous studies (e.g. Malhotra et al., 2010; Pyron et al., 2013; Wüster et al., 2008), we found “*Ovophis*” *okinavensis* to be sister to “*Trimeresurus*” *gracilis*. Both species have been suggested to be more closely related to the genus *Gloydus* than to *Ovophis* or *Trimeresurus* (e.g. Castoe and Parkinson, 2006; Malhotra and Thorpe, 2004; Pyron et al., 2013; Wüster et al., 2008). Here, we not just found a monophyletic clade comprising “*Ovophis*” *okinavensis*, “*Trimeresurus*” *gracilis* and the genus *Gloydus* but we recovered this relationship with a very high support value (>0.99) contrasting with the results from other studies (e.g. Pyron et al., 2013). Hence our results support previous ideas that those two species should be placed in a genus of its own or perhaps considered to belong to the genus *Gloydus* (Malhotra and Thorpe, 2004; Malhotra et al., 2010). However, more studies are needed in order to make a reliable decision.

Regarding viperines, the most striking difference between phylogenetic studies relates to the position of the genus *Causus*. The particular phenotype of *Causus* suggests the genus is sister to the remaining viperines in morphological analyses (see Herrmann et al., 1999). However, many molecular studies disagree from morphological ones and found different arrangements suggesting *Causus* to be either sister to *Proatheris* (present study, Figure 1), *Cerastes* (Wüster et al., 2008; Figure S2) or *Echis* (Pyron et al., 2013). Interestingly, molecular studies have found weak support for these more inclusive positions of *Causus* among viperines. Additionally, the present study corroborates Wüster et al. (2008) and Pyron et al. (2013) in recovering well-supported phylogenetic relationships between Eurasian vipers but differs from these studies as we found a strongly supported relationship between *Cerastes* and *Echis*.

Intergeneric relationships among the Crotalinae species poor genera *Tropidolaemus*, *Deinagkistrodon*, *Garthius*, *Hypnale* and *Calloselasma* remained unresolved in the present study. Although we found this clade to be monophyletic agreeing with Malhotra et al. (2010) who also

analyzed nuclear sequences, both studies recovered a weakly supported configuration. Paraphyletic configurations were frequently recovered in studies using only mitochondrial genes (Castoe and Parkinson, 2006; Wüster et al., 2008) or including nuclear genes for only one genus (Pyrón et al., 2013) but also with weak support. Despite our dataset comprising only mitochondrial genes (dataset 2) also suggested a weakly supported monophyletic configuration, posterior probability values doubled when we included nuclear sequences. That is, including nuclear sequences for other species belonging to this crotaline clade (e.g. *Tropidolaemus* species, *Garthius chaseni*) might be the key to obtain higher support values and confidently estimate the relationship between these genera connected by long-branches. Irrespective to the true configuration of *Tropidolaemus*, *Deinagkistrodon*, *Garthius*, *Hypnale* and *Calloselasma*, a strongly supported relationship between *Calloselasma* and *Hypnale* have been recovered here and in other studies (Pyrón et al., 2013; Wüster et al., 2008).

Although Malhotra et al. (2010) recovered *Gloydius* as the sister clade of NW crotalines with moderate to high support (0.87 - 0.99), relationships between the other related lineages (*Protobothrops*, *Ovophis*, “*Trimeresurus*” *gracilis* + “*Ovophis*” *okinavensis*) remained unresolved (see also Castoe and Parkinson, 2006; Pyrón et al., 2013; Wüster et al., 2008). In the present study, we recovered with great confidence a monophyletic clade as the sister group to NW crotalines comprising not only the previously mentioned clade formed by *Gloydius* and *O. okinavensis* + *T. gracilis*, but also including the genera *Protobothrops* and *Ovophis*. Moreover, great majority of the nodes in the sister clade proposed for the NW crotalines shows strong support. Given the proximity of the current geographic ranges of those lineages, predominantly in southeast to East Asia (Malhotra et al., 2011; Uetz and Hosek 2014), the NW sister clade proposed here would not be unexpected. Additionally, some *Gloydius* species also occur in more northern parts relatively close to the Beringia (Gloyd and Conant, 1990; Malhotra et al., 2010), where the invasion of the NW is thought to have taken place (Wüster et al., 2008).

Like previous studies on viper phylogenetics (e.g. Gutberlet and Harvey, 2002; Pyrón et al., 2013; Wüster et al., 2008), relationships among deeper branches in the clade of NW pit vipers were poorly resolved. This difficulty is likely to be related to the suggested rapid cladogenesis of pit vipers after their invasion in the NW (Gutberlet and Harvey, 2002; Wüster et al., 2008; see also analysis below) leading to very short branches among major NW clades (see Figure 1). However, differently from morphological analyses (e.g. Gutberlet and Harvey, 2002), molecular studies recovered a well-supported clade comprising *Porthidium*, *Cerrophidion* and *Atropoides* (Pyrón et al., 2013; Wüster et al., 2008; present study), and our results also agree with Pyrón et al. (2013) and Wüster et al. (2008) in recovering highly supported clades comprising the genera *Crotalus* + *Sistrurus*, and *Mixcoatlus* + *Ophryacus*.

Divergence time estimates

Our dating analyses suggest slightly older ages for Viperidae than those proposed by Wüster et al. (2008) (Table 1). However, our date estimates are much older than those suggested by Fenwick et al. (2012) (Table 1). Discrepancies in divergence time estimates can emerge for several reasons (see Ho and Phillips, 2009; Sauquet et al., 2012). The differences in the taxa included among studies, tree prior used (birth-death speciation in the present study vs Yule in the others) and/or the different calibration points and prior distributions chosen, could all be plausible explanations.

Although the oldest fossil known for vipers dates back to 19.5 Mya (early Miocene, Kuch et al., 2006), our date estimates agree with Wüster et al. (2008) that the ancestral of vipers diverged from its sister group during the Paleocene. Age estimates for the crown group of vipers vary widely in the literature (e.g. ~25.5 Mya in Fenwick et al., 2012, 35.6 Mya in Pyron and Burbrink, 2012, 47.4 Mya in Wüster et al., 2008) but our results suggest that vipers diversified around the late Paleocene to middle Eocene, with subfamilies crown group ages most likely dating back to the Eocene. Additionally, close to the Oligocene/Miocene boundary the subfamily Crotalinae invaded the New World. The invasion of the New World by crotalines has been proposed to have occurred as a single event via the Beringian Land Bridge (e.g. Wüster et al., 2002; Wuster et al., 2008). The Beringian Land Bridge was an important corridor for vertebrate dispersal but since its formation around 100 Mya, dispersal by many warm adapted taxa was only possible during the increased temperatures that maintained a Boreotropical forest belt throughout the Eocene during the Eocene climatic optimum (Wolfe, 1975, Townsend et al., 2011). Given the uncertainty in our date estimates, crotalines could have started to diversify both during the Eocene or the Oligocene not providing confidence on whether this lineage have dispersed to North America through the Boreotropical forest belt (Eocene) or during a cooler time period (Oligocene).

The Isthmus of Panama, a strip of land that connects North and Central America to South America (Bacon et al., 2013), has also been suggested as a mechanism of dispersion by which some NW pit vipers, such as the rattlesnake *Crotalus durissus* and some *Porthidium* species, probably invaded South America (Vanzolini and Heyer, 1985; Wüster et al., 2002). It is generally accepted that the emergence of the Isthmus occurred at about 3.5 Mya (but see Bacon et al., 2013, 2015), making the dispersion explanation for older lineages of South American pit vipers like *Bothrops* and *Bothrocophias*, more elusive. These older pit viper clades may have invaded South America by stepping-stone dispersal over the Central American archipelago, as suggested for other lineages (e.g. Jansa et al., 2013). In fact, very recently, studies have suggested that migration through the

Isthmus of Panama (not necessarily a fully continuous landmass) might in fact have occurred much earlier (Bacon et al., 2013, 2015; Farris et al., 2011; Montes et al., 2012) at some point between early Oligocene and early Miocene (31 - 16.1 Mya). Our date estimates for *Bothrops* and *Bothrocophias* are in accordance with an earlier migration than the usually accepted date for the emergence of the Isthmus of Panama of 3.5 Mya.

CONCLUSIONS

The phylogenetic relationships among vipers recovered well-supported clades with most genera being monophyletic. However, like previous studies, we draw attention to the paraphyletic position of *Atropoides picadoi*, *Ovophis okinavensis* and *Trimeresurus gracilis* respective to their current genera, and to the non-monophyletic configuration of *Bothrops* and *Bothrocophias*. Our study also clarify previously problematic relationships between Africa-Middle East taxa (e.g. *Echis* and *Cerastes*), and propose a highly supported Asian sister group for NW pit vipers. On the other hand, we could not recover well supported relationships for some deeper nodes in NW pit vipers indicating that the clade still needs further investigation. Our date estimates suggested that vipers started to diversify around the end of the middle Paleocene/middle Eocene epochs recovering older ages than previous studies. Until the early Miocene pit vipers emerged and rapidly colonized the NW. Moreover, date estimates for South American clades are in accordance with an earlier migration than the usually accepted date for the emergence of the Isthmus of Panama.

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SUPPLEMENTARY FIGURES

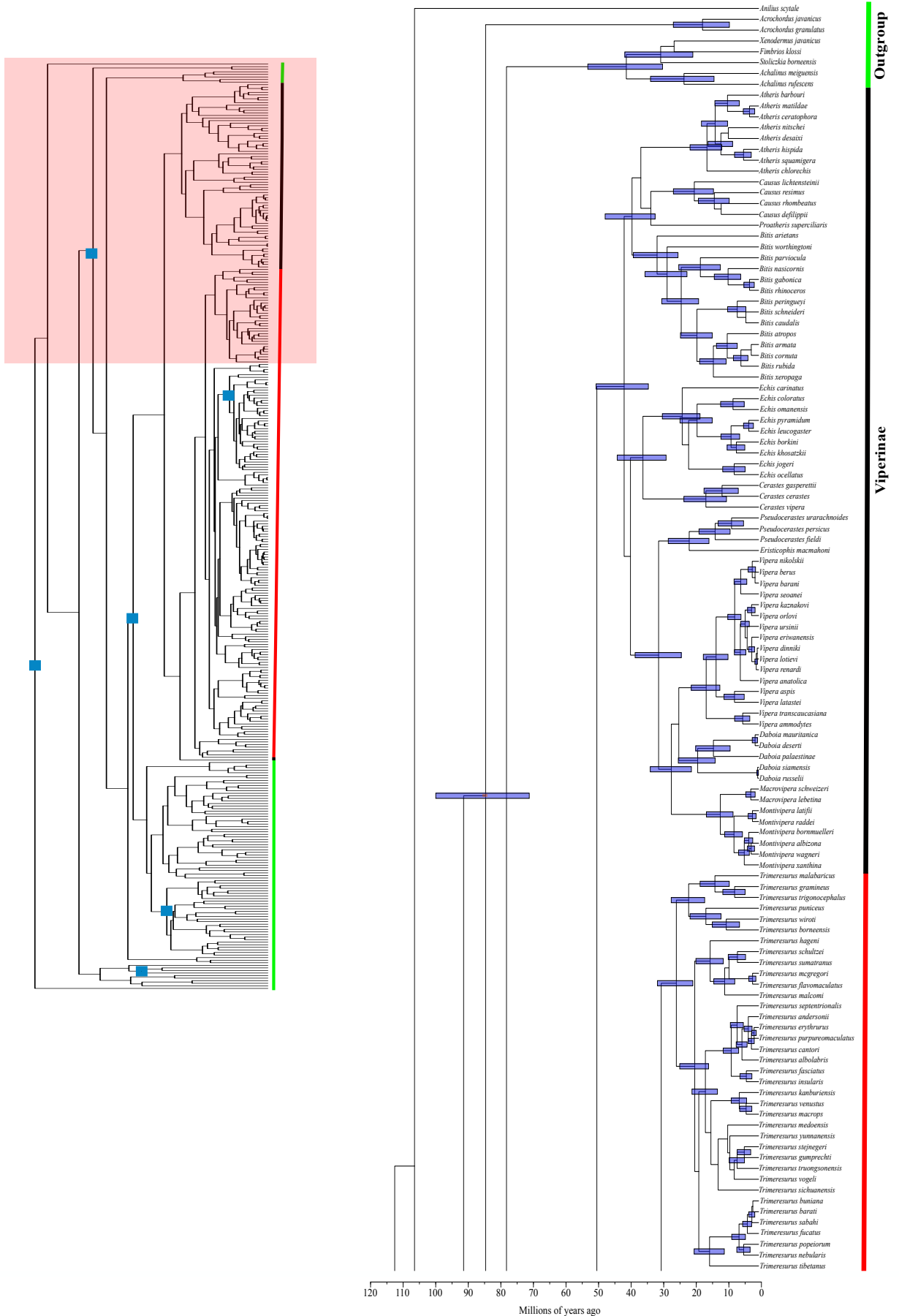


Figure S1. part1

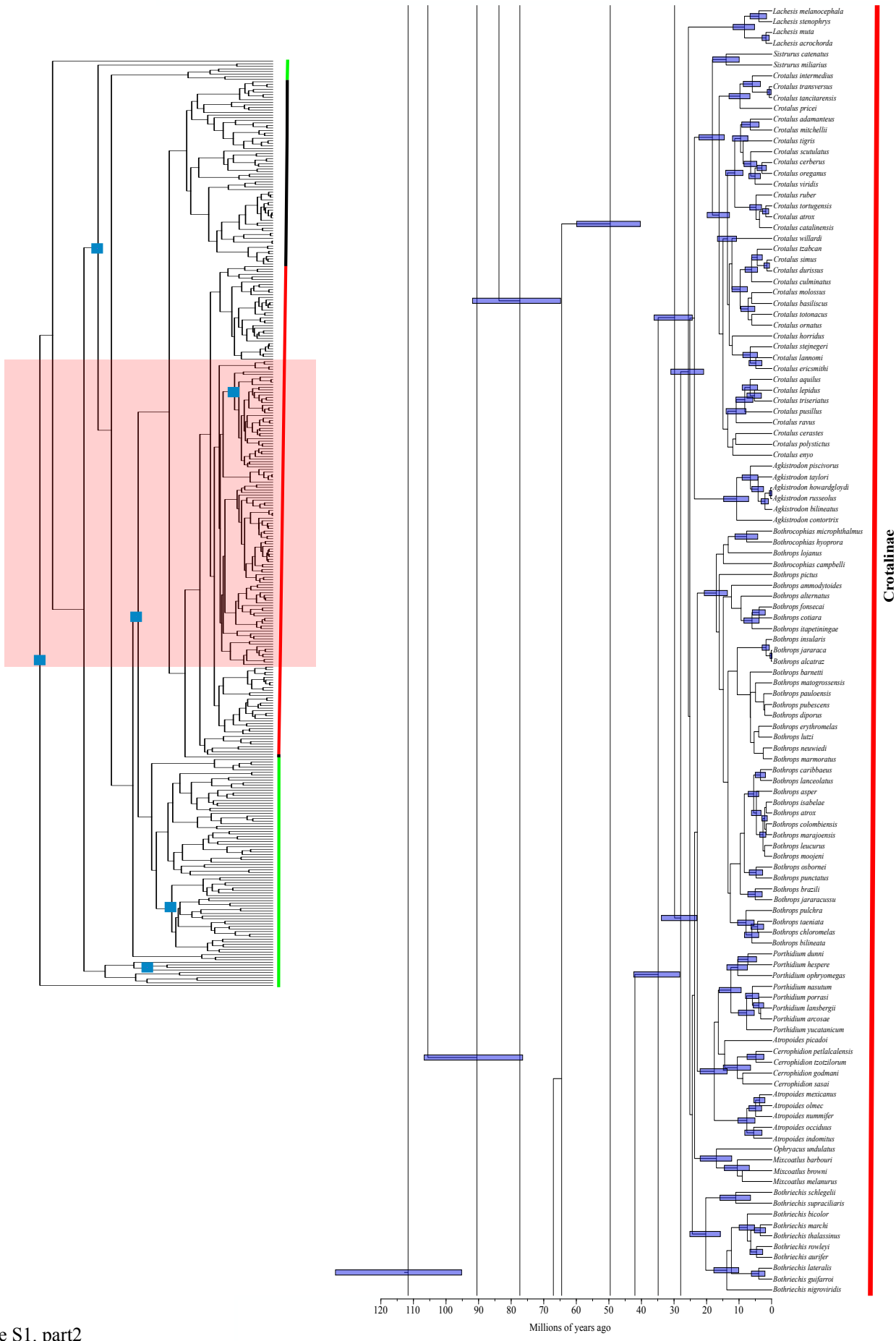


Figure S1. part2

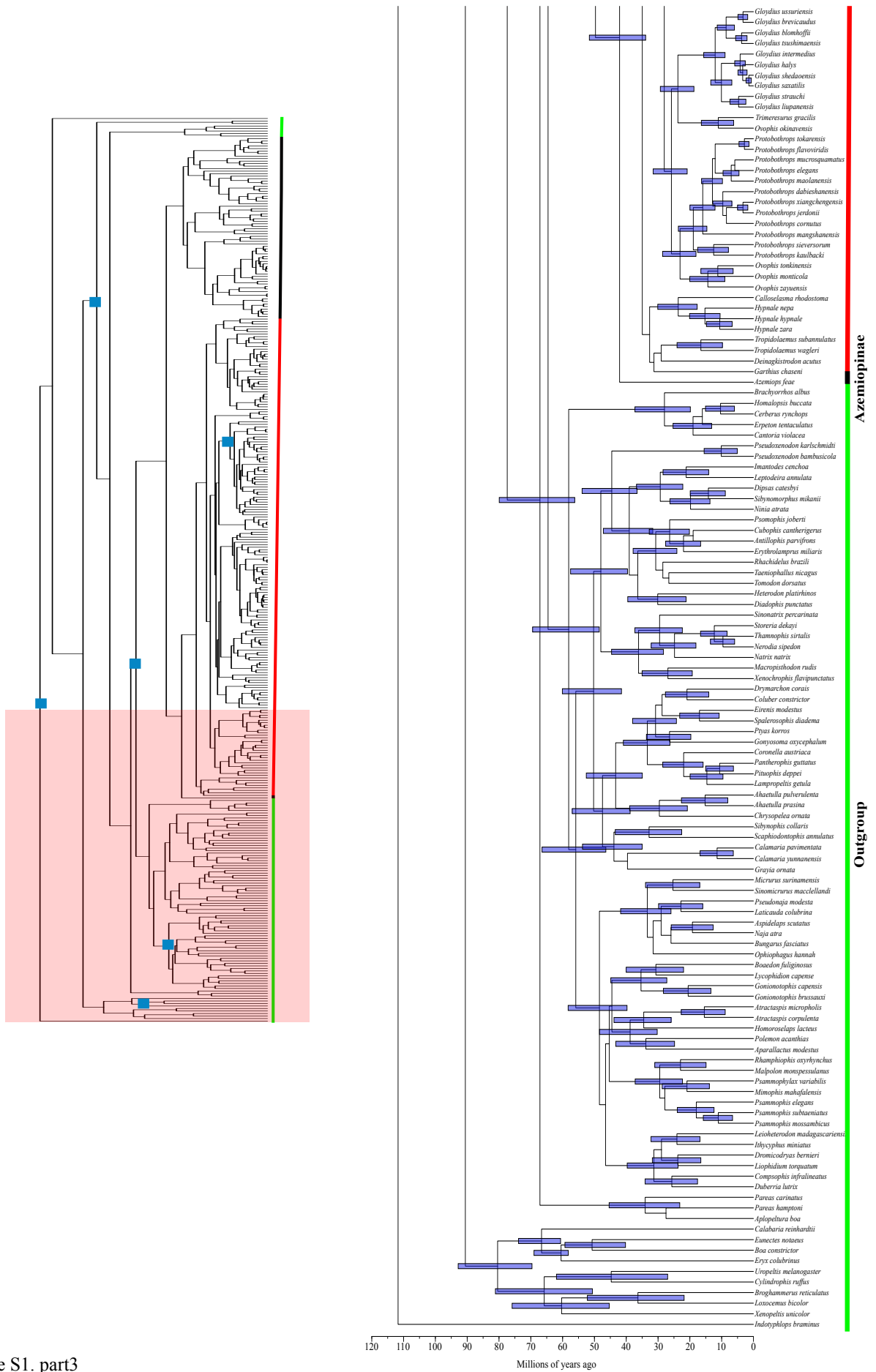


Figure S1. part3

Figure S1. Full version (including the outgroup) of the maximum credibility tree generated for the family Viperidae using the dataset comprising sequences from all 11 genes. Blue squares denote calibration points. Bars showing the 95% high posterior density interval of age estimates were added only for nodes with posterior probability values equal to or higher than 0.95. Large red squares denote the position zoomed.

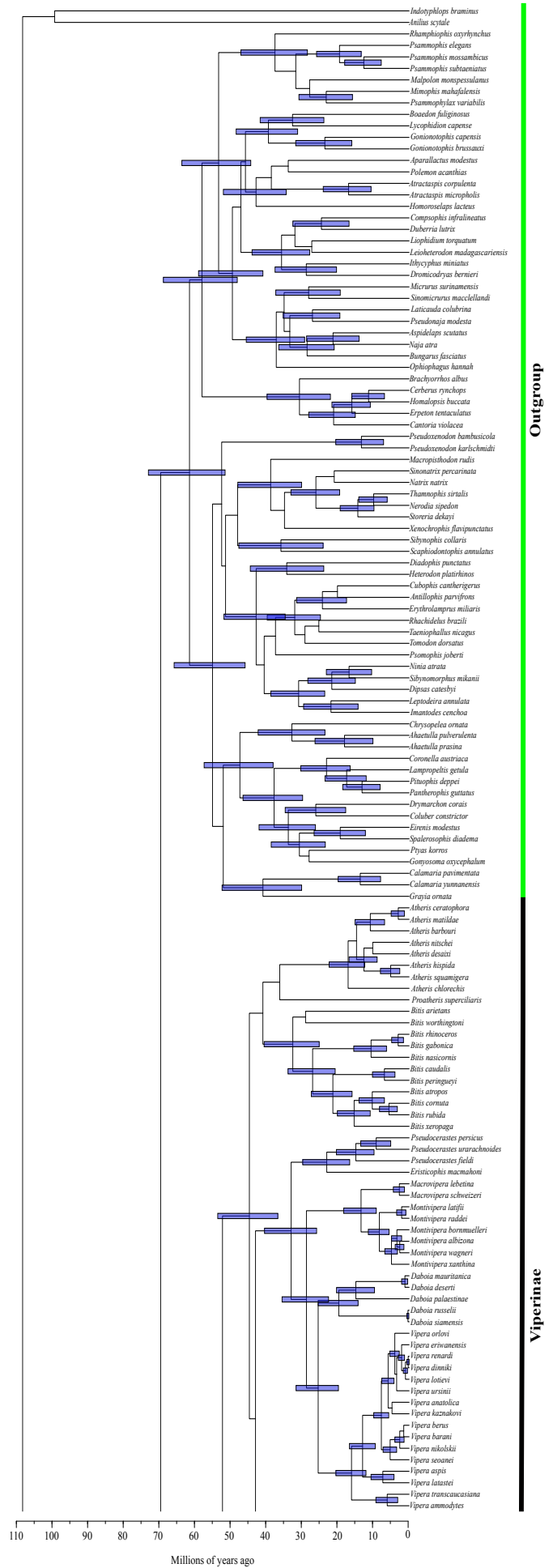
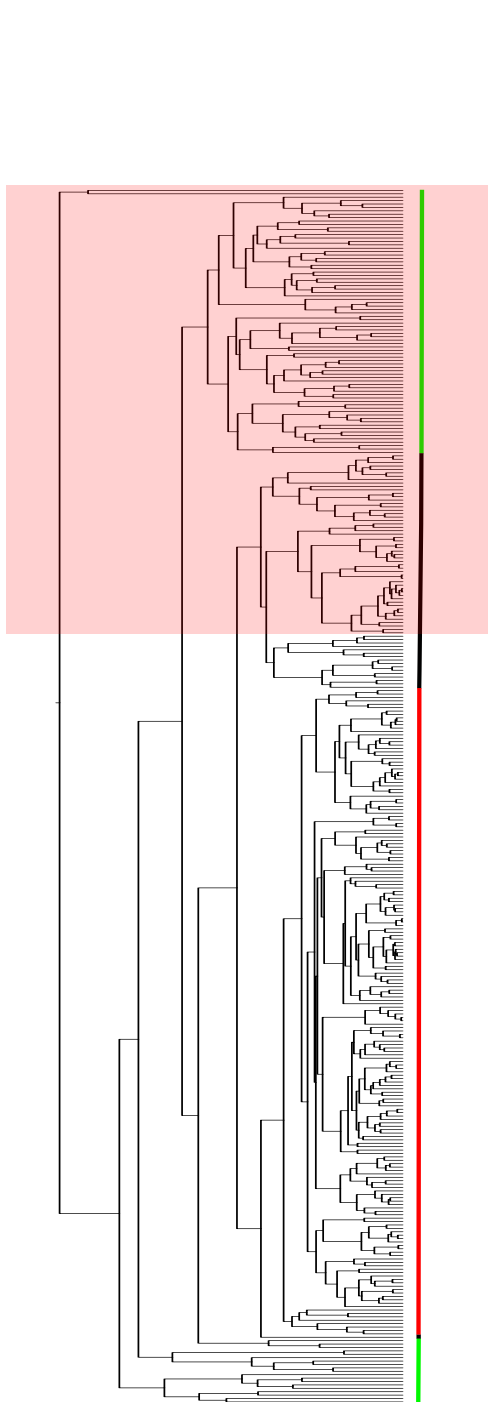


Figure S2. part1

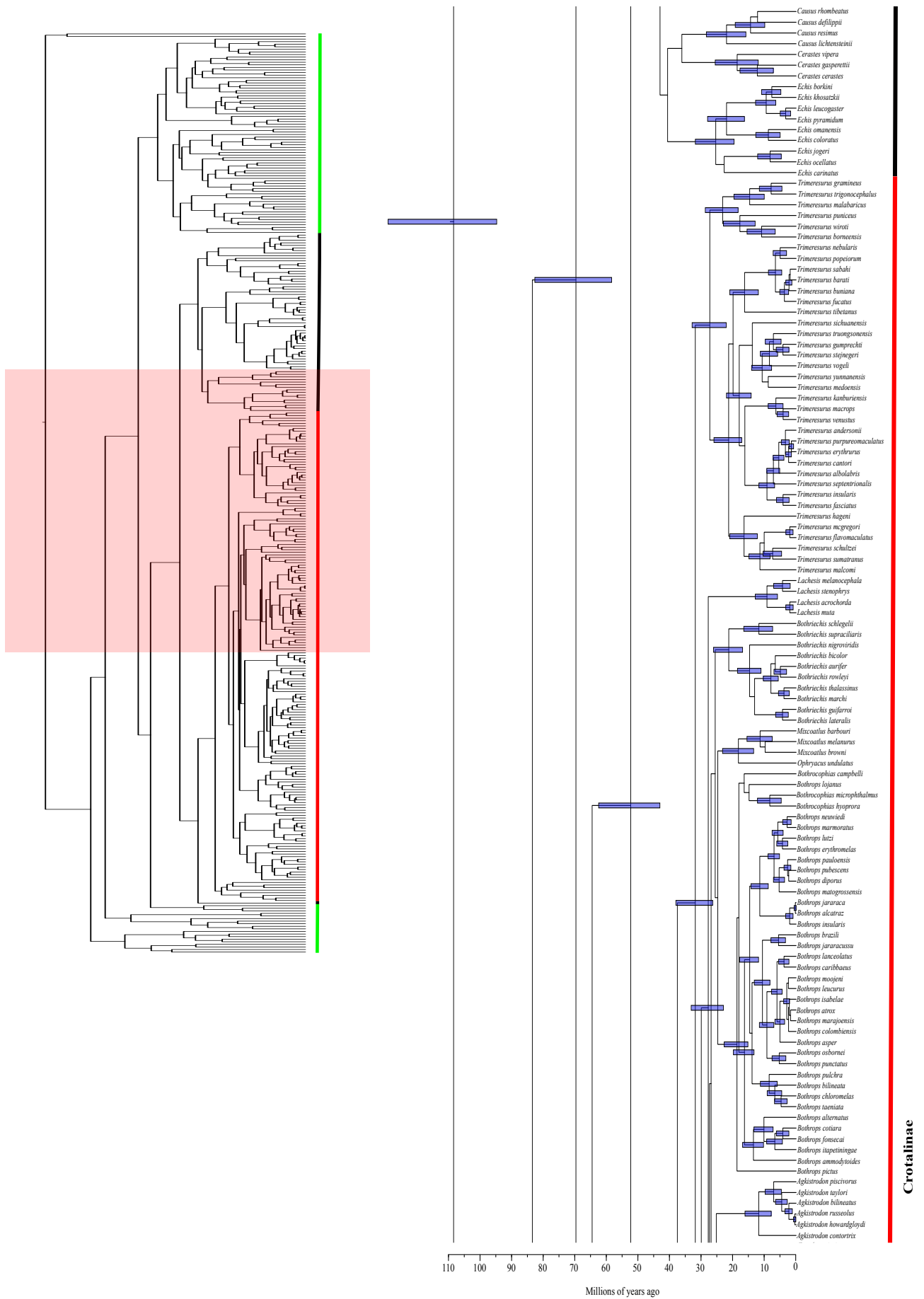


Figure S2. part2

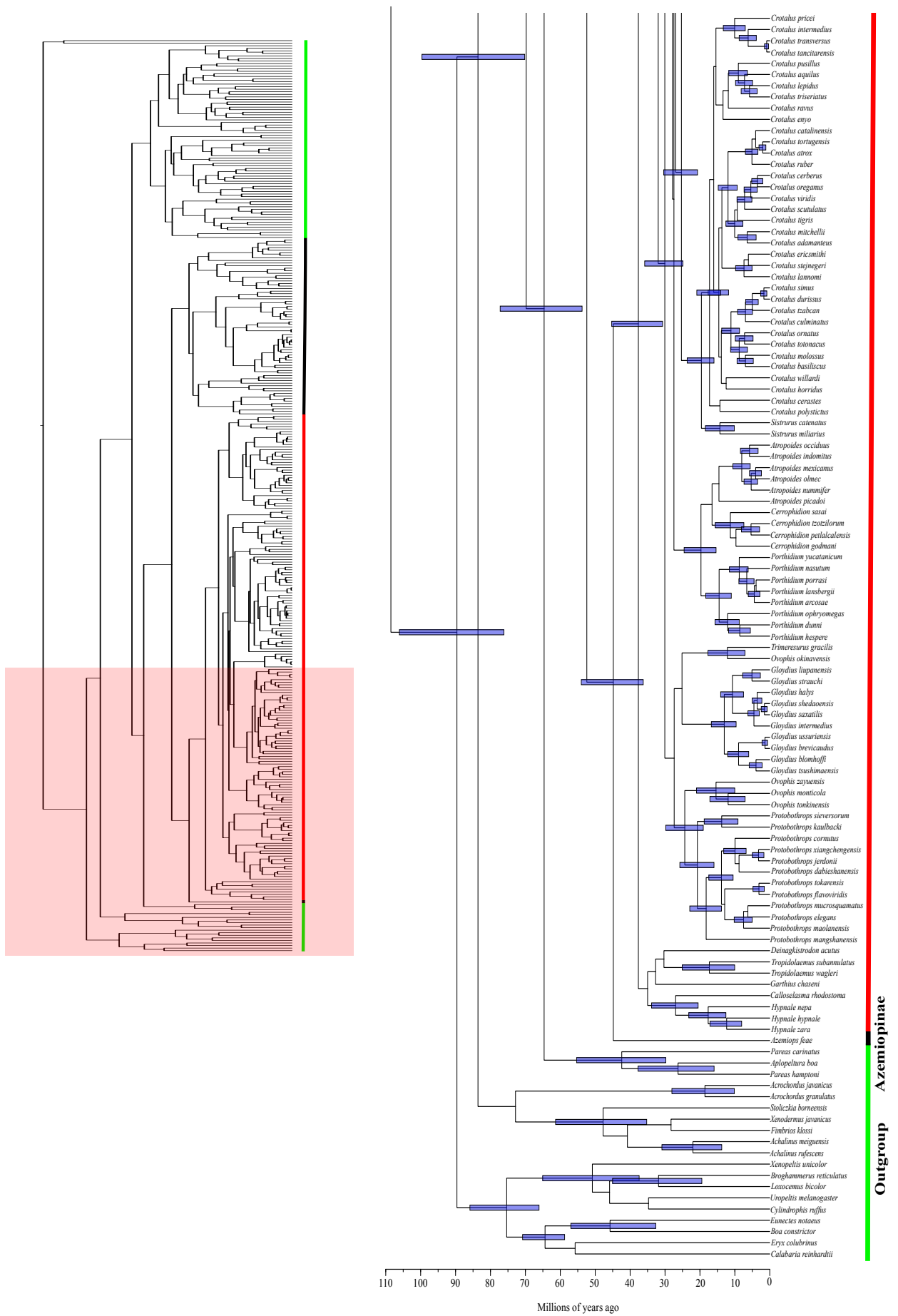


Figure S2. part3

Figure S2. Maximum credibility tree generated for the family Viperidae using the concatenated dataset comprising sequences from 12S, 16S, Cytb and NAD4 genes. These four mitochondrial genes comprise sequences representing the majority of the species included in the phylogenetic analysis. Bars showing the 95% high posterior density interval of age estimates were added only for nodes with posterior probability values equal to or higher than 0.95. Large red squares denote the position zoomed.

SUPPLEMENTARY TABLES

Table S1. GenBank and Bold accession numbers of sequences analyzed. Accession numbers in bold correspond to new sequences to be deposited in GenBank. * denotes sequences obtained from the Bold systems database. YPX = State Key Laboratory of Genetic resources and Evolution, Kunming Institute of Zoology. Murphy et al. (2011): Murphy, J.C., Mumpuni, Sanders, K.L., 2011. First molecular evidence for the phylogenetic placement of the enigmatic snake genus *Brachyorrhos* (Serpentes: Caenophidia). *Molecular Phylogenetics and Evolution* 61(3), 953-957.

Species	Mitochondrial genes						Nuclear genes				
	12S	16S	Cytb	ND4	ND2	Cox1	Cmos	Rag1	NT3	Bndf	Jun
Ingroup											
<i>Agkistrodon bilineatus</i>	AF156591	AF156570	-	AF156583	-	-	-	-	-	-	-
<i>Agkistrodon contortrix</i>	AF259224	AF259117	EU483263	AF156576	-	EANAH587-12*	-	EU402833	-	EU402623	-
<i>Agkistrodon howardgloydi</i>	AF156592	AF156571	-	AF156584	-	-	-	-	-	-	-
<i>Agkistrodon piscivorus</i>	NC_009768- YPX802	EF669477- YPX802	DQ523161	DQ523161	DQ523161	NC_009768	YPX802	YPX802	YPX802	YPX802	YPX802
<i>Agkistrodon russeolus</i>	L01766.1	AF156573	-	AF156586	-	-	-	-	-	-	-
<i>Agkistrodon taylori</i>	AF156590	AF156569	EU483477	AF156581	-	-	-	-	-	-	-
<i>Atheris barbouri</i>	-	AJ275739	AJ275686	-	-	-	-	-	-	-	-
<i>Atheris ceratophora</i>	DQ305410	DQ305433	DQ305456	DQ305474	-	-	-	-	-	-	-
<i>Atheris chlorechis</i>	YPX797	YPX797	YPX797	EU624211	-	-	YPX797	YPX797	YPX797	YPX797	YPX797
<i>Atheris desaixi</i>	-	AJ275733	AJ275680	-	-	-	-	-	-	-	-
<i>Atheris hispida</i>	-	AJ275734	AJ275681	-	-	-	-	-	-	-	-
<i>Atheris matildae</i>	-	-	JF825389	-	-	-	-	-	-	-	-
<i>Atheris nitschei</i>	AY223650	AY223663	AF471070	AY223618	-	-	AF471125	-	-	-	-
<i>Atheris squamigera</i>	AF544762	EU624279	EU624303	EU624212	-	-	AF544734	-	-	-	-
<i>Atropoides indomitus</i>	-	-	-	KC354700	-	-	-	-	-	-	-
<i>Atropoides mexicanus</i>	KC847268	KC847255	KC847271	KC847289	-	-	-	-	-	-	-
<i>Atropoides nummifer</i>	DQ305422	DQ305445	DQ061195	DQ061220	-	-	-	-	-	-	-
<i>Atropoides occiduus</i>	DQ305423	DQ305446	AY220315	AY220338	-	-	-	-	-	-	-
<i>Atropoides olmec</i>	AY223656	AY223669	AY220321	AY220344	-	-	-	-	-	-	-
<i>Atropoides picadoi</i>	AF057208	AF057255	AY223583	U41872	-	-	-	-	-	-	-
<i>Azemiops feae</i>	YPX807	AF512748- YPX807	AY352747- YPX807	AY352808	-	-	YPX807	EU402836	YPX807	EU402628	YPX807
<i>Bitis arietans</i>	YPX408	YPX408	AY235728- YPX408	EU852304	JX073288	EU852310	-	YPX408	YPX408	-	YPX408
<i>Bitis armata</i>	-	-	-	-	JX073291	-	-	-	-	-	-
<i>Bitis atropos</i>	EU624246	EU624281	AJ275691	EU624214	JX073287	-	-	-	-	-	-
<i>Bitis caudalis</i>	EU624247	EU624282	AJ275693	EU624215	JX073293	-	-	-	-	-	-
<i>Bitis cornuta</i>	EU624248	EU624283	EU624305	EU624216	-	-	-	-	-	-	-
<i>Bitis gabonica</i>	YPX816	YPX816	AJ275695- YPX816	EU624217	JX073296	-	YPX816	YPX816	YPX816	YPX816	YPX816
<i>Bitis nasicornis</i>	DQ305411	DQ305434	DQ305457	DQ305475	-	-	AY187970	-	-	-	-
<i>Bitis parviocula</i>	-	-	-	-	JX073292	-	-	-	-	-	-
<i>Bitis peringueyi</i>	DQ305412	DQ305435	DQ305458	DQ305476	-	-	-	-	-	-	-

<i>Bitis rhinoceros</i>	EU624250	EU624285	AJ275696	EU624218	-	-	-	-	-	-	-
<i>Bitis rubida</i>	EU624251	EU624286	EU624306	EU624219	JX073290	-	-	-	-	-	-
<i>Bitis schneideri</i>	-	-	-	-	JX073297	-	-	-	-	-	-
<i>Bitis worthingtoni</i>	EU624252	AJ275745	AJ275692	EU624220	JX073295	-	-	-	-	-	-
<i>Bitis xeropaga</i>	EU624253	EU624287	EU624307	EU624221	JX073294	-	-	-	-	-	-
<i>Bothriechis aurifer</i>	DQ305425	DQ305448	DQ305466	DQ305483	-	-	-	-	-	-	-
<i>Bothriechis bicolor</i>	DQ305426	DQ305449	DQ305467	DQ305484	-	-	-	-	-	-	-
<i>Bothriechis guifarroii</i>	KC847265	KC847264	KC847274	KC847287	-	-	-	-	-	-	-
<i>Bothriechis lateralis</i>	AF057211	AF057258	AY223588	U41873	-	EANAH450-12*	-	-	-	-	-
<i>Bothriechis marchi</i>	DQ305428	DQ305451	DQ305469	DQ305486	-	-	-	-	-	-	-
<i>Bothriechis nigroviridis</i>	AF057212	AF057259	AY223589	AY223635	-	-	-	-	-	-	-
<i>Bothriechis rowleyi</i>	DQ305427	DQ305450	DQ305468	DQ305485	-	-	-	-	-	-	-
<i>Bothriechis schlegelii</i>	AF057213	AF057260	AY223590	AF292611	-	EANAH419-12*	AF544680	AY487374	FJ434084	FJ433983	-
<i>Bothriechis superciliaris</i>	DQ305429	DQ305452	DQ305470	DQ305487	-	-	-	-	-	-	-
<i>Bothriechis thalassinus</i>	DQ305424	DQ305447	DQ305465	DQ305482	-	-	-	-	-	-	-
<i>Bothrocophias campbelli</i>	-	-	AF191582	AF292622	-	-	-	-	-	-	-
<i>Bothrocophias hyoprora</i>	AF057206	AF057253	AY223593	AF292614	-	-	-	-	-	-	-
<i>Bothrocophias microphthalmus</i>	AY223657	AY223670	-	AY223638	-	-	-	-	-	-	-
<i>Bothrops alcatraz</i>	-	-	AY865822	-	-	-	-	-	-	-	-
<i>Bothrops alternatus</i>	AY223660	AY223673	AY223601	AY223642	-	-	-	-	-	-	-
<i>Bothrops ammodytoides</i>	AY223658	AY223671	AY223595	AY223639	-	-	-	-	-	-	-
<i>Bothrops asper</i>	AF057218	AF057265	AY223599	U41876	-	EANAH434-12*	-	EU402838	-	EU402630	-
<i>Bothrops atrox</i>	YPX813	YPX813	AF246267-YPX813	AF292642	-	-	YPX813	YPX813	YPX813	YPX813	YPX813
<i>Bothrops barnetti</i>	-	-	-	-	-	EANAH423-12*	-	-	-	-	-
<i>Bothrops bilineatus</i>	AF057214	AF057261	AY223591	U41875	-	-	-	-	-	-	-
<i>Bothrops brazili</i>	EU867252	EU867264	AF292597	AF292635	-	EANAH874-12*	YPX812	YPX812	-	-	-
<i>Bothrops caribbaeus</i>	-	-	AF292598	AF292636	-	-	-	-	-	-	-
<i>Bothrops chloromelas</i>	DQ305430	DQ305453	DQ305471	DQ305488	-	-	-	-	-	-	-
<i>Bothrops colombiensis</i>	-	-	AF292602	AF292640	-	-	-	-	-	-	-
<i>Bothrops cotiara</i>	AF057217	AF057264	AY223597	AY223640	-	-	-	-	-	-	-
<i>Bothrops diporus</i>	DQ305431	DQ305454	DQ305472	DQ305489	-	-	-	-	-	-	-
<i>Bothrops erythromelas</i>	AF057219	AF057266	AY223600	U41877	-	-	-	-	-	-	-
<i>Bothrops fonsecai</i>	YPX411	YPX411	YPX411	AF292618	-	-	YPX411	YPX411	-	-	YPX411
<i>Bothrops insularis</i>	AF057216	AF057263	AY223596	AF292628	-	-	-	-	-	-	-
<i>Bothrops isabellae</i>	-	-	AF292603	AF292641	-	-	-	-	-	-	-
<i>Bothrops itapetiningae</i>	EU867253	EU867265	EU867277	EU867289	-	-	-	-	-	-	-
<i>Bothrops jararaca</i>	YPX412	EU867266	YPX412	AY122860	-	-	-	-	YPX412	-	-
<i>Bothrops jararacussu</i>	YPX413	YPX413	AY223602-YPX413	AF292634	-	-	YPX413	YPX413	YPX413	-	YPX413
<i>Bothrops lanceolatus</i>	-	-	AF292599	AF292637	-	-	-	-	-	-	-
<i>Bothrops leucurus</i>	EU867255	EU867267	EU867279	EU867291	-	EANAH422-12*	-	-	-	-	-
<i>Bothrops lojanus</i>	-	-	FR691566	FR691536	-	-	-	-	-	-	-
<i>Bothrops lutzi</i>	-	-	KF801131	KF801261	-	-	-	-	-	-	-
<i>Bothrops marajoensis</i>	-	-	AF292605	AF292643	-	-	-	-	-	-	-
<i>Bothrops marmoratus</i>	-	-	KF801137	KF801265	-	-	-	-	-	-	-
<i>Bothrops mattogrossensis</i>	-	-	KF801150	KF801278	-	-	-	-	-	-	-
<i>Bothrops moojeni</i>	EU867256	EU867268	EU867282	EU867293	-	-	-	-	-	-	-
<i>Bothrops neuwiedi</i>	-	-	AF292585	AF292623	-	-	-	-	-	-	-
<i>Bothrops osbornei</i>	-	-	AF292595	AF292633	-	-	-	-	-	-	-

