

1 **Moving Past Neonicotinoids and Honeybees: A Systematic Review of Existing**
2 **Research on Other Insecticides and Bees**

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10 **Abstract**

11 Synthetic pesticides are used widely in agriculture to protect crops from pests, weeds and
12 disease. However, their use also comes with a range of environmental concerns. One of which
13 is effects of insecticides on non-target organisms such as bees, who provide pollination services
14 for crops and wild plants. This systematic literature review quantifies the existing research on
15 bees and insecticides broadly, and then focuses more specifically on non-neonicotinoid
16 insecticides and non-honeybees. We find that articles on honeybees (*Apis sp.*) and insecticides
17 account for 80% of all research, with all other bees combined making up 20%. Neonicotinoids
18 were studied in 34% of articles across all bees and were the most widely studied insecticide
19 class for non-honeybees overall, with almost three times as many studies than the second most
20 studied class. Of non-neonicotinoid insecticide classes and non-honeybees; the most studied
21 were pyrethroids and organophosphates followed by carbamates, and the most widely
22 represented bee taxa were bumblebees (*Bombus*), followed by leaf-cutter bees (*Megachile*) and
23 mason bees (*Osmia*). Research has taken place across several countries, with the highest
24 numbers of articles from Brazil and the US, and with notable gaps from countries in Asia,
25 Africa and Oceania. Mortality was the most studied effect type, while sub-lethal effects such
26 as on behaviour were less studied. Few studies tested how insecticides were influenced by other
27 multiple pressures, such as climate change and co-occurring pesticides (cocktail effects). As
28 anthropogenic pressures do not occur in isolation, we suggest that future research also
29 addresses these knowledge gaps. Given the changing global patterns in insecticide use, and the
30 increasing inclusion of both non-honeybees and sub-lethal effects in pesticide risk assessment,
31 there is a need for expanding research beyond current state to ensure a strong scientific
32 evidence base for the development of risk assessment and associated policy.

33

34 **Keywords:** Synthetic insecticides, Pesticides, Risk Assessment, Plant Protection Products,
35 bumblebee

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38 **1 Introduction**

39 Since the 1940s, synthetic pesticides have been increasingly produced globally (Tilman et al.,
40 2002) and are used widely in agriculture to protect crops from pests, weeds and disease and
41 maintain yields. However, use of pesticides also comes with a range of environmental concerns
42 (Goulson, 2013) such as contamination of soils (Silva et al., 2019), water (Casado et al., 2019),
43 or non-target impacts on biodiversity (Beketov et al., 2013; Pisa et al., 2015). Bees provide
44 essential pollination services for global crops and wild plants (Klein et al., 2007; Ollerton et
45 al., 2011), but can come into contact with pesticides in a variety of ways when foraging in
46 agricultural areas or pollinating crops. While some evidence suggests that fungicides and
47 herbicides have implications for bees (Cullen et al., 2019), most concerns relate to insecticides
48 which are designed specifically to control insects by targeting aspects of their biology.

49 Insecticides are used globally to control insect pests of crops. Insecticide use varies regionally
50 and makes up the lowest weight of the main three pesticide groups (fungicide, herbicide,
51 insecticide) in Europe (12%) compared to the highest in Africa (30%) (FAO, 2021). Chemical
52 insecticides (i.e., from artificially derived substances) such as organochlorines, carbamates,
53 pyrethroids, organophosphates were in use already before the launch of the neonicotinoid
54 insecticides in the 1990s. The availability of neonicotinoids resulted in a shift and dramatic
55 increase in the use of this particular insecticide class, which soon had the highest market share
56 of any insecticide class in the world in 2008 at 24% (Elbert et al., 2008; Jeschke et al., 2011)
57 which was followed by pre-existing pyrethroids (16%), organophosphates (14%) and
58 carbamates (11%). However, the use of the neonicotinoids has been restricted in various
59 regions such as the EU, where three neonicotinoids are banned from outdoor use since 2018
60 (EC, 2013; EC, 2018a; EC, 2018b; EC, 2018c). This, alongside other factors such as pesticide
61 resistance has led to a change in insecticide usage patterns globally, and more recently, the
62 most used active substances with solely insecticidal properties, in terms of application rate,
63 again include organophosphates (malathion, chlorpyrifos, dicotophos and acephate) and the
64 pyrethroids (lambda-cyhalothrin) (Maggi et al., 2019).

65 Although there is huge variety and changing trends in insecticide usage globally, there has been
66 a strong research focus on the impacts of neonicotinoids on bees (Abati et al., 2021; Godfray
67 et al., 2015; Lundin et al., 2015). This is likely due to a number of reasons; (i) their widespread
68 use, (ii) worries among beekeepers who had noted reduced bee fitness after foraging from
69 treated crops, but also (iii) their systematic properties and persistence resulting in detectable
70 concentrations in nectar and pollen long after treatment (Botías et al., 2015; David et al., 2016).