<i>Bothrops pauloensis</i>	EU867260	EU867272	EU867284	EU867296	-	-	-	-	-	-	-
<i>Bothrops pictus</i>	-	-	AF292583	AF292621	-	-	-	-	-	-	-
<i>Bothrops pubescens</i>	JN870180	JN870192	JN870200	JN870206	-	-	-	-	-	-	-
<i>Bothrops pulchra</i>	JN870179	-	AF292593	AF292631	-	-	-	-	-	-	-
<i>Bothrops punctatus</i>	-	-	AF292594	AF292632	-	-	-	-	-	-	-
<i>Bothrops taeniata</i>	YPX414	YPX414	AY223592- YPX414	AF292629	-	-	-	-	-	-	-
<i>Calloselasma rhodostoma</i>	YPX806	YPX806	AF292569	AY352813	-	-	YPX806	YPX806	YPX806	YPX806	YPX806
<i>Causus defilippii</i>	AF057186	AF057233	AY223556	NC13479	NC13479	NC13479	-	-	-	EU402633	-
<i>Causus lichtensteini</i>	YPX796	YPX796	YPX796	-	-	-	YPX796	YPX796	YPX796	YPX796	YPX796
<i>Causus resimus</i>	AY223649	AY223662	AY223555	AY223616	-	-	AF544696	-	-	-	-
<i>Causus rhombeatus</i>	DQ305409	DQ305432	DQ305455	U41866	-	-	-	-	-	-	-
<i>Cerastes cerastes</i>	YPX800	YPX800	AF471028	EU624222	-	EU852311	AF471131-AF544679- YPX800	EU852329	YPX800	YPX800	YPX800
<i>Cerastes gasperettii</i>	JN870181	AJ275756	AJ275704	-	-	-	-	-	-	-	-
<i>Cerastes vipera</i>	-	AJ275757	AJ275705	-	-	-	-	-	-	-	-
<i>Mixcoatlus barbouri</i>	HM363639	HM363640	HM363641	HM363642	-	-	-	-	-	-	-
<i>Cerrophidion godmani</i>	DQ305419	DQ305442	AY220325	AY220348	-	-	-	-	-	-	-
<i>Cerrophidion petlalcalensis</i>	DQ305420	DQ305443	DQ061202	DQ061227	-	-	-	-	-	-	-
<i>Cerrophidion sasai</i>	-	-	-	DQ061225	-	-	-	-	-	-	-
<i>Cerrophidion tzotzilorum</i>	JN870182	JN870193	DQ061203	DQ061228	-	-	-	-	-	-	-
<i>Crotalus adamanteus</i>	AF259255- YPX793	YPX793	AF259185	U41880	-	KC750807	YPX793	JN621000	YPX793	-	YPX793
<i>Crotalus aquilus</i>	AF259232	AF259125	AF259162	HQ257782	-	-	-	-	-	-	-
<i>Crotalus atrox</i>	AF259256	AF259148	AY223608	DQ679873	AY016239	EANAH725-12*	JN620892	JN621002	KF410312	-	-
<i>Crotalus basiliscus</i>	AF259244	AF259136	AY704845	AY704894	AY704796	-	-	-	KF410313	-	-
<i>Crotalus catalinensis</i>	AF259259	AF259151	-	-	-	-	-	-	-	-	-
<i>Crotalus cerastes</i>	AF259235	AF259128	U69773	JN620962	AY016245	EANAH642-12*	JN620894	JN621003	KF410314	-	-
<i>Crotalus cerberus</i>	-	-	JN620813	JN620963	AY016225	-	JN620895	-	-	-	-
<i>Crotalus culminatus</i>	-	-	AY704830	AY704880	AY704785	-	-	-	-	-	-
<i>Crotalus durissus</i>	AF259248	AF259140	AF200223	AF292608	AY704789	JQ627372	-	-	-	-	-
<i>Crotalus enyo</i>	AF259245	AF259137	-	-	AY016246	-	-	-	-	-	-
<i>Crotalus ericsmithi</i>	-	-	KF410284	KF410290	-	-	KF410301	KF410330	KF410315	-	-
<i>Crotalus horridus</i>	AF259252	AF259144	AF337057	NC14400	NC14400	NC14400	KF410302	JN621004	KF410316	-	-
<i>Crotalus intermedius</i>	AF259238	AF259131	-	JN870208	-	-	-	-	-	-	-
<i>Crotalus lannomi</i>	-	-	KF410280	KF410292	-	-	KF410303	KF410331	KF410318	-	-
<i>Crotalus lepidus</i>	AF259230	AF259123	AF259160	U41881	-	EANAH735-12*	JN620898	JN621005	-	-	-
<i>Crotalus mitchellii</i>	AF259250	AF259142	-	JN620967	AY016241	-	JN620899	JN621006	-	-	-
<i>Crotalus molossus</i>	AF259243	AF259135	JN620845	JN620995	AY704797	EANAH740-12*	JN620924	JN621030	-	-	-
<i>Crotalus oreganus</i>	AF259253	AF259145	AF147857	AF194161	AY016221	-	AF471135	-	-	-	-
<i>Crotalus ornatus</i>	-	-	JN620820	JN620970	-	-	JN620902	JN621009	-	-	-
<i>Crotalus polystictus</i>	AF259236	AF512747	-	KF410294	-	-	KF410305	KF410332	KF410320	-	-
<i>Crotalus pricei</i>	AF259237	AF259130	-	JN022878	-	EANAH897-12*	KF410306	-	KF410321	-	-
<i>Crotalus pusillus</i>	AF259229	AF259122	AF259159	HQ257774	-	-	-	-	-	-	-
<i>Crotalus ravus</i>	AF259228	AF259121	AY223609	AY223647	-	-	-	-	-	-	-
<i>Crotalus ruber</i>	AF259261	AF259153	-	DQ679838	-	EANAH1090-13*	-	-	-	-	-
<i>Crotalus scutulatus</i>	AF259254	AF259146	AF292571	AF194167	AY016237	EANAH766-12*	-	-	KF410322	-	-
<i>Crotalus simus</i>	EU624240	EU624274	EU624302	AY704885	-	-	-	-	-	-	-
<i>Crotalus stejnegeri</i>	-	-	KF410283	KF410296	-	-	KF410308	KF410333	KF410323	-	-
<i>Crotalus tancitarenis</i>	JN022897	JN022897	-	JN022851	-	-	-	-	-	-	-

<i>Crotalus tigris</i>	AF259249	AF259141	AY223606	AF156574	AY016240	EANAH747-12*	JN620900	JN621007	GQ334665	-	-
<i>Crotalus tortugensis</i>	AF259257	AF259149	AF259187	DQ679839	-	-	-	-	-	-	-
<i>Crotalus totonacus</i>	-	-	AY704839	AY704889	AY704795	-	-	-	-	-	-
<i>Crotalus transversus</i>	AF259239	JN022895	AF259169	JN022840	-	-	-	-	-	-	-
<i>Crotalus triseriatus</i>	HQ257625	HQ257625	-	HQ257870	-	-	KF410310	-	KF410325	-	-
<i>Crotalus tzcaban</i>	-	-	AY704806	AY704856	AY704791	-	-	-	-	-	-
<i>Crotalus viridis</i>	-	-	JN620819	AF194160	AY016235	-	JN620901	JN621008	-	-	-
<i>Crotalus willardi</i>	AF259242- YPX794	AF259133- YPX794	AF259170- YPX794	JN870209	-	EANAH756-12*	YPX794	YPX794	YPX794	YPX794	-
<i>Daboia deserti</i>	-	AJ275765	AJ275712	-	-	-	-	-	-	-	-
<i>Daboia mauritanica</i>	EU624261	EU624295	EU624313	EU624229	-	-	-	-	-	-	-
<i>Daboia palaestinae</i>	JN870183	AJ275775	AJ275722	-	-	-	-	-	-	-	-
<i>Daboia russellii</i>	EU913478	NC011391	NC11391	EU913478	NC11391	NC011391	AF471156	EU402843	-	EU402636	-
<i>Daboia siamensis</i>	DQ305413	DQ305436	DQ305459	DQ305477	-	-	-	-	-	-	-
<i>Deinagkistrodon acutus</i>	DQ343647- YPX799	NC010223- YPX799	EU913476- YPX799	DQ343647	DQ343647	NC010223	YPX799	YPX799	YPX799	YPX799	YPX799
<i>Echis borkini</i>	GQ359644	GQ359730	GQ359486	GQ359563	-	-	-	-	-	-	-
<i>Echis carinatus</i>	EU852313	EU852319	EU852295	GQ359524	-	EU852307	-	EU852325	-	-	-
<i>Echis coloratus</i>	EU852315	EU852321	EU852297	EU624224	-	EU852309	-	EU852327	-	-	-
<i>Echis jogeri</i>	GQ359641	GQ359732	GQ359476	GQ359560	-	-	-	-	-	-	-
<i>Echis khosatzkii</i>	GQ359615	EU642575	EU642584	GQ359536	-	-	-	-	-	-	-
<i>Echis leucogaster</i>	GQ359622	GQ359705	GQ359456	GQ359541	-	-	-	-	-	-	-
<i>Echis ocellatus</i>	EU852312	EU852318	EU642592	GQ359510	-	EU852306	-	EU852324	-	-	-
<i>Echis omanensis</i>	GQ359631	GQ359715	GQ359466	GQ359550	-	-	-	-	-	-	-
<i>Echis pyramidum</i>	EU852314	EU852320	EU852296	EU624226	-	EU852308	-	EU852326	-	-	-
<i>Eristicophis macmahoni</i>	EU624259	EU624293	AJ275711	EU624227	-	-	-	-	-	-	-
<i>Garthius chaseni</i>	AY352791	AY352729	AY352760	AY352825	-	-	-	-	-	-	-
<i>Gloydus blomhoffi</i>	AY352780	AY352719	AY352751	AY352814	-	-	-	-	-	-	-
<i>Gloydus brevicaudus</i>	AY352781	AY352720	AY352752	AY352815	NC_011390	NC_011390	JQ687505	-	-	-	-
<i>Gloydus halyi</i>	YPX792	AF057238	AY223564	AY223621	AY662540	-	-	AY662614_ YPX782	YPX792	YPX792	-
<i>Gloydus intermedius</i>	EF012806	JN870194	JN870201	U41869	-	-	JQ687507	-	-	-	-
<i>Gloydus liupanensis</i>	-	-	JQ687491	JQ687472	-	-	JQ687510	-	-	-	-
<i>Gloydus saxatilis</i>	JN870185	JN870195	JN870202	JN870210	-	JQ798882	JQ687508	-	-	-	-
<i>Gloydus shedaensis</i>	AF057194	AF057241	AY223566	AY223623	-	-	AF435019	-	-	-	-
<i>Gloydus strauchi</i>	EF012813- YPX798	AF057239	AY223563- YPX798	EF012795	-	-	YPX798	YPX798	YPX798	YPX798	YPX798
<i>Gloydus tsushimaensis</i>	JN870186	JN870196	JN870203	JN870211	-	-	-	-	-	-	-
<i>Gloydus ussuriensis</i>	AF057193	AF057240	AY223565	AY223622	-	JQ798886	JQ687520	-	-	-	-
<i>Hypnale hypnale</i>	AY352778	AY352717	AY352750	AY352812	-	-	-	-	-	-	-
<i>Hypnale nepa</i>	KC347325	KC347362	KC347463	KC347491	-	-	KC347401	KC347439	-	-	-
<i>Hypnale zara</i>	KC347326	KC347363	KC347485	KC347513	-	-	KC347402	KC347440	-	-	-
<i>Lachesis acrochorda</i>	JN870187	JN870197	JN870204	JN870212	-	-	-	-	-	-	-
<i>Lachesis melanocephala</i>	-	-	U96018	U96028	-	-	-	-	-	-	-
<i>Lachesis muta</i>	AF057221	AF057268	AY223604	AY223644	-	-	-	-	-	-	-
<i>Lachesis stenophrys</i>	AF057220	AF057267	AY223603	EU402644	-	-	-	-	-	-	-
<i>Macrovipera lebetina</i>	YPX817	YPX817	AJ275713	DQ897729	-	-	YPX817	YPX817	YPX817	YPX817	YPX817
<i>Macrovipera schweizeri</i>	EU624262	AJ275768	AJ275715	-	-	-	-	-	-	-	-
<i>Mixcoatlus browni</i>	HM363650	HM363651	HM363652	HM363653	-	-	-	-	-	-	-
<i>Montivipera albizona</i>	EU624265	AJ275780	AJ275727	DQ897731	-	-	-	-	-	-	-
<i>Montivipera borrmuelleri</i>	-	AJ275779	AJ275726	-	-	-	-	-	-	-	-

<i>Montivipera_latifii</i>	JN870191	JN870199	JN870205	-	-	-	-	-	-	-	-
<i>Montivipera_raddei</i>	YPX818	AJ275784	AJ275730- YPX818	-	-	EANAH446-12*	YPX818	YPX818	YPX818	YPX818	YPX818
<i>Montivipera_wagneri</i>	JN870188	AJ275778	AJ275725	JN870213	-	-	-	-	-	-	-
<i>Montivipera_xanthina</i>	EU624268	AJ275777	AJ275724	EU624234	-	-	-	-	-	-	-
<i>Mixcoatlus_melanurus</i>	AF057210	AF057257	AY223587	AY223634	-	-	-	-	-	-	-
<i>Ophryacus_undulatus</i>	AF057209	AF057256	AY223586	AY223633	-	-	-	-	-	-	-
<i>Ovophis_monticola</i>	AY763191- YPX808	YPX808	AY352748-AF171907	AY352809	-	-	YPX808	YPX808	YPX808	YPX808	-
				NC_00739	NC_00739	NC_00739					
<i>Ovophis_okinavensis</i>	AB175670	AB175670	NC_007397	7	7	NC_007397	-	-	-	-	-
<i>Ovophis_tonkinensis</i>	HQ325308	HQ325092	HQ325132	HQ325209	-	-	-	-	-	-	-
<i>Ovophis_zayuensis</i>	HQ325273	HQ325089	HQ325150	HQ325208	-	-	-	-	-	-	-
<i>Porthidium_arcosae</i>	EU624241	EU624275	AF292575	AF292613	-	-	-	-	-	-	-
<i>Porthidium_dunni</i>	AY223654	AY223667	AY223581	DQ061242	-	-	-	-	-	-	-
<i>Porthidium_hesperie</i>	-	-	EU017535	EU016099	-	-	-	-	-	-	-
<i>Porthidium_lansbergii</i>	EU624242	EU624276	AY713375	AF393623	-	-	-	-	-	-	-
<i>Porthidium_nasutum</i>	YPX810	YPX810	DQ061210- YPX810	DQ061232	-	-	YPX810	-	YPX810	YPX810	YPX810
<i>Porthidium_ophryomegas</i>	AF057205	AF057252	AY223580	U41888	-	-	-	-	-	-	-
<i>Porthidium_porrasi</i>	DQ305421	DQ305444	DQ061214	DQ061237	-	-	-	-	-	-	-
<i>Porthidium_yucatanicum</i>	JN870189	JN870198	DQ061215	DQ061244	-	-	-	-	-	-	-
<i>Proatheris_superciliaris</i>	EU624263	EU624296	AJ275685	EU624230	-	-	-	-	-	-	-
					NC_02269						
<i>Protobothrops_cornutus</i>	AY294276	AY294267	AY294272	AY294262	5	NC_022695	-	-	-	-	-
				NC_02247	3	NC_022473	-	-	-	-	-
<i>Protobothrops_dabieshanensis</i>	NC_022473	NC_022473	NC_022473	3	3	NC_022473	-	-	-	-	-
<i>Protobothrops_elegans</i>	AF057201	AF057248	AY223575	U41893	-	-	-	-	-	-	-
<i>Protobothrops_flavoviridis</i>	AY352792	AY352730	AY223574	AY352826	-	-	-	-	-	-	-
					NC_02140						
<i>Protobothrops_jerdonii</i>	EF012807	AY763193	AY294274	AY294264	2	NC_021402	-	-	-	-	-
<i>Protobothrops_kaulbacki</i>	DQ666056	DQ666055	DQ666060	DQ666057	-	-	-	-	-	-	-
<i>Protobothrops_mangshanensis</i>	AY352787	AY352726	AY352758	AY352821	-	JX233620	JQ687524	-	-	-	-
<i>Protobothrops_maolanensis</i>	JN799405	JN799398	JN799401	JN799409	-	-	-	-	-	-	-
					NC_02141						
<i>Protobothrops_mucrosquamatus</i>	AY223653	AY223666	AY223577	AY223629	2	NC_021412	-	-	-	-	-
<i>Protobothrops_sieversorum</i>	DQ305414	DQ305437	AY352753	DQ305478	-	-	-	-	-	-	-
<i>Protobothrops_tokarensis</i>	AF057202	AF057249	AY223576	AY223628	-	-	-	-	-	-	-
<i>Protobothrops_xiangchengensis</i>	AY763189	AY763207	DQ666062	DQ666058	KF460436	KF460436	-	-	-	-	-
<i>Pseudocerastes_persicus</i>	-	AJ275770	AJ275717	-	-	-	-	-	-	-	-
<i>Pseudocerastes_urarachnoides</i>	-	-	KF314714	-	-	-	-	-	-	-	-
<i>Pseudocerastes_fieldi</i>	EU624264	AJ275769	AJ275716	-	-	-	-	-	-	-	-
<i>Sistrurus_catenatus</i>	DQ464268	AF259119	AY223610	HQ257759	GQ359800	EANAH863-12*	-	-	GQ334686	-	-
<i>Sistrurus_miliarius</i>	YPX795	AF259120- YPX795	EU483385	HQ257760	GQ359815	EANAH768-12*	-	-	-	-	-
<i>Trimeresurus_albolabris</i>	AF057195	AF057242	AY223567	U41890	KF311102	KF311102	-	-	-	-	-
<i>Trimeresurus_andersonii</i>	AY352801	AY352740	-	AY352835	-	-	-	-	-	-	-
<i>Trimeresurus_barati</i>	AY371753	AY371769	AY371801	AY371837	-	-	-	-	-	-	-
<i>Trimeresurus_borneensis</i>	AY352783	AY352722	AY352754	AY352817	-	-	-	-	-	-	-
<i>Trimeresurus_buniana</i>	AY371752	AY371778	AY371818	AY371853	-	-	-	-	-	-	-
<i>Trimeresurus_cantori</i>	AY352802	AY352741	AF171889	AY352836	-	-	-	-	-	-	-
<i>Trimeresurus_erythrus</i>	AF517161	AF517174	AF171900	AF517217	-	-	-	-	-	-	-
<i>Trimeresurus_fasciatus</i>	GQ428492	GQ428466	GQ428475	GQ428482	-	-	-	-	-	-	-
<i>Trimeresurus_flavomaculatus</i>	AY059535	AY059551	AF171916	AY059584	-	-	-	-	-	-	-
<i>Trimeresurus_fucatus</i>	AY059537	AY059553	AY371796	AY059588	-	-	-	-	-	-	-

<i>Trimeresurus gracilis</i>	DQ305415	DQ305438	AF171913	AY352823	-	-	-	-	-	-	-
<i>Trimeresurus gramineus</i>	AY352793	AY352731	AY352762	AY352827	-	-	-	-	-	-	-
<i>Trimeresurus gumprechtii</i>	EU443791	EU443792	AY321489	EU443787	-	KC171172	-	-	-	-	-
<i>Trimeresurus hageni</i>	AY059536	AY059552	AY059567	AY059585	-	-	-	-	-	-	-
<i>Trimeresurus insularis</i>	AY059534	AY352738	AY059568	AY352833	-	-	-	-	-	-	-
<i>Trimeresurus kanburiensis</i>	AY289219	AY352737	AY289225	AY289231	-	-	-	-	-	-	-
<i>Trimeresurus macrops</i>	AF517163	AF517176	AF517184	AF517219	-	-	-	-	-	-	-
<i>Trimeresurus malabaricus</i>	AY059548	AY059564	AY059569	AY059587	-	-	-	-	-	-	-
<i>Trimeresurus malcolmi</i>	AY371758	AY371793	AY371832	AY371861	-	-	-	-	-	-	-
<i>Trimeresurus mcgregori</i>	AY371756	AY371795	AY371831	AY371858	-	-	-	-	-	-	-
<i>Trimeresurus medoensis</i>	AY352797	AY352735	AY352765	AY352831	-	-	-	-	-	-	-
<i>Trimeresurus nebularis</i>	AY371737	AY371774	AY371814	AY371839	-	-	-	-	-	-	-
<i>Trimeresurus popeiorum</i>	AY059538	AY059554	AY059571	AY059590	-	-	-	-	-	-	-
<i>Trimeresurus puniceus</i>	AF517164	AF517177	AF517192	AF517220	-	-	-	-	-	-	-
<i>Trimeresurus purpureomaculatus</i>	AF517162	AF517175	AF517188	AF517218	-	-	-	-	-	-	-
<i>Trimeresurus sabahi</i>	AY371736	AY371771	AY371815	AY371842	-	-	-	-	-	-	-
<i>Trimeresurus schultzei</i>	AY352785	AY352725	AY352756	AY352819	-	-	-	-	-	-	-
<i>Trimeresurus septentrionalis</i>	AY059543	AY352724	AY352755	AY352818	-	-	-	-	-	-	-
<i>Trimeresurus sichuanensis</i>	HQ850445	HQ850446	HQ850448	HQ850450	-	-	-	-	-	-	-
<i>Trimeresurus stejnegeri</i>	NC012146- YPX805	FJ752492- YPX805	YPX805	FJ752492	NC_01214	NC_012146	YPX805	YPX805	YPX805	YPX805	YPX805
<i>Trimeresurus sumatranus</i>	AY371759	AY371788	AY371828	AY371866	6	-	-	-	-	-	-
<i>Trimeresurus tibetanus</i>	AY352776	AY352715	AY352749	AY352810	-	-	-	-	-	-	-
<i>Trimeresurus trigonocephalus</i>	AY059549	AY059565	AF171890	AY059597	-	-	KC347412	-	-	-	-
<i>Trimeresurus truongsongensis</i>	EU443817	EU443818	EU443815	EU443816	-	-	-	-	-	-	-
<i>Trimeresurus venustus</i>	AY293931	AY352723	AF171914	AY289228	-	-	-	-	-	-	-
<i>Trimeresurus vogeli</i>	AF517170	AF517183	AY059574	EU443808	-	-	-	-	-	-	-
<i>Trimeresurus wiroti</i>	-	-	DQ646788	-	-	-	-	-	-	-	-
<i>Trimeresurus yunnanensis</i>	EU443813	EU443814	EF597523	EF597527	-	KC171173	-	-	-	-	-
<i>Tropidolaemus subannulatus</i>	AF057198	AF057245	AY223571	AY223625	-	-	-	-	-	-	-
<i>Tropidolaemus wagleri</i>	AF517167	AF517180	AF517191	AF517223	-	-	-	-	-	-	-
<i>Vipera ammodytes</i>	EU624266	EU624297	AM944799	EU624232	AM944799	-	-	-	-	-	-
<i>Vipera anatolica</i>	-	-	KC316113	-	-	KC122762	-	-	-	-	-
<i>Vipera aspis</i>	JN870190	-	AM944742	-	AM944744	-	-	-	-	-	-
<i>Vipera barani</i>	-	-	AY321092	-	-	-	-	-	-	-	-
<i>Vipera berus</i>	EU543221	DQ186081	DQ186032	DQ897728	EU625373	KC122718	-	-	-	-	-
<i>Vipera dinniki</i>	-	AJ275773	AJ275720	-	-	KC122721	-	-	-	-	-
<i>Vipera erivanensis</i>	YPX803	YPX803	YPX803	FR727021	-	KC122733	YPX803	-	YPX803	YPX803	YPX803
<i>Vipera kaznakovi</i>	-	-	AY321093	FR727034	-	KC122724	-	-	-	-	-
<i>Vipera latastei</i>	-	-	AY321094	JX649572	AY321074	-	-	-	-	-	-
<i>Vipera lotievi</i>	-	-	JN204717	FR727030	-	KC122732	-	-	-	-	-
<i>Vipera nikolskii</i>	YPX814	YPX814	YPX814	-	EU625372	EANAH876-12*	YPX814	YPX814	-	YPX814	YPX814
<i>Vipera orlovi</i>	-	-	KC176746	-	-	KC122734	-	-	-	-	-
<i>Vipera renardi</i>	-	-	HQ845740	FR727033	-	KC122761	-	-	-	-	-
<i>Vipera seoanei</i>	-	AJ275782	DQ186030	FR727035	AY321071	KC122736	-	-	-	-	-
<i>Vipera transcaucasiana</i>	-	-	DQ186479	-	AY321070	-	-	-	-	-	-
<i>Vipera ursinii</i>	AF236687	-	AY311383	FR726981	AY321069	KC122737	-	-	-	-	-

Outgroup

<i>Achalinus_meiguensis</i>	FJ424614	FJ424614	FJ424614	FJ424614	FJ424614	NC_011576	-	-	-	-	-
<i>Achalinus_rufescens</i>	YPX105	YPX105	YPX105	U49319	-	-	-	YPX105	YPX105	YPX105	YPX105
<i>Acrochordus_granulatus</i>	AB177879	NC007400	-	NC_007400	-	-	YPX464	AY487388_EU402831	-	-	EF144042
<i>Acrochordus_javanicus</i>	YPX001	YPX001	YPX001	HM234055	-	-	-	YPX001	YPX001	YPX001	YPX001
<i>Ahaetulla_prasina</i>	YPX543	YPX543	-	-	-	-	YPX543	-	YPX543	YPX543	YPX543
<i>Ahaetulla_pulverulenta</i>	KC347304	-	KC347454	KC347512	-	-	KC347378	KC347416	-	-	-
<i>Anilius_scytale</i>	FJ755180	-	FJ755180	FJ755180	NC_014343	AF544722_AY099965	NC_014343	AY988072_AY487382	AY988055	FJ433961	FJ434012
<i>Antillophis_parvifrons</i>	YPX108	YPX108	FJ416740_YPX108	FJ416814	FJ416778	-	-	YPX108	YPX108	YPX108	YPX108
<i>Aparallactus_modestus</i>	FJ404130	AY611824	AY612007	FJ404332	-	-	AY611916_FJ404235	-	-	-	-
<i>Aplopeltura_boa</i>	AF544761	AF544787	-	U49312	-	-	AF544715	-	FJ434085	FJ433984	EF144044
<i>Aspidelaps_scutatus</i>	YPX589	YPX589	AY188007_YPX589	AY058969	-	-	YPX589	YPX589	YPX589	YPX589	YPX589
<i>Atractaspis_corpulenta</i>	Z46597	AY611837	AY612020	FJ404335	-	-	AY611929_FJ404238	DQ993174	-	-	-
<i>Atractaspis_micropholis</i>	AF544740	AY611823	AY612006	FJ404336	NC_007398	AY611915_AF544677	-	-	FJ434095	FJ433994	EF144053
<i>Boa_constrictor</i>	NC_007398	NC_007398	AF471036	NC_007398	NC_007398	AF471115_AF544676_AF471115	NC_007398	AY487351_EU402837	AY988047	AY988030	-
<i>Boaedon_fuliginosus</i>	AY122681	AY188079	AF471060	HQ207151	-	AY122663	DQ486163_AF544686	AY487378_EU402849	FJ434094	-	EF144052
<i>Brachyorrhos_albus</i>	Murphy et al. (2011)	Murphy et al. (2011)	Murphy et al. (2011)	-	-	-	-	-	-	-	-
<i>Brogammerus_reticulatus</i>	Z46448	-	U69860_FJ717481	-	-	-	AF544675	EU624119_AY487396	FJ434074	FJ433969	FJ434020
<i>Bungarus_fasciatus</i>	NC_011393	NC_011393	EU579523	EU579523	NC_011393	AF544732_AY058924_YPX591	NC_011393	AY487389	YPX591	YPX591	YPX591
<i>Calabaria_reinhardtii</i>	Z46464	-	AY099985	AF302943	-	AF544682_AY099978	-	EU402839_AY487391	AY988058	AY988041	FJ434023
<i>Calamaria_pavimentata</i>	YPX548	YPX548	YPX548	-	-	-	YPX548	EF144092	YPX548	YPX548	YPX548
<i>Calamaria_yunnanensis</i>	YPX503	YPX503	YPX503	-	-	-	YPX503	YPX503	YPX503	YPX503	YPX503
<i>Cantoria_yiolacea</i>	AF499292_YPX087	YPX087	EF395897_YPX087	-	-	-	EF395922	YPX087	YPX087	YPX087	YPX087
<i>Cerberus_rynchops</i>	AF499289_YPX500	YPX500	YPX500	U49327	-	-	YPX500	YPX500	YPX500	YPX500	YPX500
<i>Chrysopelea_ornata</i>	YPX063	YPX063	YPX063	KC347496	-	-	KC347393	YPX063	YPX063	YPX063	YPX063
<i>Coluber_constrictor</i>	U96794_AY122819_YPX528	YPX528	EU180486	AF138746	AY487002	AY122735	YPX528	YPX528	YPX528	YPX528	YPX528
<i>Compsophis_infralineatus</i>	FJ404157_YPX97	YPX097	EF203990	FJ404359	-	-	EF203996_EF204000_FJ404259	YPX097	YPX097	YPX097	-
<i>Coronella_austriaca</i>	AY122836_YPX537	YPX537	AY486930_YPX537	AY487065	AY487026	AY122752	AY486954	YPX537	YPX537	YPX537	-
<i>Cubophis_cantherigerus</i>	AF158405	AF158475	AF544669	FJ416818	FJ416782	-	AF544694	AY487376	FJ434100	FJ433999	EF144057
<i>Cylindrophis_ruffus</i>	NC_007401	-	NC_007401	NC_007401	NC_007401	NC_007401	AF471133_AF544698	AY662613	AY988054	EU402635	YPX157
<i>Diadophis_punctatus</i>	AY577015	YPX089	EU193843	EU194025	-	-	AF471122_AF544705_YPX089	AY487403	YPX089	EU402637	YPX089
<i>Dipsas_catesbyi</i>	Z46459	YPX119	EF078537_YPX119	EF078558	-	-	JQ598977	YPX119	YPX119	YPX119	YPX119
<i>Dromicodryas_bernieri</i>	YPX096	YPX096	DQ979987	FJ404353	-	-	AY187975	YPX096	YPX096	YPX096	YPX096
<i>Drymarchon_corais</i>	HM565758	YPX068	AF471064_YPX68	DQ902314	DQ902207	-	AF471137	YPX068	YPX068	YPX068	YPX068
<i>Duberria_lutrix</i>	FJ404154_YPX090	YPX090	DQ486413_YPX90	DQ486308	-	-	FJ387207_DQ486161_AF471138	YPX090	YPX090	YPX090	YPX090
<i>Eirenis_modestus</i>	AY039143	YPX533	YPX533	AY487072	AY487033	AY039181	YPX533	YPX533	YPX533	YPX533	-
<i>Erpeton_tentaculatus</i>	AF499286_YPX527	YPX527	EF395913_YPX527	-	-	-	AF544700_EF395936_YPX527	YPX527	YPX527	YPX527	YPX527
<i>Erythrolamprus_militaris</i>	AF158409_YPX129	YPX129	YPX129	-	-	-	JQ598982	YPX129	YPX129	YPX129	YPX129
<i>Eryx_colubrinus</i>	AF544747	-	U69811	-	-	-	AF544716_DQ465568	EU402845_AY487392	FJ434078	DQ465570	FJ434027
<i>Eunectes_notaeus</i>	AM236347	-	AM236347	AM236347	AM236347	AM236347	YPX491	AY988063	AY988046	AY988029	YPX491
<i>Fimbrios_klossi</i>	-	-	YPX545	-	-	-	-	YPX545	-	YPX545	-
<i>Gonionotophis_brussauxi</i>	FJ404156	YPX054	AY612043	FJ404358	-	-	AY611952_FJ404258	YPX054	YPX054	YPX054	YPX054
<i>Gonionotophis_capensis</i>	AF544769	YPX060	HQ207116	AF544665	-	-	AF544703	AY487379	FJ434096	FJ433995	EF144054
<i>Gonyosoma_oxycephalum</i>	AY122678_YPX558	YPX558	AF471084_YPX558	DQ902309	DQ902241	AY122660	YPX558	YPX558	YPX558	YPX558	YPX558
<i>Grayia_ornata</i>	AF158434	AF158503	YPX055	AF544663	-	-	YPX055	EF144091	YPX055	YPX055	EF144061
<i>Heterodon_platirhinos</i>	AY577019	YPX568	YPX568	AF402659	FJ416750	-	YPX568	-	-	YPX568	-
<i>Homalopsis_buccata</i>	AF499288_YPX560	EF395868_YPX560	EF395917	-	-	-	EF395940_AF544701_YPX560	-	YPX560	YPX560	YPX560

<i>Homoroselaps lacteus</i>	FJ404135_YPX158	AY611809_YPX158	AY612026_YPX158	AY058976 NC_01398	- NC_01398	- NC_01398	FJ404241_AY611935	-	YPX158	YPX158	YPX158
<i>Imantodes cenchoa</i>	EU728586_YPX127	EU728586_YPX127	EU728586_YPX127	8	8	NC_01398	GQ457865	YPX127	YPX127	YPX127	-
<i>Ithycyphus miniatius</i>	YPX098	YPX098	AY188019_YPX098	-	-	-	AY187980	YPX098	YPX098	YPX098	YPX127
<i>Lamprozelotes getula</i>	AY122821_YPX511	YPX511	FJ997820_YPX511	AF138759	-	AY122738	YPX511	EU402848_YPX511	YPX511	YPX511	YPX511
<i>Laticauda colubrina</i>	U96799	EU547138	EU547040	AY058977	-	-	AY058932_UJ366446	AY487404_EU402850	FJ434091	EU402647	EF144050
<i>Leioheterodon madagascariensis</i>	AF544768_YPX502	AY188061	AY188022_YPX502	FJ404371	-	-	YPX502_AF544685_AY187983	AY487377	YPX502	YPX502	YPX502
<i>Leptodeira annulata</i>	GQ457806	GQ457746	EF078516	FJ416787	FJ416749	-	AF544690_GQ457866	AY487375	FJ434099	FJ433998	EF144056
<i>Liophidium torquatum</i>	YPX100	YPX100	DQ979984_YPX100	-	-	-	AY187984	YPX100	YPX100	YPX100	YPX100
<i>Loxocemus bicolor</i>	Z46456	-	AY099993	-	-	-	AF544730_AY444035	EU402854_AY487406	FJ434072	FJ433967	FJ434018
<i>Lycophidion capense</i>	FJ404178	AY611893	AY612075	FJ404376	-	-	DQ486168_FJ404279	EU402855	-	EU402652	-
<i>Macropisthodon rudis</i>	YPX091	YPX091	YPX091	U49326	JQ687458	-	JQ687452	YPX091	YPX091	YPX091	YPX091
<i>Malpolon monspessulanus</i>	YPX570	YPX570	AY058965_YPX570	FJ404320	-	-	YPX570	YPX570	YPX570	YPX570	YPX570
<i>Micrurus surinamensis</i>	AF544770	AF544799	EF137415	EF137407	-	-	EF137422_AF544708	AY487411	FJ434092	FJ433991	EF144051
<i>Mimophis mahafalensis</i>	CTMZ-04140	-	CTMZ-04140	DQ486202	-	-	AY187993	CTMZ-04140	CTMZ-04140	CTMZ-04140	CTMZ-04140
<i>Naja atra</i>	YPX829	EU913475_YPX829	EU913475_YPX829	EU921898	EU921898	NC_011389	YPX829	-	YPX829	YPX829	YPX829
<i>Natrix natrix</i>	YPX538	-	AY866541_YPX538	AY873716	AY870624	AY122664	AF471121_AF544697_YPX539	EU402858	-	EU402655	YPX538
<i>Nerodia sipedon</i>	AF402630_YPX525	YPX525	AF402913_YPX052	-	DQ915150	GQ278935	YPX052	YPX052	YPX052	YPX052	YPX052
<i>Ninia atrata</i>	GQ457814	YPX131	YPX131	GQ334659	-	-	GQ457874	YPX131	YPX131	YPX131	YPX131
<i>Ophiophagus hannah</i>	EU921899_YPX830	EU921899_YPX830	EU921899_YPX830	EU921899	NC_01139	NC_011394	YPX830	YPX830	YPX830	YPX830	YPX830
<i>Pantherophis guttatus</i>	AY122814	AM236349_YPX071	AM236349	AM236349	DQ902218	AY122730	YPX071	-	-	YPX071	YPX071
<i>Pareas carinatus</i>	AF544773	AF544802	JQ598940	-	-	-	AF544692	-	FJ434086	FJ433985	EF144045
<i>Pareas hamptoni</i>	YPX562	YPX562	AY425809_YPX562	-	-	-	YPX562	EU402860_YPX562	YPX562	YPX562	YPX562
<i>Pituophis deppei</i>	YPX515	YPX515	FJ627818_YPX515	AF138765	FJ627848	AY122741	YPX515	YPX515	YPX515	YPX515	YPX515
<i>Polemon acanthias</i>	FJ404138	AY611848	AY612031	FJ404341	-	-	FJ404243_AY611940	-	-	-	-
<i>Psammophis elegans</i>	YPX530	YPX530	-	EU526862	-	-	YPX530	-	YPX530	YPX530	YPX530
<i>Psammophis mossambicus</i>	YPX094	YPX094	FJ404314_YPX094	DQ486283	-	-	DQ486185_FJ404224	YPX094	-	YPX094	YPX094
<i>Psammophis subtaeniatus</i>	-	YPX095	DQ486358	DQ486253	-	-	-	YPX095	YPX095	YPX095	YPX095
<i>Psammophylax variabilis</i>	AF544774	AY611864	AY612046	EU526859	-	-	AF544709_AY611955	AY487380	FJ434097	FJ433996	EF144055
<i>Pseudonaja modesta</i>	EU547098	YPX588	EU547049	DQ098490	-	-	YPX588	-	YPX588	YPX588	YPX588
<i>Pseudoxenodon bambusicola</i>	YPX551	YPX551	-	-	-	-	YPX551	EF144090	YPX551	YPX551	YPX551
<i>Pseudoxenodon karlschmidti</i>	YPX564	YPX564	YPX564	-	-	-	YPX564	YPX564	YPX564	YPX564	YPX564
<i>Psomophis joberti</i>	GQ457829	GQ457768	YPX137	-	-	-	GQ457889	YPX137	YPX137	YPX137	YPX137
<i>Ptyas korros</i>	AF236680_YPX542	YPX542	AY486929_YPX542	AY487062 NC_01019	AY487023 NC_01019	AY122652	YPX542	-	YPX542	YPX542	YPX542
<i>Indotyphlops braminus</i>	DQ343649	-	DQ343649	6	6	DQ343649	AF544717_AY099980	AY444062_AY487410	GU902556	FJ433959	FJ434010
<i>Rhachidelus brazili</i>	JQ598837	JQ598897	YPX139	-	-	-	-	YPX139	YPX139	YPX139	YPX139
<i>Rhamphiophis oxyrhynchus</i>	YPX523	YPX523	-	-	-	-	FJ387213_AF544710_YPX523	-	YPX523	YPX523	YPX523
<i>Scaphiodontophis annulatus</i>	-	-	GQ927323	-	-	-	GQ927318	YPX080	YPX080	YPX080	YPX080
<i>Sibynomorphus mikanii</i>	GQ457832	GQ457771	YPX141	-	-	-	GQ457892	YPX141	YPX141	YPX141	YPX141
<i>Sibynophis collaris</i>	YPX535	YPX535	KC000121	JN211315	KC000133	-	YPX535	-	YPX535	YPX535	YPX535
<i>Sinomicrocurus macclellandi</i>	YPX590	YPX590	YPX590	EF137410	-	-	YPX590	YPX590	YPX590	YPX590	YPX590
<i>Sinonatrix percarinata</i>	YPX552	YPX552	YPX552	JQ687414	JQ687455	-	YPX552	YPX552	YPX552	YPX552	YPX552
<i>Spalerosophis diadema</i>	AY039148_YPX526	YPX526	AY486926_YPX526	AY487059	AY487020	AY039186	YPX526	YPX526	YPX526	YPX526	YPX526
<i>Stoliczka borneensis</i>	AF544779	AF544808	-	-	-	-	AF544721	AY487398	FJ434083	FJ433982	EF144043
<i>Storeria dekayi</i>	AF402639_YPX567	YPX567	AF471050_YPX567	EF417365	EF417460	EF417389	YPX567	-	YPX567	YPX567	-
<i>Taeniophallus nicagus</i>	YPX146	YPX146	YPX146	-	-	-	CTMZ0483	YPX146	YPX146	YPX146	-
<i>Thamnophis sirtalis</i>	YPX539	YPX539	AF420193	AF420196	DQ995365	-	DQ902094	YPX539	YPX539	YPX539	YPX539

<i>Tomodon dorsatus</i>	GQ457838	GQ457777	YPX148	-	-	-	GQ457897	YPX148	YPX148	YPX148	YPX148
<i>Uropeltis melanogaster</i>	AF512739	-	-	-	-	-	-	AY487399	FJ434070	FJ433965	FJ434016
<i>Xenochrophis flavipunctatus</i>	AF544780	AF544809	-	-	FJ416748	-	AF544714	-	FJ434102	FJ434001	EF144060
<i>Xenodermus javanicus</i>	AF544781	AF544810	AY425810	U49320	-	-	AF544711	-	-	EU402667	-
<i>Xenopeltis unicolor</i>	NC_007402	-	NC_007402	NC_007402	NC_007402	NC_007402	AF544689 AY099977	EU402870 AY487400	DQ465562	EU402668	FJ434019

Table S2. Partition schemes and nucleotide substitution models suggested by PartitionFinder for each dataset. Dataset 1 corresponds to our main dataset and dataset 2 corresponds to the sequences from 16S, 12S, cytb and nd4 genes (see text for details).

Scenarios	Partition	Best Model
Dataset 1	12S, 16S	GTR+I+G
	Codon 1st and 2nd pos of bndf and 2nd pos of rag1	HKY+I+G
	Codon 3rd pos of bndf and jun	K80+G
	Codon 1st pos of cox1	SYM+I+G
	Codon 2nd pos of cox1	GTR+I+G
	Codon 3rd pos of cox1 and cytb	GTR+I+G
	Codon 1st pos of cmos, jun and rag1, and 2nd pos of cmos and jun	HKY+I+G
	Codon 3rd pos of cmos and rag	GTR+G
	Codon 1st pos of Cytb and ND4	GTR+I+G
	Codon 2nd pos of Cytb, ND2, and ND4	GTR+I+G
	Codon 1st pos of nd2	GTR+I+G
	Codon 3rd pos of nd2 and nd4	GTR+G
	Codon of 1st and 2nd pos of Nt3	SYM+I+G
	Codon 3rd position of Nt3	GTR+G
Dataset 2	12S, 16S	GTR+I+G
	Codon 1st pos of cytb and nd4	GTR+I+G
	Codon 2nd pos of cytb and nd4	GTR+I+G
	Codon 3rd pos of cytb	GTR+G
	Codon 3rd pos of nd4	GTR+G

DIVERSIFICATION DYNAMICS IN VIPERS

Alencar, L.R.V., Quental, T., Grazziotin, F.G., Alfaro, M.L., Martins, M., Venzon, M., and Zaher, H.

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Abstract. Snakes of the family Viperidae comprise around 329 venomous species distributed worldwide with the three sub-families showing striking heterogeneity in species richness. While the subfamily Azemiopinae comprises only two species, 70% of all viper species are arranged in the subfamily Crotalinae, also known as the pit vipers. The radiation of the pit vipers was marked by the evolution of the heat-sensing pits, which has been suggested to be a key innovation for the successful diversification of the group. Additionally, only crotalines were able to successfully colonize the New World. Given the disparity in species number among the lineages of vipers and the unique events that characterized the evolution of pit vipers, we explored the hypothesis that crotalines have undergone an explosive radiation. Diversification analyses suggested a shift in speciation rates during the radiation of a sub-clade of pit vipers where speciation rates rapidly increased but slowed down toward the present. Our results suggest that the evolution of the loreal pits alone does not seem to explain their explosive speciation rates. We suggest that climatic and geological changes in Asia and the invasion of the New World may have also contributed to the speciation shift found in vipers.

INTRODUCTION

The disparity in species richness across the tree of life is an intriguing large-scale pattern. For example, the order Squamata comprise around 9.900 species whereas its sister lineage, the Rhynchocephalia (tuataras), comprise only two (Hedges and Vidal, 2009, Uetz and Hosek, 2014). In order to explain such patterns scientists need to understand the underlying mechanisms driving species richness among lineages (Rabosky et al., 2012). Present-day heterogeneities in species richness can reflect different temporal dynamics of species diversification during the evolutionary history of lineages. While groups with higher species richness could be associated with higher diversification rates, low diversification rates could characterize the poor species clades like the

Tuatara (Alfaro et al., 2009). However, the fossil record indicates that species richness dramatically changed through time and lineages comprising few species in the present used to be much more diverse in the past (e.g. Quental and Marshall, 2010). This in turn suggests that diversification rates might also have changed through time. Thus, exploring the temporal patterns of species diversification can yield significant insights into the processes that might regulate diversity.

Although paleontological data are ideal to explore patterns and processes regulating diversity through time (Quental and Marshall, 2010, Silvestro et al., 2015), most groups of organisms exhibit a poor fossil record. However, in the past years several methods have been proposed to investigate the dynamics in species diversification using solely extant species by inferring their history through molecular phylogenies (e.g. Magallon and Sanderson, 2001, Alfaro et al., 2009). Although those methods were heavily criticized (Quental and Marshall, 2010, Rabosky, 2010), newer methods promise to account for speciation and extinction rate variation among lineages and along time (Morlon et al., 2011, Rabosky et al., 2013, Rabosky, 2014) and seem to better describe diversification dynamics (e.g. Morlon et al., 2011). These methods together with an increase in the availability of time-calibrated molecular phylogenies for several groups opened numerous possibilities to understand the diversification dynamics across the tree of life.

The cosmopolitan group of vipers (Serpentes: Viperidae) shows a striking heterogeneity in diversity among lineages, and a number of hypothesis have been proposed to explain differential success within the group. The early radiation of these snakes is associated with the evolution of a highly derived venom system, which may have allowed the invasion of new niches (Greene, 1997; Pyron and Burbrink, 2012). Furthermore, the evolutionary history of the subfamily Crotalinae is marked by the evolution of a pair of heat-sensing pits in each side of their heads between the eye and the nostril (“loreal pits”) (Goris, 2011; Roelke and Childress, 2007), which have been suggested as a key innovation (e.g. Greene, 1997) even though not directly tested. Crotalines also invaded the New World, an event frequently associated to an explosive radiation (Burbrink et al., 2012a; Wüster et al., 2002, 2008).

Explosive radiations or “early bursts” have been frequently reported in molecular phylogenies (e.g. Harmon et al., 2003; Morlon et al., 2012) and are usually characterized by very high diversification rates during the early radiation of a lineage followed by a decrease towards the present. The emergence of a key innovation (e.g. Glor, 2010; Losos and Mahler, 2010) and/or the invasion of new areas might allow a lineage to explore previously unavailable niches (Burbrink et al., 2012a,b) and, in theory, could be associated with explosive radiations. Although the predominant explanation regarding diversification slowdowns rely upon speciation mediated by niche differentiation and the subsequent decrease in ecological opportunities (e.g. Burbrink et al.,

2012a; Rabosky and Lovette, 2008), recent studies have suggested alternative processes that can also underlie diversification rates slowdowns (see Moen and Morlon, 2014).

In this paper we assembled the most complete time-calibrated molecular dataset of Viperidae to investigate the diversification dynamics of vipers. We also explicitly explored the hypothesis that crotalines, the most diverse subfamily, undergone an explosive radiation (Greene 1997; Wüster et al., 2008) by assessing if diversification rates significantly increased during their radiation comparing with those from the remaining vipers.

MATERIAL AND METHODS

To perform diversification analyses while incorporating phylogenetic uncertainty we randomly sampled 100 trees from the posterior distribution of dataset 1 (our main dataset, see Chapter I). We performed all diversification analyses on those 100 trees after removing the outgroup. To visually inspect the pattern of diversification across time we generated lineage through time plots (LTT, i.e. the cumulative number of lineages through time, Nee et al., 1992) for the 100 posterior trees. To test if there were temporal changes in diversification rates we first used the gamma statistic while controlling for the statistic bias imposed by species under-sampling (Pybus and Harvey, 2000). For that we simulated 1000 phylogenetic trees with the known number of viper species and then removed species in order to represent our sampling scheme. All those analysis were done both for Viperidae as a whole (329 known species plus three subspecies considered here as full species and 264 terminals in our sampling scheme) as well as for Viperinae (98 known species and 72 terminals in our sampling scheme) and Crotalinae (229 known species plus three subspecies considered here as full species and 191 terminals in our sampling scheme). These analyses were performed using the packages Ape (Paradis et al., 2004) and TreeSim (Hartmann et al., 2010; Stadler, 2014) implemented in R (R Development Core Team, 2013).

Given the limitations of the gamma statistic (Quental and Marshall 2010, 2011) we also used a novel approach named BAMM (Bayesian Analysis of Macroevolution Mixtures, Rabosky, 2014; Rabosky et al., 2013; Rabosky et al., 2014a), which estimates speciation and extinction rates throughout the different branches in a phylogenetic tree. BAMM is based on the premise that phylogenetic trees are often shaped by heterogeneous mixtures of distinct processes (see Figure 1 in Rabosky, 2014). This rate heterogeneity across lineages and time has been shown to be the case in numerous phylogenetic trees (e.g. Etienne et al., 2012) and failure to accommodate such variation can cause serious bias in results and interpretations (Rabosky, 2010; Rabosky, 2012). We note that for our data and proposed hypothesis BAMM might be specially interesting because it: 1 - accounts for incomplete taxon sampling; 2 - allows extinction rates to exceed speciation rates therefore

allowing to characterize diversification dynamics where clades are in decline; and 3 - is designed to detect rate shifts across the tree.

In this study we used BAMM to explore if the extant diversity of vipers is the result of a single or multiple diversification regimes, and how speciation and extinction rates varied throughout their evolutionary history using the phylogenetic framework generated in Chapter I. With this approach we also explicitly explored the hypothesis that the subfamily Crotalinae has a different diversification dynamic characterizing an explosive radiation. We ran BAMM on the 100 randomly sampled posterior trees for 25 million generations sampling every 10.000 generations. Given that incomplete sampling can bias analyses of diversification and that vipers are not randomly sampled in the phylogeny, we informed BAMM specific sampling fractions of clades included (Table S1). To evaluate how many distinct macroevolutionary regimes characterized the radiation of vipers, we first compiled the overall number of rate shifts found among all trees. We then generated for each tree the rate shift configuration most frequently sampled in BAMM posterior, taking into account the prior probability using a Bayes Factor criterion of 5. By doing this, we are analyzing only those rate shifts that are supported by the data and not by prior alone. Through the shift configuration it is possible to visualize when, during the radiation of vipers, diversification rates significantly changed. Finally, we analyzed how speciation and extinction rates varied through time in the whole family and in crotalines separately. BAMM output was analyzed using the R package BAMMtools (Rabosky, 2014; Rabosky et al., 2014b).

RESULTS

Lineage through time plots show no clear slowdown in lineage accumulation except very close to the present both in Viperidae and in the subfamily Crotalinae respectively (Figure 1). The negative gamma statistics estimated for the whole family and for crotalines alone suggest a slowdown in lineage diversification, looking even more evident among crotalines (Figure 1). However, the LTT plot for the subfamily Viperinae suggests that lineage accumulation decelerated right after their initial diversification but seems to have remained constant toward the present (Figure 1). The gamma statistics test after correcting for incompleteness did not show any evidence for slowdowns among viperines (Figure 1).

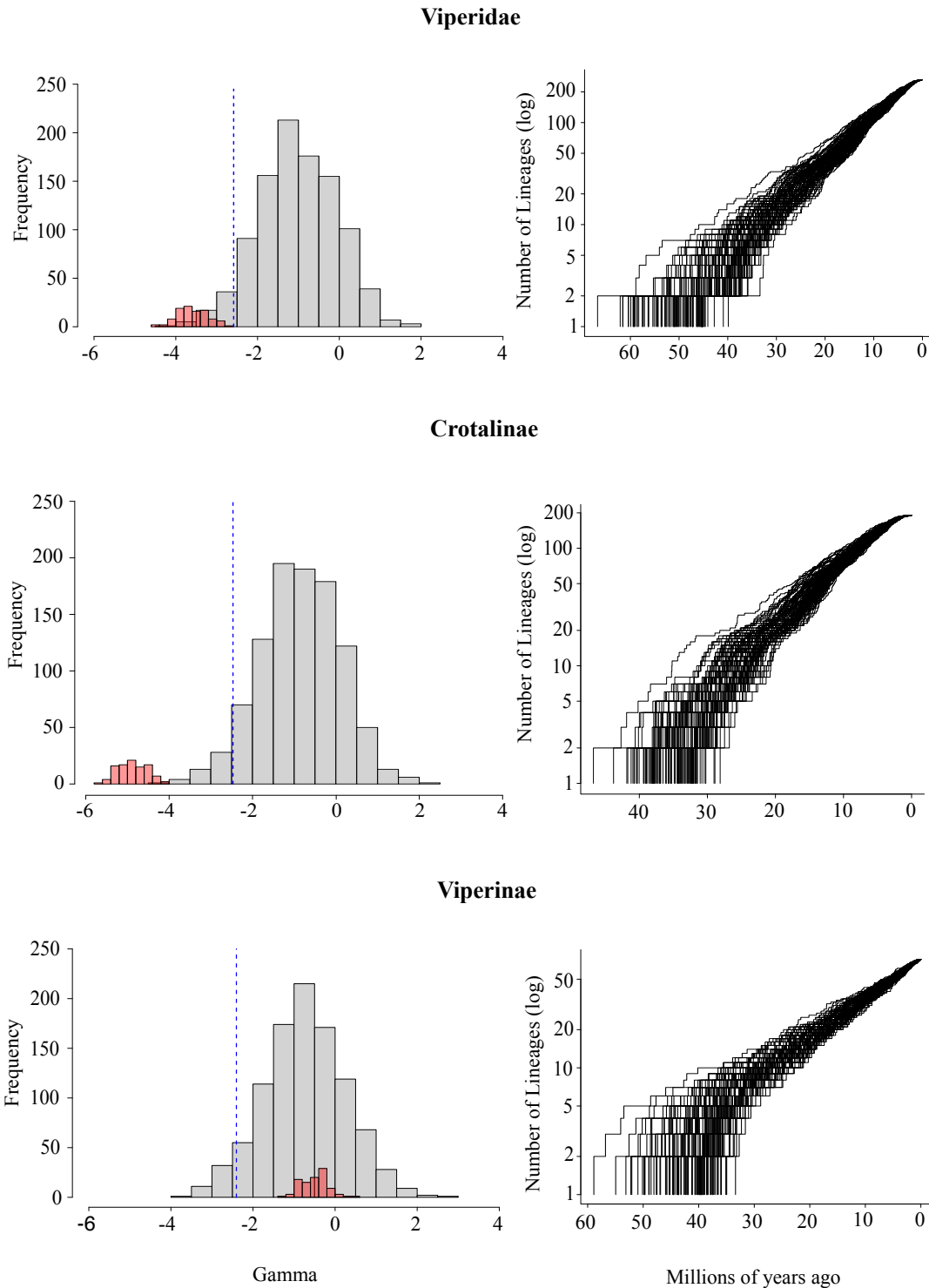


Figure 1. Simulated (shaded bars) gamma values under the null model of constant diversification rates vs empirical (red bars) gamma values, and lineage through time plots estimated for all Viperidae and subfamilies Crotalinae and Viperinae separately. The gamma values for empirical trees were calculated for 100 trees from the posterior distribution of trees (see Chapter I) and therefore incorporate phylogenetic uncertainty. Blue dashed lines represent the threshold where the null hypothesis of constant diversification rates is rejected at a 5% level.

When looking for dynamic heterogeneity, BAMM indicated that models with one or two shifts in macroevolutionary rate regimes were chosen more frequently during MCMC analyses (Figure 2). However, after discounting the shifts likely to be attributed to the prior alone, the most frequently sampled shift configuration estimated in BAMM for 83 trees (out of 100) comprised only one shift (Figure 3). Fourteen trees showed no shifts as the most frequent shift configuration, and three trees showed two shifts (Figure 4A and B). Configurations with one shift (Figure 3) suggest an increase in speciation rates after the first divergence event among crotalines, thus comprising part of the OW crotalines and the whole NW crotaline clade, with decay in speciation rates towards the present. Configurations with two shifts (Figure 4B) include the shift described above plus a second shift representing an increase in speciation rates that varies in the exact position but always among the genera *Vipera*, *Macrovipera*, *Montivipera*, *Daboia*, *Eristicophis* and *Pseudocerastes*.

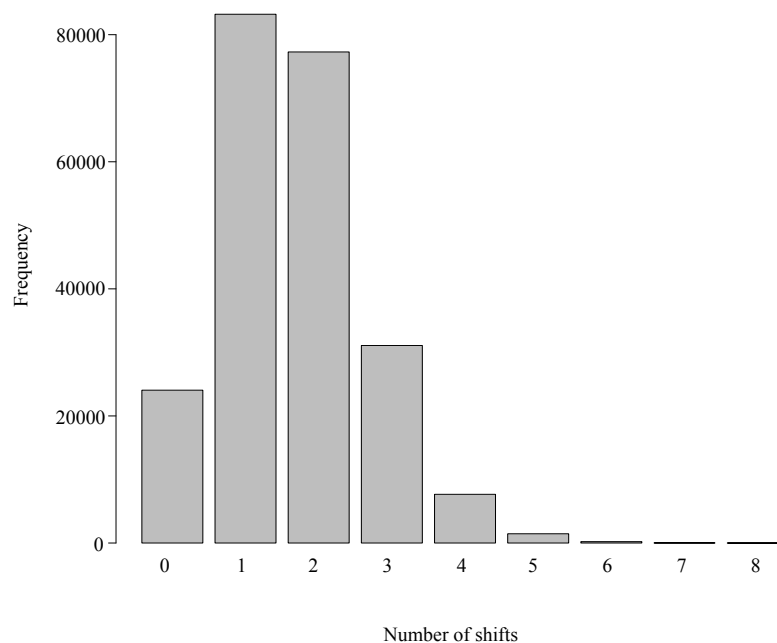


Figure 2. Number of shifts in macroevolutionary regimes found by BAMM during MCMC for the 100 posterior trees analyzed. A model with one shift was visited more frequently. Note the histogram shows the number of shifts before eliminating those shifts that were probable given the prior alone.

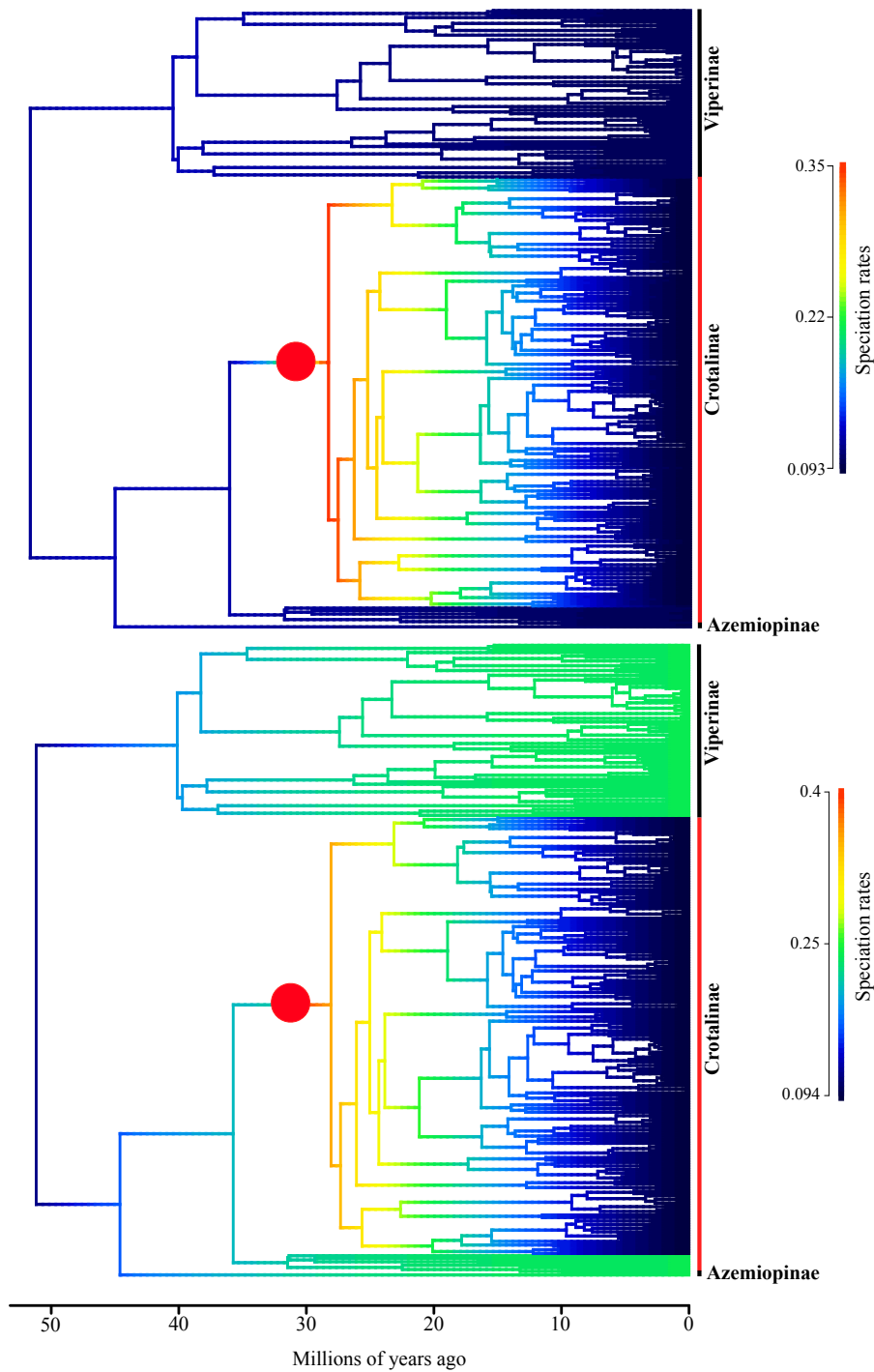


Figure 3. Best shift configuration sampled by BAMM for 83 phylogenetic trees. For each tree, speciation rates are calculated as follow: in the upper figure instantaneous rates for each posterior sample are calculated along each branch, and these instantaneous rates are then averaged. In the bottom figure BAMM posterior samples corresponding to this best shift configuration are pooled and the mean is taken for each parameter and rates are estimated. Red circle indicates the place where rate shift might have occurred.

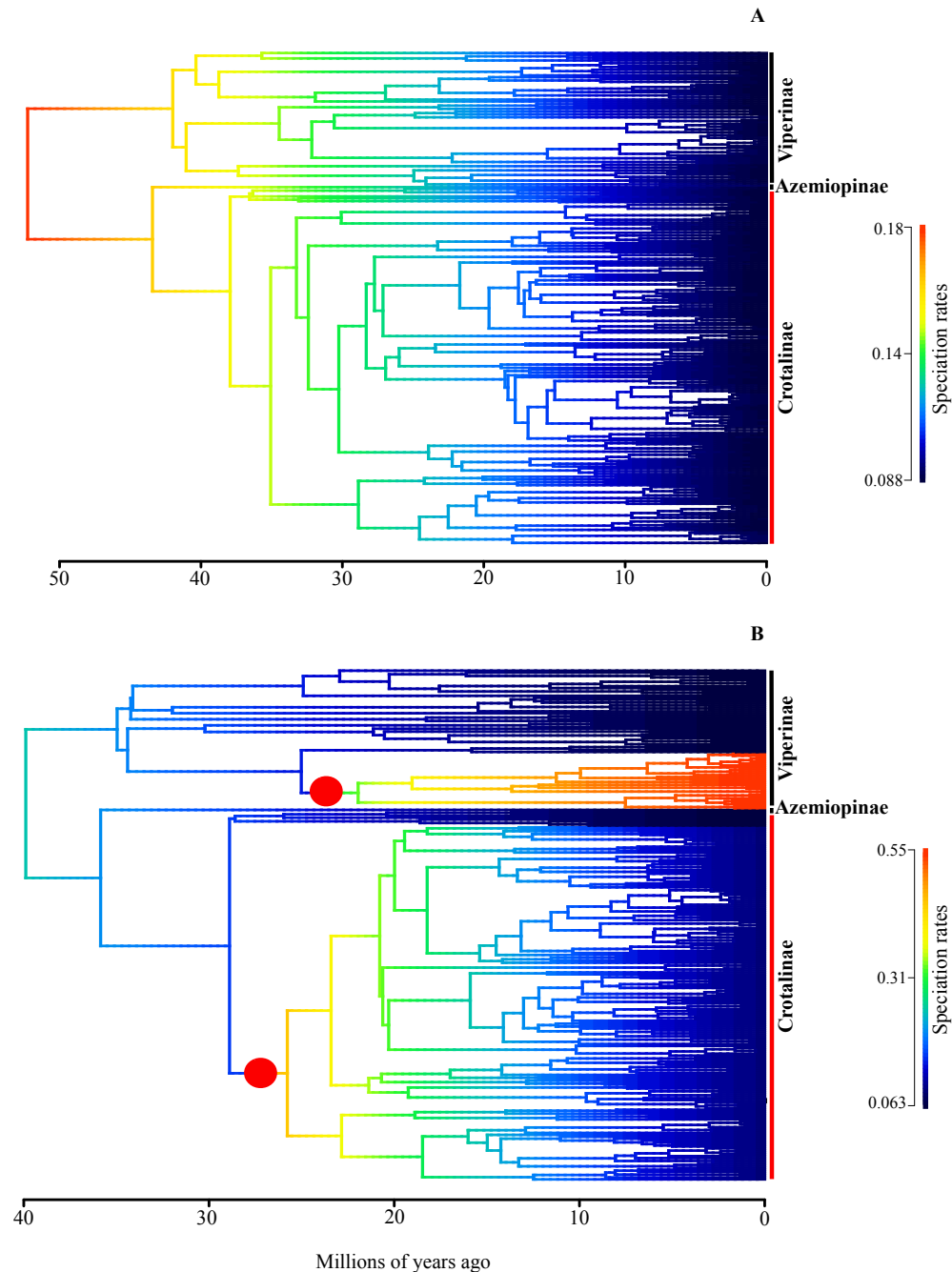


Figure 4. A) Best shift configuration sampled by BAMM for 14 phylogenetic trees. B) Best shift configuration sampled by BAMM for three phylogenetic trees. Red circles indicate the place where rate shifts might have occurred.

In summary, the great majority of BAMM runs suggest that a distinct macroevolutionary regime characterize not the whole subfamily Crotalinae but a sub-group of crotaline lineages. Although there is uncertainty on the absolute values and in the general trend in speciation rates for the background regime (that is, the lineages not part of the clade covered by the rate shift) (Figure

3), background lineages comprise a different speciation regime compared to the one inferred after the shift. The uncertainty here is due to two optimum parameter combinations, one with initial high speciation rates associated with a slightly decay on it and another with small initial values of speciation and a gentle increase on it (Figure 5A). This suggests an increase or a decrease in speciation rates and even constant speciation rates during the background diversification.

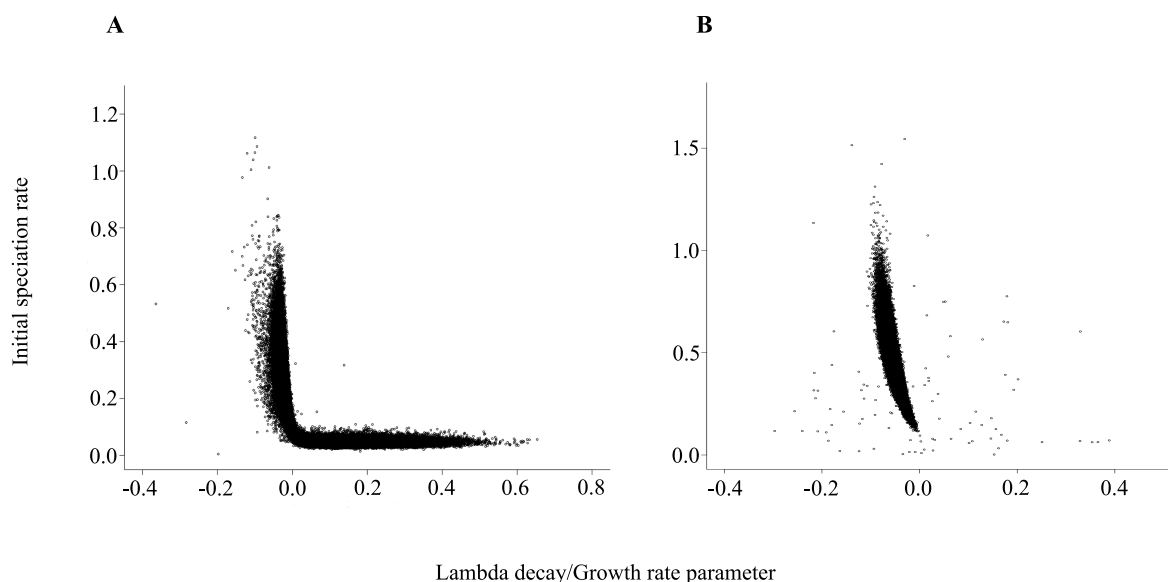


Figure 5. A) Two optimum parameter combinations estimated by BAMM for the background regime leading to uncertainty in rate estimates. B) Parameter combination estimated by BAMM for the pos-shift regime indicating confidence for a scenario where rates slow down toward the present.

Rate through time plots show that initially speciation rates slightly decreases but after a given amount of time (Figure 6A) experience a considerable increase. After this burst, speciation starts to decrease. On the other hand, extinction rates seem to remain roughly constant for most of the time (Figure 6B). Speciation through time plots generated for the Crotalinae subfamily and the remaining vipers separately (Figure 6C) showed that at the early radiation of crotalines speciation rates do not differ between the two groups. However, speciation rates among crotalines reach much higher values after this initial radiation. Extinction rates do not show any significant variation along time when plotted for crotalines and remaining vipers separately (Figure 6D).

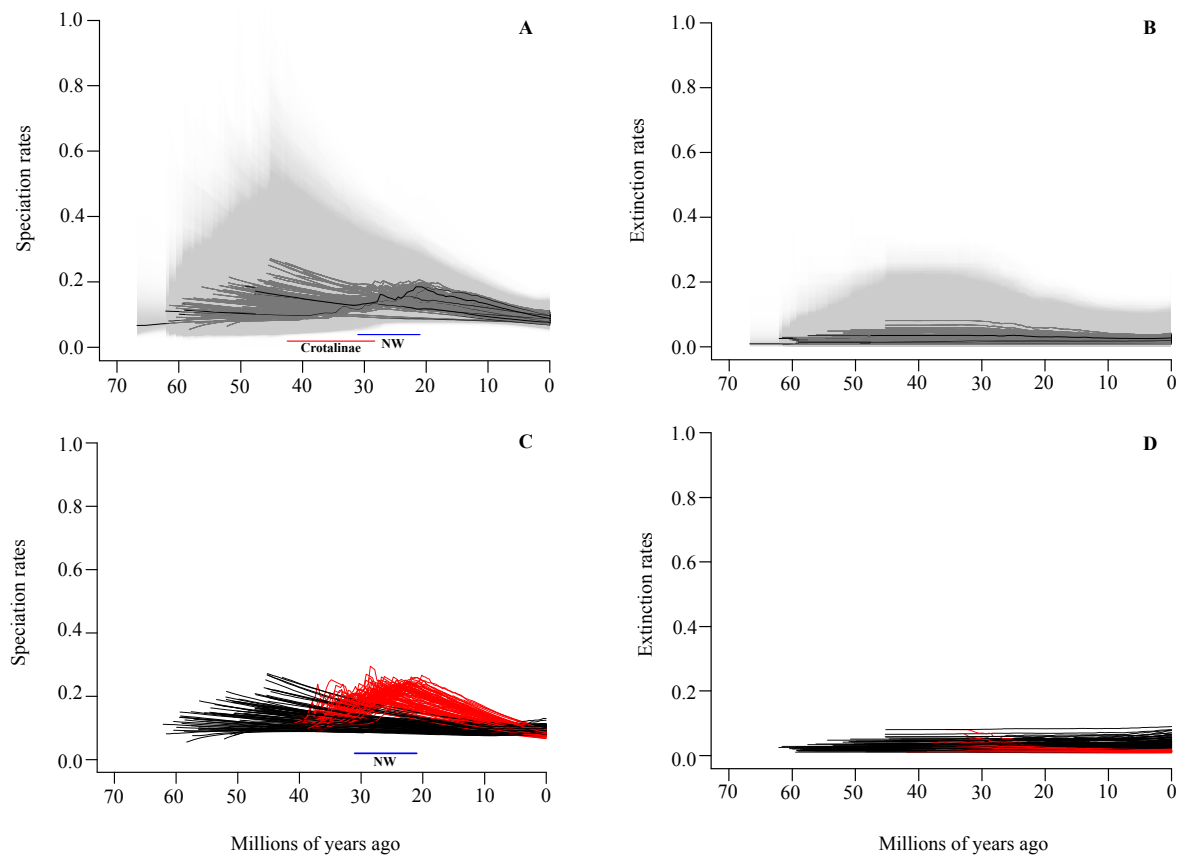


Figure 6. A) Speciation and B) extinction rates through time estimated for 100 phylogenetic trees for all vipers. Shaded regions denote the 95% credible interval of the posterior distribution of rates at a given point in time. C) Speciation and D) extinction rates through time estimated for the 83 phylogenetic trees showing the one shift configuration as the most frequent diversification scenario (see Results for details), with the subfamily Crotalinae (red lines) and the remaining vipers (black lines) plotted separately. We preferred to omit the 95% credible intervals in figures 6C and 6D for better visualization. The red and blue bars indicate the posterior density interval (dataset 1) of the crown group ages of the Crotalinae subfamily and the NW crotalines respectively.

DISCUSSION

Two distinct diversification regimes characterized the radiation of vipers. The diversification regime suggested by BAMM for the background lineages are inconclusive and prevent us from discussing which possible scenario took place after the initial radiation of the family Viperidae. This background regime basically reflects the speciation and extinction dynamics of the subfamily Viperinae, plus the species-poor subfamily Azemiopinae, and few species from the Crotalinae subfamily. BAMM results indicated that it is either likely that speciation rates increased, declined

or remained constant as soon as vipers started to diversify. Additionally, gamma values estimated for the subfamily Viperinae, which comprise the majority of lineages in the background regime, does not indicate any sign of decreasing diversification rate. Given that the deceleration signature depicted by gamma statistics might deteriorate as time goes by (Liow et al. 2010, Quental and Marshall, 2011) it is indeed difficult to infer what is the most likely dynamics for the lineages that comprises the background diversification regime, and it is likely that only a good fossil record could tease those apart (Quental and Marshall, 2010). Regardless of the real scenario underlying the early diversification of vipers and of the background lineages as a whole, it becomes evident that a different diversification regime took place during the diversification of crotalines when speciation rates started to rapidly increase (Figure 3 and 6 A, C). Surprisingly, this shift in speciation rates took place after the first divergence event among crotalines and not at the crown group of pit vipers as expected. The shift did not comprise the five very species-poor genera *Calloselasma*, *Deinagkistrodon*, *Garthius*, *Hypnale* and *Tropidolaemus* which are connected by long branches in our phylogenetic trees (see Figure 1 in Chapter I), suggesting infrequent speciation and/or frequent extinction events during their radiation. One could argue that a species undersampling effect on this clade could have biased the shift position in BAMM analyses. However, besides the fact that BAMM takes into account possible species undersampling effects (see Material and Methods), 73% of the clade species were included in our analyses. Thus, the distinct diversification regime found comprises only part of the OW crotalines and the whole NW clade (see Figure 1 in Chapter I and Figure 3 this chapter). Our results suggest that the emergence of the loreal pits in crotalines did not immediately trigger an explosive radiation, the diversification pattern usually associated with the evolution of key innovations (Glor, 2010).

It is important to highlight, however, that our results do not mean that the evolution of loreal pits was not important for the increase in speciation rates shown by the sub-group of crotalines. Studies suggest that the loreal pits evolved not only for enhancing prey finding, but also as an efficient defense mechanism and as a powerful tool in the search for optimal thermoregulation sites (e.g. Goris, 2011; Krochmal et al., 2004; Krochmal and Bakken, 2003). That is, this change in the life history of vipers may have allowed crotalines to invade different habitats facing new selective pressures. We suggest that the evolution of loreal pits in combination with other factors might have triggered the rapid speciation events found among most crotalines.

The remarkable shift in the diversification regime of vipers might have started to take place either during the middle Eocene until the Oligocene with speciation rates remaining relatively high until the Miocene. These time periods were characterized by several geographical events that occurred in Asia and in fact, pit vipers are thought to have originated in this continent (Malhotra et al., 2010; Wüster et al., 2008). Almost all extant OW pit vipers are restricted to Asia (see Figure 1

in Chapter I) mainly in the southeast and China (Uetz and Hosek 2014). The Asian continent experienced a complex geological history during the Cenozoic, and two orogenic events, the India-Asia collision and the Asia-Australia collision, drastically affected its climate, topography, and vegetation (Bruyn et al., 2014; Morley, 2012; Wang, 2004). Despite the uncertainty regarding the true time period of the collision between India and Asia (70-50 Mya), the effects of this collision are thought to have propagated throughout the Oligocene and Miocene comprising the uplift of the Tibetan plateau and of several mountain ranges in central Asia (Favre et al., 2015; Wang, 2004). These deformations together with the subsequent Asia-Australia collision had a dramatic impact on the Asian climate characterizing a significant climate shift toward much wetter climates (Bruyn et al., 2014; Morley, 2012; Wang, 2004). Since the Eocene to Miocene, arid regions retracted to northwest China and forests expanded in Southeast Asia to south and eastern China (Sun and Wang, 2005).

These climate and geological changes described may be linked to the shifts in speciation rates found among crotalines. We hypothesize that forest expansion allowed crotaline ancestral lineages to expand their ranges culminating in rapid subsequent speciation events. In fact, the great majority of OW crotalines are restricted to forested habitats (Gumprecht et al. 2004) as probably were their ancestors. Speciation events could have occurred as a consequence of the colonization of different areas imposing different selective pressures as well as due to vicariant events (Rosenzweig, 1995). Forest expansion may have also favored arboreality in Asian pit vipers with the subsequent diversification of arboreal lineages contributing to the increase in speciation rates. Interestingly, the first lineage split after the speciation shift gave rise to the arboreal and species rich genus *Trimeresurus*. We think that forest expansion in the Oligocene and Miocene might have provided a significant increase in the availability of niches and opportunities for OW crotalines to speciate.

During the late Oligocene to early Miocene pit vipers also colonized the New World, perhaps favored by the forest expansion and wetter climate during this time period in East Asia. The invasion of the NW has been suggested as a potential driver of species diversification in different groups of organisms (e.g. Barker et al., 2015; Burbrink and Pyron, 2009) and may have contributed to the increase in speciation rates observed in crotalines as already suggested in the literature (Burbrink et al., 2012a; Wüster et al., 2008). The invasion of a new area free of competitors and/or predators can lead to explosive radiations due to the greater ecological opportunities available (Glor, 2010; Rabosky and Lovette, 2008). When pit vipers invaded the NW, the snake fauna at the region was far less diverse than it is today (Burbrink and Pyron, 2009; Holman, 2000), what may have contributed to the diversification opportunities faced by NW vipers (Wüster et al., 2002).

The ecological opportunity hypothesis is usually evoked to explain explosive radiations in spatially limited systems such as islands or lakes (e.g. Darwin's finches, Schluter, 2000; Cichlids

fishes, Seehausen, 2006). When colonizing these closed systems, lineages rapidly speciate in sympatry with significant phenotypic change reflecting niche differentiation (e.g. different beaks associated with different diets among Darwin's Finches, Grant, 1981). Continental radiations may be more complex because geographical limits are much broader compared to islands (Derryberry et al., 2011; Price et al., 2014). Instead of diversifying in sympatry, the recently arrived continental lineages can disperse and occupy a wider variety of environments. The first NW pit viper populations could have easily spread into different environments due to the availability of new adequate potential habitats and the scarcity of competitors/predators. As populations reached a very large area, distinct selective pressures imposed by each environment would bring populations closer to distinct and spatially distributed adaptive optima increasing genetic and ecological divergence between them (see Cornell, 2013; Nosil, 2012, Sacks et al., 2008). Under this scenario, divergent natural selection would pull populations apart turning them into new species increasing speciation rates during the early diversification of NW pit vipers. Differently from the processes underlying radiations on islands, here, ecological opportunities were spatially distributed and sympatric speciation might have had a secondary role in generating new lineages. The fast speciation events after the invasion of the NW by pit vipers could thus be related to rapid dispersion followed by diversification into different environments characterized by distinct abiotic/biotic conditions (see Sacks et al., 2008).

After the remarkable increase in speciation rates among crotalines however, rates started to decline. Decreases in speciation rates are commonly found both on molecular phylogenies (McPeck 2008; Morlon et al., 2010) and the fossil record (Alroy, 1996; Quental and Marshall, 2013; Silvestro et al., 2015). These rate slowdowns are usually attributed to diversity dependent diversification dynamics where ecological niches are filled with new taxa, the available ecological space shrinks, and opportunities to speciate decrease (Burbrink and Pyron, 2009; Phillimore and Price, 2008; Rabosky and Lovette, 2008; but see Quental and Marshall, 2010). We suspect that in the case of vipers, the relevant driver is a geographic space that becomes saturated in a short time period. For example, after pit vipers invaded the NW rapidly dispersing through the continent, populations occupied different environments culminating in the accelerated formation of several new lineages not long after the invasion. Speciation rates would thus slowdown because the presence of a recently adapted lineage in an area might inhibit the invasion of another (niche pre-emption, see Wiens, 2011), and a large set of geographically spaced and distinct environments were no longer available all at the same time as they once were. At this time, inter-specific interactions could have started to play a major role in shaping the diversification dynamics of NW pit vipers, contributing to the decline in speciation rates. We should keep in mind, however, that most snakes, especially vipers, seem to have a very low-cost lifestyle compared to endotherms (e.g., Greene, 1983, 1986;

Nagy et al., 1999) and given the low food intake by vipers (Fitch, 1982; Greene, 1986; Martins et al. 2002), competition for food might not necessarily be the main reason why one viper species would avoid an environment already occupied by another.

A similar scenario could explain speciation slowdowns among OW crotalines with new opportunities emerging with the forest expansion in Asia. Dispersion of OW pit vipers would have been restricted to the new forested areas because differently from the NW, competitors and predators were already established in other areas. After OW pit vipers dispersed through the forests occupying the newly available opportunities, the speed of species formation would have decreased reflecting a drop in speciation rate. It is important to note that we do not mean that after speciation slowdowns environments remained stable or that new opportunities to speciate stopped to emerge. Instead, environments are always changing and new opportunities should always be available (Van Valen, 1973). However, these new opportunities will emerge in longer time intervals and/or in much more local scales compared to previous time periods. Although speculative, the scenarios above provide several predictions that can be easily tested in future studies. For such scenarios to produce a burst in speciation followed by a deceleration, the time to fully occupy a new environment or continent would have to be a lot faster than the time it takes for a new species to be formed.

A slowdown in speciation, however, does not imply an association with niche divergence (Moen and Morlon, 2014). The increase in geographic ranges of lineages after their invasion of new areas (new forested areas and/or the NW) could also be subject to successive vicariant events, and speciation due to isolation may increase speciation rates (Moen and Morlon, 2014). In fact, lineages with larger geographic distributions are more likely to be dissected by geographic barriers (Rosenzweig, 1995) and at least in theory, geographic isolation can lead to an increase in diversification rates (Pigot et al., 2010). The successive division of the larger ranges would produce many new species with smaller ranges, which in turn would be less likely to be subdivided by potential barriers causing speciation rates to decline (Moen and Morlon, 2014). That is, bursts of speciation with subsequent slowdowns could also be caused by purely geographic factors.

CONCLUSIONS

Our results would suggest that vipers are still growing in species number because speciation rates are higher relative to extinction rates. However, extinction rates were frequently estimated to be very low and seem roughly constant through time. Extinction rates estimated from molecular phylogenies are typically very low and extreme caution should be taken when interpreting those estimates (Quental and Marshall, 2010; Rabosky, 2010). Moreover, our results could not recover

with confidence the diversification scenario of the background regime, mostly comprised by the subfamily Viperinae. However, irrespective of what is the true extinction rates and diversity trajectory for vipers, our results undoubtedly suggest that after the first split event among crotalines, pit vipers experienced a very high speciation rate that was followed by a considerable decline. The evolution of the loreal pits alone does not seem to explain this explosive speciation rates. We suggest that the evolution of the loreal pits coupled with the dispersion to new niches emerged as a consequence of climatic and geological changes in Asia and the invasion of the NW may have spurred the increase in speciation rates. Moreover, the drop in speciation rates in this successful group of snakes may be a consequence of the decrease in opportunities to speciate after their rapid dispersion through the several available environments. Our results evoke a series of interesting specific hypothesis, but directly testing those ideas is beyond the scope of the present chapter. Additionally, further investigation is needed to fully understand the role of adaptation to spatially distributed environments and/or vicariance events during the radiation of pit vipers, which will certainly improve our understanding of the different potential regulators of biodiversity in vipers.

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SUPPLEMENTARY TABLE

Table S1. Clade specific sampling probabilities used in BAMM

speciesName	cladeName	samplingFraction
Agkistrodon_bilineatus	Agkistrodon	1
Agkistrodon_contortrix	Agkistrodon	1
Agkistrodon_howardgloydi	Agkistrodon	1
Agkistrodon_piscivorus	Agkistrodon	1
Agkistrodon_russeolus	Agkistrodon	1
Agkistrodon_taylori	Agkistrodon	1
Atheris_barbouri	Viperinae	0.73
Atheris_ceratophora	Viperinae	0.73
Atheris_chlorechis	Viperinae	0.73
Atheris_desaixi	Viperinae	0.73
Atheris_hispida	Viperinae	0.73
Atheris_matildae	Viperinae	0.73
Atheris_nitschei	Viperinae	0.73
Atheris_squamigera	Viperinae	0.73
Atropoides_indomitus	Atro_cerro_porth	0.95
Atropoides_mexicanus	Atro_cerro_porth	0.95
Atropoides_nummifer	Atro_cerro_porth	0.95
Atropoides_occiduus	Atro_cerro_porth	0.95
Atropoides_olmec	Atro_cerro_porth	0.95
Atropoides_picadoi	Atro_cerro_porth	0.95
Azemiops_feae	Azemiopinae	0.5
Bitis_arietans	Viperinae	0.73
Bitis_armata	Viperinae	0.73
Bitis_atropos	Viperinae	0.73
Bitis_caudalis	Viperinae	0.73
Bitis_cornuta	Viperinae	0.73
Bitis_gabonica	Viperinae	0.73
Bitis_nasicornis	Viperinae	0.73
Bitis_parviocula	Viperinae	0.73
Bitis_peringueyi	Viperinae	0.73
Bitis_rhinoceros	Viperinae	0.73
Bitis_rubida	Viperinae	0.73
Bitis_schneideri	Viperinae	0.73
Bitis_worthingtoni	Viperinae	0.73
Bitis_xeropaga	Viperinae	0.73
Bothriechis_aurifer	Bothriechis	1
Bothriechis_bicolor	Bothriechis	1
Bothriechis_guifarroi	Bothriechis	1
Bothriechis_lateralis	Bothriechis	1
Bothriechis_marchi	Bothriechis	1
Bothriechis_nigroviridis	Bothriechis	1

Bothriechis_rowleyi	Bothriechis	1
Bothriechis_schlegelii	Bothriechis	1
Bothriechis_superciliaris	Bothriechis	1
Bothriechis_thalassinus	Bothriechis	1
Bothrocophias_campbelli	Bothrocophias_Bothrops	0.75
Bothrocophias_hyoprora	Bothrocophias_Bothrops	0.75
Bothrocophias_microphthalmus	Bothrocophias_Bothrops	0.75
Bothrops_alcatraz	Bothrocophias_Bothrops	0.75
Bothrops_alternatus	Bothrocophias_Bothrops	0.75
Bothrops_ammodytoides	Bothrocophias_Bothrops	0.75
Bothrops_asper	Bothrocophias_Bothrops	0.75
Bothrops_atrox	Bothrocophias_Bothrops	0.75
Bothrops_barnettii	Bothrocophias_Bothrops	0.75
Bothrops_bilineatus	Bothrocophias_Bothrops	0.75
Bothrops_brazili	Bothrocophias_Bothrops	0.75
Bothrops_caribbaeus	Bothrocophias_Bothrops	0.75
Bothrops_chloromelas	Bothrocophias_Bothrops	0.75
Bothrops_colombiensis	Bothrocophias_Bothrops	0.75
Bothrops_cotiara	Bothrocophias_Bothrops	0.75
Bothrops_diporus	Bothrocophias_Bothrops	0.75
Bothrops_erythromelas	Bothrocophias_Bothrops	0.75
Bothrops_fonsecai	Bothrocophias_Bothrops	0.75
Bothrops_insularis	Bothrocophias_Bothrops	0.75
Bothrops_isabelae	Bothrocophias_Bothrops	0.75
Bothrops_itapetiningae	Bothrocophias_Bothrops	0.75
Bothrops_jararaca	Bothrocophias_Bothrops	0.75
Bothrops_jararacussu	Bothrocophias_Bothrops	0.75
Bothrops_lanceolatus	Bothrocophias_Bothrops	0.75
Bothrops_leucurus	Bothrocophias_Bothrops	0.75
Bothrops_lojanus	Bothrocophias_Bothrops	0.75
Bothrops_lutzi	Bothrocophias_Bothrops	0.75
Bothrops_marajoensis	Bothrocophias_Bothrops	0.75
Bothrops_marmoratus	Bothrocophias_Bothrops	0.75
Bothrops_mattogrossensis	Bothrocophias_Bothrops	0.75
Bothrops_moojeni	Bothrocophias_Bothrops	0.75
Bothrops_neuwiedi	Bothrocophias_Bothrops	0.75
Bothrops_osbornei	Bothrocophias_Bothrops	0.75
Bothrops_pauloensis	Bothrocophias_Bothrops	0.75
Bothrops_pictus	Bothrocophias_Bothrops	0.75
Bothrops_pubescens	Bothrocophias_Bothrops	0.75
Bothrops_pulchra	Bothrocophias_Bothrops	0.75
Bothrops_punctatus	Bothrocophias_Bothrops	0.75
Bothrops_taeiniata	Bothrocophias_Bothrops	0.75
Calloselasma_rhodostoma	Calloselasma	1
Causus_defilippii	Viperinae	0.73
Causus_lichtensteini	Viperinae	0.73
Causus_resimus	Viperinae	0.73
Causus_rhombeatus	Viperinae	0.73

Cerastes_cerastes	Viperinae	0.73
Cerastes_gasperettii	Viperinae	0.73
Cerastes_vipera	Viperinae	0.73
Cerrophidion_barbouri	Atro_cerro_porth	0.95
Cerrophidion_godmani	Atro_cerro_porth	0.95
Cerrophidion_petlalcalensis	Atro_cerro_porth	0.95
Cerrophidion_sasai	Atro_cerro_porth	0.95
Cerrophidion_tzotzilorum	Atro_cerro_porth	0.95
Crotalus_adamanteus	crot_sist	0.90
Crotalus_aquilus	crot_sist	0.90
Crotalus_atrox	crot_sist	0.90
Crotalus_basiliscus	crot_sist	0.90
Crotalus_catalinensis	crot_sist	0.90
Crotalus_cerastes	crot_sist	0.90
Crotalus_cerberus	crot_sist	0.90
Crotalus_culminatus	crot_sist	0.90
Crotalus_durissus	crot_sist	0.90
Crotalus_enyo	crot_sist	0.90
Crotalus_ericsmithi	crot_sist	0.90
Crotalus_horridus	crot_sist	0.90
Crotalus_intermedius	crot_sist	0.90
Crotalus_lannomi	crot_sist	0.90
Crotalus_lepidus	crot_sist	0.90
Crotalus_mitchellii	crot_sist	0.90
Crotalus_molossus	crot_sist	0.90
Crotalus_oreganus	crot_sist	0.90
Crotalus_ornatus	crot_sist	0.90
Crotalus_polystictus	crot_sist	0.90
Crotalus_pricei	crot_sist	0.90
Crotalus_pusillus	crot_sist	0.90
Crotalus_ravus	crot_sist	0.90
Crotalus_ruber	crot_sist	0.90
Crotalus_scutulatus	crot_sist	0.90
Crotalus_simus	crot_sist	0.90
Crotalus_stejnegeri	crot_sist	0.90
Crotalus_tancitarensis	crot_sist	0.90
Crotalus_tigris	crot_sist	0.90
Crotalus_tortugensis	crot_sist	0.90
Crotalus_totonacus	crot_sist	0.90
Crotalus_transversus	crot_sist	0.90
Crotalus_triseriatus	crot_sist	0.90
Crotalus_tzcaban	crot_sist	0.90
Crotalus_viridis	crot_sist	0.90
Crotalus_willardi	crot_sist	0.90
Daboia_deserti	Viperinae	0.73
Daboia_mauritanica	Viperinae	0.73
Daboia_palaestinae	Viperinae	0.73
Daboia_russellii	Viperinae	0.73

Daboia_siamensis	Viperinae	0.73
Deinagkistrodon_acutus	Deinagkistrodon	1
Echis_borkini	Viperinae	0.73
Echis_carinatus	Viperinae	0.73
Echis_coloratus	Viperinae	0.73
Echis_jogeri	Viperinae	0.73
Echis_khosatzkii	Viperinae	0.73
Echis_leucogaster	Viperinae	0.73
Echis_ocellatus	Viperinae	0.73
Echis_omanensis	Viperinae	0.73
Echis_pyramidum	Viperinae	0.73
Eristicophis_macmahoni	Viperinae	0.73
Garthius_chaseni	Garthius	1
Gloydius_blomhoffi	Gloydius	0.8
Gloydius_brevicaudus	Gloydius	0.8
Gloydius_halys	Gloydius	0.8
Gloydius_intermedius	Gloydius	0.8
Gloydius_liupanensis	Gloydius	0.8
Gloydius_saxatilis	Gloydius	0.8
Gloydius_shedaoensis	Gloydius	0.8
Gloydius_strauchi	Gloydius	0.8
Gloydius_tsushimaensis	Gloydius	0.8
Gloydius_ussuriensis	Gloydius	0.8
Hypnale_hypnale	Hypnale	1
Hypnale_nepa	Hypnale	1
Hypnale_zara	Hypnale	1
Lachesis_acrochorda	Lachesis	1
Lachesis_melanocephala	Lachesis	1
Lachesis_muta	Lachesis	1
Lachesis_stenophrys	Lachesis	1
Macrovipera_lebetina	Viperinae	0.73
Macrovipera_schweizeri	Viperinae	0.73
Mixcoatlus_browni	Oph_mix	0.75
Montivipera_albizona	Viperinae	0.73
Montivipera_bornmuelleri	Viperinae	0.73
Montivipera_latifii	Viperinae	0.73
Montivipera_raddei	Viperinae	0.73
Montivipera_wagneri	Viperinae	0.73
Montivipera_xanthina	Viperinae	0.73
Ophryacus_melanurus	Oph_mix	0.75
Ophryacus_undulatus	Oph_mix	0.75
Ovophis_monticola	Ovophis	0.6
Ovophis_okinavensis	Gloydius	0.8
Ovophis_tonkinensis	Ovophis	0.6
Ovophis_zayuensis	Ovophis	0.6
Porthidium_arcosae	Atro_cerro_porth	0.95
Porthidium_dunni	Atro_cerro_porth	0.95
Porthidium_hesperere	Atro_cerro_porth	0.95

Porthidium_lansbergii	Atro_cerro_porth	0.95
Porthidium_nasutum	Atro_cerro_porth	0.95
Porthidium_ophryomegas	Atro_cerro_porth	0.95
Porthidium_porrasi	Atro_cerro_porth	0.95
Porthidium_yucatanicum	Atro_cerro_porth	0.95
Proatheris_superciliaris	Viperinae	0.73
Protobothrops_cornutus	Protobothrops	0.85
Protobothrops_dabieshanensis	Protobothrops	0.85
Protobothrops_elegans	Protobothrops	0.85
Protobothrops_flavoviridis	Protobothrops	0.85
Protobothrops_jerdonii	Protobothrops	0.85
Protobothrops_kaulbacki	Protobothrops	0.85
Protobothrops_mangshanensis	Protobothrops	0.85
Protobothrops_maolanensis	Protobothrops	0.85
Protobothrops_mucrosquamatus	Protobothrops	0.85
Protobothrops_sieversorum	Protobothrops	0.85
Protobothrops_tokarensis	Protobothrops	0.85
Protobothrops_xiangchengensis	Protobothrops	0.85
Pseudocerastes_persicus	Viperinae	0.73
Pseudocerastes_urarachnoides	Viperinae	0.73
Pseudocerastesfieldi	Viperinae	0.73
Sistrurus_catenatus	crot_sist	0.90
Sistrurus_miliarius	crot_sist	0.90
Trimeresurus_albolabris	Trimeresurus	0.75
Trimeresurus_andersonii	Trimeresurus	0.75
Trimeresurus_barati	Trimeresurus	0.75
Trimeresurus_borneensis	Trimeresurus	0.75
Trimeresurus_buniana	Trimeresurus	0.75
Trimeresurus_cantori	Trimeresurus	0.75
Trimeresurus_erythrurus	Trimeresurus	0.75
Trimeresurus_fasciatus	Trimeresurus	0.75
Trimeresurus_flavomaculatus	Trimeresurus	0.75
Trimeresurus_fucatus	Trimeresurus	0.75
Trimeresurus_gracilis	Gloydius	0.8
Trimeresurus_gramineus	Trimeresurus	0.75
Trimeresurus_gumprechtii	Trimeresurus	0.75
Trimeresurus_hageni	Trimeresurus	0.75
Trimeresurus_insularis	Trimeresurus	0.75
Trimeresurus_kanburiensis	Trimeresurus	0.75
Trimeresurus_macrops	Trimeresurus	0.75
Trimeresurus_malabaricus	Trimeresurus	0.75
Trimeresurus_malcolmi	Trimeresurus	0.75
Trimeresurus_mcgregori	Trimeresurus	0.75
Trimeresurus_medoensis	Trimeresurus	0.75
Trimeresurus_nebularis	Trimeresurus	0.75
Trimeresurus_popeiorum	Trimeresurus	0.75
Trimeresurus_puniceus	Trimeresurus	0.75
Trimeresurus_purpureomaculatus	Trimeresurus	0.75

Trimeresurus_sabahi	Trimeresurus	0.75
Trimeresurus_schultzei	Trimeresurus	0.75
Trimeresurus_septentrionalis	Trimeresurus	0.75
Trimeresurus_sichuanensis	Trimeresurus	0.75
Trimeresurus_stejnegeri	Trimeresurus	0.75
Trimeresurus_sumatranus	Trimeresurus	0.75
Trimeresurus_tibetanus	Trimeresurus	0.75
Trimeresurus_trigonocephalus	Trimeresurus	0.75
Trimeresurus_truongsonensis	Trimeresurus	0.75
Trimeresurus_venustus	Trimeresurus	0.75
Trimeresurus_vogeli	Trimeresurus	0.75
Trimeresurus_wiroti	Trimeresurus	0.75
Trimeresurus_yunnanensis	Trimeresurus	0.75
Tropidolaemus_subannulatus	Tropidolaemus	0.4
Tropidolaemus_wagleri	Tropidolaemus	0.4
Vipera_ammodytes	Viperinae	0.73
Vipera_anatolica	Viperinae	0.73
Vipera_aspis	Viperinae	0.73
Vipera_barani	Viperinae	0.73
Vipera_berus	Viperinae	0.73
Vipera_dinniki	Viperinae	0.73
Vipera_erivanensis	Viperinae	0.73
Vipera_kaznakovi	Viperinae	0.73
Vipera_latastei	Viperinae	0.73
Vipera_lotievi	Viperinae	0.73
Vipera_nikolskii	Viperinae	0.73
Vipera_orlovi	Viperinae	0.73
Vipera_renardi	Viperinae	0.73
Vipera_seoanei	Viperinae	0.73
Vipera_transcaucasiana	Viperinae	0.73
Vipera_ursinii	Viperinae	0.73

ARBOREAL HABITATS CONSTRAIN MORPHOLOGICAL BUT NOT SPECIES DIVERSIFICATION IN VIPERS

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Abstract. A shift into novel environmental conditions faced by a lineage is frequently suggested to be associated with an increase in ecological opportunities, which consequently will spur both species and morphological diversification. Although most adaptive radiation studies have relied on this tenet, it is indeed feasible that lineages might sometimes face the opposite situation, where changing environmental conditions will prevent lineages to diversify. Here we investigate the potential association between habitat and diversification regime in vipers. We hypothesize that arboreal habitats impose stronger limitations in morphological evolution consequently decreasing the speciation opportunities among arboreal compared to terrestrial vipers. Using recently developed trait-evolution models and a Bayesian species diversification method while incorporating phylogenetic and habitat reconstruction uncertainty, we show that although arboreal habitats strongly limit morphological evolution in vipers, speciation rates remain similar between habitats. We suggest two distinct scenarios underlying our results: (1) speciation could be independent of morphological evolution in vipers; or (2) geographic isolation would be an important mechanism underlying species diversification in arboreal lineages offsetting decreases in speciation opportunities potentially related to the selective pressures imposed by the arboreal environment.

INTRODUCTION

The exceptional diversity found in many groups of organisms has long been associated with factors that promote an increase in the ecological opportunities faced by a lineage (Schluter, 2000, Simpson, 1949, 1953, Yoder et al., 2010). In its simplest form, ecological opportunity can be defined as “a wealth of evolutionary accessible resources little used by competing taxa” (Schluter 2000). The invasion of new habitats, the emergence of a key innovation, the extinction of a predator or competitor are all considered as possible triggers of a rapid increase in ecological opportunities (Simpson, 1949, Yoder et al., 2010). The emergence of new ecological conditions can therefore trigger divergent selection, leading populations to adapt into the several new niches and ultimately

promote speciation (Lapiedra et al., 2013, Nosil, 2012, Schluter, 2001). Given that the morphology of an organism reflects its ecology (Wainwright and Reilly, 1994), the emergence of ecological opportunities is, under this scenario, expected to generate an increase in both species and morphological diversification of a lineage experiencing such opportunities.

In fact ecological opportunity is frequently suggested as the main cause of adaptive radiations (Glor, 2010, Yoder et al., 2010), which are usually characterized by a higher rate of speciation and morphological evolution. Classic adaptive radiations such as Darwin's Finches (e.g. Grant and Grant, 2008), Caribbean *Anolis* lizards (Losos, 2009), and coral reef fishes (Alfaro et al., 2007, Price et al., 2011), all comprise examples of such scenario, where the invasion of previously unoccupied regions (e.g. islands) rapidly generated new morpho-types, presumably reflecting niche partitioning, (e.g. diet, microhabitat), ultimately leading to the rapid formation of new species (Schluter, 2000). Although adaptive radiations constitute remarkable examples of how shifts in ecological conditions led to a great increase in both species and morphological diversification of a lineage, adaptive radiations are considered an exception in nature (see Figure 15.4 in Losos and Mahler, 2010).

Several processes underlying the diversification patterns of different groups of organisms might not end up generating an impressive number of species or morphological variation (e.g. Kozak et al., 2006). In fact, it is easy to imagine a scenario where the invasion of a new habitat would constrain, instead of spur, both species and morphological diversification. Certain habitats might impose strong selective pressures, and evolutionary forces reducing trait variance like stabilizing selection (see Hansen, 1997) might allow only lineages with a specific combination of phenotypes to persist favoring evolutionary stasis and niche conservatism (Lapiedra et al., 2013). In an adaptive landscape framework (i.e. mapping fitness onto trait combinations, Laughlin and Messier, 2015), this specific combination of traits would represent a sharp adaptive peak comprising the higher fitness values in those habitats. Here, divergent selection would have a limited role in speciation, and diversification would be the result of mechanisms that generate species without necessarily generating significant morphological variation (e.g. geographic isolation followed by drift, Coyne and Orr, 2004, Nosil, 2012). In contrast, a scenario conducive to an adaptive radiation (i.e. those comprising ecological opportunities) might be better represented by multiple adaptive peaks favoring lineages with different morphological combinations to persist. In this case, divergent selection and speciation rate are assumed to be intertwined.

Snakes comprise a cosmopolitan group known to have radiated into a diverse array of habitats facing different ecological conditions during their evolutionary history (Greene, 1997, Uetz and Hosek, 2014). Although diversification rates vary considerably among snakes (e.g. Pyron and Burbrink, 2012), we do not know if this variation results from the interaction with different

environments. Of particular interest is the invasion of the arboreal habitat, which took place multiple times during the evolutionary history of snakes (Lillywhite and Henderson, 1993). Although many species are terrestrial, some are considered to be fully arboreal (see Greene, 1997). Given its inherent limitations of being elongated and limbless organisms, arboreal snakes are known to possess a suite of behavioral, morphological and physiological adaptations allowing them to deal with the challenges of living in a structurally complex environment such as the arboreal habitats (e.g. Martins et al., 2001, Pizzatto et al., 2007, Sheehy et al., 2016). Moreover, distantly related arboreal lineages tend to show similar phenotypes and arboreal life style in snakes is usually considered to result in “morphological syndromes” (Greene, 1997, Lillywhite and Henderson, 1993). Thus, arboreal snake lineages comprise a striking example of evolutionary convergence (*sensu* Losos, 2011). In contrast, terrestrial snakes are known to have a much more generalized morphology (Greene, 1997). The similar phenotype among arboreal snakes and a much more varied one among terrestrial ones suggest that arboreal and terrestrial habitats might represent distinct adaptive landscape scenarios. In this sense, the invasion of the arboreal habitat by a terrestrial lineage of snakes might be accompanied by strong constraints in their morphology which could potentially affect species diversification dynamics.

Vipers (family Viperidae) are ideal models to explore the effect that habitat might impose in morphological and species diversification. Currently there are 329 viper species (Uetz and Hosek, 2014) and arboreality independently evolved multiple times, perhaps as many as 13 times in eight different genera (Figure 1). Additionally, arboreal lineages are not restricted to a single continent and instead are distributed in Central and South America, Africa and Asia (Campbell and Lamar, 2004, Orlov et al., 2002, Phelps, 2010, Figure 1). Differently from other snake groups, non-arboreal vipers do not comprise fossorial or aquatic species and are for the most part strictly terrestrial (Campbell and Lamar, 2004, Phelps, 2010 but see Material and Methods). Moreover, Hendry et al. (2014) recently suggested that changes in habitat use might be associated with changes in speciation rates in New World vipers. Thus, vipers constitute an ideal group to investigate how the invasion of a new habitat, and its new selective pressures, might affect diversification of both morphological and species diversification.

In the present study, our goal was to investigate the potential association between habitat and diversification regimes in vipers. Specifically, we explored if the invasion of arboreal habitats led to changes in morphological and species diversification. We first investigated the tempo and mode of morphological evolution associated with the different habitats. Given the potential limitations that arboreal life might impose on a snake (Lillywhite and Henderson, 1997, Sheehy et al., 2016) we hence expected arboreal vipers to have a more constrained morphology when compared to terrestrial ones. We also expected that lineages switching from habitats comprising

multiple adaptive peaks (terrestrial) to others characterized by only one main peak (arboreal) would be associated with smaller speciation rates given the decrease in speciation opportunities related to the phenotypic constraint that might be associated with these new environments.

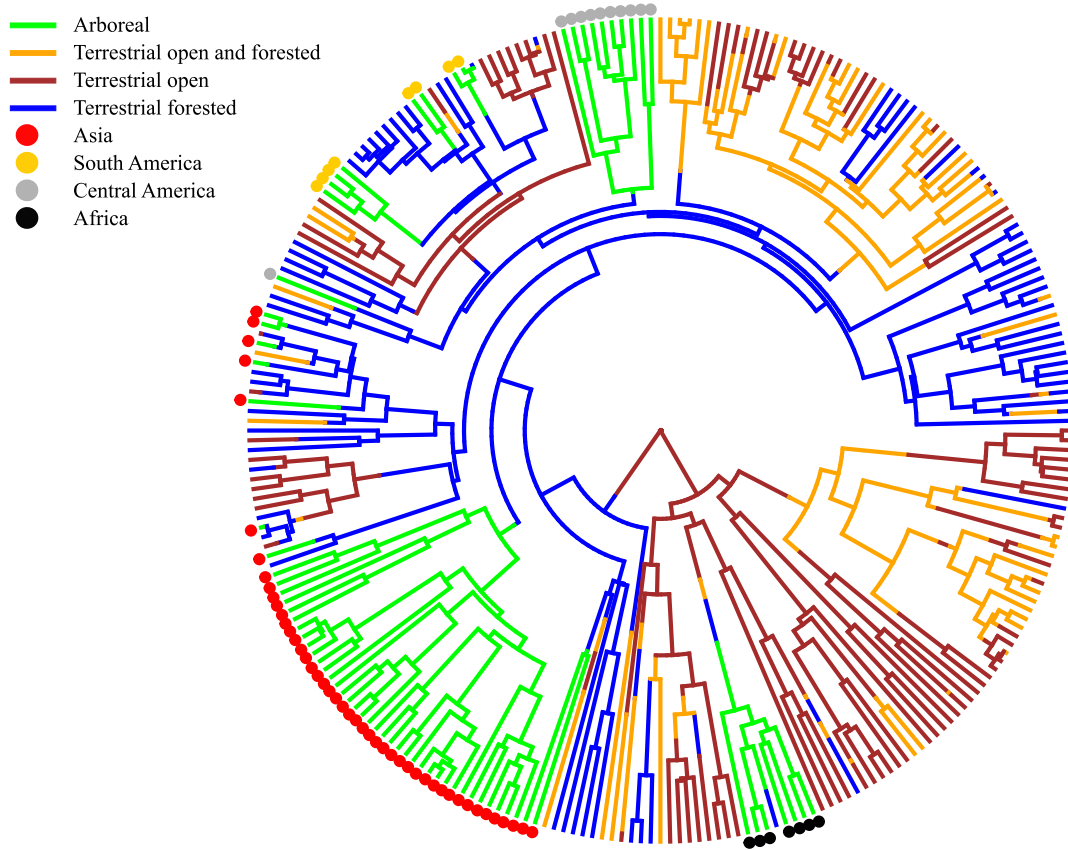


Figure 1. Example of a stochastic character mapping (simmap) on a viper phylogeny from the 1000 analyzed. Colors on branches denote the evolution of different habitats in vipers. Circles at tips indicate the predominant geographical occurrence of arboreal lineages.

MATERIAL AND METHODS

Phylogenies

In the present study, we used 100 comprehensive time-calibrated phylogenetic trees of extant vipers randomly obtained from the posterior distribution of BEAST analyses performed elsewhere (Chapter I). Phylogenetic reconstructions comprised 264 viper species and combined newly collected sequences and others previously available in GenBank and BOLD databases obtained for six mitochondrial (12S, 16S, cytb, nad4, nad2, cox1) and five nuclear (bndf, c-mos, jun, nt3, rag1) genes (for detailed information regarding phylogenetic reconstruction see Chapter I).

Thus, the phylogenetic scenario used here represents the most complete phylogenetic dataset for vipers to date, and includes almost 80% of all described species and all except one genus of vipers (Uetz and Hosek, 2014).

Habitat information

We characterized the habitat of each viper species present in the phylogeny using published literature and observations of researchers with recognized expertise in this snake group (Table S1). We categorized vipers as being: 1- arboreal; 2 - terrestrial restricted to open habitats; 3 - terrestrial restricted to forests; 4 - terrestrial occurring in both open and forested habitats. We chose to subdivide terrestrial vipers into three categories because while arboreal vipers are predominantly restricted to forests, terrestrial vipers occur not only in forested but also in open habitats (e.g. Campbell and Lamar, 2004, Phelps, 2010) and it is reasonable to expect that such different habitats (close forested vs open habitat) might impose very different selective regimes. By using such subdivision we are also being more conservative from the perspective of our proposed hypothesis given that using a single terrestrial category including both species restricted to forested and open habitats would more easily render evidence in favor of a more restricted morphology and lower speciation rates in arboreal species.

Some viper species occurring in arboreal habitats can also be frequently found on the ground and thus could be considered as “semi-arboreal”. However, arboreal and semi-arboreal vipers seem to be closely related in their morphospace (Figure S1, 11 semi-arboreal species shown) suggesting they share the same morphological syndromes, and thus, that potential selective pressures imposed by the arboreal environment might also affect semi-arboreal lineages. Moreover, semi-arboreal vipers comprise only 17 species in the phylogeny and would represent a very species poor habitat category, which would negatively affect the performance of methodological approaches to study the phenotypic and species diversification between different ecological regimes (e.g. Davis et al., 2013). Therefore, we decided to include “semi-arboreal” vipers into the “arboreal” category. Lastly, *Agkistrodon piscivorus* is sometimes considered as “semi-aquatic” (e.g. Vincent et al., 2005), but given that it is only one species we chose to include it in the terrestrial category.

Ancestral state reconstructions

We used stochastic character mapping (Simmmap, R package Phytools, Revell 2012) to reconstruct the evolutionary history of habitat in vipers assigning the habitat categories to the internal branches of phylogenetic trees. We first fitted three different continuous-time Markov

models of trait evolution (“equal rates”, “symmetrical rates” or “all rates different”, Harmon et al., 2008) across our 100 posterior randomly chosen trees and species habitat information to evaluate which one best explain our dataset (Table S2). We used the function “make.simmap” to estimate the habitat evolutionary history across our 100 trees. For each one of the posterior trees we set the transition rates matrix (Q) to be empirically estimated: first the previously chosen continuous-time Markov model (“equal rates”, “symmetrical rates” or “all rates different”) is fitted to our datasets (phylogenetic trees and habitat information), a single value of Q is estimated and used to simulate character histories (Revell, 2012). In each of the 100 posterior trees, we simulated 10 maps to incorporate uncertainty in the ancestral state reconstruction. However, for seven out of a 100 trees the “make.simmap” function could not estimate character histories under the best Markov model chosen (“all rates different”). Excluding these seven trees could remove part of the phylogenetic uncertainty incorporated in our analyses, and thus, in those cases we chose to use a simpler Markov model (“symmetrical rates”, Table S2) instead. In total we generated 1000 SIMMAP trees to be used in trait evolution analyses incorporating phylogenetic and ancestral estimates uncertainty.

Morphological data

We used snout-vent length (distance from the tip of the snout to the cloaca) as a proxy for body size, which is considered to be a relevant trait to many physiological and ecological adaptations across different animals (Bonner, 2011, Feldman et al., 2015). We chose to use body size instead of body mass because the later cannot be measured with accuracy in preserved specimens. In snakes, body size is frequently related to different life-history aspects (see Greene, 1997, Lillywhite and Henderson, 1993), including thermoregulation, fecundity, and diet. Apart from correlating allometrically with numerous other morphological traits body size explains a greater proportion of trait variance in vipers (83%, present study, results not shown, see França et al. 2008 for another example with snakes). Additionally, previous studies using independent contrasts (Felsenstein, 1985) either found an association between arboreality and a decrease in body size (e.g. Sheehy et al., 2016) or no association between this trait and the evolution of arboreal habits in snakes (e.g. Martins et al., 2001, Pizzatto et al., 2007). Hendry et al. (2014) suggested that arboreality is related to a decrease in body size in New World vipers but using total length as a proxy for body size. However, none of these studies investigated if the evolution of arboreality could impose constraining evolutionary force on morphological evolution.

Body size measurements were taken from several preserved specimens of vipers deposited in the National Museum of Rio de Janeiro (MNRJ), Butantan Institute (IBSP), Museum of Zoology of the University of São Paulo (MZUSP), Natural History Museum of London (BMNH), National

Museum of Natural History of Paris (MNHNP), Raffles Museum of Biodiversity Research (ZRC), University of Texas at Arlington (UTA), California Academy of Sciences (CAS), American Museum of Natural History (AMNH), Field Museum of Natural History (FMNH), and Smithsonian National Museum of Natural History (USNM) (Appendix I). We were able to get measurements for 236 viper species and to complement our morphological database, we also added literature records for 13 species not represented in the scientific collections visited (Table S1). In total we were able to get body size measurements for 94% of the species included in the phylogeny representing 75% of all viper species. Table S1 provides a summary of body size data gathered for the present study.

In the following analyses we used mostly measurements of adult males. However, for 16 out of 249 species we only had access to morphological measurements for adult females (Table S1). We chose to include female body size data instead of excluding these 16 species from our analyses and are confident that including these data have not biased our results (see Appendix II). We attempted to measure several individuals for each species but because snakes and especially vipers are frequently damaged when killed, we were not able to include measurements for several preserved specimens. Thus, the number of specimens varies among species. When more than one individual was measured, we used the average body size of each species. All measurements were log transformed (base 10) prior to analyses.

Models of trait evolution

To explore if the invasion of arboreal habitats is associated with changes in the morphological diversification of vipers we used different models of character evolution implemented in a model-fitting framework, namely Brownian Motion “BM” and Ornstein-Uhlenbeck “OU” models (Beaulieu et al., 2012, Price and Hopkins 2015). The OU process is ideal to model changes in selection regimes, such as differences in the phenotypic optima (Θ), phenotypic rate (σ^2 , also known as stochastic motion parameter) or strength of selection (α) between habitats (Ho and Ané, 2014). The phenotypic optima describes a hypothetical phenotype toward which populations are evolving when under an OU process; the phenotypic rate (stochastic motion parameter) measures the intensity of stochastic fluctuations in the evolutionary process, in other words, the rate of stochastic evolution away from the phenotypic optima; the strength of selection indicates the attraction toward the phenotypic optima measuring how fast a trait evolves to the associated phenotypic optima (Butler and King, 2004, Beaulieu et al., 2012, Hansen, 1997). When $\alpha = 0$ constraint to stay near the optima disappears and the OU process reduces to Brownian motion (see Butler and King, 2004, Beaulieu et al., 2012, Ho and Ané, 2014). We expect that the phenotypic rate will be smaller and/or the strength of selection will be larger in arboreal habitats

compared to terrestrial ones suggesting that arboreal vipers might be evolving under a distinct and possibly more constrained selection regime.

We fitted seven different models to the body size of vipers and simmap trees using the R package OUwie (Beaulieu et al., 2012), after pruning species according to body size data. The models used were: (1) BM1, a single rate Brownian motion model; (2) OU1, a single optimum model; both (1) and (2) suggest that phenotype is evolving independently of habitat; (3) BMS, a multi-rate Brownian motion model that allows phenotypic rate to differ among habitats; (4) OUM, assumes distinct phenotypic optima for each habitat while having a single phenotypic rate and strength of selection for all habitats; (5) OUMV, allows a different phenotypic rate related to different phenotypic optima for each habitat; (6) OUMA, allows a different strength of selection and a different phenotypic optima for each habitat; (7) OUMVA, allows both phenotypic rate and strength of selection, as well as different phenotypic optima for each habitat.

After performing OUwie analyses, we followed Price and Hopkins (2015) in first checking if the eigenvalues of the Hessian matrix were positive, which indicates that parameters were reliably estimated (Beaulieu et al., 2012, Price and Hopkins, 2015). We also checked if the phenotypic optima estimated were biologically feasible. We considered as biologically feasible optimum values between one and five (log₁₀) (values within or close to the interval of the empirical log₁₀ values for body size, see Figure 2). Using this interval we excluded analyses returning very unrealistic optima (optimum values equal or higher than five would represent lineages with body size evolving to 100 meters or more). When we found negative eigenvalues and unreliable phenotypic optima, we excluded the results for the corresponding model and tree combination. We calculated the Akaike criterion for small sample size (AICc) (Burnham and Anderson, 2002) followed by the Δ AICc to investigate which model best explains body size evolution in each simmap tree. A model was considered significantly better when the Δ AICc difference between the second and the best model was higher than two. Quantifying model fitting for each simmap tree combination separately allowed us to assess the sensitivity of model selection to alternative habitat evolutionary histories and phylogenetic scenarios combinations (e.g. Collar et al., 2011).

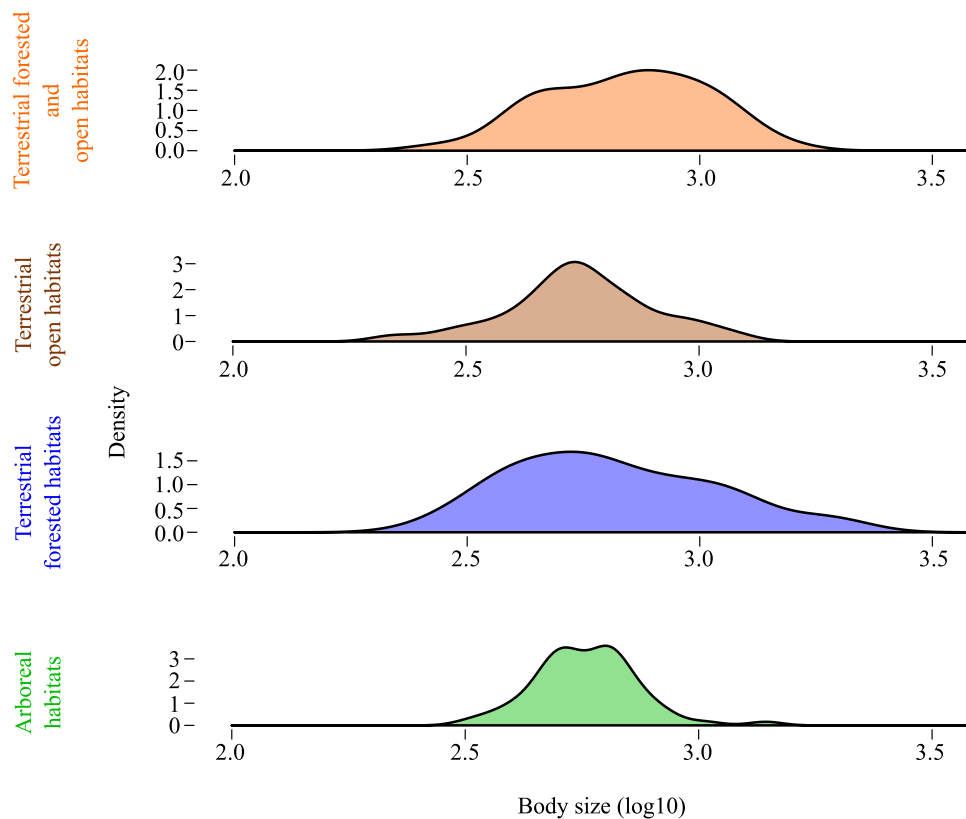


Figure 2. Body size distribution in vipers across different habitats.

Fitting all seven models for a given simmap tree was rarely possible (26% of the simmaps analyzed in OUwie were able to fit all seven models; see Results) mainly because fitting different α for each habitat (OUMA and OUMVA) frequently resulted on inaccurate estimates. These inaccurate fittings were frequently associated with negative eigenvalues (parameters not reliably estimated) or more occasionally with unrealistic estimated optima (e.g. values falling far beyond the empirical body size values). Failing to adequately fit α varying models in certain simmaps have been reported before (Moen et al., 2015, Price and Hopkins, 2015) and are likely to be related to very complex estimated character histories, which renders OUwie with not enough data to infer body size evolution under these models (e.g. limited number of tips in the phylogenetic tree for very complex histories).

Previous studies (Moen et al. 2015, Price and Hopkins 2015) have simply removed complex models from the analysis. But because estimating the α parameter would greatly improve the evaluation of our trait evolution hypothesis, we decided to keep these complex models in our analyses whenever possible. We note that different simmap tree combinations allowed different models to be fit and that at least one complex model (e.g. OUMA, OUMV or OUMVA) could always be fit. Thus we expected to be able to at least discern if OUMA, OUMV or OUMVA were best models over BM, BMS, OU1 and OUM. By not restricting ourselves to the very simple models

we expect to be able to investigate the possibility of a more constrain evolution in the phenotype of arboreal when compared to terrestrial vipers. A detailed investigation regarding the limitations of OUwie in fitting complex models is beyond the scope of this study. However, information regarding how many simmaps were able to fit each of the seven trait evolution models is available in Table 1.

Finally, we performed additional analyses to test for model adequacy. First, we randomly selected 300 simmap trees used in OUwie analyses for which the best model was one of the three more complex models (100 simmap trees for OUMVA, 100 for OUMV and 100 for OUMA). Collectively those explain the vast majority of our estimates. We then simulated 100 continuous traits for each of the 300 simmap trees using the corresponding empirical parameter values (Θ , σ^2 and α) estimated by OUwie under the best model chosen for each tree (OUMVA, OUMV or OUMA) and compared the resulting body size distribution to the empirical one. By doing this we explored if simulated values (estimated based on empirical parameters) recovered empirical body size values and hence allowing us to evaluate the adequacy of the best chosen models.

Species diversification

In order to investigate if habitat evolution has an impact on species diversification in vipers we used the Multiple State Speciation Extinction (MuSSE) model (FitzJohn, 2012) fitted to 100 randomly chosen phylogenetic trees obtained from BEAST posterior distribution (Chapter I). The MuSSE models were implemented in a Bayesian MCMC framework accounting for phylogenetic and rate value uncertainties to be naturally incorporated in the analysis. We performed MuSSE analyses in the R environment using the diversitree package (FitzJohn, 2012) and a script implementing the MCMC approach (available at <https://github.com/dsilvestro/mcmc-diversitree>), allowing all parameters to be independently estimated (no constrains). We ran MuSSE for 2,000,000 steps to achieve convergence of parameters discarding estimates from 13 trees that failed to converge.

We combined the posterior distributions of parameters from the 87 phylogenetic trees analyzed into one posterior distribution for each parameter estimated by MuSSE (speciation and extinction in each habitat and transition rates among habitats). We calculated speciation minus extinction for each iteration of the MCMC to build the posterior distribution for net diversification rates in each habitat. To explore if arboreal lineages have smaller speciation rates compared to terrestrial ones, we calculated the differences between speciation rates estimated for arboreal lineages and those estimated for the three terrestrial categories for each MCMC step. By doing this we built posterior distributions of differences and considered speciation rates of arboreal lineages to

be smaller and distinct from the remaining when values were negative and 0 was not included inside the 95% highest posterior density (HPD) calculated for each posterior difference. Although our original hypothesis relates to differences in speciation rates between arboreal and terrestrial vipers, we also present results for net diversification, extinction and transitions rates.

Recently, the performance of the state-dependent speciation extinction models (xxSSE models) has been criticized (Maddison and FitzJohn, 2014, Rabosky and Goldberg, 2015). The major concerns of xxSSE models are related to these models showing high type I error. Even though our results do not show any associations between diversification rates and habitat (see Results), our parameter estimates could still be compromised by model inadequacy (Rabosky and Goldberg, 2015). Following Price et al. (2012) and Burin et al. (*accepted*) we performed posterior predictive simulations in order to evaluate how reliable our results are. We simulated 1000 trees using the empirical number of states (four) and the median of the posterior distribution of each parameter estimated in MuSSE analyses using the trait-dependent simulation function `tree.musse` (diversitree R package, FitzJohn, 2012). We set the root states in the simulations proportionally to the empirical proportion of each state (i.e. each habitat) observed in the empirical dataset. We then calculated the proportion of states (arboreal and the three terrestrial categories) recovered in each simulated tree and compared it to the empirical distribution. By doing this we could check if the proportions of empirical states lie within the values delimited by the simulated proportions, i.e. if simulated proportions can generate realistic predictions about trait state proportions.

RESULTS

Trait evolution

We were able to successfully fit all seven trait evolution models in 261 simmaps from the 1000 analyzed (26% of all simmaps). Among these 261, 243 (93%) suggested OUMVA as the best model (ΔAICc difference between the second and the best model was higher than two). Although most simmaps were not able to fit all seven models (returning negative eigenvalues or unrealistic phenotypic optima) when taking all of the 1000 analyzed, 89% suggested OU models with multiple θ and different α and/or σ^2 parameters as best models (OUMVA: 436 simmaps, OUMV: 347 simmaps, OUMA: 111 simmaps, Table 1). Looking to the fittings of each model separately (each line in Table 1), OUMVA was the best model in 93% of the simmaps it was fitted (436 from 465). Considering the simmaps in which OUMV was fitted (983), 35% suggested OUMV as the best model (347 from 983) but 44% suggested OUMVA (434 from 983). However, it is worth mentioning that from the 983 simmaps fitting OUMV, only 465 were also able to fit OUMVA,

probably underestimating the real proportion of simmaps suggesting OUMVA as the best model (Table 1). Among the simmaps in which OUMA was fitted (493), 22% suggested it as the best model (111 from 493) whereas 49% suggested OUMVA (245 from 493). Similar to OUMV fittings, from the 493 simmaps that fitted OUMA, only 263 also fitted OUMVA, also probably underestimating the real proportion of simmaps suggesting OUMVA as the best model (Table 1). In general, our results suggest that OUMVA, the most complex model, is chosen as the best model in most of the simmaps it was fitted with OUMV and OUMA being chosen when OUMVA could not be fitted (Table 1).

Table 1. Number of OUwie fittings for each model (number at the top, total of 1000 simmaps) in trait evolution analyses, and the respective number each model was chosen as the best model (number at the bottom). Each line represents fitting and best model frequency relative to the sample of simmaps fitted for each model separately (shaded). For example, in the first line 465 simmaps analyzed by OUwie fitted OUMVA and 436 suggested it as the best model. From these 465 simmaps, 463 fitted both OUMVA and OUMV, but only 11 suggested OUMV as the best model. Best model frequency in each line does not sum the number of fittings of each reference model (in the first line example, 465) because some simmaps analyzed did not suggest a best model ($\Delta AICc$ difference between the second and best model smaller or equal to 2).

	BM1	BMS	OU1	OUM	OUMA	OUMV	OUMVA
Fittings	465	464	465	465	263	463	465
Best model	0	0	0	0	2	11	436
	BM1	BMS	OU1	OUM	OUMA	OUMV	OUMVA
Fittings	983	982	983	982	490	983	463
Best model	0	0	12	1	110	347	434
	BM1	BMS	OU1	OUM	OUMA	OUMV	OUMVA
Fittings	493	493	493	493	493	490	263
Best model	0	0	1	0	111	96	245
	BM1	BMS	OU1	OUM	OUMA	OUMV	OUMVA
Fittings	999	998	999	999	493	982	465
Best model	0	0	12	15	111	347	436
	BM1	BMS	OU1	OUM	OUMA	OUMV	OUMVA
Fittings	1000	999	1000	999	493	983	465
Best model	0	0	12	15	111	347	436
	BM1	BMS	OU1	OUM	OUMA	OUMV	OUMVA
Fittings	999	999	999	998	493	982	464
Best model	0	0	12	15	111	347	435
	BM1	BMS	OU1	OUM	OUMA	OUMV	OUMVA
Fittings	1000	999	1000	999	493	983	465
Best model	0	0	0	15	111	347	436

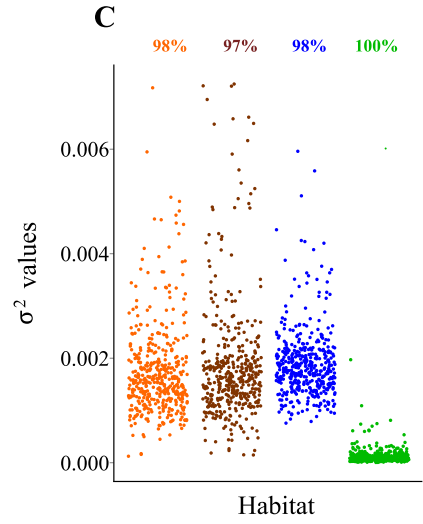
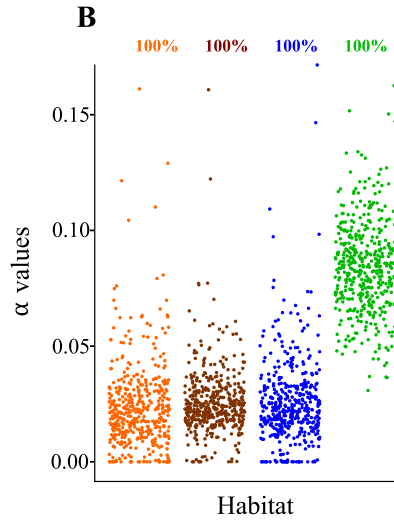
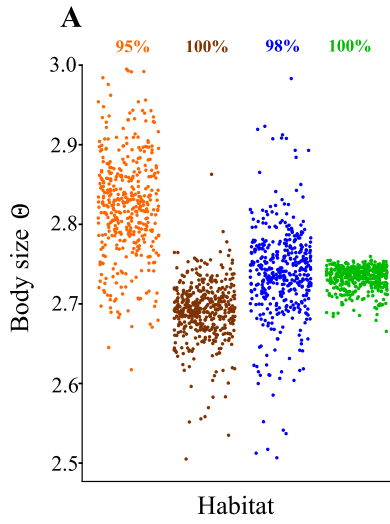
Because OUMVA could not be fitted in somewhat more than half of the simmaps, we also present parameter estimates for the other two frequently chosen models, OUMV and OUMA (Figure 3, Figure S2). When looking at the parameter estimates it is clear that a very similar

scenario is depicted independently of the best model chosen (OUMVA, OUMV or OUMA, Figure 3 and Figure S2). Parameter estimates usually show similar values among the different three best models (exceptions are the large σ^2 values in Figure 3E). Independently of the best model, body size optima (Θ) tend to show larger values for terrestrial vipers occurring in both forested and open habitats or restricted to forested habitats compared to terrestrial vipers restricted to open habitats or occurring in arboreal habitats (Figures 3A, D, G). Interestingly, body size optima estimates show less variation in arboreal compared to terrestrial habitats (Figure 3, Figure S2).

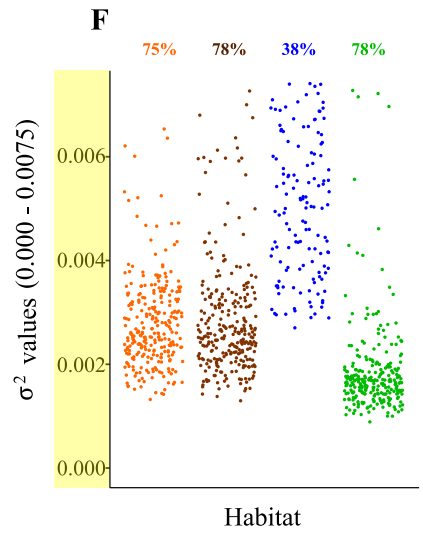
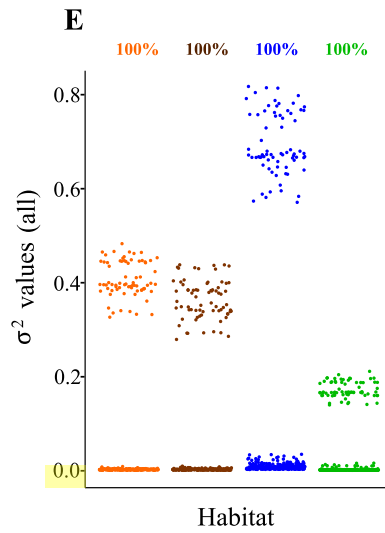
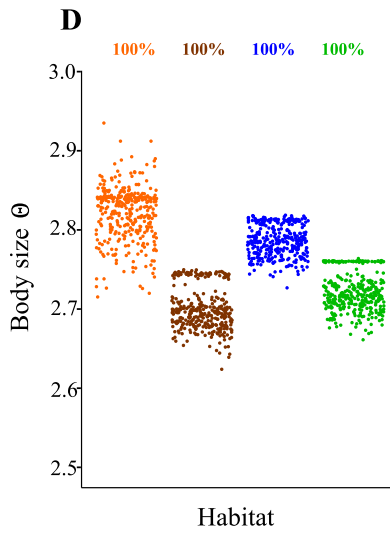
When α was allowed to vary between categories (models OUMVA and OUMA), arboreal habitats show higher α suggesting that after a lineage invades the arboreal environment, body size might evolve faster toward the optima compared to terrestrial habitats (Figures 3B, H, and Figure S2). Some analyses even suggested α values very close to zero for terrestrial categories indicating that body size could be evolving under a Brownian motion instead of an Ornstein-Uhlenbeck process in those habitats (Figure 3B, H, Figure S2). In contrast, when σ^2 was allowed to vary between habitats (models OUMVA and OUMV), arboreal habitats showed smaller σ^2 compared to terrestrial categories (Figures 3C, E, F, Figure S2), with OUMVA analyses suggesting even smaller σ^2 in arboreal habitats, frequently approaching zero (Figure 3C, Figure S2). Smaller σ^2 suggest that body size evolves under a much smaller rate of stochastic fluctuation in arboreal habitats compared to terrestrial ones. It is important to note that although σ^2 estimates under the OUMV model comprise two sets of values with different magnitudes, each one represent estimates from a distinct sample of simmaps that is the same across habitats. Thus, both σ^2 estimates still suggest smaller σ^2 in arboreal relative to terrestrial habitats.

Simulations performed using empirical parameters estimated under the most frequently chosen OUwie models (OUMA, OUMV, OUMVA) recovered body size values similar to empirical ones (Figure S3). Thus, simulations suggest that parameters were reliably estimated in our OUwie analyses.

OUMVA



OUMV



OUMA

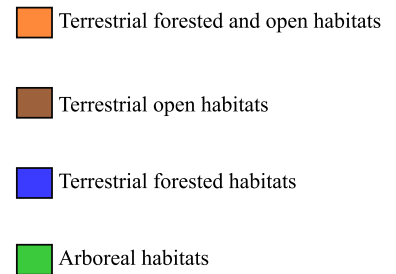
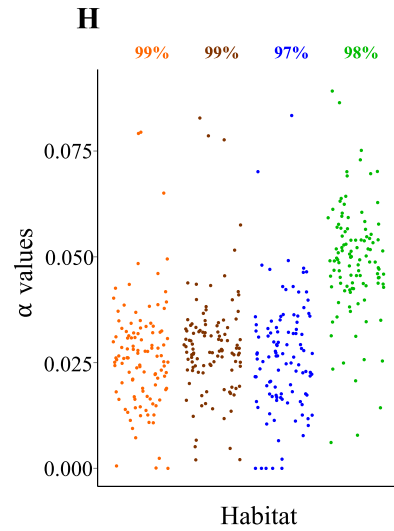
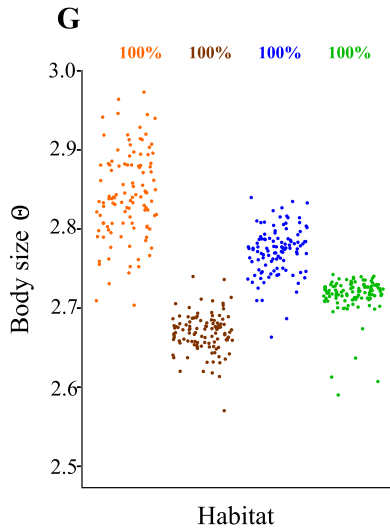


Figure 3. Parameters estimates under OUMVA (A-C), OUMV (D-F), and OUMA (G, H) when chosen as best models. Dots represent OUwie analyses performed in distinct simmaps. For better visualization we excluded extreme values and numbers on the top represent the proportion of values included. Because sigma values estimated under OUMV model have two orders of magnitude, we presented the smaller values separately in graph F. However, it is important to point out that these two orders of magnitude estimated for σ^2 represent distinct sample of simmaps and thus both still recovered the same pattern of σ^2 estimated under OUMVA with arboreal vipers associated with smaller sigma values.

Species diversification

Contrary to our expectations, speciation rates do not differ among habitats (Figure 4). If speciation rates of arboreal lineages were smaller than terrestrial ones, we would expect the 95% HPD intervals of posterior distributions of differences (arboreal minus terrestrial) to comprise only negative values. However, although medians fall either in positive or negative values depending on the terrestrial category being compared, the 95% HPD intervals include zero in all comparisons (Figure 4B). Similarly, extinction and net diversification rates do not differ among habitats (Figure S4 and S5).

Transition rates estimated by MuSSE suggest that during the diversification of vipers, transitions between habitats predominantly occurred among terrestrial ones (Figure 4C). Higher transition rates occurred between lineages from open habitats and those occurring in both open and forested habitats, with higher rates into open habitats (Figure 4C). Interestingly, the open and forested habitats seems to act as an “intermediate stage” for shifts between more restricted terrestrial categories (terrestrial forested habitats and terrestrial open habitats, Figure 4C). Transitions into arboreal habitats seem to have been mostly limited to occur from terrestrial lineages restricted to forested habitats. Finally, transition estimates suggest that once a lineage becomes arboreal, it will rarely become terrestrial again (Figure 4C).

Our posterior predictive simulations indicate that the rate values obtained in MuSSE analyses represent a plausible and reliable evolutionary scenario, since the distribution of trait state proportions from simulated trees encompass the trait state proportions observed in our data (Figure S6).

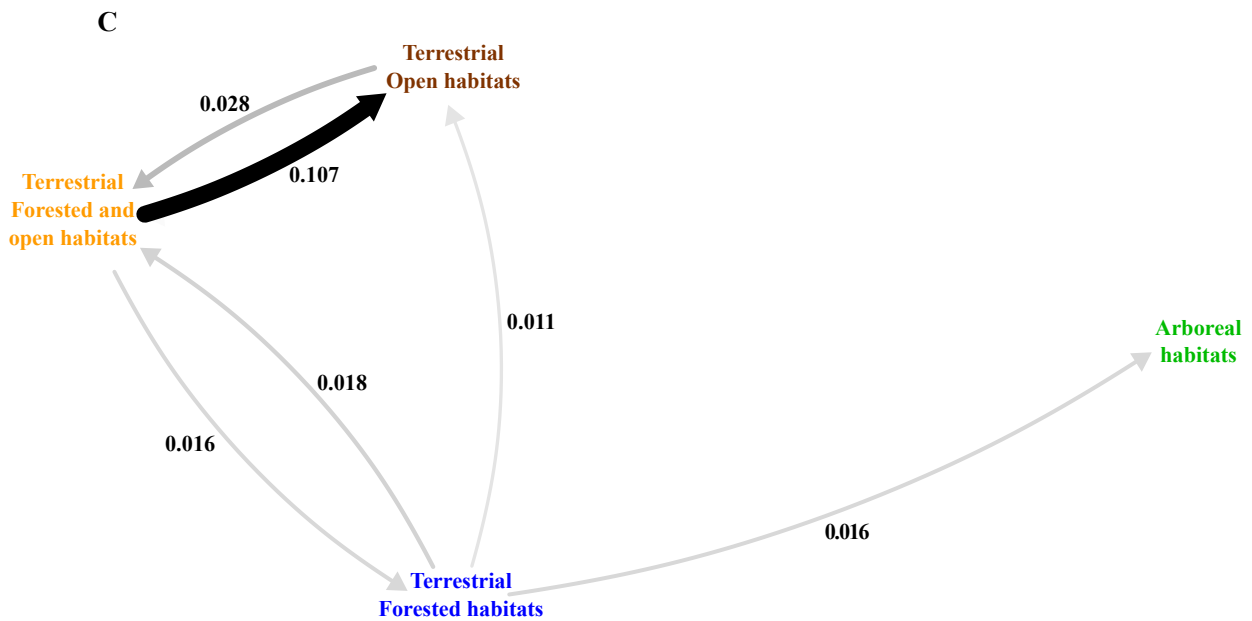
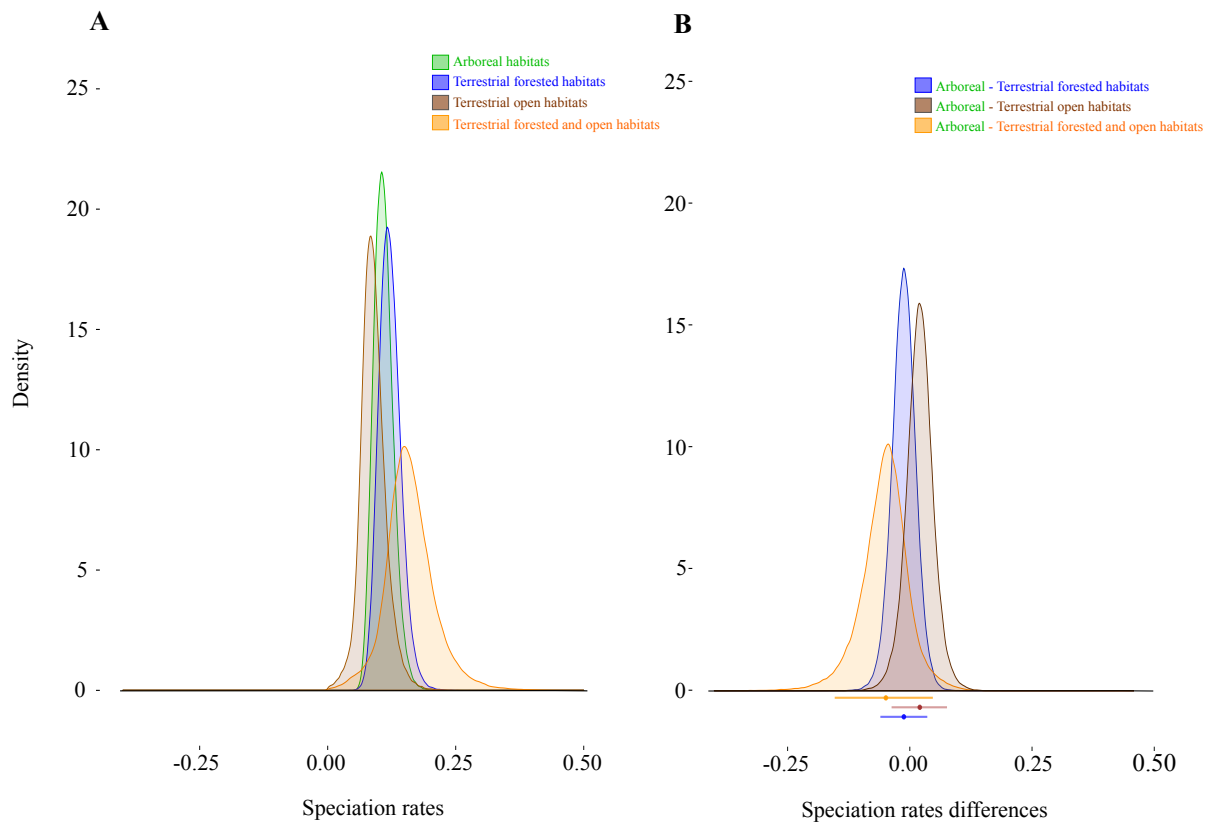


Figure 4. (A) Posterior distributions of speciation rates estimated for arboreal and terrestrial lineages. (B) Posterior distributions of differences between speciation rates of arboreal and terrestrial lineages. Bars represent 95% HPD interval and dots represent medians. (C) Median of the estimated transition rates between lineages from distinct habitats. Transitions rates less than or equal to 0.01 are not shown.

DISCUSSION

Body size evolves differently between arboreal and terrestrial habitats in vipers. Trait evolution analyses suggest that habitats are associated with distinct body size optima, and in accordance with our expectation, arboreal habitats seem to impose stronger evolutionary constraints in phenotypic diversification. Surprisingly, MuSSE analyses did not find any substantial difference between speciation rates among habitats. That is, phenotypic and species diversification seem to be decoupled, suggesting that phenotypic constraints associated to the arboreal habitat do not affect species diversification in vipers.

Being elongated and limbless organisms, snakes face several challenges in arboreal habitats, such as moving through the discontinuous physical environment and the gravitational effects on blood circulation (see Lillywhite and Henderson, 1993, Pizzatto et al., 2007, Sheehy et al., 2016). As a consequence, living in arboreal habitats demand a much more specialized morphology compared to other environments (as suggested in the literature by the “morphological syndromes”, Greene, 1997, Lillywhite and Henderson, 1993), potentially preventing further phenotypic diversification, which might be more relaxed in terrestrial habitats. Although other studies explored potential associations between arboreal habitats and body size evolution among snakes (e.g. Martins et al., 2001, Pizzatto et al., 2007, Sheehy et al., 2016), these studies solely investigated correlational trends (i.e. evolution of arboreal habits associated with a decrease in body size) finding contrasting results. Despite extensively suggested in the literature (e.g. Greene, 1997, Lillywhite and Henderson, 1993, Lillywhite et al., 2012), this is the first study to find evidence for distinct morphological evolutionary regimes among habitats in snakes with arboreal ones imposing stronger limits on morphological evolution. Interestingly, limitations on morphological diversification imposed by arboreal habitats are not exclusive among vipers but also seem to strongly constrain morphology among pigeons and doves (Lapiedra et al., 2013). Given that a change to arboriality might in fact impose novel environmental challenges, the evolution of arboreality might frequently characterize a remarkable event during the radiation of a lineage where shifts into this functional demanding ecological regime might promote a strong effect in phenotypic diversification.

The distribution of body sizes in arboreal vipers tends to be intermediate and more clumped when compared to terrestrial lineages (Figure 2). Indeed trait evolution models suggested that arboreal vipers might be evolving toward an intermediate body size optima compared to terrestrial habitats. In arboreal habitats, having larger body sizes could increase gravitational disturbance in blood circulation, especially in elongate organisms assuming vertically oriented postures such as snakes (e.g. Lillywhite and Henderson, 1993, Sheehy et al., 2016). This could be especially true for vipers because differently from other arboreal snakes, some physiological adaptations to arboreality

suggested to potentially overcome gravitational impacts on blood circulation (e.g. anterior heart position), did not evolve among arboreal vipers (Lillywhite et al., 2012). Thus, longer body sizes would be even more problematic for arboreal vipers. Moreover, although arboreal snakes have a smaller mass/body size ratio compared to terrestrial ones (e.g. Feldman and Meiri, 2013, Guyer and Donnelly, 1990), there might be a limit in decreasing mass relative to an increase in body size. Therefore, an increase in body size would also increase mass bringing difficulties for more heavy bodied snakes in moving along small and fragile tree branches (Ray, 2012, Sheehy et al., 2016). In contrast, limits to the evolution of smaller body sizes in arboreal snakes have not been extensively explored in the literature. Studying NW vipers, Hendry et al. (2014) suggest that despite the overall decrease in body size in arboreal lineages there may be a lower bound on female size possibly imposed by fecundity selection. Our results indicate however that a lower bound in body size evolution in fact exists among arboreal vipers, at least in males which were the commonly measured sex in this study. We thus argue that being too small may also prevent snakes in spanning gaps between branches leading to an inefficient locomotion along trees. Thus, it seems completely feasible that neither a body size that is too large nor too small characterizes the optimum for vipers leaving in arboreal habitats. Different body size optima associated with arboreal relative to other habitats have also been suggested for lizards (Collar et al., 2011) and pigeons and doves (Lapiedra et al., 2013).

Trait evolution analyses also show that body size optimum in arboreal habitats seems to be more confidently estimated than for terrestrial habitats (Figure 3). This greater variation found among the optima estimated for terrestrial habitats might reflect its greater environmental heterogeneity comprising several opportunities for terrestrial lineages to adapt driving them to distinct adaptive peaks. We interpret the observed variation on the model selection analysis seen for terrestrial habitats as a result of models with different estimated optima equally explaining body size evolution in terrestrial vipers. We note though that the suggestion of terrestrial habitats comprising multiple adaptive peaks relative to arboreal ones has also been made in studies focusing in other animal groups (e.g. lizards, Collar et al., 2011, Birds, Lapiedra et al., 2013). This idea is, in fact, in accordance with the differences in α and σ^2 estimates found between habitats.

Alpha and σ^2 are very important to the interpretations of our results because they describe the possible phenotypic evolutionary trajectories found among habitats. The three most frequently chosen trait evolution models suggested the same general pattern regarding α and σ^2 : lineages evolving in arboreal habitats tend to be associated with higher α and smaller σ^2 compared to those evolving in the three terrestrial habitats. Taken together, higher α and smaller σ^2 mean weaker phylogenetic correlation, that is, ancient adaptations are less important relative to the adaptation to new conditions (Hansen, 1997), faster evolution toward the optima and weaker stochastic noise (or

variance accumulation rate) (Butler and King, 2004, Hansen, 1997, Ho and Ané, 2014). Thus, these two parameter estimates are in accordance with a scenario where vipers might be evolving toward a broader range of morphological variants in terrestrial habitats. On the other hand, limited morphotypes are able to persist in arboreal habitats and once a lineage invades the arboreal environment, strong selective pressures rapidly drive morphology toward this optima. It is interesting to note that arboreal snakes tend to evolve from lineages that inhabit terrestrial forested habitats (Figure 4C), which depending on the model are characterized by either a slightly larger body size optima than arboreal habitats (Figure 3D and Figure 3G), or in the case of the most complex model, to have a broader distribution of body size that encompasses larger and smaller species (Figure 3A). This is indeed in accordance with a scenario that suggests a very strong selection towards evolving smaller or larger body size (the higher alpha value) and a strong phenotypic constraint once the arboreal optima is attained (smaller sigma).

One could argue that α values recovered in the present study might be too small compared to other studies (e.g. Lapedra et al., 2013) possibly suggesting that processes described here do not differ from a Brownian-like process (see Cooper et al., 2015) and that the differences found between arboreal and terrestrial habitats would be negligible. However, as suggested by Cressler et al. (2015), interpretation and comparison of α and σ^2 between studies are complicated because of their dependence on time and trait being studied. Snakes have much less varied morphologies compared to other animals and thus, magnitude and directions that their body size have to evolve will undoubtedly be much more limited. To get an idea if the estimated α correspond to meaningful values, it is more interesting to interpret α values relative to the height of the phylogenetic tree used in the study by estimating the phylogenetic half-life ($t_{1/2}$): $\ln(2)/\alpha$ (Cooper et al., 2015, Ho and Ané, 2014). Phylogenetic half-life measures the time an ancestral phenotype takes to evolve halfway toward the optimum after a lineage invade a new selective regime (in our case, a different habitat). Thus, if $t_{1/2}$ is extremely large relative to the tree height it suggests that the OU process may be extremely weak (Cooper et al., 2015). Among vipers, $t_{1/2}$ of body size in arboreal habitats comprise much shorter times (8.23 or 14.25 mya, Table 2) than the stem age of most arboreal lineages (~25-35 mya, see Chapter I) and thus, are even shorter compared to the tree height of vipers (~50 mya crown group, see Chapter I). Phylogenetic half-lives in terrestrial habitats are comparatively larger suggesting that body size evolves at least two times faster toward the optimum in arboreal habitats than in terrestrial ones (Table 2). This difference is even more pronounced for phylogenetic half-lives estimated when OUMVA was the best model (Table 2). That is, α values in the present study seem to comprise plausible estimates undoubtedly showing evidence that different evolutionary processes are associated to the evolution of body size in arboreal when compared to terrestrial vipers.

Table 2. Median of phylogenetic half lives ($\ln(2)/\alpha$) estimated for each habitat category across simmaps suggesting OUMVA or OUMA as the best trait evolution models. Mya: million of years ago.

	Terrestrial forested and open habitats	Terrestrial open habitats	Terrestrial forested habitats	Arboreal habitats
OUMVA	32.69 mya	29.56 mya	30.56 mya	8.23 mya
OUMA	26.51 mya	24.65 mya	26.07 mya	14.05 mya

Because morphology usually reflects the ecology of an organism (Wainwright and Reilly, 1994) and selection to different niches is known to underlie species diversification in several groups (e.g. Grant and Grant, 2008, Losos, 2009), speciation is frequently coupled with morphological changes (see Nosil, 2012). Thus, an ecological regime imposing strong constraints in the morphological evolution of a lineage is expected to comprise limited speciation opportunities for that lineage to diversify. However, this was not found in vipers. Although trait evolution analyses suggest that morphological evolution is more constrained among lineages evolving in arboreal habitats, differently from what we expected, this limited adaptive landscape does not seem to have affected the speciation opportunities faced by arboreal relative to terrestrial vipers; in other words, speciation rates remain similar between both types of habitats. In fact, the present study is in accordance with a decoupling between morphological and species diversification in vipers, bringing light into the speciation dynamics that are driving the diversity of the group.

A decoupling of morphological and taxonomic diversity has been found in the fossil record (Benton, 2015, Ruta et al., 2013) indicating that, over geological time scales, patterns of constraints in morphology might change and species diversity can accumulate without increasing morphological diversity (Foote, 1993). In fact, although underappreciated, there are several mechanisms of morphologically static cladogenesis (i.e. speciation not accompanied by significantly morphological change) (Bickford et al., 2006). For example, after two populations become geographically isolated random changes in gene frequencies might take place due to genetic drift or bottleneck effects ultimately causing hybrid sterility or inviability (Coyne and Orr, 2004, Nosil, 2012). Geographically isolated populations could also give rise to new species through processes involving selection and still do not show significant morphological differences. As an example, in processes like mutation-order speciation, different but incompatible mutations are fixed in separated populations adapting to similar selective regimes (Butlin et al., 2012, Nosil and Flaxman, 2011, Nosil, 2012). Similar to drift, reproductive isolation would emerge, for example, due to genetic conflicts between populations (e.g. Coyne and Orr, 2004). Moreover, mating

preferences divergence on chemical signals between isolated populations could also underlie species diversification without leading to morphological changes. Indeed, differential recognition with subsequently preference for different chemical signals is suggested to play a significant role during the speciation process in different organisms (Review in Smadja and Butlin, 2009), especially among those relying more upon chemical senses than vision (Bickford et al., 2006), like snakes (Fornasiero et al., 2011, Shine et al., 2002). It is worth mentioning that mutation-driven or sexual chemical selection do not only occur in the presence of geographic isolation, but these processes might be significantly enhanced when populations are geographically isolated (e.g. Fornasiero et al., 2011, Nosil and Flexman, 2011).

At first glance, our results could suggest that mechanisms underlying species diversification in vipers might agree with a scenario where speciation events are not coupled with morphological changes. That is, selective pressures imposed on body size, even if constraining it, might not have affected the probability of speciation among habitats rendering speciation rates similar. The morphological diversity observed among terrestrial lineages would have arrived due to natural selection mostly after speciation is completed and not as drivers of reproductive isolation (Rundell and Price, 2009, Svensson, 2012). Interesting, the formation of cryptic species, usually related to species diversification without morphological change, has been described among both terrestrial and arboreal vipers (e.g. Jadin et al., 2012, Sanders et al., 2006). Alternatively, if morphological differentiation indeed occur during the speciation process in vipers, the similar speciation rates found among habitats showing distinct morphological regimes require an additional explanation. In fact, we could argue that after populations are geographically isolated, completion of the speciation process might occur more frequently in arboreal compared to terrestrial habitats potentially offsetting any decrease in speciation opportunities mediated by divergent selection in the arboreal environment.

The evolutionary transitions rates found here suggest that once arboreality evolved in vipers, reversals to terrestrial habitats are rare (Figure 4C). In fact this pattern seems to predominate among arboreal lineages in general (e.g. Lapiedra et al., 2013). In contrast, our results show that transitions between terrestrial habitats are more common (Figure 4C). Vipers comprise several genera including solely arboreal species (and thus, without any reversions to terrestriality, see Figure 1), such as the species-rich *Trimeresurus* comprising around 50 species and estimated to have ~35 million years (Uetz and Hosek, 2014, Chapter I). As proposed by Wiens (2004) the tendency of lineages to retain their niches over long evolutionary time scales (i.e. phylogenetic niche conservatism), and thus, failing to adapt to new ecological conditions, is a key factor during speciation following geographic isolation (see also Kozak and Wiens, 2006). Under this scenario it is possible that individuals of arboreal lineages would be unable to persist in new environmental

conditions rendering their geographic range to become fragmented (Wiens, 2004). Assuming such mechanism in our case would suggest that speciation rates would remain similar between habitats because the decrease in speciation opportunities due to divergent selection in arboreal habitats would be offset by the higher frequency of speciation mediated by geographic isolation in lineages inhabiting those environments.

A scenario where an increase in the opportunities to speciate via geographic isolation offset the decrease in divergent selection opportunities in arboreal habitats might indeed leave a detectable signal in the geographical ranges of lineages. If geographical isolation is the predominant process underlying species divergence among arboreal lineages, terrestrial ones are expected to show higher spatial overlap. However, if no differences are detected, geographic isolation might be a common pattern among vipers despite of their habitat. If this is the case, the different morphological trajectories but similar speciation rates found among habitats would possibly suggest that speciation mechanisms among vipers might not be followed by morphological differentiation.

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SUPPLEMENTARY FIGURES

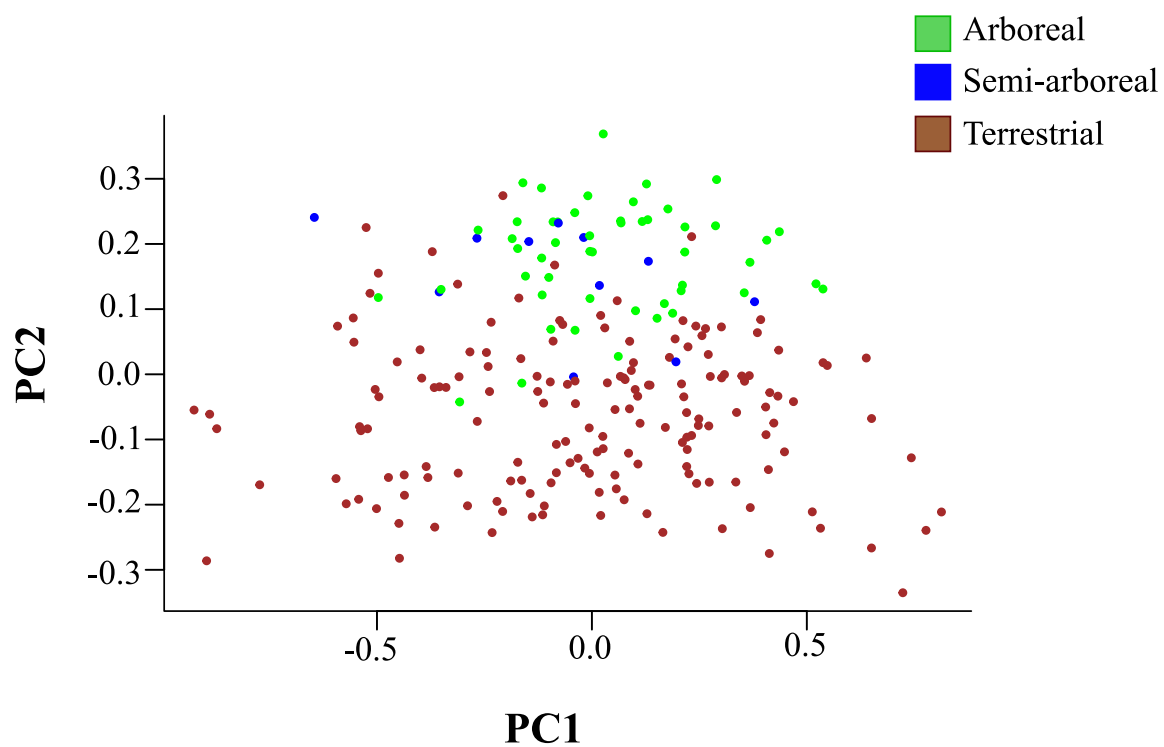


Figure S1. Morphospace of arboreal, semi-arboreal and terrestrial vipers.

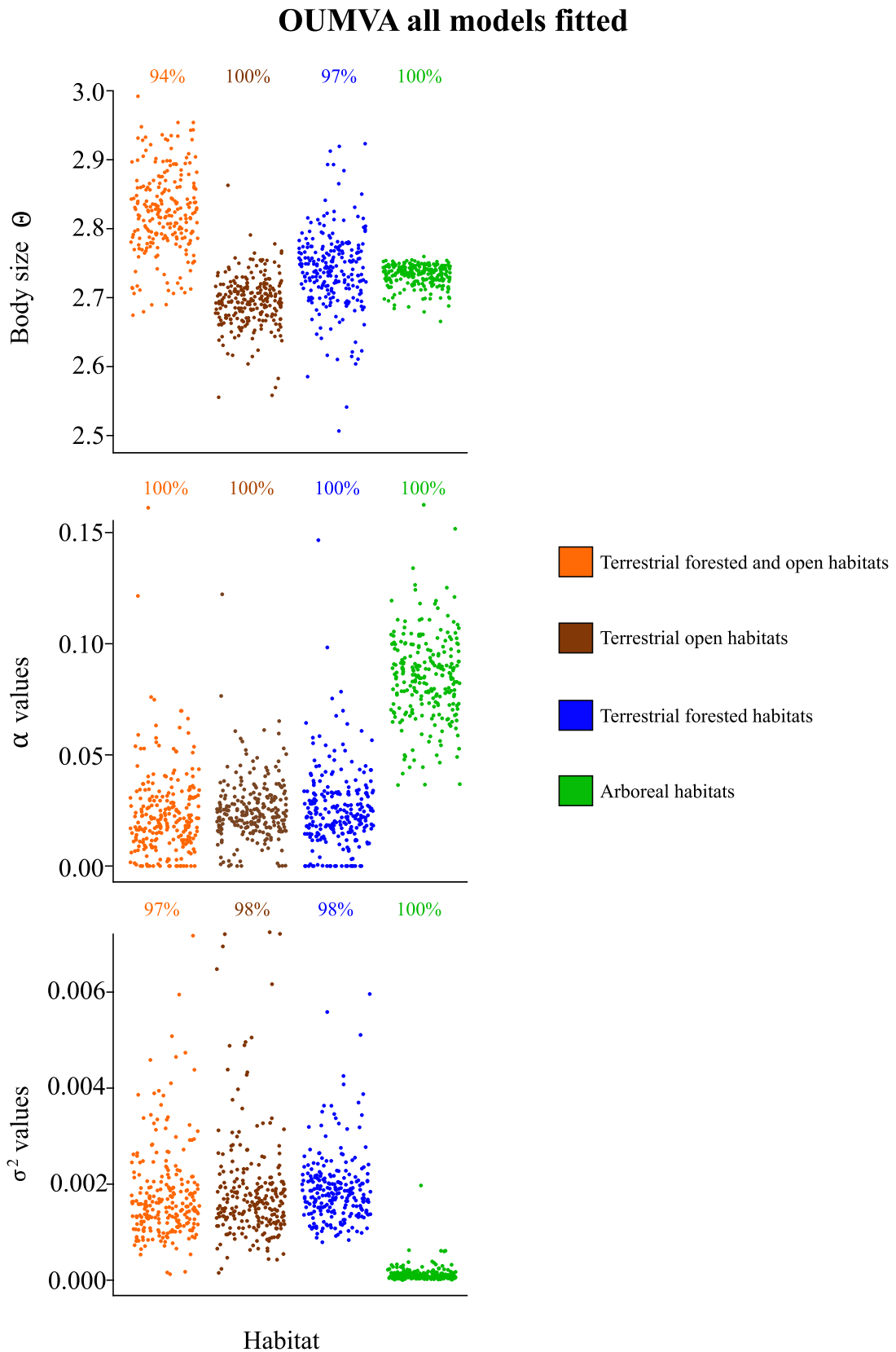


Figure S2. Parameters estimates under OUMVA when chosen as the best model in OUwie analyses where all models could be fitted (261 simmaps). Dots represent OUwie analyses performed in distinct simmaps. For better visualization we excluded extreme values and numbers on the top represent the proportion of values included.

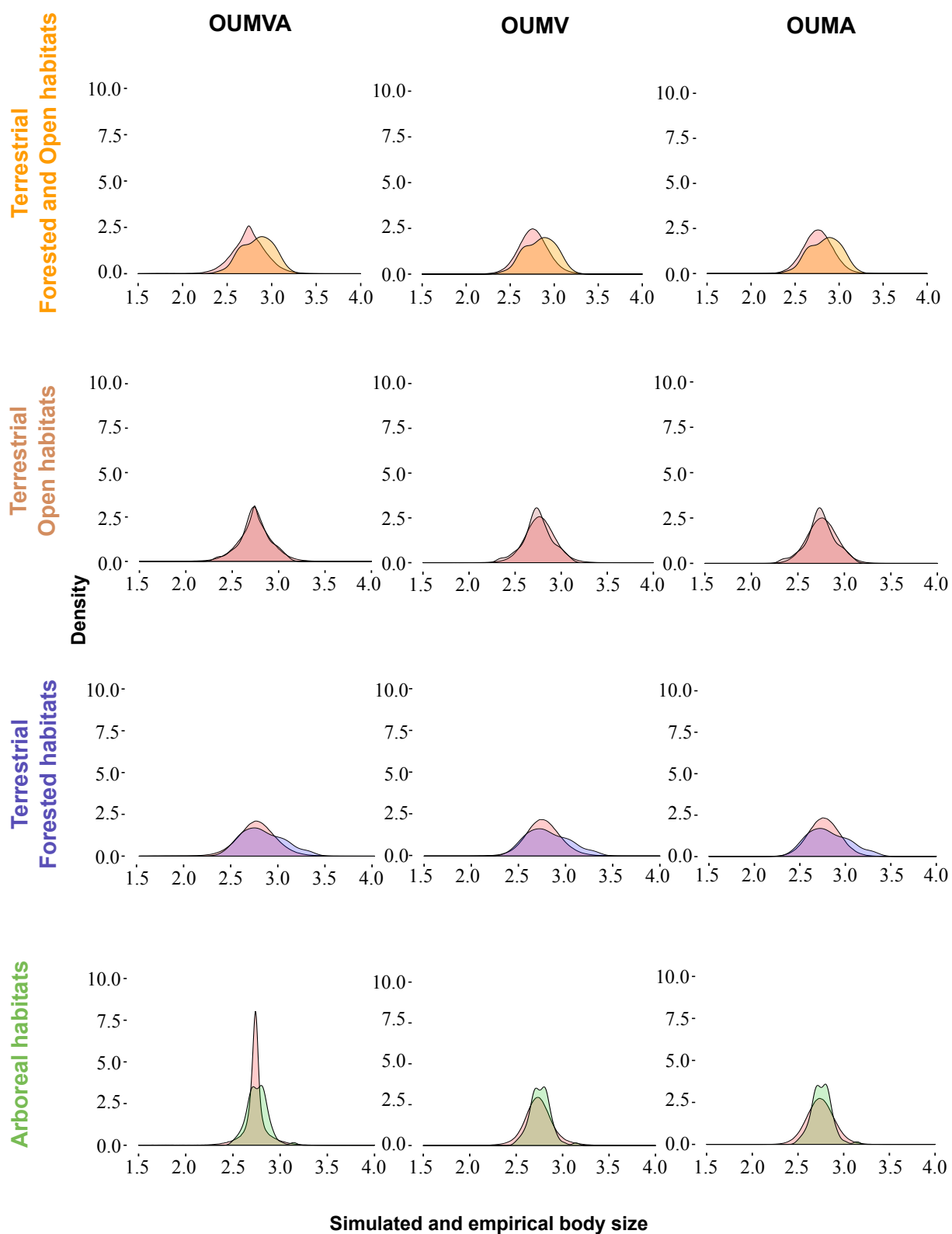


Figure S3. Simulated (light red curves) body size values using parameters estimated under each of the most frequently chosen models (OUMVA, OUMV and OUMA) (see methods for details on simulations) compared to empirical body size values in each habitat category.

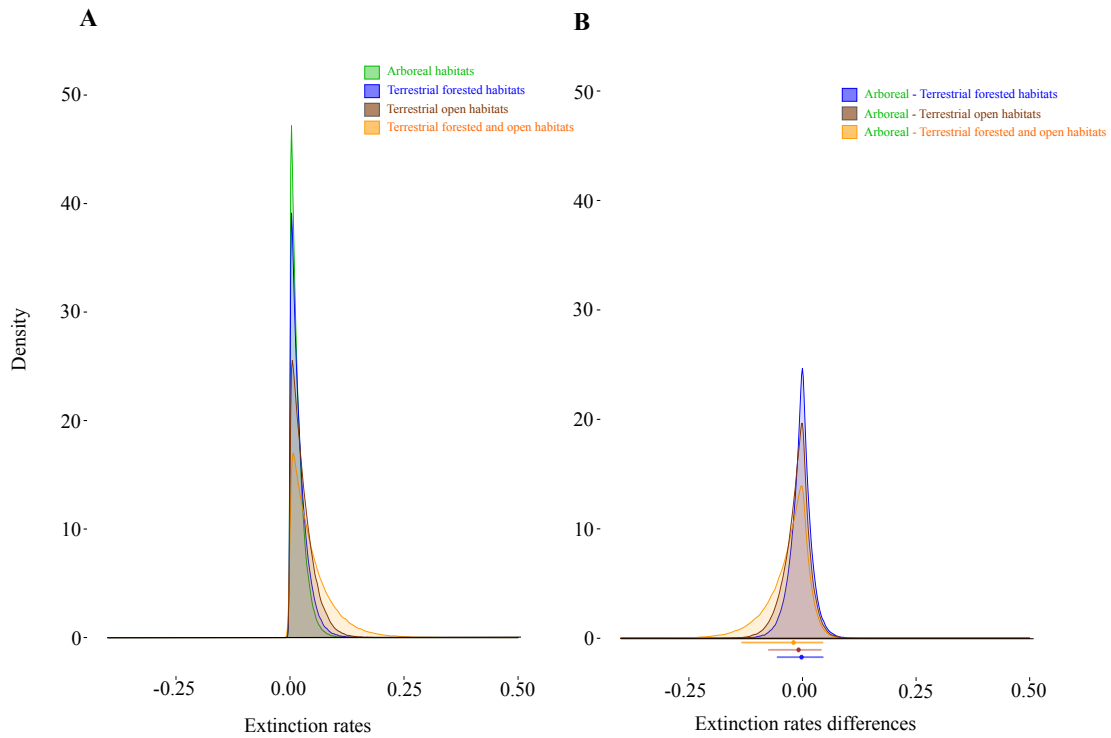


Figure S4. (A) Posterior distributions of extinction rates estimated for arboreal and terrestrial lineages. (B) Posterior distributions of differences between extinction rates of arboreal and terrestrial lineages. Bars represent 95% HPD interval and dots represent medians.

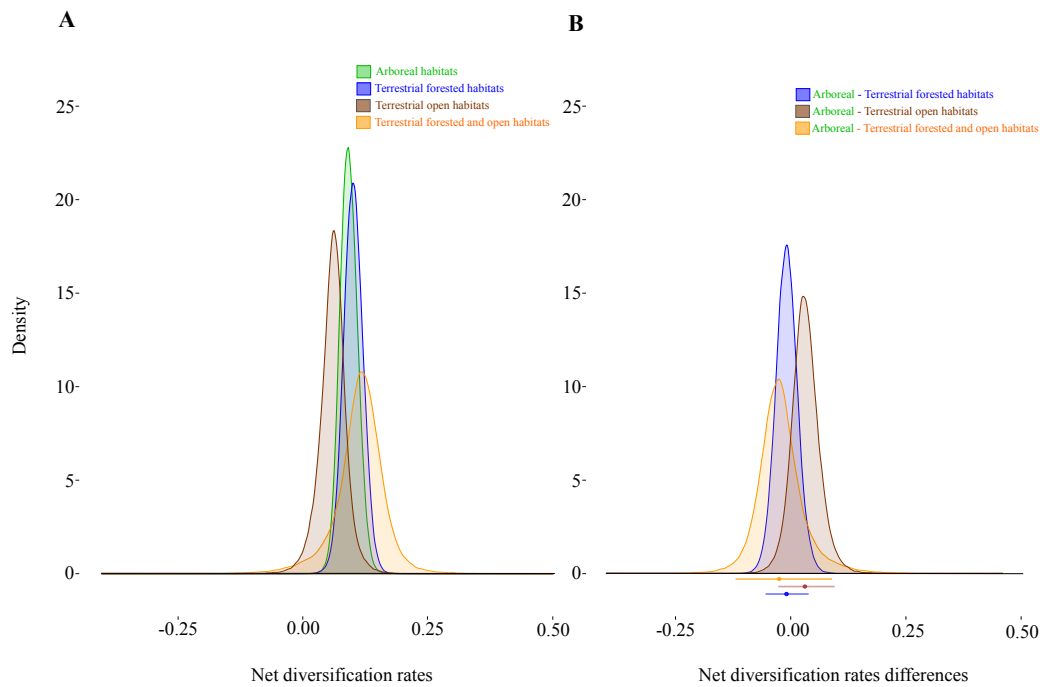


Figure S5. (A) Posterior distributions of net diversification rates estimated for arboreal and terrestrial lineages. (B) Posterior distributions of differences between net diversification rates of arboreal and terrestrial lineages. Bars represent 95% HPD interval and dots represent medians.

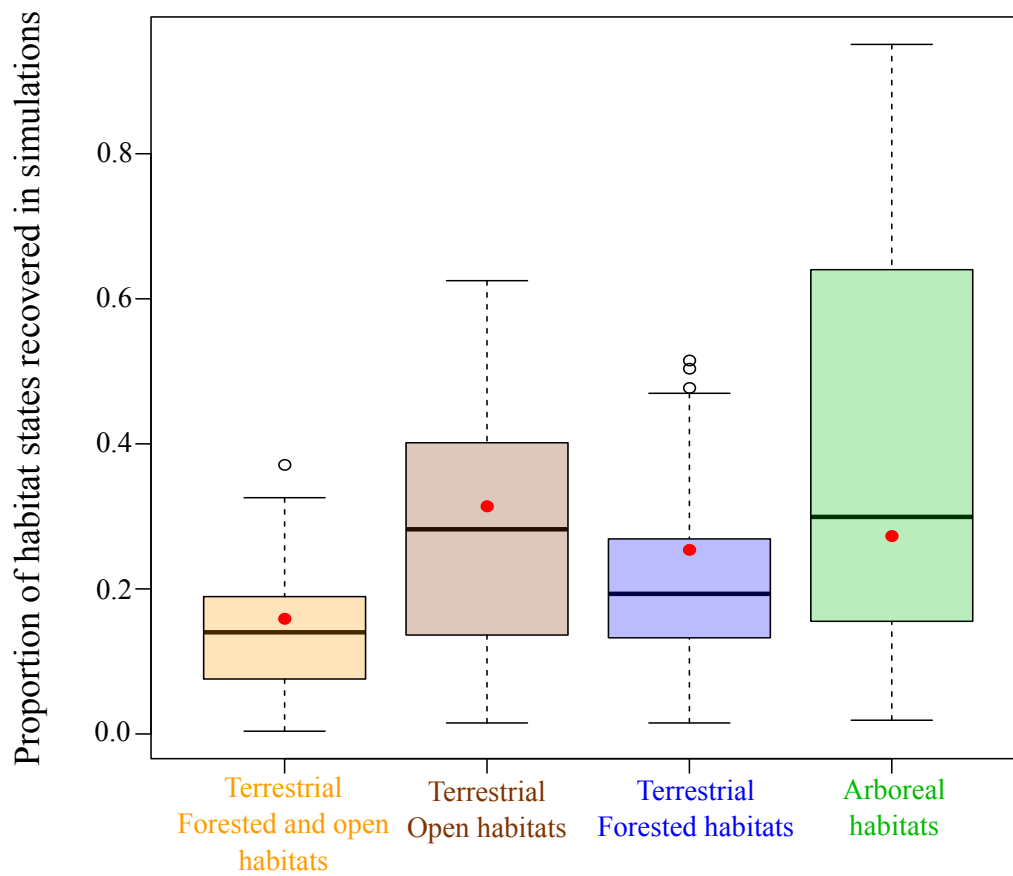


Figure S6. Posterior predictive simulations showing the proportion of each habitat state recovered by simulated trees and the empirical proportion found in vipers (red dot).

SUPPLEMENTARY TABLES

Table S1. Average body size (mm) and habitat use for viper species included in the present study. N = number of individuals measured for each species, 1 = terrestrial open and forested habitats, 2 = terrestrial open habitats, 3 = terrestrial forested habitats, 4 = arboreal habitats. * denotes species for which females measurements were used.

Species	Mean adult body size	N	Source	Habitat	Source
<i>Agkistrodon bilineatus</i>	855.7	4	<i>This study</i>	1	Gloyd and Conant (1990), Campbell and Lamar (2004)
<i>Agkistrodon contortrix</i>	958.2	5	<i>This study</i>	1	Gloyd and Conant (1990), Campbell and Lamar (2004)
<i>Agkistrodon howardgloydi</i>	750	1	<i>This study</i>	1	Solórzano et al. (1999), Campbell and Lamar (2004)
<i>Agkistrodon piscivorus</i>	889.5	6	<i>This study</i>	1	Gloyd and Conant (1990), Campbell and Lamar (2004)
<i>Agkistrodon russeolus</i>	-	-	-	1	Campbell and Lamar (2004)
<i>Agkistrodon taylori</i>	648	3	<i>This study</i>	1	Deufel and Cundall (2006), Lavin et al. (2007)
<i>Atheris barbouri</i>	282*	2	<i>This study</i>	3	Mallow et al. (2003), Phelps (2010), Menegon et al. (2011)
<i>Atheris ceratophora</i>	365	3	<i>This study</i>	4	Branch (2005), Phelps (2010)
<i>Atheris chlorechis</i>	459.5	2	<i>This study</i>	4	Mallow et al. (2003), Phelps (2010)
<i>Atheris desaixi</i>	473.5	4	<i>This study</i>	4	Mallow et al. (2003), Branch (2005)
<i>Atheris hispida</i>	474.6	5	<i>This study</i>	4	Branch (2005)
<i>Atheris matildae</i>	539.5	2	Menegon et al. (2011)	4	Menegon et al. (2011)
<i>Atheris nitschei</i>	522.7	6	<i>This study</i>	4	Mallow et al. (2003), Branch (2005), Phelps (2010)
<i>Atheris squamigera</i>	427.3	8	<i>This study</i>	4	Mallow et al. (2003), Branch (2005), Phelps (2010)
<i>Atropoides indomitus</i>	577	1	<i>This study</i>	3	Smith and Ferrari-Castro (2008),
<i>Atropoides mexicanus</i>	743.3	9	<i>This study</i>	1	Campbell and Lamar (2004)

<i>Atropoides nummifer</i>	589.5	4	<i>This study</i>	3	Campbell and Lamar (2004), Lopez-Luna and Canseco-Márquez (2007)
<i>Atropoides occiduus</i>	665.8	4	<i>This study</i>	3	Campbell and Lamar (2004), Hampton (2011)
<i>Atropoides olmec</i>	583.3	3	<i>This study</i>	3	Hampton (2011), Calderón Mandujano and Lopez-Luna (2014)
<i>Atropoides picadoi</i>	1078	5	<i>This study</i>	3	Campbell and Lamar (2004), Hampton (2011)
<i>Azemiops feae</i>	571	6	<i>This study</i>	1	Orlov et al. (2002), Leviton et al. (2003), Mallow et al. (2003), Phelps (2010)
<i>Bitis arietans</i>	821.8	9	<i>This study</i>	1	Branch (2005), Deufel and Cundall (2006), Phelps (2010)
<i>Bitis armata</i>	223	1	<i>This study</i>	2	Phelps (2010)
<i>Bitis atropos</i>	492.8	5	<i>This study</i>	2	Mallow et al. (2003)
<i>Bitis caudalis</i>	295	5	<i>This study</i>	2	Branch (2001), Mallow et al. (2003), Phelps (2010)
<i>Bitis cornuta</i>	436.5	4	<i>This study</i>	2	Branch (2001), Mallow et al. (2003), Phelps (2010)
<i>Bitis gabonica</i>	1143.8	6	<i>This study</i>	3	Branch (2001) and (2005), Mallow et al. (2003), Phelps (2010)
<i>Bitis nasicornis</i>	824.2	6	<i>This study</i>	3	Branch (2005), Deufel and Cundall (2006), Phelps (2010)
<i>Bitis parviocula</i>	-	-	-	1	Largen and Rasmussen (1993), Mallow et al. (2003)
<i>Bitis peringueyi</i>	259.7	6	<i>This study</i>	2	Branch (2001), Mallow et al. (2003), Phelps (2010)
<i>Bitis rhinoceros</i>	1352	1	<i>This study</i>	3	Phelps (2010)
<i>Bitis rubida</i>	-	-	-	2	Phelps (2010)
<i>Bitis schneideri</i>	217	2	<i>This study</i>	2	Branch (2001), Phelps (2010)
<i>Bitis worthingtoni</i>	357.3	3	<i>This study</i>	2	Mallow et al. (2003), Phelps (2010)
<i>Bitis xeropaga</i>	390	2	<i>This study</i>	2	Branch (2001), Phelps (2010)
<i>Bothriechis aurifer</i>	736.7	7	<i>This study</i>	4	Campbell and Lamar (2004)
<i>Bothriechis bicolor</i>	647.8	4	<i>This study</i>	4	Campbell and Lamar (2004)
<i>Bothriechis guifarroi</i>	653	1	<i>This study</i>	4	Townsend et al. (2013)

<i>Bothriechis lateralis</i>	679.4	9	<i>This study</i>	4	Campbell and Lamar (2004)
<i>Bothriechis marchi</i>	656.5	11	<i>This study</i>	4	McCranie and Castañeda (2006)
<i>Bothriechis nigroviridis</i>	404.8	6	<i>This study</i>	4	Campbell and Lamar (2004), Deufel and Cundall (2006)
<i>Bothriechis rowleyi</i>	738.1	9	<i>This study</i>	4	Campbell and Lamar (2004)
<i>Bothriechis schlegelii</i>	516	5	<i>This study</i>	4	Campbell and Lamar (2004), Deufel and Cundall (2006)
<i>Bothriechis supraciliaris</i>	320	1	<i>This study</i>	4	Solórzano (1998)
<i>Bothriechis thalassinus</i>	674.3	3	<i>This study</i>	4	Campbell and Smith (2000)
<i>Bothrocophias campbelli</i>	675*	1	<i>This study</i>	3	Gutberlet and Campbell (2001), Campbell and Lamar (2004), Cisneros-Heredia et al. (2006), Valencia et al. (2008)
<i>Bothrocophias hyoprora</i>	405.9	9	<i>This study</i>	3	Campbell and Lamar (2004), Cisneros-Heredia et al. (2006)
<i>Bothrocophias microphthalmus</i>	446	2	<i>This study</i>	3	Campbell and Lamar (2004), Cisneros-Heredia et al. (2006)
<i>Bothrops alcatraz</i>	462.3	3	<i>This study</i>	4	Campbell and Lamar (2004)
<i>Bothrops alternatus</i>	775.6	9	<i>This study</i>	2	Martins et al. (2001), Fenwick et al. (2009)
<i>Bothrops ammodytoides</i>	451	4	<i>This study</i>	2	Campbell and Lamar (2004), Fenwick et al. (2009)
<i>Bothrops asper</i>	1241.5	4	<i>This study</i>	3	Campbell and Lamar (2004), Fenwick et al. (2009)
<i>Bothrops atrox</i>	913.5	6	<i>This study</i>	3	Campbell and Lamar (2004), Fenwick et al. (2009)
<i>Bothrops barnetti</i>	600.8	4	<i>This study</i>	2	Schmith and Walker (1943), Campbell and Lamar (2004), Kwiatkowski and Burt (2011)
<i>Bothrops bilineata</i>	538.3	3	<i>This study</i>	4	Martins et al. (2001), Campbell and Lamar (2004), Harvey et al. (2005)
<i>Bothrops brazili</i>	842	2	<i>This study</i>	3	Martins et al. (2001), Campbell and Lamar (2004), Fenwick et al. (2009)
<i>Bothrops caribbaeus</i>	1096.3	4	<i>This study</i>	3	Campbell and Lamar, 2004, Henderson (2004), Deufel and Cundall (2006)
<i>Bothrops chloromelas</i>	571*	1	<i>This study</i>	4	Campbell and Lamar (2004), Hampton (2011)
<i>Bothrops colombiensis</i>	-	-	-	3	Silva et al. (1985), Rivas et al. (2012)
<i>Bothrops cotiara</i>	728.7	7	<i>This study</i>	1	Campbell and Lamar (2004), Fenwick et al. (2009)

<i>Bothrops diporus</i>	503	1	<i>This study</i>	2	Silva and Rodrigues (2008), Fenwick et al. (2009)
<i>Bothrops erythromelas</i>	479.3	6	<i>This study</i>	2	Campbell and Lamar (2004), Fenwick et al. (2009)
<i>Bothrops fonsceai</i>	775	6	<i>This study</i>	1	Martins et al. (2001), Fenwick et al. (2009)
<i>Bothrops insularis</i>	660.5	8	<i>This study</i>	4	Amaral (1921a,b)
<i>Bothrops isabelae</i>	-	-	-	3	Campbell and Lamar (2004), Fenwick et al. (2009), Rivas et al. (2012)
<i>Bothrops itapetiningae</i>	392	2	<i>This study</i>	2	Martins et al. (2001), Campbell and Lamar (2004)
<i>Bothrops jararaca</i>	833	26	<i>This study</i>	3	Schmith and Walker (1943), Campbell and Lamar (2004)
<i>Bothrops jararacussu</i>	995	6	<i>This study</i>	3	Martins et al. (2001), Campbell and Lamar (2004), Fenwick et al. (2009)
<i>Bothrops lanceolatus</i>	1189.8	4	<i>This study</i>	3	Gosner (1987), Campbell and Lamar (2004), Henderson (2004)
<i>Bothrops leucurus</i>	772.5	8	<i>This study</i>	3	Martins et al. (2001), Campbell and Lamar (2004), Fenwick et al. (2009)
<i>Bothrops lojanus</i>	514.4	5	<i>This study</i>	3	Campbell and Lamar (2004), Harvey et al. (2005)
<i>Bothrops lutzi</i>	550.5	2	<i>This study</i>	2	Campbell and Lamar (2004), Silva and Rodrigues (2008), Fenwick et al. (2009)
<i>Bothrops marajoensis</i>	-	-	-	3	Campbell and Lamar (2004), Fenwick et al. (2009), M. Martins pers. comm.
<i>Bothrops marmoratus</i>	521.5	2	<i>This study</i>	2	Silva and Rodrigues (2008), Carrasco et al. (2012)
<i>Bothrops matogrossensis</i>	636	4	<i>This study</i>	2	Martins et al. (2001), Silva and Rodrigues (2008), Fenwick et al. (2009)
<i>Bothrops moojeni</i>	863.3	20	<i>This study</i>	3	Campbell and Lamar (2004), Fenwick et al. (2009), M. Martins pers. comm.
<i>Bothrops neuwiedi</i>	681.9	8	<i>This study</i>	2	Martins et al. (2001), Campbell and Lamar (2004), Silva and Rodrigues (2008)
<i>Bothrops osbornei</i>	-	-	-	4	Campbell and Lamar (2004), Fenwick et al. (2009)
<i>Bothrops pauloensis</i>	661.8	4	<i>This study</i>	2	Martins et al. (2001), Campbell and Lamar (2004), Silva and Rodrigues (2008), Fenwick et al. (2009)
<i>Bothrops pictus</i>	691.8	4	<i>This study</i>	2	Campbell and Lamar (2004), Kwiatkowski and Burt (2011)
<i>Bothrops pubescens</i>	698	2	<i>This study</i>	3	Martins et al. (2001), Fenwick et al. (2009), M. Hartmann pers. comm.
<i>Bothrops pulchra</i>	589	1	<i>This study</i>	4	Gutberlet and Campbell (2001)

<i>Bothrops punctatus</i>	770	2	<i>This study</i>	4	Campbell and Lamar (2004)
<i>Bothrops taeniata</i>	706*	1	<i>This study</i>	4	Martins et al. (2001), Campbell and Lamar (2004)
<i>Calloselasma rhodostoma</i>	644.7	3	<i>This study</i>	3	Orlov et al. (2002)
<i>Causus defilippii</i>	297.6	9	<i>This study</i>	2	Branch (2001), Mallow et al. (2003), Phelps (2010)
<i>Causus lichtensteini</i>	413.6	7	<i>This study</i>	3	Branch (2005), Mallow et al. (2003), Phelps (2010)
<i>Causus resimus</i>	532.3	8	<i>This study</i>	2	Branch (2005), Mallow et al. (2003), Phelps (2010)
<i>Causus rhombeatus</i>	559.6	7	<i>This study</i>	2	Branch (2005), Mallow et al. (2003), Phelps (2010)
<i>Cerastes cerastes</i>	534.2	5	<i>This study</i>	2	Mallow et al. (2003), Phelps (2010)
<i>Cerastes gasperettii</i>	542	5	<i>This study</i>	2	Mallow et al. (2003), Phelps (2010)
<i>Cerastes vipera</i>	343.2	6	<i>This study</i>	2	Mallow et al. (2003), Phelps (2010)
<i>Cerrophidion godmani</i>	643.5	6	<i>This study</i>	1	Campbell and Lamar (2004), Jadin et al. (2012)
<i>Cerrophidion petlalcalensis</i>	345	1	López-luna et al. (1999)	3	López-Luna et al. (1999), Campbell and Lamar (2004)
<i>Cerrophidion sasai</i>	583	3	<i>This study</i>	3	Jadin et al. (2012)
<i>Cerrophidion tzotzilorum</i>	370	2	<i>This study</i>	3	Campbell and Lamar (2004)
<i>Crotalus adamanteus</i>	1494.7	6	<i>This study</i>	1	Campbell and Lamar (2004)
<i>Crotalus aquilus</i>	655	7	<i>This study</i>	2	Campbell and Lamar (2004), Mociño-Deloya et al. (2008)
<i>Crotalus atrox</i>	1184.3	6	<i>This study</i>	2	Armstrong and Murphy (1979), Campbell and Lamar (2004)
<i>Crotalus basiliscus</i>	1265.4	5	<i>This study</i>	3	Campbell and Lamar (2004)
<i>Crotalus catalinensis</i>	996	1	<i>This study</i>	2	Campbell and Lamar (2004), Arnaud et al., 2008, Martins et al., 2008
<i>Crotalus cerastes</i>	517.4	8	<i>This study</i>	2	Campbell and Lamar (2004)
<i>Crotalus cerberus</i>	916	3	<i>This study</i>	2	Campbell and Lamar (2004)
<i>Crotalus culminatus</i>	1223.3	3	<i>This study</i>	2	Campbell and Lamar (2004)

<i>Crotalus durissus</i>	962	14	<i>This study</i>	2	Campbell and Lamar (2004)
<i>Crotalus enyo</i>	705.8	8	<i>This study</i>	2	Taylor and Price (2001), Campbell and Lamar (2004)
<i>Crotalus ericsmithi</i>	505	1	<i>This study</i>	3	Campbell and Flores-Villela (2008), J. Reyes-Velasco pers. comm.
<i>Crotalus horridus</i>	1247.7	6	<i>This study</i>	3	Campbell and Lamar (2004)
<i>Crotalus intermedius</i>	521.7	9	<i>This study</i>	3	Campbell and Lamar (2004)
<i>Crotalus lannomi</i>	493	2	Reyes-Velasco et al. (2010)	3	Campbell and Lamar (2004), Reyes-Velasco et al. (2010)
<i>Crotalus lepidus</i>	659.3	7	<i>This study</i>	1	Campbell and Lamar (2004)
<i>Crotalus mitchellii</i>	882.7	6	<i>This study</i>	2	Campbell and Lamar (2004)
<i>Crotalus molossus</i>	1130.8	6	<i>This study</i>	1	Campbell and Lamar (2004)
<i>Crotalus oreganus</i>	817.4	5	<i>This study</i>	1	Campbell and Lamar (2004)
<i>Crotalus ornatus</i>	1135	1	<i>This study</i>	1	Campbell and Lamar (2004)
<i>Crotalus polystictus</i>	778.3	8	<i>This study</i>	2	Campbell and Lamar (2004), Meik et al. (2012)
<i>Crotalus pricei</i>	496.6	8	<i>This study</i>	1	Campbell and Lamar (2004)
<i>Crotalus pusillus</i>	564.9	7	<i>This study</i>	2	Campbell and Lamar (2004)
<i>Crotalus ravus</i>	686.3	6	<i>This study</i>	1	Campbell and Lamar (2004)
<i>Crotalus ruber</i>	1064.1	8	<i>This study</i>	2	Klauber (1956), Douglas et al. (2006),
<i>Crotalus scutulatus</i>	740.6	5	<i>This study</i>	2	Campbell and Lamar (2004)
<i>Crotalus simus</i>	1174.6	5	<i>This study</i>	1	Campbell and Lamar (2004)
<i>Crotalus stejnegeri</i>	592	1	<i>This study</i>	3	Campbell and Lamar (2004), J. Reyes-Velasco pers. comm.
<i>Crotalus tancitarensis</i>	406.5*	2	<i>This study</i>	3	Campbell (2007)
<i>Crotalus tigris</i>	666.6	10	<i>This study</i>	2	Campbell and Lamar (2004), Douglas et al. (2006)
<i>Crotalus tortugensis</i>	1078	1	<i>This study</i>	2	Campbell and Lamar (2004)

<i>Crotalus totonacus</i>	1096	3	<i>This study</i>	3	Campbell and Lamar (2004)
<i>Crotalus transversus</i>	309	1	<i>This study</i>	2	Campbell and Lamar (2004)
<i>Crotalus triseriatus</i>	532.2	8	<i>This study</i>	1	Campbell and Lamar (2004)
<i>Crotalus tzcaban</i>	1082.7	3	<i>This study</i>	1	Campbell and Lamar (2004)
<i>Crotalus viridis</i>	950.8	5	<i>This study</i>	1	Campbell and Lamar (2004)
<i>Crotalus willardi</i>	625	8	<i>This study</i>	1	Campbell and Lamar (2004)
<i>Daboia deserti</i>	897	5	<i>This study</i>	2	Mallow et al. (2003), Phelps (2010)
<i>Daboia mauritanica</i>	851.8	4	<i>This study</i>	2	Mallow et al. (2003), Phelps (2010)
<i>Daboia palaestinae</i>	771.8	5	<i>This study</i>	3	Phelps (2010)
<i>Daboia russelii</i>	1076.8	5	<i>This study</i>	1	Mallow et al. (2003), Phelps (2010)
<i>Daboia siamensis</i>	921.4	5	<i>This study</i>	1	Phelps (2010), Wogan (2012)
<i>Deinagkistrodon acutus</i>	978	8	<i>This study</i>	1	Orlov et al. (2002)
<i>Echis borkini</i>	-	-	-	2	G. Vogel pers. comm.
<i>Echis carinatus</i>	512.8	5	<i>This study</i>	2	Mallow et al. (2003), Phelps (2010)
<i>Echis coloratus</i>	571.6	8	<i>This study</i>	2	Mallow et al. (2003), Phelps (2010)
<i>Echis jogeri</i>	280*	1	<i>This study</i>	1	Phelps (2010)
<i>Echis khosatzkii</i>	422	1	<i>This study</i>	2	Mallow et al. (2003), Els and Rasbi (2012)
<i>Echis leucogaster</i>	470.4	7	<i>This study</i>	2	Mallow et al. (2003), Phelps (2010)
<i>Echis ocellatus</i>	473.2	7	<i>This study</i>	1	Mallow et al. (2003), Phelps (2010)
<i>Echis omanensis</i>	519	5	<i>This study</i>	2	Phelps (2010)
<i>Echis pyramidum</i>	461.7	6	<i>This study</i>	2	Mallow et al. (2003), Phelps (2010)
<i>Eristicophis macmahoni</i>	534.2	5	<i>This study</i>	2	Mallow et al. (2003), Phelps (2010)

<i>Garthius chaseni</i>	420	1	<i>This study</i>	3	Das (2012)
<i>Gloydus blomhoffii</i>	475.6	8	<i>This study</i>	2	Gloyd and Conant (1990), Hampton (2011), O'Shea (2011)
<i>Gloydus brevicaudus</i>	532.3	7	<i>This study</i>	2	Gloyd and Conant (1990)
<i>Gloydus halys</i>	555.7	3	<i>This study</i>	2	Orlov et al. (2002)
<i>Gloydus intermedius</i>	569.8	4	<i>This study</i>	3	Orlov et al. (2002)
<i>Gloydus liupanensis</i>	-	-	-	2	Gloyd and Conant (1990), P. David pers. comm.
<i>Gloydus saxatilis</i>	635.4	5	<i>This study</i>	3	Gloyd and Conant (1990), Deufel and Cundall (2006)
<i>Gloydus shedaensis</i>	639.8	4	Shine et al. (2002), (2003)	4	Shine et al. (2002)
<i>Gloydus strauchi</i>	416.8	4	<i>This study</i>	2	Gloyd and Conant (1990), Hampton (2011), P. David pers. comm.
<i>Gloydus tsushimaensis</i>	-	-	-	3	Gloyd and Conant (1990)
<i>Gloydus ussuriensis</i>	453.6	5	<i>This study</i>	2	Orlov et al. (2002)
<i>Hypnale hypnale</i>	308.4	5	<i>This study</i>	3	Maduwage et al. (2009), Sawant et al. (2010)
<i>Hypnale nepa</i>	264.7	3	<i>This study</i>	3	Maduwage et al. (2009)
<i>Hypnale zara</i>	325	1	<i>This study</i>	3	Maduwage et al. (2009)
<i>Lachesis acrochorda</i>	1800.5	2	<i>This study</i>	3	Campbell and Lamar (2004)
<i>Lachesis melanocephala</i>	1925.5*	2	<i>This study</i>	3	Campbell and Lamar (2004), Pla et al. (2013)
<i>Lachesis muta</i>	1932.3	9	<i>This study</i>	3	Martins and Oliveira (1998), Campbell and Lamar (2004)
<i>Lachesis stenophrys</i>	1929.8	5	<i>This study</i>	3	Campbell and Lamar (2004), Deufel and Cundall (2006)
<i>Macrovipera lebetina</i>	956.8	6	<i>This study</i>	2	Mallow et al. (2003), Phelps (2010)
<i>Macrovipera schweizeri</i>	557.5	2	<i>This study</i>	2	Mallow et al. (2003), Phelps (2010)
<i>Mixcoatlus barbouri</i>	441.7	3	<i>This study</i>	3	Campbell and Lamar (2004), Jadin et al. (2011)
<i>Mixcoatlus browni</i>	351	1	<i>This study</i>	3	Campbell and Lamar (2004), Jadin et al. (2011)

<i>Mixcoatlus melanurus</i>	442.3	6	<i>This study</i>	1	Campbell and Lamar (2004), Jadin et al. (2011)
<i>Montivipera albizona</i>	644.7	6	Mulder (1994)	2	Phelps (2010), G. Nilson pers. comm., P. David pers. comm.
<i>Montivipera bornmuelleri</i>	505.7	3	<i>This study</i>	2	Mallow et al. (2003), Phelps (2010)
<i>Montivipera latifii</i>	622	4	<i>This study</i>	2	Mallow et al. (2003), Phelps (2010)
<i>Montivipera raddei</i>	730.3	4	<i>This study</i>	2	Phelps (2010), Hampton (2011)
<i>Montivipera wagneri</i>	575	1	<i>This study</i>	2	Phelps (2010), G. Nilson pers. comm., P. David pers. comm.
<i>Montivipera xanthina</i>	800.1	7	<i>This study</i>	2	Böhme et al. (2009), Wilson and Grillitsch (2009), Phelps (2010), Hampton (2011)
<i>Ophryacus undulatus</i>	639.4	9	<i>This study</i>	4	Campbell and Lamar (2004), Hampton (2011), Jadin et al. (2011)
<i>Ovophis monticola</i>	535.3	6	<i>This study</i>	2	Orlov et al. (2002), Leviton et al. (2003)
<i>Ovophis okinavensis</i>	544.8	5	<i>This study</i>	3	Kadota (2011)
<i>Ovophis tonkinensis</i>	529	2	<i>This study</i>	3	Orlov et al. (2002)
<i>Ovophis zayuensis</i>	626	1	<i>This study</i>	3	Orlov et al. (2002), Guo et al. (2012)
<i>Porthidium arcosae</i>	466*	1	<i>This study</i>	3	Campbell and Lamar (2004), Valencia et al. (2010)
<i>Porthidium dunni</i>	374.6	18	<i>This study</i>	3	Campbell and Lamar (2004)
<i>Porthidium hespere</i>	605*	1	<i>This study</i>	3	Campbell (1976), Campbell and Lamar (2004)
<i>Porthidium lansbergii</i>	375.8	8	<i>This study</i>	3	Campbell and Lamar (2004)
<i>Porthidium nasutum</i>	359.2	5	<i>This study</i>	3	Campbell and Lamar (2004)
<i>Porthidium ophryomegas</i>	485.1	9	<i>This study</i>	3	Campbell and Lamar (2004)
<i>Porthidium porrasi</i>	402.3*	3	<i>This study</i>	3	Lamar and Sasa (2003), Campbell and Lamar (2004)
<i>Porthidium yucatanicum</i>	423.4	5	<i>This study</i>	3	Campbell and Lamar (2004)
<i>Proatheris superciliaris</i>	458.5	2	<i>This study</i>	2	Branch (2001), Phelps (2010)
<i>Protobothrops cornutus</i>	349	1	<i>This study</i>	4	Orlov et al. (2002)

<i>Protobothrops dabieshanensis</i>	696*	1	Huang et al. (2012)	1	Huang et al. (2012)
<i>Protobothrops elegans</i>	927.7	6	<i>This study</i>	3	G. Vogel pers. comm.
<i>Protobothrops flavoviridis</i>	1391.3	6	<i>This study</i>	4	Aird et al. (2013), P. David pers. comm.
<i>Protobothrops jerdonii</i>	865.3	7	<i>This study</i>	4	Orlov et al. (2002), Leviton et al. (2003)
<i>Protobothrops kaulbacki</i>	1219.8	4	<i>This study</i>	1	Leviton et al. (2003), Guo et al. (2012), P. David pers. comm.
<i>Protobothrops mangshanensis</i>	-	-	-	4	Zhou (2012), Gong et al. (2013)
<i>Protobothrops maolanensis</i>	588	3	Yang et al. (2011)	2	Yang et al. (2011), Wang and Lau (2012)
<i>Protobothrops mucrosquamatus</i>	815.1	8	<i>This study</i>	3	Orlov et al. (2002)
<i>Protobothrops sieversorum</i>	1060	1	Herrmann et al. (2002)	3	Herrmann et al. (2002)
<i>Protobothrops tokarensis</i>	818.5	2	<i>This study</i>	4	P. David pers. comm.
<i>Protobothrops xiangchengensis</i>	-	-	-	2	Guo and Jiang (2012)
<i>Pseudocerastes persicus</i>	735.8	4	<i>This study</i>	2	Mallow et al. (2003), Phelps (2010)
<i>Pseudocerastes urarachnoides</i>	582*	1	<i>This study</i>	2	Anderson and Papenfuss (2009), Phelps (2010)
<i>Pseudocerastes fieldi</i>	673.8	5	<i>This study</i>	2	Phelps (2010)
<i>Sistrurus catenatus</i>	663.3	9	<i>This study</i>	2	Holycross and Mackessy (2002), Campbell and Lamar (2004)
<i>Sistrurus miliarius</i>	525.3	6	<i>This study</i>	1	May et al. (1996), Campbell and Lamar (2004)
<i>Trimeresurus albolabris</i>	526.4	5	<i>This study</i>	4	Vijayakumar and David (2006), Whitaker (2008)
<i>Trimeresurus andersonii</i>	502.3	3	<i>This study</i>	4	Whitaker (2008)
<i>Trimeresurus barati</i>	511.5	2	<i>This study</i>	4	Iskandar et al. (2012)
<i>Trimeresurus borneensis</i>	472.5	2	<i>This study</i>	4	David et al. (2006), Deufel and Cundall (2006)
<i>Trimeresurus buniana</i>	649	1	<i>This study</i>	4	Grismer (2012)
<i>Trimeresurus cantori</i>	582.3	4	<i>This study</i>	4	Vijayakumar and David (2006), Whitaker (2008)

<i>Trimeresurus erythrurus</i>	534.2	6	<i>This study</i>	4	Orlov et al. (2002)
<i>Trimeresurus fasciatus</i>	437	1	<i>This study</i>	4	Deufel and Cundall (2006), P. David pers. comm.
<i>Trimeresurus flavomaculatus</i>	709.3	3	<i>This study</i>	4	Sanders et al. (2004), Deufel and Cundall (2006)
<i>Trimeresurus fucatus</i>	616.3	3	<i>This study</i>	4	Chanhome et al. (2011), Wogan et al. (2012)
<i>Trimeresurus gracilis</i>	397	1	Lin and Tu (2008)	4	Zhao (2006), Jiang and Lau (2012)
<i>Trimeresurus gramineus</i>	531.7	7	<i>This study</i>	4	Whitaker (2008), Sawant and Jadhav (2013)
<i>Trimeresurus gumprechtii</i>	574.8	4	<i>This study</i>	4	Chanhome et al. (2011), Stuart et al. (2012)
<i>Trimeresurus hageni</i>	650.8	6	<i>This study</i>	4	Ryabov et al. (2002), Sanders et al. (2004)
<i>Trimeresurus insularis</i>	513.3	4	<i>This study</i>	4	Auliya (2010), G. Vogel pers. comm.
<i>Trimeresurus kanburiensis</i>	428	1	<i>This study</i>	4	Chanhome et al. (2011)
<i>Trimeresurus macrops</i>	471.9	7	<i>This study</i>	4	Stuart et al. (2012)
<i>Trimeresurus malabaricus</i>	455.8	5	<i>This study</i>	4	Whitaker (2008), Sawant and Jadhav (2013)
<i>Trimeresurus malcolmi</i>	1030.5*	2	<i>This study</i>	4	Sanders et al. (2004)
<i>Trimeresurus mcgregori</i>	588.8	5	<i>This study</i>	4	Sanders et al. (2004)
<i>Trimeresurus medoensis</i>	493.2	9	David et al. (2001)	4	Leviton et al. (2003), Whitaker (2008)
<i>Trimeresurus nebularis</i>	782	1	<i>This study</i>	4	Grismer (2012)
<i>Trimeresurus popeiorum</i>	584.2	6	<i>This study</i>	4	Orlov et al. (2002)
<i>Trimeresurus puniceus</i>	501.5	6	<i>This study</i>	4	Orlov et al. (2002)
<i>Trimeresurus purpureomaculatus</i>	655.2	6	<i>This study</i>	4	Leviton et al. (2003), Vogel et al. (2012)
<i>Trimeresurus sabahi</i>	654	2	<i>This study</i>	4	Orlov et al. (2002)
<i>Trimeresurus schultzei</i>	746	2	<i>This study</i>	4	Sanders et al. (2004)
<i>Trimeresurus septentrionalis</i>	634*	3	<i>This study</i>	4	Orlov et al. (2002)

<i>Trimeresurus sichuanensis</i>	-	-	-	4	Guo and Wang (2011), P. Guo pers. comm.
<i>Trimeresurus stejnegeri</i>	512	9	<i>This study</i>	4	Gumpretchi (1997), Creer et al. (2002), Lin et al. (2006)
<i>Trimeresurus sumatranus</i>	862.5	2	<i>This study</i>	4	Ryabov et al. (2002), Sanders et al. (2004)
<i>Trimeresurus tibetanus</i>	528	1	<i>This study</i>	4	Orlov et al. (2002)
<i>Trimeresurus trigonocephalus</i>	594	6	<i>This study</i>	4	Gumpretchi (1999), Deufel and Cundall (2006)
<i>Trimeresurus truongsongensis</i>	498.5	4	Orlov et al. (2004)	4	Orlov et al. (2004), Ziegler et al. (2006)
<i>Trimeresurus venustus</i>	390	1	<i>This study</i>	4	Chanhome et al. (2011)
<i>Trimeresurus vogeli</i>	722.6	5	<i>This study</i>	4	Chanhome et al. (2011), Stuart and Nguyen (2012)
<i>Trimeresurus wiroti</i>	485	1	<i>This study</i>	4	Chanhome et al. (2011), Grismer and Chan-Ard (2012), G. Vogel pers. comm.
<i>Trimeresurus yunnanensis</i>	660.8	4	<i>This study</i>	4	Leviton et al. (2003), P. David pers. comm.
<i>Tropidolaemus subannulatus</i>	613.7*	6	<i>This study</i>	4	Auliya et al. (2012)
<i>Tropidolaemus wagleri</i>	587.7	3	<i>This study</i>	4	Orlov et al. (2002), Grismer and Chan-Ard (2012)
<i>Vipera ammodytes</i>	608.8	11	<i>This study</i>	2	Phelps (2010)
<i>Vipera anatolica</i>	-	-	-	1	Tok et al. (2009), Phelps (2010)
<i>Vipera aspis</i>	541.5	12	<i>This study</i>	2	Luiselli et al. (2007), Phelps (2010), Hampton (2011)
<i>Vipera barani</i>	449.3	3	Franzes and Heckes (2000), Avci et al. (2004)	1	Mallow et al. (2003), Tok et al. (2009), Phelps (2010), Hampton (2011)
<i>Vipera berus</i>	456.8	12	<i>This study</i>	2	Phelps (2010), Hampton (2011)
<i>Vipera dinniki</i>	367*	1	<i>This study</i>	2	Mallow et al. (2003), Phelps (2010)
<i>Vipera eriwanensis</i>	327.5	4	<i>This study</i>	2	Tuniyev et al. (2009), Phelps (2010)
<i>Vipera kaznakovi</i>	409.5	4	<i>This study</i>	1	Mallow et al. (2003), Tuniyev et al. (2009), Phelps (2010)
<i>Vipera latastei</i>	453.4	5	<i>This study</i>	1	Brito (2003), Phelps (2010)
<i>Vipera lotievi</i>	-	-	-	1	Mallow et al. (2003), Tuniyev et al. (2009), Phelps (2010)

<i>Vipera nikolskii</i>	410	1	<i>This study</i>	1	Zinenko (2006), Phelps (2010)
<i>Vipera orlovi</i>	380	1	Tuniyev and Ostrovskikh (2001)	1	Tuniyev and Ostrovskikh (2001), Tuniyev et al. (2012), Tuniyev pers. comm.
<i>Vipera renardi</i>	484.44	9	<i>This study</i>	2	Höggren et al. (1993), Phelps (2010)
<i>Vipera seoanei</i>	428.6	7	<i>This study</i>	1	Mallow et al. (2003), Phelps (2010), Brito pers. comm.
<i>Vipera transcaucasiana</i>	571	1	<i>This study</i>	2	Mallow et al. (2003), Phelps (2010), Hampton (2011)
<i>Vipera ursinii</i>	388.2	5	<i>This study</i>	2	Filippi and Luiselli (2004), Luiselli et al. (2006), Phelps (2010)

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Table S2. Δ AICc calculated for continuous-time Markov models of trait evolution fitted to each simmap used in the analyses. ER = equal rates model, SYM = symmetrical rates model, ARD = all rates different model. * denotes simmaps for which we used the second best Markov model fitted.

SIMMAP	Δ AICc		
	ER	SYM	ARD
1	35.84	5.75	0.00
2	35.88	6.96	0.00
3	34.54	5.70	0.00
4	36.94	4.89	0.00
5	32.50	3.85	0.00
6	38.08	5.37	0.00
7	37.51	5.71	0.00
8	35.54	4.32	0.00
9	35.39	5.20	0.00
10	32.97	6.07	0.00
11	31.24	3.76	0.00
12	36.32	4.79	0.00
13	35.07	4.26	0.00
14	34.37	5.75	0.00
15	33.27	4.68	0.00
16	33.74	4.23	0.00
17	36.64	4.32	0.00
18	33.25	5.47	0.00
19*	30.96	4.75	0.00
20	31.51	3.23	0.00
21	30.36	4.48	0.00
22	37.38	5.34	0.00
23	33.19	5.74	0.00
24	35.51	5.99	0.00
25	34.27	4.55	0.00
26	35.31	7.63	0.00
27	32.93	5.22	0.00
28	30.46	5.27	0.00
29	28.56	3.42	0.00
30	34.83	5.08	0.00
31	36.20	4.86	0.00
32	34.64	5.19	0.00
33	31.87	4.00	0.00
34	39.14	6.38	0.00
35	32.18	5.10	0.00
36	34.97	4.71	0.00
37	38.46	6.48	0.00
38	35.48	5.35	0.00
39*	36.69	5.36	0.00
40	32.59	4.01	0.00

41	32.19	4.24	0.00
42	35.28	6.07	0.00
43	35.50	5.23	0.00
44*	32.08	4.10	0.00
45	33.78	4.31	0.00
46	38.02	6.03	0.00
47	32.40	4.02	0.00
48*	34.42	5.33	0.00
49	35.87	4.11	0.00
50	32.16	3.44	0.00
51	35.86	7.67	0.00
52	35.49	4.86	0.00
53	31.02	4.35	0.00
54	36.59	6.17	0.00
55	36.75	5.59	0.00
56	32.62	4.84	0.00
57	32.18	4.05	0.00
58	32.48	4.91	0.00
59	33.28	5.33	0.00
60	30.05	5.32	0.00
61	26.22	1.92	0.00
62	30.40	4.03	0.00
63	36.64	5.36	0.00
64	28.69	0.88	0.00
65*	36.90	6.63	0.00
66	31.50	5.22	0.00
67	29.70	3.64	0.00
68	33.41	3.55	0.00
69	35.29	4.67	0.00
70	32.68	4.67	0.00
71	27.30	0.00	4.23
72	31.83	3.82	0.00
73	33.22	4.19	0.00
74	35.18	6.16	0.00
75	34.64	4.86	0.00
76	34.89	5.12	0.00
77	32.75	4.85	0.00
78	34.26	5.33	0.00
79*	34.50	5.03	0.00
80	31.76	2.43	0.00
81	33.27	3.47	0.00
82	32.40	4.38	0.00
83	30.90	4.61	0.00
84	36.51	6.45	0.00
85	33.29	2.97	0.00
86	31.96	4.63	0.00
87	36.45	6.88	0.00
88	35.63	6.00	0.00

89	35.78	5.06	0.00
90	36.21	6.69	0.00
91	37.63	6.19	0.00
92	36.93	7.05	0.00
93	31.62	4.46	0.00
94	33.70	5.44	0.00
95	38.00	4.49	0.00
96*	38.35	6.63	0.00
97	35.61	3.61	0.00
98	34.90	5.88	0.00
99	32.05	4.50	0.00
100	32.74	4.77	0.00

APPENDIX I

Museum specimens analyzed in the present study. UTA = University of Texas at Arlington; CAS = California Academy of Sciences; FMNH = Field Museum of Natural History; AMNH = American Museum of Natural History; USNM = Smithsonian National Museum of Natural History; BMNH = Natural History Museum of London; MNHN = Muséum national d'Histoire naturelle; FMNH = Florida Museum of Natural History; MNRJ = Museu Nacional do Rio de Janeiro; IBSP = Instituto Butantan; MZUSP = Museu de Zoologia da Universidade de São Paulo; ZRC = The Raffles Museum of Biodiversity Research of the National University of Singapore.

Species	Museum	Catalog number
<i>Agkistrodon bilineatus</i>	UTA	R6359
<i>Agkistrodon bilineatus</i>	UTA	R2449
<i>Agkistrodon bilineatus</i>	CAS	204153
<i>Agkistrodon bilineatus</i>	FMNH	39093
<i>Agkistrodon contortrix</i>	AMNH	128682
<i>Agkistrodon contortrix</i>	AMNH	128678
<i>Agkistrodon contortrix</i>	AMNH	95950
<i>Agkistrodon contortrix</i>	AMNH	134673
<i>Agkistrodon contortrix</i>	AMNH	146527
<i>Agkistrodon howardgloydi</i>	UTA	R59855
<i>Agkistrodon piscivorus</i>	USNM	203553
<i>Agkistrodon piscivorus</i>	USNM	203565
<i>Agkistrodon piscivorus</i>	USNM	203566
<i>Agkistrodon piscivorus</i>	AMNH	107946
<i>Agkistrodon piscivorus</i>	AMNH	90128
<i>Agkistrodon piscivorus</i>	AMNH	161861
<i>Agkistrodon taylori</i>	UTA	R8117
<i>Agkistrodon taylori</i>	FMNH	28794
<i>Agkistrodon taylori</i>	AMNH	124948
<i>Atheris ceratophora</i>	BMNH	1960.1.6.54
<i>Atheris ceratophora</i>	CAS	173812
<i>Atheris ceratophora</i>	CAS	162615
<i>Atheris chlorechis</i>	BMNH	1907.4.18.6
<i>Atheris chlorechis</i>	UTA	R58204
<i>Atheris desaixi</i>	BMNH	1969.2419
<i>Atheris desaixi</i>	BMNH	979
<i>Atheris desaixi</i>	BMNH	980
<i>Atheris desaixi</i>	UTA	R50162
<i>Atheris hispida</i>	CAS	141750
<i>Atheris hispida</i>	CAS	122747
<i>Atheris hispida</i>	CAS	147906
<i>Atheris hispida</i>	FMNH	154900
<i>Atheris hispida</i>	AMNH	11873

<i>Atheris_nitschei</i>	BMNH	1954.1.12.65-66
<i>Atheris_nitschei</i>	MNHParis	1980.1278
<i>Atheris_nitschei</i>	MNHParis	1981.476
<i>Atheris_nitschei</i>	MNHParis	1934.14
<i>Atheris_nitschei</i>	CAS	201654
<i>Atheris_nitschei</i>	CAS	201655
<i>Atheris_squamigera</i>	BMNH	1934.12.1.35
<i>Atheris_squamigera</i>	BMNH	1901.6.24.65-66
<i>Atheris_squamigera</i>	BMNH	1908.5.30.35
<i>Atheris_squamigera</i>	BMNH	1908.5.30.33
<i>Atheris_squamigera</i>	BMNH	1969.2554-2556
<i>Atheris_squamigera</i>	BMNH	1908.10.20.25
<i>Atheris_squamigera</i>	BMNH	1907.5.22.66
<i>Atheris_squamigera</i>	BMNH	1906.5.28.84
<i>Atropoides_indomitus</i>	UTA	R52952
<i>Atropoides_mexicanus</i>	UTA	R6738
<i>Atropoides_mexicanus</i>	UTA	R12943
<i>Atropoides_mexicanus</i>	UTA	R43610
<i>Atropoides_mexicanus</i>	UTA	R45500
<i>Atropoides_mexicanus</i>	USNM	578902
<i>Atropoides_mexicanus</i>	FLMNH	119441
<i>Atropoides_mexicanus</i>	FLMNH	141977
<i>Atropoides_mexicanus</i>	FLMNH	142074
<i>Atropoides_mexicanus</i>	FLMNH	144757
<i>Atropoides_nummifer</i>	UTA	R24842
<i>Atropoides_nummifer</i>	UTA	R6206
<i>Atropoides_nummifer</i>	UTA	R12562
<i>Atropoides_nummifer</i>	UTA	R51731
<i>Atropoides_occiduus</i>	UTA	R34158
<i>Atropoides_occiduus</i>	UTA	R12785
<i>Atropoides_occiduus</i>	UTA	R9089
<i>Atropoides_occiduus</i>	UTA	R31979
<i>Atropoides_olmec</i>	UTA	R16964
<i>Atropoides_olmec</i>	UTA	R25113
<i>Atropoides_olmec</i>	UTA	R18691
<i>Atropoides_picadoi</i>	BMNH	1913.7.19.1.47
<i>Atropoides_picadoi</i>	UTA	R18215
<i>Atropoides_picadoi</i>	UTA	R24834
<i>Atropoides_picadoi</i>	UTA	R24835
<i>Atropoides_picadoi</i>	UTA	R32080
<i>Azemiops_feae</i>	MNHParis	1953.463
<i>Azemiops_feae</i>	UTA	R35599
<i>Azemiops_feae</i>	UTA	R32069
<i>Azemiops_feae</i>	UTA	R18701
<i>Azemiops_feae</i>	FMNH	218627
<i>Azemiops_feae</i>	FMNH	152987
<i>Bitis_arietans</i>	BMNH	1936.7.3.39
<i>Bitis_arietans</i>	BMNH	1916.3.20.1

<i>Bitis_arietans</i>	BMNH	1906.8.25.3
<i>Bitis_arietans</i>	BMNH	90.2.26.15-17
<i>Bitis_arietans</i>	BMNH	1916.3.20.1
<i>Bitis_arietans</i>	BMNH	1975.1113
<i>Bitis_arietans</i>	BMNH	1975.1097
<i>Bitis_arietans</i>	BMNH	1948.1.2.35
<i>Bitis_arietans</i>	MNHParis	1904.205
<i>Bitis_armata</i>	FMNH	206280
<i>Bitis_atropos</i>	BMNH	1908.3.26.2
<i>Bitis_atropos</i>	MNHParis	323
<i>Bitis_atropos</i>	UTA	R46188
<i>Bitis_atropos</i>	UTA	R7325
<i>Bitis_atropos</i>	CAS	135749
<i>Bitis_caudalis</i>	BMNH	1969.2639
<i>Bitis_caudalis</i>	BMNH	1988.583
<i>Bitis_caudalis</i>	BMNH	65.5.4.153A
<i>Bitis_caudalis</i>	BMNH	65.5.4.153B
<i>Bitis_caudalis</i>	CAS	214702
<i>Bitis_cornuta</i>	UTA	R34491
<i>Bitis_cornuta</i>	UTA	R38536
<i>Bitis_cornuta</i>	UTA	R34492
<i>Bitis_cornuta</i>	UTA	R26613
<i>Bitis_gabonica</i>	BMNH	1969.526
<i>Bitis_gabonica</i>	BMNH	1908.3.4.2
<i>Bitis_gabonica</i>	MNHParis	1988.2441
<i>Bitis_gabonica</i>	MNHParis	1988.50
<i>Bitis_gabonica</i>	FMNH	216460
<i>Bitis_gabonica</i>	FMNH	30874
<i>Bitis_nasicornis</i>	S/N	
<i>Bitis_nasicornis</i>	BMNH	1955.1.11.68-69
<i>Bitis_nasicornis</i>	BMNH	1932.2.3.1
<i>Bitis_nasicornis</i>	MNHParis	1899
<i>Bitis_nasicornis</i>	MNHParis	1988.3925
<i>Bitis_nasicornis</i>	MNHParis	1988.3923
<i>Bitis_peringueyi</i>	BMNH	1976.263
<i>Bitis_peringueyi</i>	UTA	R25696
<i>Bitis_peringueyi</i>	UTA	R25700
<i>Bitis_peringueyi</i>	UTA	R25697
<i>Bitis_peringueyi</i>	UTA	R39081
<i>Bitis_peringueyi</i>	CAS	111963
<i>Bitis_rhinoceros</i>	MNHParis	1986.105
<i>Bitis_schneideri</i>	BMNH	1967.594
<i>Bitis_schneideri</i>	UTA	R25701
<i>Bitis_xeropaga</i>	BMNH	1976.87
<i>Bitis_xeropaga</i>	FMNH	233126
<i>Bitis_worthingtoni</i>	BMNH	1952.1.2.30-31
<i>Bitis_worthingtoni</i>	BMNH	1952.1.2.30-31
<i>Bitis_worthingtoni</i>	FMNH	58141

<i>Bothriechis_aurifer</i>	UTA	R39070
<i>Bothriechis_aurifer</i>	UTA	R50585
<i>Bothriechis_aurifer</i>	UTA	R51728
<i>Bothriechis_aurifer</i>	UTA	R6525
<i>Bothriechis_aurifer</i>	UTA	R10434
<i>Bothriechis_aurifer</i>	UTA	R26574
<i>Bothriechis_aurifer</i>	UTA	R36211
<i>Bothriechis_bicolor</i>	UTA	R24759
<i>Bothriechis_bicolor</i>	UTA	R34156
<i>Bothriechis_bicolor</i>	UTA	R2274
<i>Bothriechis_bicolor</i>	FMNH	20162
<i>Bothriechis_guifarroi</i>	UTA	R60303
<i>Bothriechis_lateralis</i>	BMNH	96.10.8.18-19
<i>Bothriechis_lateralis</i>	UTA	R51155
<i>Bothriechis_lateralis</i>	UTA	R51154
<i>Bothriechis_lateralis</i>	UTA	R7634
<i>Bothriechis_lateralis</i>	UTA	R16351
<i>Bothriechis_lateralis</i>	FLMNH	54148
<i>Bothriechis_lateralis</i>	FLMNH	39820
<i>Bothriechis_lateralis</i>	FLMNH	70572
<i>Bothriechis_lateralis</i>	FLMNH	85322
<i>Bothriechis_marchi</i>	UTA	R13034
<i>Bothriechis_marchi</i>	UTA	R7159
<i>Bothriechis_marchi</i>	UTA	R13617
<i>Bothriechis_marchi</i>	UTA	R32068
<i>Bothriechis_marchi</i>	UTA	R50571
<i>Bothriechis_marchi</i>	FMNH	37217
<i>Bothriechis_marchi</i>	FMNH	21892
<i>Bothriechis_marchi</i>	AMNH	46957
<i>Bothriechis_marchi</i>	FLMNH	144679
<i>Bothriechis_marchi</i>	FLMNH	52555
<i>Bothriechis_marchi</i>	FLMNH	52554
<i>Bothriechis_nigroviridis</i>	UTA	R10433
<i>Bothriechis_nigroviridis</i>	UTA	R24841
<i>Bothriechis_nigroviridis</i>	FMNH	154530
<i>Bothriechis_nigroviridis</i>	FMNH	154515
<i>Bothriechis_nigroviridis</i>	USNM	32580
<i>Bothriechis_nigroviridis</i>	AMNH	142203
<i>Bothriechis_rowleyi</i>	UTA	R23954
<i>Bothriechis_rowleyi</i>	UTA	R6207
<i>Bothriechis_rowleyi</i>	UTA	R22243
<i>Bothriechis_rowleyi</i>	UTA	R13619
<i>Bothriechis_rowleyi</i>	UTA	R25703
<i>Bothriechis_rowleyi</i>	UTA	R6636
<i>Bothriechis_rowleyi</i>	AMNH	102895
<i>Bothriechis_rowleyi</i>	AMNH	102894
<i>Bothriechis_rowleyi</i>	FLMNH	52553
<i>Bothriechis_schlegelii</i>	MNRJ	8794

<i>Bothriechis schlegelii</i>	UTA	R12957
<i>Bothriechis schlegelii</i>	USNM	562913
<i>Bothriechis schlegelii</i>	AMNH	35780
<i>Bothriechis schlegelii</i>	FLMNH	157577
<i>Bothriechis superciliaris</i>	UTA	R35246
<i>Bothriechis thalassinus</i>	UTA	R42259
<i>Bothriechis thalassinus</i>	UTA	R38891
<i>Bothriechis thalassinus</i>	UTA	R44438
<i>Bothrocophias campbelli</i>	AMNH	22094
<i>Bothrocophias hyoprora</i>	FMNH	165849
<i>Bothrocophias hyoprora</i>	FMNH	56171
<i>Bothrocophias hyoprora</i>	USNM	S/N
<i>Bothrocophias hyoprora</i>	AMNH	54601
<i>Bothrocophias hyoprora</i>	AMNH	53369
<i>Bothrocophias hyoprora</i>	AMNH	52808
<i>Bothrocophias hyoprora</i>	AMNH	52265
<i>Bothrocophias hyoprora</i>	AMNH	52098
<i>Bothrocophias hyoprora</i>	AMNH	113642
<i>Bothrocophias microphthalmus</i>	BMNH	58.7.25.13
<i>Bothrocophias microphthalmus</i>	AMNH	52470
<i>Bothrops alcatraz</i>	IBSP	85710
<i>Bothrops alcatraz</i>	IBSP	57105
<i>Bothrops alcatraz</i>	IBSP	55791
<i>Bothrops alternatus</i>	MNRJ	8608
<i>Bothrops alternatus</i>	MNRJ	22192
<i>Bothrops alternatus</i>	MNRJ	8602
<i>Bothrops alternatus</i>	MNRJ	22094
<i>Bothrops alternatus</i>	MNRJ/AL	5553
<i>Bothrops alternatus</i>	IBSP	85656
<i>Bothrops alternatus</i>	IBSP	85659
<i>Bothrops alternatus</i>	IBSP	84796
<i>Bothrops alternatus</i>	IBSP	84797
<i>Bothrops ammodytoides</i>	BMNH	72.11.4.2
<i>Bothrops ammodytoides</i>	FMNH	10832
<i>Bothrops ammodytoides</i>	USNM	73421
<i>Bothrops ammodytoides</i>	AMNH	110143
<i>Bothrops asper</i>	MNRJ	8757
<i>Bothrops asper</i>	MNRJ	8758
<i>Bothrops asper</i>	MNRJ	8759
<i>Bothrops asper</i>	UTA	R50466
<i>Bothrops atrox</i>	MNRJ	15003
<i>Bothrops atrox</i>	MNRJ	16821
<i>Bothrops atrox</i>	MNRJ	16825
<i>Bothrops atrox</i>	MNRJ	230
<i>Bothrops atrox</i>	MNRJ	8577
<i>Bothrops atrox</i>	MNRJ	974
<i>Bothrops barnetti</i>	BMNH	1946.1.18.81
<i>Bothrops barnetti</i>	BMNH	1946.1.18.71

<i>Bothrops barnetti</i>	FMNH	9787
<i>Bothrops barnetti</i>	FMNH	41603
<i>Bothrops bilineata</i>	MNRJ	3434
<i>Bothrops bilineata</i>	MNRJ	3424
<i>Bothrops bilineata</i>	MNRJ	3044
<i>Bothrops brazili</i>	USNM	316671
<i>Bothrops brazili</i>	USNM	345184
<i>Bothrops caribbaeus</i>	UTA	R16311
<i>Bothrops caribbaeus</i>	USNM	15082
<i>Bothrops caribbaeus</i>	AMNH	134681
<i>Bothrops caribbaeus</i>	AMNH	134683
<i>Bothrops chloromelas</i>	AMNH	104298
<i>Bothrops cotiara</i>	MNRJ	10078
<i>Bothrops cotiara</i>	MNRJ/AL	5401A
<i>Bothrops cotiara</i>	IBSP	1100
<i>Bothrops cotiara</i>	IBSP	811
<i>Bothrops cotiara</i>	IBSP	1143
<i>Bothrops cotiara</i>	IBSP	1031
<i>Bothrops cotiara</i>	IBSP	1034
<i>Bothrops diporus</i>	IBSP	84180
<i>Bothrops erythromelas</i>	MNRJ	3068
<i>Bothrops erythromelas</i>	MNRJ	7528
<i>Bothrops erythromelas</i>	MNRJ	3067
<i>Bothrops erythromelas</i>	MNRJ	7527
<i>Bothrops erythromelas</i>	MNRJ	16925
<i>Bothrops erythromelas</i>	MNRJ	9074
<i>Bothrops lutzi</i>	IBSP	552
<i>Bothrops lutzi</i>	IBSP	563
<i>Bothrops fonsecai</i>	MNRJ	19280
<i>Bothrops fonsecai</i>	MNRJ	8601
<i>Bothrops fonsecai</i>	MNRJ	19280
<i>Bothrops fonsecai</i>	MNRJ	8333
<i>Bothrops fonsecai</i>	IBSP	31019
<i>Bothrops fonsecai</i>	IBSP	53962
<i>Bothrops insularis</i>	IBSP	18491
<i>Bothrops insularis</i>	IBSP	1969
<i>Bothrops insularis</i>	IBSP	1882
<i>Bothrops insularis</i>	IBSP	1965
<i>Bothrops insularis</i>	IBSP	1972
<i>Bothrops insularis</i>	IBSP	1960
<i>Bothrops insularis</i>	IBSP	1951
<i>Bothrops insularis</i>	IBSP	18484
<i>Bothrops leucurus</i>	MNRJ	22731
<i>Bothrops leucurus</i>	MNRJ	3943
<i>Bothrops leucurus</i>	MNRJ	9801
<i>Bothrops leucurus</i>	MNRJ	2516
<i>Bothrops leucurus</i>	IBSP	20666
<i>Bothrops leucurus</i>	IBSP	50505

<i>Bothrops_leucurus</i>	IBSP	20663
<i>Bothrops_leucurus</i>	IBSP	20659
<i>Bothrops_itapetiningae</i>	MNRJ	19345
<i>Bothrops_itapetiningae</i>	IBSP	S/N
<i>Bothrops_moojeni</i>	MNRJ	5396
<i>Bothrops_moojeni</i>	MNRJ	7169
<i>Bothrops_moojeni</i>	MNRJ	7120
<i>Bothrops_moojeni</i>	MNRJ	15302
<i>Bothrops_moojeni</i>	MNRJ	7113
<i>Bothrops_moojeni</i>	MNRJ	14029
<i>Bothrops_moojeni</i>	MNRJ	9020
<i>Bothrops_moojeni</i>	MNRJ	8226
<i>Bothrops_moojeni</i>	MNRJ	21876
<i>Bothrops_moojeni</i>	MNRJ	14994
<i>Bothrops_moojeni</i>	MNRJ	7109
<i>Bothrops_moojeni</i>	MNRJ	7122
<i>Bothrops_moojeni</i>	MNRJ	5396
<i>Bothrops_moojeni</i>	MNRJ	10046
<i>Bothrops_moojeni</i>	MNRJ	7499
<i>Bothrops_moojeni</i>	MNRJ	20280
<i>Bothrops_moojeni</i>	MNRJ	7118
<i>Bothrops_moojeni</i>	IBSP	44905
<i>Bothrops_moojeni</i>	IBSP	20249
<i>Bothrops_moojeni</i>	IBSP	44920
<i>Bothrops_jararacussu</i>	MZUSP	3100
<i>Bothrops_jararacussu</i>	MNRJ	17830
<i>Bothrops_jararacussu</i>	MNRJ	21859
<i>Bothrops_jararacussu</i>	MNRJ	20918
<i>Bothrops_jararacussu</i>	MNRJ	14041
<i>Bothrops_jararacussu</i>	IBSP	11559
<i>Bothrops_jararaca</i>	MZUSP	4496
<i>Bothrops_jararaca</i>	MZUSP	2255
<i>Bothrops_jararaca</i>	MZUSP	9479
<i>Bothrops_jararaca</i>	MZUSP	9436
<i>Bothrops_jararaca</i>	MZUSP	8064
<i>Bothrops_jararaca</i>	MZUSP	7572
<i>Bothrops_jararaca</i>	MNRJ	13108
<i>Bothrops_jararaca</i>	MNRJ	20347
<i>Bothrops_jararaca</i>	MNRJ	21094
<i>Bothrops_jararaca</i>	MNRJ	12754
<i>Bothrops_jararaca</i>	MNRJ	10260
<i>Bothrops_jararaca</i>	MNRJ	10592
<i>Bothrops_jararaca</i>	MNRJ	7929
<i>Bothrops_jararaca</i>	MNRJ	932
<i>Bothrops_jararaca</i>	MNRJ	18921
<i>Bothrops_jararaca</i>	MNRJ	17834
<i>Bothrops_jararaca</i>	MNRJ	10077
<i>Bothrops_jararaca</i>	MNRJ	20347

<i>Bothrops_jararaca</i>	MNRJ	7900
<i>Bothrops_jararaca</i>	MNRJ	19948
<i>Bothrops_jararaca</i>	IBSP	82274
<i>Bothrops_jararaca</i>	IBSP	77469
<i>Bothrops_jararaca</i>	IBSP	85415
<i>Bothrops_jararaca</i>	IBSP	85414
<i>Bothrops_jararaca</i>	IBSP	85418
<i>Bothrops_jararaca</i>	IBSP	85417
<i>Bothrops_pauloensis</i>	MNRJ	22694
<i>Bothrops_pauloensis</i>	MNRJ	22697
<i>Bothrops_pauloensis</i>	MNRJ	18557
<i>Bothrops_pauloensis</i>	IBSP	63402
<i>Bothrops_neuwiedi</i>	MNRJ	20696
<i>Bothrops_neuwiedi</i>	MNRJ	19957
<i>Bothrops_neuwiedi</i>	MNRJ	19958
<i>Bothrops_neuwiedi</i>	MNRJ	18955
<i>Bothrops_neuwiedi</i>	MNRJ	16950
<i>Bothrops_neuwiedi</i>	MNRJ	17860
<i>Bothrops_neuwiedi</i>	MNRJ	19398
<i>Bothrops_neuwiedi</i>	IBSP	7990
<i>Bothrops_lanceolatus</i>	MNHParis	1988.104
<i>Bothrops_lanceolatus</i>	MNHParis	1988.108
<i>Bothrops_lanceolatus</i>	MNHParis	1988.110
<i>Bothrops_lanceolatus</i>	MNHParis	1988.106
<i>Bothrops_lojanus</i>	BMNH	1930.10.12.26-27
<i>Bothrops_lojanus</i>	BMNH	1930.10.12.26-27
<i>Bothrops_lojanus</i>	BMNH	1931.11.3.12-14
<i>Bothrops_lojanus</i>	BMNH	1935.11.3.115-119
<i>Bothrops_lojanus</i>	USNM	98935
<i>Bothrops_marmoratus</i>	IBSP	57153
<i>Bothrops_marmoratus</i>	IBSP	57155
<i>Bothrops_matogrossensis</i>	MNRJ	20961
<i>Bothrops_matogrossensis</i>	MNRJ	20963
<i>Bothrops_matogrossensis</i>	MNRJ	20960
<i>Bothrops_matogrossensis</i>	MNRJ	8227
<i>Bothrops_pictus</i>	BMNH	19.11.12.20.12
<i>Bothrops_pictus</i>	MNHParis	6410
<i>Bothrops_pictus</i>	MNHParis	1543
<i>Bothrops_pictus</i>	FMNH	5662
<i>Bothrops_pubescens</i>	MNRJ	12138
<i>Bothrops_pubescens</i>	MNRJ	10079
<i>Bothrops_pulchra</i>	BMNH	78.1.25.54
<i>Bothrops_punctatus</i>	BMNH	1914.5.21.71
<i>Bothrops_punctatus</i>	AMNH	107914
<i>Bothrops_taeniata</i>	MNRJ	10836
<i>Calloselasma_rhodostoma</i>	MNHParis	1988.69
<i>Calloselasma_rhodostoma</i>	MNHParis	1988.70
<i>Calloselasma_rhodostoma</i>	MNHParis	1990.3824

<i>Causus_deflippii</i>	BMNH	1975.1120
<i>Causus_defilippii</i>	BMNH	1952.1.10.81
<i>Causus_defilippii</i>	BMNH	1952.1.10.85
<i>Causus_defilippii</i>	BMNH	1952.1.10.86
<i>Causus_defilippii</i>	BMNH	93.10.26.71-72
<i>Causus_defilippii</i>	BMNH	1983.1089
<i>Causus_defilippii</i>	BMNH	95.2.26.3
<i>Causus_defilippii</i>	MNHParis	1896.596
<i>Causus_defilippii</i>	MNHParis	1896.595
<i>Causus_lichtensteini</i>	BMNH	1919.8.16.99
<i>Causus_lichtensteini</i>	BMNH	1950.1.2.11-12
<i>Causus_lichtensteini</i>	BMNH	1950.1.2.11-12
<i>Causus_lichtensteini</i>	BMNH	1934.12.1.34
<i>Causus_lichtensteini</i>	BMNH	1907.5.22.63-5
<i>Causus_lichtensteini</i>	BMNH	1907.5.22.63-5
<i>Causus_lichtensteini</i>	MNHParis	1973.1245
<i>Causus_resimus</i>	BMNH	1924.2.19.7
<i>Causus_resimus</i>	BMNH	1907.12.2.29
<i>Causus_resimus</i>	BMNH	1912.10.14.7
<i>Causus_resimus</i>	BMNH	1924.5.21.23
<i>Causus_resimus</i>	BMNH	1911.7.8.19
<i>Causus_resimus</i>	MNHParis	1989.39
<i>Causus_resimus</i>	MNHParis	1904.204
<i>Causus_resimus</i>	MNHParis	1988.2115
<i>Causus_rhombeatus</i>	BMNH	48.3.13.3-4
<i>Causus_rhombeatus</i>	BMNH	1929.4.14.16-18
<i>Causus_rhombeatus</i>	BMNH	1954.1.12.54
<i>Causus_rhombeatus</i>	BMNH	95.5.3.53-5
<i>Causus_rhombeatus</i>	BMNH	58.4.11.4
<i>Causus_rhombeatus</i>	MNHParis	1912.382
<i>Causus_rhombeatus</i>	MNHParis	1921.491
<i>Cerastes_cerastes</i>	BMNH	1970.1760
<i>Cerastes_cerastes</i>	BMNH	1912.11.30.130-138
<i>Cerastes_cerastes</i>	BMNH	97.10.26.633
<i>Cerastes_cerastes</i>	MNHParis	4022
<i>Cerastes_cerastes</i>	AMNH	116016
<i>Cerastes_gasperettii</i>	BMNH	1986.351
<i>Cerastes_gasperettii</i>	BMNH	1985.654
<i>Cerastes_gasperettii</i>	BMNH	1994.133
<i>Cerastes_gasperettii</i>	USNM	129454
<i>Cerastes_gasperettii</i>	USNM	129881
<i>Cerastes_vipera</i>	BMNH	1912.11.9.135
<i>Cerastes_vipera</i>	BMNH	1987.2362
<i>Cerastes_vipera</i>	MNHParis	4014
<i>Cerastes_vipera</i>	CAS	138623
<i>Cerastes_vipera</i>	FMNH	81217
<i>Cerastes_vipera</i>	FMNH	22927
<i>Mixcoatlus_barbouri</i>	UTA	R15558

<i>Mixcoatlus_barbouri</i>	UTA	R4450
<i>Mixcoatlus_barbouri</i>	FMNH	38503
<i>Cerrophidion_godmani</i>	UTA	R37293
<i>Cerrophidion_godmani</i>	UTA	R7143
<i>Cerrophidion_godmani</i>	UTA	R42230
<i>Cerrophidion_godmani</i>	UTA	R38272
<i>Cerrophidion_godmani</i>	UTA	R24765
<i>Cerrophidion_godmani</i>	USNM	319985
<i>Cerrophidion_sasai</i>	UTA	R51399
<i>Cerrophidion_sasai</i>	UTA	R51403
<i>Cerrophidion_sasai</i>	UTA	R51402
<i>Cerrophidion_tzotzilorum</i>	UTA	R9641
<i>Cerrophidion_tzotzilorum</i>	AMNH	103133
<i>Crotalus_adamanteus</i>	USNM	250
<i>Crotalus_adamanteus</i>	USNM	104562
<i>Crotalus_adamanteus</i>	AMNH	81997
<i>Crotalus_adamanteus</i>	AMNH	63830
<i>Crotalus_adamanteus</i>	AMNH	160660
<i>Crotalus_adamanteus</i>	AMNH	815537
<i>Crotalus_aquilus</i>	UTA	R12595
<i>Crotalus_aquilus</i>	UTA	R22525
<i>Crotalus_aquilus</i>	UTA	R17904
<i>Crotalus_aquilus</i>	UTA	R6180
<i>Crotalus_aquilus</i>	UTA	R18341
<i>Crotalus_aquilus</i>	UTA	R25407
<i>Crotalus_aquilus</i>	AMNH	64604
<i>Crotalus_atrox</i>	CAS	65086
<i>Crotalus_atrox</i>	CAS	65085
<i>Crotalus_atrox</i>	CAS	33656
<i>Crotalus_atrox</i>	CAS	63879
<i>Crotalus_atrox</i>	CAS	14366
<i>Crotalus_atrox</i>	CAS	65698
<i>Crotalus_basiliscus</i>	CAS	159399
<i>Crotalus_basiliscus</i>	CAS	24095
<i>Crotalus_basiliscus</i>	CAS	74403
<i>Crotalus_basiliscus</i>	AMNH	64254
<i>Crotalus_basiliscus</i>	AMNH	31886
<i>Crotalus_catalinensis</i>	UTA	R32129
<i>Crotalus_cerastes</i>	CAS	183159
<i>Crotalus_cerastes</i>	CAS	192596
<i>Crotalus_cerastes</i>	CAS	223518
<i>Crotalus_cerastes</i>	CAS	182489
<i>Crotalus_cerastes</i>	CAS	182478
<i>Crotalus_cerastes</i>	CAS	101442
<i>Crotalus_cerastes</i>	CAS	12779
<i>Crotalus_cerastes</i>	USNM	205522
<i>Crotalus_cerberus</i>	CAS	17539
<i>Crotalus_cerberus</i>	FMNH	51766

<i>Crotalus_cerberus</i>	FMNH	18423
<i>Crotalus_culminatus</i>	UTA	R41028
<i>Crotalus_culminatus</i>	CAS	71761
<i>Crotalus_culminatus</i>	FMNH	38500
<i>Crotalus_durissus</i>	MNRJ	7133
<i>Crotalus_durissus</i>	MNRJ	1362
<i>Crotalus_durissus</i>	MNRJ	8597
<i>Crotalus_durissus</i>	MNRJ	22949
<i>Crotalus_durissus</i>	MNRJ	8753
<i>Crotalus_durissus</i>	MNRJ	15353
<i>Crotalus_durissus</i>	MNRJ	14995
<i>Crotalus_durissus</i>	MNRJ	21744
<i>Crotalus_durissus</i>	MNRJ	14996
<i>Crotalus_durissus</i>	MNRJ	21850
<i>Crotalus_durissus</i>	MNRJ	21873
<i>Crotalus_durissus</i>	IBSP	83928
<i>Crotalus_durissus</i>	UTA	R21912
<i>Crotalus_durissus</i>	UTA	R12673
<i>Crotalus_enyo</i>	UTA	R8082
<i>Crotalus_enyo</i>	UTA	R7470
<i>Crotalus_enyo</i>	UTA	R7500
<i>Crotalus_enyo</i>	CAS	143853
<i>Crotalus_enyo</i>	CAS	14022
<i>Crotalus_enyo</i>	CAS	103469
<i>Crotalus_enyo</i>	CAS	90317
<i>Crotalus_enyo</i>	CAS	143987
<i>Crotalus_ericsmithi</i>	UTA	R55372
<i>Crotalus_horridus</i>	AMNH	46392
<i>Crotalus_horridus</i>	AMNH	128165
<i>Crotalus_horridus</i>	AMNH	129137
<i>Crotalus_horridus</i>	AMNH	4082
<i>Crotalus_horridus</i>	AMNH	129420
<i>Crotalus_horridus</i>	AMNH	146519
<i>Crotalus_intermedius</i>	UTA	R6232
<i>Crotalus_intermedius</i>	UTA	R4539
<i>Crotalus_intermedius</i>	UTA	R4538
<i>Crotalus_intermedius</i>	UTA	R9354
<i>Crotalus_intermedius</i>	CAS	93777
<i>Crotalus_intermedius</i>	AMNH	102477
<i>Crotalus_intermedius</i>	AMNH	102909
<i>Crotalus_intermedius</i>	AMNH	90799
<i>Crotalus_intermedius</i>	AMNH	72476
<i>Crotalus_lepidus</i>	UTA	R8343
<i>Crotalus_lepidus</i>	UTA	R8331
<i>Crotalus_lepidus</i>	UTA	R17836
<i>Crotalus_lepidus</i>	UTA	R32144
<i>Crotalus_lepidus</i>	UTA	R16285
<i>Crotalus_lepidus</i>	UTA	R7434

<i>Crotalus_lepidus</i>	AMNH	138254
<i>Crotalus_mitchellii</i>	CAS	52842
<i>Crotalus_mitchellii</i>	CAS	103470
<i>Crotalus_mitchellii</i>	CAS	182495
<i>Crotalus_mitchellii</i>	CAS	40092
<i>Crotalus_mitchellii</i>	CAS	85698
<i>Crotalus_mitchellii</i>	CAS	45887
<i>Crotalus_molossus</i>	CAS	48048
<i>Crotalus_molossus</i>	CAS	48047
<i>Crotalus_molossus</i>	CAS	48042
<i>Crotalus_molossus</i>	AMNH	67305
<i>Crotalus_molossus</i>	AMNH	26025
<i>Crotalus_molossus</i>	AMNH	138246
<i>Crotalus_oreganus</i>	CAS	238534
<i>Crotalus_oreganus</i>	CAS	223732
<i>Crotalus_oreganus</i>	CAS	209200
<i>Crotalus_oreganus</i>	USNM	209457
<i>Crotalus_oreganus</i>	AMNH	66076
<i>Crotalus_ornatus</i>	USNM	103738
<i>Crotalus_polystictus</i>	MNHParis	1975.145
<i>Crotalus_polystictus</i>	MNHParis	5092
<i>Crotalus_polystictus</i>	MNHParis	1888.284
<i>Crotalus_polystictus</i>	UTA	R4499
<i>Crotalus_polystictus</i>	UTA	R6927
<i>Crotalus_polystictus</i>	AMNH	161898
<i>Crotalus_polystictus</i>	AMNH	19861
<i>Crotalus_polystictus</i>	AMNH	19858
<i>Crotalus_pricei</i>	UTA	R6251
<i>Crotalus_pricei</i>	UTA	R45019
<i>Crotalus_pricei</i>	UTA	R57636
<i>Crotalus_pricei</i>	CAS	169788
<i>Crotalus_pricei</i>	CAS	48037
<i>Crotalus_pricei</i>	CAS	121003
<i>Crotalus_pricei</i>	CAS	48036
<i>Crotalus_pricei</i>	CAS	48034
<i>Crotalus_pusillus</i>	UTA	R4530
<i>Crotalus_pusillus</i>	UTA	R9358
<i>Crotalus_pusillus</i>	UTA	R6119
<i>Crotalus_pusillus</i>	FMNH	37048
<i>Crotalus_pusillus</i>	FMNH	39120
<i>Crotalus_pusillus</i>	FMNH	39113
<i>Crotalus_pusillus</i>	FMNH	39112
<i>Crotalus_ravus</i>	UTA	R12610
<i>Crotalus_ravus</i>	UTA	R12607
<i>Crotalus_ravus</i>	UTA	R12608
<i>Crotalus_ravus</i>	UTA	R12625
<i>Crotalus_ravus</i>	UTA	R25859
<i>Crotalus_ravus</i>	UTA	R13633

<i>Crotalus_ruber</i>	CAS	223537
<i>Crotalus_ruber</i>	CAS	169526
<i>Crotalus_ruber</i>	CAS	195596
<i>Crotalus_ruber</i>	CAS	103467
<i>Crotalus_ruber</i>	CAS	53717
<i>Crotalus_ruber</i>	CAS	51258
<i>Crotalus_ruber</i>	CAS	146940
<i>Crotalus_ruber</i>	CAS	104358
<i>Crotalus_scutulatus</i>	AMNH	69936
<i>Crotalus_scutulatus</i>	AMNH	68345
<i>Crotalus_scutulatus</i>	AMNH	68321
<i>Crotalus_scutulatus</i>	AMNH	68958
<i>Crotalus_scutulatus</i>	AMNH	88819
<i>Crotalus_simus</i>	UTA	R5252
<i>Crotalus_simus</i>	UTA	R21911
<i>Crotalus_simus</i>	UTA	R33133
<i>Crotalus_simus</i>	UTA	R52497
<i>Crotalus_simus</i>	UTA	R7701
<i>Crotalus_stejnegeri</i>	UTA	R6234
<i>Crotalus_tancitarensis</i>	UTA	R52401
<i>Crotalus_tancitarensis</i>	FMNH	39115
<i>Crotalus_tigris</i>	UTA	R31060
<i>Crotalus_tigris</i>	UTA	R32248
<i>Crotalus_tigris</i>	CAS	92265
<i>Crotalus_tigris</i>	CAS	92266
<i>Crotalus_tigris</i>	CAS	192769
<i>Crotalus_tigris</i>	CAS	102913
<i>Crotalus_tigris</i>	USNM	156808
<i>Crotalus_tigris</i>	AMNH	64273
<i>Crotalus_tigris</i>	AMNH	26024
<i>Crotalus_tigris</i>	AMNH	02537
<i>Crotalus_tortugensis</i>	FMNH	18426
<i>Crotalus_totonacus</i>	UTA	R7222
<i>Crotalus_totonacus</i>	UTA	R7224
<i>Crotalus_totonacus</i>	USNM	209856
<i>Crotalus_transversus</i>	UTA	R3988
<i>Crotalus_triseriatus</i>	MNHParis	1898.254
<i>Crotalus_triseriatus</i>	MNHParis	1641
<i>Crotalus_triseriatus</i>	UTA	R14516
<i>Crotalus_triseriatus</i>	UTA	R12601
<i>Crotalus_triseriatus</i>	UTA	R59041
<i>Crotalus_triseriatus</i>	UTA	R32788
<i>Crotalus_triseriatus</i>	FMNH	39110
<i>Crotalus_triseriatus</i>	AMNH	106588
<i>Crotalus_tzcaban</i>	UTA	R6802
<i>Crotalus_tzcaban</i>	UTA	R6732
<i>Crotalus_tzcaban</i>	USNM	46400
<i>Crotalus_viridis</i>	CAS	10419

<i>Crotalus_viridis</i>	CAS	93802
<i>Crotalus_viridis</i>	CAS	93803
<i>Crotalus_viridis</i>	USNM	313396
<i>Crotalus_viridis</i>	AMNH	121442
<i>Crotalus_willardi</i>	UTA	R9356
<i>Crotalus_willardi</i>	UTA	R17849
<i>Crotalus_willardi</i>	UTA	R32081
<i>Crotalus_willardi</i>	UTA	R22526
<i>Crotalus_willardi</i>	UTA	R6942
<i>Crotalus_willardi</i>	UTA	R17855
<i>Crotalus_willardi</i>	UTA	R21922
<i>Crotalus_willardi</i>	UTA	R18229
<i>Daboia_deserti</i>	MNHParis	1988.27
<i>Daboia_deserti</i>	MNHParis	1988.2497
<i>Daboia_deserti</i>	MNHParis	1988.4000
<i>Daboia_deserti</i>	MNHParis	1990.4197
<i>Daboia_deserti</i>	UTA	R8226
<i>Daboia_mauritanica</i>	BMNH	1906.10.31.11
<i>Daboia_mauritanica</i>	BMNH	1906.10.31.13
<i>Daboia_mauritanica</i>	MNHParis	1989.99
<i>Daboia_mauritanica</i>	MNHParis	1961.327
<i>Daboia_palaestinae</i>	BMNH	1937.9.4.1
<i>Daboia_palaestinae</i>	BMNH	1984.1220
<i>Daboia_palaestinae</i>	MNHParis	1989.70
<i>Daboia_palaestinae</i>	MNHParis	1986.123A
<i>Daboia_palaestinae</i>	MNHParis	1986.123B
<i>Daboia_russelii</i>	MNHParis	1962.259
<i>Daboia_russelii</i>	MNHParis	1962.258
<i>Daboia_russelii</i>	MNHParis	3187A
<i>Daboia_russelii</i>	AMNH	84442
<i>Daboia_russelii</i>	AMNH	161845
<i>Daboia_siamensis</i>	BMNH	1929.10.10.1-2
<i>Daboia_siamensis</i>	BMNH	1987.456
<i>Daboia_siamensis</i>	BMNH	1987.454
<i>Daboia_siamensis</i>	BMNH	1987.451
<i>Daboia_siamensis</i>	MNHParis	4020
<i>Deinagkistrodon_acutus</i>	MNHParis	1993.3464
<i>Deinagkistrodon_acutus</i>	MNHParis	1969.8
<i>Deinagkistrodon_acutus</i>	UTA	R32133
<i>Deinagkistrodon_acutus</i>	UTA	R8190
<i>Deinagkistrodon_acutus</i>	CAS	18765
<i>Deinagkistrodon_acutus</i>	CAS	121007
<i>Deinagkistrodon_acutus</i>	FMNH	127212
<i>Deinagkistrodon_acutus</i>	AMNH	83994
<i>Echis_carinatus</i>	BMNH	1983.716
<i>Echis_carinatus</i>	BMNH	1903.3.6.54
<i>Echis_carinatus</i>	BMNH	1964.259
<i>Echis_carinatus</i>	MNHParis	1988.3993

<i>Echis_carinatus</i>	AMNH	88472
<i>Echis_coloratus</i>	BMNH	1986.451
<i>Echis_coloratus</i>	BMNH	1986.456
<i>Echis_coloratus</i>	BMNH	1985.732
<i>Echis_coloratus</i>	MNHParis	1988.97
<i>Echis_coloratus</i>	MNHParis	1997.3446
<i>Echis_coloratus</i>	MNHParis	1990.4194
<i>Echis_coloratus</i>	MNHParis	1990.4192
<i>Echis_coloratus</i>	CAS	139767
<i>Echis_jogeri</i>	MNHParis	1993.0144
<i>Echis_khosatzkii</i>	BMNH	97.3.11.117
<i>Echis_leucogaster</i>	BMNH	1961.1849
<i>Echis_leucogaster</i>	BMNH	1961.1850
<i>Echis_leucogaster</i>	BMNH	1906.5.7.8
<i>Echis_leucogaster</i>	MNHParis	2002.0779
<i>Echis_leucogaster</i>	MNHParis	1985.454
<i>Echis_leucogaster</i>	CAS	136088
<i>Echis_leucogaster</i>	USNM	161948
<i>Echis_ocellatus</i>	BMNH	1976.688
<i>Echis_ocellatus</i>	BMNH	1976.725
<i>Echis_ocellatus</i>	BMNH	1975.704
<i>Echis_ocellatus</i>	BMNH	1975.706
<i>Echis_ocellatus</i>	MNHParis	1977.602
<i>Echis_ocellatus</i>	MNHParis	1977.596
<i>Echis_ocellatus</i>	MNHParis	1977.603
<i>Echis_omanensis</i>	BMNH	1977.84
<i>Echis_omanensis</i>	BMNH	1972.712
<i>Echis_omanensis</i>	BMNH	1971.127
<i>Echis_omanensis</i>	BMNH	1978.353
<i>Echis_omanensis</i>	CAS	250912
<i>Echis_pyramidum</i>	BMNH	1954.1.15.53-54
<i>Echis_pyramidum</i>	BMNH	1938.8.4.85-87
<i>Echis_pyramidum</i>	BMNH	51.7.17.157
<i>Echis_pyramidum</i>	BMNH	11.1.3a
<i>Echis_pyramidum</i>	MNHParis	1981.482
<i>Echis_pyramidum</i>	MNHParis	1990.4203
<i>Eristicophis_macmahoni</i>	BMNH	S/N
<i>Eristicophis_macmahoni</i>	UTA	R16392
<i>Eristicophis_macmahoni</i>	UTA	R16391
<i>Eristicophis_macmahoni</i>	UTA	R15665
<i>Eristicophis_macmahoni</i>	UTA	R16390
<i>Garthius_chaseni</i>	FMNH	71860
<i>Gloydus_blomhoffi</i>	BMNH	1870.1.14.2
<i>Gloydus_blomhoffi</i>	BMNH	1980.6.25.10
<i>Gloydus_blomhoffi</i>	BMNH	1893.4.20.28
<i>Gloydus_blomhoffi</i>	MNHParis	1999.7969
<i>Gloydus_blomhoffi</i>	MNHParis	1999.7970
<i>Gloydus_blomhoffi</i>	MNHParis	1888.37

<i>Gloydus_bломhoffi</i>	MNHParis	1888.36
<i>Gloydus_bломhoffi</i>	CAS	26732
<i>Gloydus_brevicaudus</i>	MNHParis	1902.74
<i>Gloydus_brevicaudus</i>	MNHParis	831
<i>Gloydus_brevicaudus</i>	FMNH	188988
<i>Gloydus_brevicaudus</i>	FMNH	188975
<i>Gloydus_brevicaudus</i>	AMNH	108495
<i>Gloydus_brevicaudus</i>	AMNH	108496
<i>Gloydus_brevicaudus</i>	AMNH	108498
<i>Gloydus_halys</i>	MNHParis	865
<i>Gloydus_halys</i>	AMNH	119011
<i>Gloydus_halys</i>	AMNH	123300
<i>Gloydus_intermedius</i>	MNHParis	1896.206
<i>Gloydus_intermedius</i>	CAS	121001
<i>Gloydus_intermedius</i>	USNM	203374
<i>Gloydus_intermedius</i>	USNM	203380
<i>Gloydus_saxatilis</i>	CAS	31540
<i>Gloydus_saxatilis</i>	CAS	31543
<i>Gloydus_saxatilis</i>	CAS	31528
<i>Gloydus_saxatilis</i>	AMNH	108502
<i>Gloydus_saxatilis</i>	AMNH	108505
<i>Gloydus_strauchi</i>	FMNH	15171
<i>Gloydus_strauchi</i>	FMNH	15134
<i>Gloydus_strauchi</i>	USNM	81980
<i>Gloydus_strauchi</i>	AMNH	58080
<i>Gloydus_ussuriensis</i>	CAS	14586
<i>Gloydus_ussuriensis</i>	FMNH	188999
<i>Gloydus_ussuriensis</i>	FMNH	S/N
<i>Gloydus_ussuriensis</i>	FMNH	11480
<i>Gloydus_ussuriensis</i>	FMNH	11481
<i>Hypnale_hypnale</i>	BMNH	1931.5.13.98
<i>Hypnale_hypnale</i>	CAS	17270
<i>Hypnale_hypnale</i>	AMNH	102462
<i>Hypnale_hypnale</i>	AMNH	99384
<i>Hypnale_hypnale</i>	AMNH	132250
<i>Hypnale_nepa</i>	BMNH	1931.5.13.100
<i>Hypnale_nepa</i>	BMNH	1937.8.1.1
<i>Hypnale_nepa</i>	BMNH	1894.9.11.39
<i>Hypnale_zara</i>	BMNH	1946.1.19.96
<i>Lachesis_acrochorda</i>	UTA	R56349
<i>Lachesis_acrochorda</i>	UTA	R7418
<i>Lachesis_melanocephala</i>	UTA	R7417
<i>Lachesis_melanocephala</i>	UTA	R10347
<i>Lachesis_muta</i>	MNRJ	13256
<i>Lachesis_muta</i>	MNRJ	3039
<i>Lachesis_muta</i>	MNRJ	241
<i>Lachesis_muta</i>	MNRJ	18048
<i>Lachesis_muta</i>	UTA	25601

<i>Lachesis_muta</i>	UTA	R25600
<i>Lachesis_muta</i>	MNRJ	3544
<i>Lachesis_muta</i>	MNRJ	4760
<i>Lachesis_muta</i>	MNRJ	8028
<i>Lachesis_stenophrys</i>	UTA	R9323
<i>Lachesis_stenophrys</i>	UTA	R25597
<i>Lachesis_stenophrys</i>	UTA	R18211
<i>Lachesis_stenophrys</i>	USNM	165008
<i>Lachesis_stenophrys</i>	AMNH	172358
<i>Macrovipera_lebetina</i>	BMNH	1900.1.19.3
<i>Macrovipera_lebetina</i>	BMNH	79.8.15.34
<i>Macrovipera_lebetina</i>	MNHParis	1229
<i>Macrovipera_lebetina</i>	MNHParis	7737
<i>Macrovipera_lebetina</i>	CAS	218070
<i>Macrovipera_lebetina</i>	CAS	179736
<i>Macrovipera_schweizeri</i>	BMNH	1936.1.4.2-6
<i>Macrovipera_schweizeri</i>	BMNH	1936.1.4.2-7
<i>Mixcoatlus_browni</i>	CAS	135274
<i>Montivipera_bornmuelleri</i>	BMNH	1957.1.13.80
<i>Montivipera_bornmuelleri</i>	BMNH	1957.1.13.72
<i>Montivipera_bornmuelleri</i>	BMNH	1984.1227
<i>Montivipera_latifii</i>	BMNH	1976.555
<i>Montivipera_latifii</i>	BMNH	1976.554
<i>Montivipera_latifii</i>	BMNH	1976.552
<i>Montivipera_latifii</i>	AMNH	109617
<i>Montivipera_raddei</i>	BMNH	1976.556
<i>Montivipera_raddei</i>	BMNH	1976.557
<i>Montivipera_raddei</i>	UTA	R7165
<i>Montivipera_raddei</i>	CAS	135746
<i>Montivipera_wagneri</i>	MNHParis	1988.9011
<i>Montivipera_xanthina</i>	BMNH	1984.1226
<i>Montivipera_xanthina</i>	MNHParis	1989.128
<i>Montivipera_xanthina</i>	MNHParis	1989.71
<i>Montivipera_xanthina</i>	MNHParis	1988.3867
<i>Montivipera_xanthina</i>	MNHParis	1988.22
<i>Montivipera_xanthina</i>	MNHParis	1988.26
<i>Montivipera_xanthina</i>	CAS	135748
<i>Mixcoatlus_melanurus</i>	UTA	R6818
<i>Mixcoatlus_melanurus</i>	UTA	R9610
<i>Mixcoatlus_melanurus</i>	UTA	R6817
<i>Mixcoatlus_melanurus</i>	UTA	R22451
<i>Mixcoatlus_melanurus</i>	UTA	R16309
<i>Mixcoatlus_melanurus</i>	UTA	R12559
<i>Ophryacus_undulatus</i>	UTA	R4518
<i>Ophryacus_undulatus</i>	UTA	R4642
<i>Ophryacus_undulatus</i>	UTA	R4108
<i>Ophryacus_undulatus</i>	UTA	R4640
<i>Ophryacus_undulatus</i>	UTA	R25115

<i>Ophryacus_undulatus</i>	UTA	R4835
<i>Ophryacus_undulatus</i>	UTA	R4832
<i>Ophryacus_undulatus</i>	UTA	R32426
<i>Ophryacus_undulatus</i>	UTA	R4858
<i>Ovophis_monticola</i>	BMNH	1930.11.16.13
<i>Ovophis_monticola</i>	BMNH	1930.11.16.14
<i>Ovophis_monticola</i>	MNHParis	2011.02.49
<i>Ovophis_monticola</i>	MNHParis	1935.115
<i>Ovophis_monticola</i>	FMNH	232827
<i>Ovophis_monticola</i>	FMNH	258632
<i>Ovophis_okinavensis</i>	BMNH	1906.8.16.45
<i>Ovophis_okinavensis</i>	MNHParis	1986.103
<i>Ovophis_okinavensis</i>	UTA	R22391
<i>Ovophis_okinavensis</i>	UTA	R28192
<i>Ovophis_okinavensis</i>	CAS	21930
<i>Ovophis_tonkinensis</i>	MNHParis	1948.105
<i>Ovophis_tonkinensis</i>	MNHParis	2011.0246
<i>Ovophis_zayuensis</i>	CAS	233203
<i>Porthidium_arcosae</i>	UTA	R55938
<i>Porthidium_dunni</i>	UTA	R17905
<i>Porthidium_dunni</i>	UTA	R4367
<i>Porthidium_dunni</i>	FMNH	104957
<i>Porthidium_dunni</i>	AMNH	91108
<i>Porthidium_dunni</i>	AMNH	19874
<i>Porthidium_dunni</i>	AMNH	66859
<i>Porthidium_dunni</i>	AMNH	66868
<i>Porthidium_dunni</i>	AMNH	66866
<i>Porthidium_dunni</i>	AMNH	66862
<i>Porthidium_dunni</i>	AMNH	66854
<i>Porthidium_dunni</i>	AMNH	66870
<i>Porthidium_dunni</i>	AMNH	66867
<i>Porthidium_dunni</i>	AMNH	65171
<i>Porthidium_dunni</i>	AMNH	65167
<i>Porthidium_dunni</i>	AMNH	65169
<i>Porthidium_dunni</i>	AMNH	65166
<i>Porthidium_dunni</i>	AMNH	68076
<i>Porthidium_dunni</i>	AMNH	65874
<i>Porthidium_hespere</i>	UTA	R4443
<i>Porthidium_lansbergii</i>	BMNH	71.2.7.4
<i>Porthidium_lansbergii</i>	BMNH	S/N
<i>Porthidium_lansbergii</i>	UTA	R3677
<i>Porthidium_lansbergii</i>	CAS	116159
<i>Porthidium_lansbergii</i>	FMNH	166032
<i>Porthidium_lansbergii</i>	FMNH	166080
<i>Porthidium_lansbergii</i>	FMNH	166097
<i>Porthidium_lansbergii</i>	FMNH	166073
<i>Porthidium_nasutum</i>	UTA	R31057
<i>Porthidium_nasutum</i>	UTA	R14181

<i>Porthidium_nasutum</i>	UTA	R23065
<i>Porthidium_nasutum</i>	UTA	R26408
<i>Porthidium_nasutum</i>	UTA	R39207
<i>Porthidium_ophryomegas</i>	UTA	R39755
<i>Porthidium_ophryomegas</i>	UTA	R28565
<i>Porthidium_ophryomegas</i>	UTA	R39217
<i>Porthidium_ophryomegas</i>	UTA	R28563
<i>Porthidium_ophryomegas</i>	UTA	R45885
<i>Porthidium_ophryomegas</i>	UTA	R45887
<i>Porthidium_ophryomegas</i>	UTA	R7722
<i>Porthidium_ophryomegas</i>	AMNH	161915
<i>Porthidium_ophryomegas</i>	AMNH	64407
<i>Porthidium_porrasi</i>	UTA	R59119
<i>Porthidium_porrasi</i>	UTA	R52898
<i>Porthidium_porrasi</i>	UTA	R30830
<i>Porthidium_yucatanicum</i>	UTA	R16960
<i>Porthidium_yucatanicum</i>	FMNH	36185
<i>Porthidium_yucatanicum</i>	FMNH	36176
<i>Porthidium_yucatanicum</i>	FMNH	36187
<i>Porthidium_yucatanicum</i>	FMNH	36182
<i>Proatheris_superciliaris</i>	BMNH	1934.4.6.36
<i>Proatheris_superciliaris</i>	BMNH	1929.11.2.7
<i>Protobothrops_cornutus</i>	MNHParis	1937.35
<i>Protobothrops_elegans</i>	CAS	21958
<i>Protobothrops_elegans</i>	CAS	21954
<i>Protobothrops_elegans</i>	CAS	21962
<i>Protobothrops_elegans</i>	CAS	21967
<i>Protobothrops_elegans</i>	CAS	21947
<i>Protobothrops_elegans</i>	CAS	21948
<i>Protobothrops_flavoviridis</i>	BMNH	87.10.5.1-2
<i>Protobothrops_flavoviridis</i>	BMNH	1906.8.16.35-36
<i>Protobothrops_flavoviridis</i>	BMNH	1906.8.16.35-36
<i>Protobothrops_flavoviridis</i>	MNHParis	1893.11
<i>Protobothrops_flavoviridis</i>	MNHParis	1895.142
<i>Protobothrops_flavoviridis</i>	USNM	121469
<i>Protobothrops_jerdonii</i>	BMNH	1974.910
<i>Protobothrops_jerdonii</i>	BMNH	91.6.13.1
<i>Protobothrops_jerdonii</i>	BMNH	1940.6.5.96
<i>Protobothrops_jerdonii</i>	BMNH	1940.6.5.95
<i>Protobothrops_jerdonii</i>	BMNH	1940.6.5.94
<i>Protobothrops_jerdonii</i>	MNHParis	1935.0114
<i>Protobothrops_jerdonii</i>	MNHParis	1938.0151
<i>Protobothrops_kaulbacki</i>	BMNH	1936.7.4.42
<i>Protobothrops_kaulbacki</i>	BMNH	1946.1.19.23
<i>Protobothrops_kaulbacki</i>	BMNH	1946.1.19.24
<i>Protobothrops_kaulbacki</i>	CAS	22430
<i>Protobothrops_mucrosquamatus</i>	MNHParis	1935.118
<i>Protobothrops_mucrosquamatus</i>	MNHParis	1897.100

<i>Protobothrops_mucrosquamatus</i>	MNHParis	1948.108
<i>Protobothrops_mucrosquamatus</i>	CAS	224365
<i>Protobothrops_mucrosquamatus</i>	CAS	238906
<i>Protobothrops_mucrosquamatus</i>	CAS	244954
<i>Protobothrops_mucrosquamatus</i>	CAS	18920
<i>Protobothrops_mucrosquamatus</i>	CAS	18923
<i>Protobothrops_tokarensis</i>	MNHParis	1988.2106
<i>Protobothrops_tokarensis</i>	USNM	348518
<i>Pseudocerastes_persicus</i>	MNHParis	1990.4.156
<i>Pseudocerastes_persicus</i>	MNHParis	1990.4.155
<i>Pseudocerastes_persicus</i>	MNHParis	1990.4.151
<i>Pseudocerastes_persicus</i>	MNHParis	1990.4.157
<i>Pseudocerastes_urarachnoides</i>	FMNH	170929
<i>Pseudocerastes_fieldi</i>	UTA	R8178
<i>Pseudocerastes_fieldi</i>	FMNH	11061
<i>Pseudocerastes_fieldi</i>	AMNH	94418
<i>Pseudocerastes_fieldi</i>	AMNH	107688
<i>Pseudocerastes_fieldi</i>	AMNH	94419
<i>Sistrurus_catenatus</i>	UTA	R6809
<i>Sistrurus_catenatus</i>	UTA	R1335
<i>Sistrurus_catenatus</i>	UTA	R11290
<i>Sistrurus_catenatus</i>	UTA	R54126
<i>Sistrurus_catenatus</i>	UTA	R53622
<i>Sistrurus_catenatus</i>	UTA	R30852
<i>Sistrurus_catenatus</i>	FMNH	135026
<i>Sistrurus_catenatus</i>	FMNH	95447
<i>Sistrurus_catenatus</i>	AMNH	8700
<i>Sistrurus_miliarius</i>	UTA	R25121
<i>Sistrurus_miliarius</i>	UTA	R22378
<i>Sistrurus_miliarius</i>	UTA	R56016
<i>Sistrurus_miliarius</i>	USNM	234410
<i>Sistrurus_miliarius</i>	USNM	234405
<i>Sistrurus_miliarius</i>	USNM	491
<i>Trimeresurus_albolabris</i>	MNHParis	1999.9474
<i>Trimeresurus_albolabris</i>	UTA	ENS 14106
<i>Trimeresurus_albolabris</i>	USNM	163960
<i>Trimeresurus_albolabris</i>	USNM	163965
<i>Trimeresurus_albolabris</i>	USNM	72077
<i>Trimeresurus_andersonii</i>	BMNH	1937.3.1.23
<i>Trimeresurus_andersonii</i>	BMNH	1937.3.1.23
<i>Trimeresurus_andersonii</i>	BMNH	1940.3.9.37
<i>Trimeresurus_barati</i>	BMNH	1915.12.2.42
<i>Trimeresurus_barati</i>	UTA	ENS 15785
<i>Trimeresurus_borneensis</i>	BMNH	94.8.37
<i>Trimeresurus_borneensis</i>	BMNH	94.6.30.67
<i>Trimeresurus_buniana</i>	BMNH	2005.1609
<i>Trimeresurus_cantori</i>	BMNH	1940.3.9.41
<i>Trimeresurus_cantori</i>	BMNH	1940.3.9.42

<i>Trimeresurus_cantori</i>	BMNH	1940.3.9.40
<i>Trimeresurus_cantori</i>	USNM	29445
<i>Trimeresurus_erythrurus</i>	BMNH	61.10.2.6
<i>Trimeresurus_erythrurus</i>	UTA	R29684
<i>Trimeresurus_erythrurus</i>	CAS	243175
<i>Trimeresurus_erythrurus</i>	CAS	239502
<i>Trimeresurus_erythrurus</i>	CAS	240036
<i>Trimeresurus_erythrurus</i>	AMNH	63897
<i>Trimeresurus_fasciatus</i>	MNHParis	2000.401
<i>Trimeresurus_flavomaculatus</i>	BMNH	2002.18
<i>Trimeresurus_flavomaculatus</i>	MNHParis	1900.426
<i>Trimeresurus_flavomaculatus</i>	FMNH	270194
<i>Trimeresurus_fucatus</i>	MNHParis	1900.4283
<i>Trimeresurus_fucatus</i>	MNHParis	1900.4284
<i>Trimeresurus_fucatus</i>	MNHParis	1991.295
<i>Trimeresurus_gramineus</i>	MNRJ	13269
<i>Trimeresurus_gramineus</i>	MNHParis	4058
<i>Trimeresurus_gramineus</i>	CAS	14979
<i>Trimeresurus_gramineus</i>	CAS	18751
<i>Trimeresurus_gramineus</i>	CAS	18748
<i>Trimeresurus_gramineus</i>	CAS	18752
<i>Trimeresurus_gramineus</i>	CAS	18747
<i>Trimeresurus_gumprechtii</i>	BMNH	2002.53
<i>Trimeresurus_gumprechtii</i>	MNHParis	1987.3835
<i>Trimeresurus_gumprechtii</i>	MNHParis	1999.9072
<i>Trimeresurus_gumprechtii</i>	CAS	234873
<i>Trimeresurus_hageni</i>	BMNH	1988.859
<i>Trimeresurus_hageni</i>	BMNH	1988.861
<i>Trimeresurus_hageni</i>	ZRC	22932
<i>Trimeresurus_hageni</i>	ZRC	25397
<i>Trimeresurus_hageni</i>	ZRC	22935
<i>Trimeresurus_hageni</i>	UTA	R55257
<i>Trimeresurus_insularis</i>	MNHParis	4057
<i>Trimeresurus_insularis</i>	USNM	579386
<i>Trimeresurus_insularis</i>	USNM	579784
<i>Trimeresurus_insularis</i>	USNM	573677
<i>Trimeresurus_kanburiensis</i>	BMNH	1992.535
<i>Trimeresurus_macrops</i>	BMNH	1988.876
<i>Trimeresurus_macrops</i>	BMNH	1988.1877
<i>Trimeresurus_macrops</i>	BMNH	1988.1057
<i>Trimeresurus_macrops</i>	BMNH	1988.1053
<i>Trimeresurus_macrops</i>	BMNH	1988.1052
<i>Trimeresurus_macrops</i>	BMNH	1988.865
<i>Trimeresurus_macrops</i>	MNHParis	1885.389
<i>Trimeresurus_malabaricus</i>	BMNH	74.4.29.136-139.1027
<i>Trimeresurus_malabaricus</i>	MNHParis	1913.5
<i>Trimeresurus_malabaricus</i>	CAS	17273
<i>Trimeresurus_malabaricus</i>	FMNH	217680

<i>Trimeresurus_malabaricus</i>	FMNH	30763
<i>Trimeresurus_malcolmi</i>	BMNH	1929.12.22.116
<i>Trimeresurus_malcolmi</i>	FMNH	239948
<i>Trimeresurus_mcgregori</i>	BMNH	2002.06
<i>Trimeresurus_mcgregori</i>	BMNH	2002.4
<i>Trimeresurus_mcgregori</i>	BMNH	2002.9
<i>Trimeresurus_mcgregori</i>	BMNH	2002.8
<i>Trimeresurus_mcgregori</i>	MNHParis	1900.427
<i>Trimeresurus_nebularis</i>	BMNH	2003.1
<i>Trimeresurus_popeiorum</i>	BMNH	1924.5.20.38
<i>Trimeresurus_popeiorum</i>	BMNH	60.3.19.1300B
<i>Trimeresurus_popeiorum</i>	BMNH	72.4.17.137
<i>Trimeresurus_popeiorum</i>	CAS	240640
<i>Trimeresurus_popeiorum</i>	FMNH	178656
<i>Trimeresurus_popeiorum</i>	FMNH	265805
<i>Trimeresurus_puniceus</i>	BMNH	85.12.31.32-33
<i>Trimeresurus_puniceus</i>	BMNH	85.12.31.32-34
<i>Trimeresurus_puniceus</i>	MNHParis	1991.3286
<i>Trimeresurus_puniceus</i>	FMNH	183785
<i>Trimeresurus_puniceus</i>	FMNH	138662
<i>Trimeresurus_puniceus</i>	FMNH	131846
<i>Trimeresurus_purpureomaculatus</i>	BMNH	1988.397
<i>Trimeresurus_purpureomaculatus</i>	MNHParis	1990.4282
<i>Trimeresurus_purpureomaculatus</i>	MNHParis	1990.3880
<i>Trimeresurus_purpureomaculatus</i>	MNHParis	1999.6588
<i>Trimeresurus_purpureomaculatus</i>	AMNH	63899
<i>Trimeresurus_purpureomaculatus</i>	AMNH	63900
<i>Trimeresurus_sabahi</i>	FMNH	251048
<i>Trimeresurus_sabahi</i>	USNM	130253
<i>Trimeresurus_schultzei</i>	FMNH	53561
<i>Trimeresurus_schultzei</i>	USNM	37871
<i>Trimeresurus_septentrionalis</i>	BMNH	1950.1.5.64
<i>Trimeresurus_septentrionalis</i>	BMNH	1955.1.13.82
<i>Trimeresurus_septentrionalis</i>	CAS	135750
<i>Trimeresurus_stejnegeri</i>	BMNH	1953.1.2.88-92
<i>Trimeresurus_stejnegeri</i>	BMNH	S/N
<i>Trimeresurus_stejnegeri</i>	BMNH	S/N
<i>Trimeresurus_stejnegeri</i>	BMNH	S/N
<i>Trimeresurus_stejnegeri</i>	MNHParis	1912.353
<i>Trimeresurus_stejnegeri</i>	MNHParis	1912.352
<i>Trimeresurus_stejnegeri</i>	MNHParis	1912.354
<i>Trimeresurus_stejnegeri</i>	MNHParis	1999.9064
<i>Trimeresurus_stejnegeri</i>	MNHParis	1999.9055
<i>Trimeresurus_sumatranus</i>	BMNH	84.1.8.47
<i>Trimeresurus_sumatranus</i>	USNM	134125
<i>Trimeresurus_tibetanus</i>	CAS	177677
<i>Trimeresurus_trigonocephalus</i>	MNHParis	245
<i>Trimeresurus_trigonocephalus</i>	UTA	R40461

<i>Trimeresurus_trigonocephalus</i>	UTA	R8191
<i>Trimeresurus_trigonocephalus</i>	UTA	R34536
<i>Trimeresurus_trigonocephalus</i>	UTA	R32527
<i>Trimeresurus_trigonocephalus</i>	USNM	254789
<i>Trimeresurus_venustus</i>	BMNH	1988.384
<i>Trimeresurus_vogeli</i>	FMNH	180274
<i>Trimeresurus_vogeli</i>	FMNH	180273
<i>Trimeresurus_vogeli</i>	FMNH	180272
<i>Trimeresurus_vogeli</i>	FMNH	180260
<i>Trimeresurus_vogeli</i>	FMNH	180256
<i>Trimeresurus_wiroti</i>	UTA	R38540
<i>Trimeresurus_yunnanensis</i>	CAS	215141
<i>Trimeresurus_yunnanensis</i>	CAS	234261
<i>Trimeresurus_yunnanensis</i>	FMNH	7065
<i>Trimeresurus_yunnanensis</i>	AMNH	62255
<i>Tropidolaemus_subannulatus</i>	BMNH	1946.1.19.33
<i>Tropidolaemus_subannulatus</i>	MNHParis	2011.02.52
<i>Tropidolaemus_subannulatus</i>	MNHParis	5821
<i>Tropidolaemus_subannulatus</i>	MNHParis	1891.0082
<i>Tropidolaemus_subannulatus</i>	CAS	8679
<i>Tropidolaemus_subannulatus</i>	FMNH	3566
<i>Tropidolaemus_wagleri</i>	BMNH	97.12.28.58
<i>Tropidolaemus_wagleri</i>	BMNH	1.2.9.0
<i>Tropidolaemus_wagleri</i>	BMNH	86.12.28.24
<i>Vipera_ammodytes</i>	MNRJ	21214
<i>Vipera_ammodytes</i>	BMNH	1931.12.15.3
<i>Vipera_ammodytes</i>	BMNH	1987.1848
<i>Vipera_ammodytes</i>	BMNH	1988.52
<i>Vipera_ammodytes</i>	BMNH	1920.1.2.2828
<i>Vipera_ammodytes</i>	MNHParis	7732
<i>Vipera_ammodytes</i>	MNHParis	1918.138
<i>Vipera_ammodytes</i>	MNHParis	1918.136
<i>Vipera_ammodytes</i>	MNHParis	1904-
<i>Vipera_ammodytes</i>	MNHParis	1904.595
<i>Vipera_ammodytes</i>	MNHParis	1904.593
<i>Vipera_aspis</i>	BMNH	1920.1.20.250
<i>Vipera_aspis</i>	BMNH	90.5.17.17-18
<i>Vipera_aspis</i>	BMNH	809
<i>Vipera_aspis</i>	MNHParis	1905.411
<i>Vipera_aspis</i>	MNHParis	1905.453
<i>Vipera_aspis</i>	MNHParis	1905.452
<i>Vipera_aspis</i>	MNHParis	8485
<i>Vipera_aspis</i>	MNHParis	1939.25
<i>Vipera_aspis</i>	MNHParis	1922.296
<i>Vipera_aspis</i>	MNHParis	9861
<i>Vipera_aspis</i>	MNHParis	9858
<i>Vipera_aspis</i>	MNHParis	1993.1776
<i>Vipera_berus</i>	MNRJ	20774

<i>Vipera_berus</i>	MNRJ	237
<i>Vipera_berus</i>	MNHParis	1904.581
<i>Vipera_berus</i>	MNHParis	1904.582
<i>Vipera_berus</i>	MNHParis	1904.586
<i>Vipera_berus</i>	MNHParis	1898.407
<i>Vipera_berus</i>	MNHParis	1900.178
<i>Vipera_berus</i>	MNHParis	3999A
<i>Vipera_berus</i>	MNHParis	1904.578
<i>Vipera_berus</i>	MNHParis	1884.579
<i>Vipera_berus</i>	CAS	14579
<i>Vipera_berus</i>	CAS	26101
<i>Vipera_dinniki</i>	FMNH	236002
<i>Vipera_eriwanensis</i>	MNHParis	1989.54
<i>Vipera_eriwanensis</i>	MNHParis	1989.52
<i>Vipera_eriwanensis</i>	MNHParis	1989.58
<i>Vipera_eriwanensis</i>	MNHParis	1989.57
<i>Vipera_kaznakovi</i>	MNHParis	1997.3302
<i>Vipera_kaznakovi</i>	FMNH	223694
<i>Vipera_kaznakovi</i>	FMNH	223695
<i>Vipera_kaznakovi</i>	AMNH	161837
<i>Vipera_latastei</i>	BMNH	1904.2.12.1-4
<i>Vipera_latastei</i>	BMNH	92.11.21.1
<i>Vipera_latastei</i>	BMNH	94.6.8.2 / 94.11.10.1
<i>Vipera_latastei</i>	MNHParis	5758
<i>Vipera_latastei</i>	FMNH	168015
<i>Vipera_nikolskii</i>	FMNH	257149
<i>Vipera_renardi</i>	BMNH	91.2.23.1
<i>Vipera_renardi</i>	BMNH	79.11.14.270
<i>Vipera_renardi</i>	BMNH	95.3.1.4-5
<i>Vipera_renardi</i>	BMNH	95.3.1.4-5
<i>Vipera_renardi</i>	BMNH	93.12.30.7
<i>Vipera_renardi</i>	MNHParis	1919.124
<i>Vipera_renardi</i>	FMNH	83965
<i>Vipera_renardi</i>	AMNH	84028
<i>Vipera_renardi</i>	AMNH	160760
<i>Vipera_seoanei</i>	BMNH	1914.5.26.35
<i>Vipera_seoanei</i>	BMNH	1920.1.20.2827
<i>Vipera_seoanei</i>	BMNH	95.4.30.35-37
<i>Vipera_seoanei</i>	MNHParis	1884.266
<i>Vipera_seoanei</i>	MNHParis	1889.584
<i>Vipera_seoanei</i>	MNHParis	1889.585
<i>Vipera_seoanei</i>	MNHParis	1971.278
<i>Vipera_transcaucasiana</i>	UTA	R18219
<i>Vipera_ursinii</i>	BMNH	92.4.4.4 / 92.9.13.12-13 / 93.2.10.1-3
<i>Vipera_ursinii</i>	MNHParis	1898.334
<i>Vipera_ursinii</i>	MNHParis	1966.1069
<i>Vipera_ursinii</i>	AMNH	161847
<i>Vipera_ursinii</i>	AMNH	90453

APPENDIX II

In the present study we used body size measurements mostly taken from male specimens. However, for 16 species out of 249 we only had access to female specimens, and we choose to keep their body size data in trait evolution analysis. We argue that including these female measurements did not biased our results for several reasons.

First, these 16 species are not restricted to a single habitat but are distributed across all habitat categories: two are terrestrial occurring in both forested and open habitats, two are terrestrial restricted to open habitats, seven are terrestrial restricted to forests and five are arboreal species. Additionally, we estimated sexual size dimorphism for body size in vipers we had information for at least two males and two females of each species using mean snout-vent length as a proxy for body size (calculation done following Hendry et al., 2014).

Although male-biased sexual dimorphism (MBSSD) occur among vipers, differences between the sexes are concentrated mostly below 10% meaning that for most species females are at best only 10% smaller than males (Figure S7). Additionally, higher degrees of MBSSD (> 0.1) mostly concentrate in the genus *Crotalus* (Figure S7, also suggested by Hendry et al. 2014 for total length) which represents only one species in the female sample used in trait evolution analyses. In contrast, female-biased sexual size dimorphism (FBSSD) is more widespread in vipers and suggested to predominantly occur among arboreal NW vipers (Hendry et al., 2014), but most values are concentrated below 0.2. Additionally, extreme values (> 0.2) seem to occur among species attaining female body sizes larger than 500 mm (Figure S8). Six of the 16 species we included in our analyses are small in size (smaller than 500 mm, see Table S1) and probably would not show extreme values of female-biased sexual size dimorphism. Although 10 species comprise female body sizes larger than 500 mm they are not restricted to a single habitat category: one is terrestrial occurring in both forested and open habitats, one is terrestrial and restricted to open habitats, three are terrestrials restricted to forests and five are arboreals.

Finally, using a two-sample Wilcoxon test we evaluated if body size significantly differs between the sexes in species phylogenetically related to the 16 for which we used female measurements based on the phylogenetic tree presented in Chapter I. From these 16 species, 10 are phylogenetic related to species not presenting significant body size differences between the sexes ($p = 1.00 - 0.15$). Although six are phylogenetically related to species presenting significant intersexual differences in body size ($p = 0.030 - 0.002$), all being female-biased, three are terrestrials restricted to forests and the other three are arboreals and thus are not restricted to a single habitat category.

Given the above considerations we suggest that is very unlikely that using female measurements for these 16 species would have biased our results toward a more constrained body size in arboreal species.

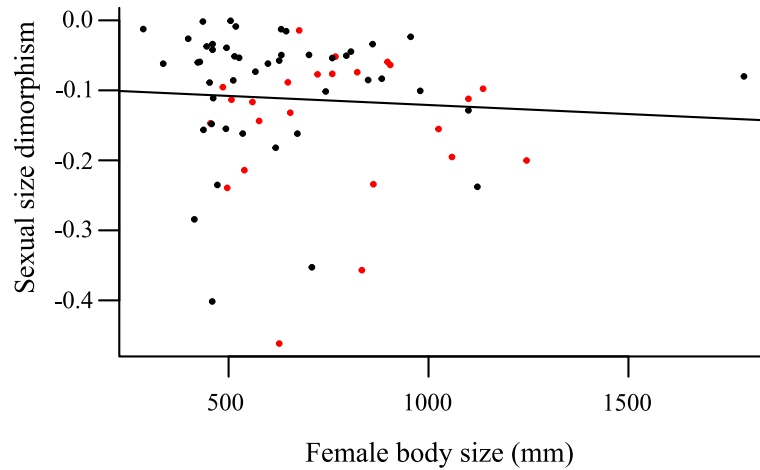


Figure S7. Male-biased sexual size dimorphism among vipers (MBSSD = $-\left[\frac{\text{larger sex}}{\text{smaller sex}} - 1\right]$) relative to female body size ($R^2 = 0.004$, $p = 0.56$). Red circles indicate species belonging to the genus *Crotalus*. Variables were log10 transformed prior to the analysis. This plot was constructed using only species where males are larger than females.

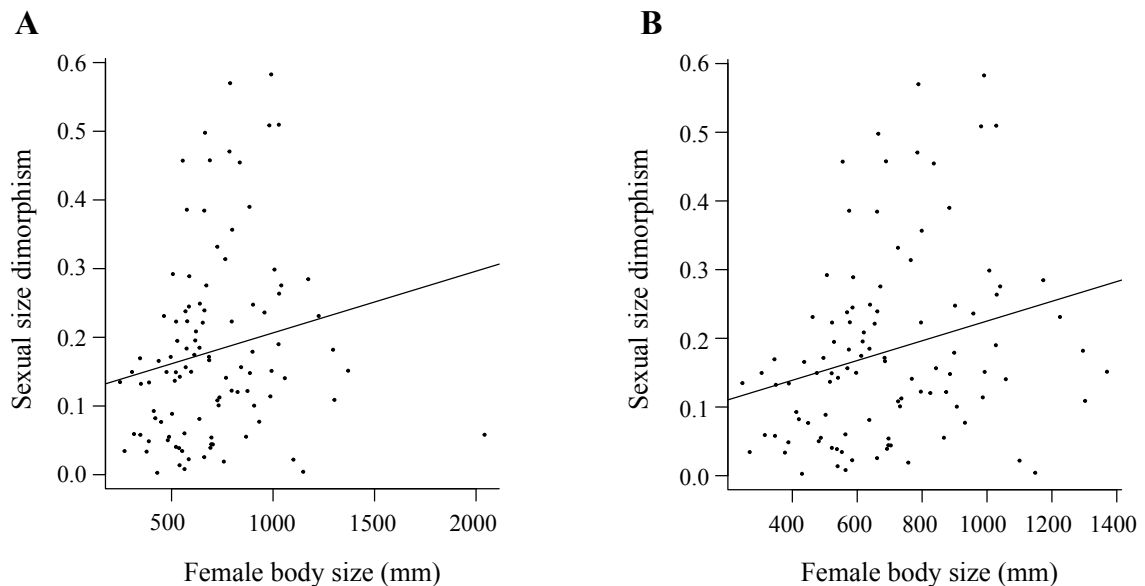


Figure S8. Female-biased sexual size dimorphism among vipers (FBSSD = $\left[\frac{\text{larger sex}}{\text{smaller sex}} - 1\right]$) relative to female body size. A) Including the larger species in our sample *Lachesis stenophrys* ($R^2 = 0.03$, $p = 0.04$). B) Excluding *Lachesis stenophrys* ($R^2 = 0.04$, $p = 0.01$).

Variables were log₁₀ transformed prior to the analysis. This plot was constructed using only species where female are larger than males.

References

Hendry, C.R., Guiher, T.J., Pyron, R.A., 2014. Ecological divergence and sexual selection drive sexual size dimorphism in new world pitvipers (Serpentes: Viperidae). *Journal of Evolutionary Biology* 27(4), 760-771.

Usando informações das espécies viventes, os capítulos que compõe a presente tese visaram explorar quais os padrões que caracterizaram a radiação dos viperídeos, buscando entender os possíveis processos responsáveis por moldar tais padrões ao longo dos seus ~50 milhões de anos:

- Utilizando sequências de 11 genes nucleares e mitocondriais para 264 espécies de viperídeos (79% de todas as espécies viventes), foi possível reconstruir relações filogenéticas robustas para diferentes clados que compõe o grupo. Os tempos de divergência obtidos indicam que o grupo deve ter começado a diversificar em meados do Paleoceno tardio/meio do Eoceno inferindo idades pouco mais tardias que o encontrado em estudos anteriores. Além disso, as idades estimadas permitiram estabelecer possíveis conexões entre a diversificação do grupo e determinadas mudanças geológicas e climáticas que ocorreram na Terra há milhões de anos atrás.
- A radiação dos viperídeos foi marcada por eventos importantes como, por exemplo, o surgimento das fossetas loreais na subfamília Crotalinae. Os resultados obtidos nesta tese indicam que um aumento nas taxas de especiação envolvendo parte do clado dos crotalíneos ocorreu provavelmente em decorrência não só da evolução das fossetas loreais mas também como resultado da expansão das florestas do sudeste asiático e da invasão do novo mundo. Após este rápido aumento nas taxas de especiação, porém, estas passam a desacelerar em direção ao presente. Com base em nossos resultados propomos um cenário para explicar como deve ter acontecido esta rápida multiplicação de linhagens e o porquê da sua desaceleração.
- A interação com diferentes habitats também pode afetar a dinâmica de diversificação de um grupo. Durante sua evolução, os viperídeos ocuparam diversos habitats incluindo o habitat arborícola. O ambiente arborícola parece impor pressões seletivas importantes nas linhagens de serpentes em geral e, os resultados aqui sugerem que nos viperídeos, a evolução morfológica foi de fato diferente nas linhagens arborícolas quando comparadas àquelas ocorrendo em habitats terrestres. Entretanto, a conquista do habitat arborícola que parece ter limitado a evolução morfológica dos viperídeos não parece ter afetado as suas taxas de especiação. Com esses resultados, relacionamos os padrões macroevolutivos encontrados com diferentes e plausíveis mecanismos de especiação que seriam importantes em gerar espécies nos diferentes habitats, concluindo que: (1) a especiação pode acontecer

independente de mudanças morfológicas nos viperídeos, ou (2) o isolamento geográfico pode ter sido o mecanismo de especiação predominante entre os viperídeos arborícolas.

O presente trabalho contribuiu para desvendar como se deu a evolução de um dos grupos de serpentes mais notáveis do mundo, os viperídeos. Indo além, ressaltamos a importância de entendermos mais sobre árvores filogenéticas para podermos integrar o uso destas em trabalhos de ecologia evolutiva. Nossa abordagem nos permitiu recuperar padrões e inferir processos que moldaram a radiação dos viperídeos e que poderiam possivelmente ter moldado a diversificação de outros grupos de organismos. Além disso, nossos resultados sugerem cenários e hipóteses a serem futuramente explorados e que deverão contribuir ainda mais para melhor entendermos os padrões e mecanismos que geram e mantêm a biodiversidade.