71 Although the concentrations of neonicotinoids that bees usually encounter when foraging in
72 real landscapes are far from lethal they can have sublethal effects on bees such as on homing
73 ability and foraging (Stanley et al., 2016). Given the widespread use of other insecticide
74 classes, particularly in areas where neonicotinoids have been restricted, a major question arises
75 - what scientific research has been carried out on insecticides and bees more broadly and where
76 do the knowledge gaps lie? This is essential in understanding the hazards and risks of
77 insecticides to bees, informing regulatory testing, and designing effective mitigation measures.

78 In addition to a focus on the study of neonicotinoids and bees, another potential bias in the
79 literature is around the species studied. Although there are 20,000 species of bees globally
80 (Michener, 2007), most research on pesticides has focussed on one key species, the honeybee
81 (*Apis mellifera*; Abati et al., 2021; Cullen et al., 2019; Lundin et al., 2015; Tosi et al., 2022).
82 The honeybee is a key crop pollinator globally (Kleijn et al., 2015) and its domestication and
83 management by beekeepers, as well as its inclusion in ecotoxicological testing for registration
84 of pesticides, has led to it being the focus of much research attention. However, other bee
85 species are also key pollinators of crops and wild plants (Garibaldi et al., 2013; Kleijn et al.,
86 2015; Winfree et al., 2007) and can be exposed to a range of pesticides in the environment in
87 a variety of ways (Main et al., 2020). Evidence suggests that different bee species may be
88 differentially impacted by pesticide use (Cresswell et al., 2012; Rundlöf et al., 2015; Woodcock
89 et al., 2017), and given that most bee species are not managed by beekeepers, they could be
90 more susceptible to pesticides as no explicit interventions are made to protect their
91 health (Straw & Stanley *under review*). Thus, it is crucial to understand how insecticides may
92 impact non-honeybees in order to be able to make accurate recommendations around the
93 hazards and risks they may pose, to inform policy and management.

94 Here, we use a systematic review to quantify what research exists on bees and insecticides.
95 First, we quantify at a high level how much literature has focussed on neonicotinoids in relation
96 to other insecticide classes, and on honeybees (*Apis sp.*) in comparison to other bees. We then
97 focus specifically on non-neonicotinoid insecticides and non-honeybees and evaluate what
98 compounds and bee taxa have been most widely studied, where research has taken place, and
99 the range of methodological approaches used.

100

101

102 2 Methods

103 All steps in this systematic review followed the ROSES systematic review protocol (Haddaway
104 et al., 2018). The software CADIMA (2017) was used for the title/abstract-screening step of
105 the review (see ROSES Flow Diagram in Supplementary material, S1). We carried out
106 consistency checks at the beginning of each stage of the process (abstract screen, full-text
107 screen and data extraction) by cross-checking with co-authors.

108 2.1 Search

109 To capture all literature on bees and pesticides for initial screening, we used the search string
110 (insecticide* OR pesticide*) AND (*bee OR *bees). This allowed us to capture all literature on
111 both honeybees (*Apis sp.*) and other (non-honeybee) bees, across all insecticide classes. As the
112 terms pesticide and insecticide are often used interchangeably, both were included in our search
113 string. Searches were first performed on 17 August 2020, and later updated with a second
114 search 17 February 2022 to capture any additional literature published in that time. Searches
115 were run in three databases: Web of Science Core Collection (all databases), Scopus and
116 PubMed.

117 Following this initial search, duplicates were removed and then titles and abstracts were
118 screened for relevance to our review criteria (see review protocol, S1). Articles looking at
119 honeybees only and/or neonicotinoid insecticides only, as well as those investigating
120 biopesticides, were recorded for high-level quantification, but did not pass to the next stage of
121 screening as they did not meet our criteria for detailed review. At title and abstract screening
122 stage a conservative approach was taken, where if the focus of the articles was not clear they
123 were included for the full-text screening stage. At the full-text screening stage, only articles
124 that met our pre-defined inclusion criteria for detailed review were included.

125 2.2 Inclusion criteria

126 To be included in the detailed data extraction for this review, articles had to assess the effect
127 of at least one (synthetic) non-neonicotinoid insecticide(s) on at least one non-honeybee (wild
128 or domesticated) bee species. Reasons for exclusion were categorised as follows (see also flow
129 diagram, S1); (abstract or full-text) not in English, research on biopesticides (e.g. essential oils,
130 etc), non-insecticide pesticides (such as; herbicides, fungicides, acaricides), pesticide not
131 specified, farming practice only (including organic farming, cohorts of unspecified pesticides,
132 etc), neonicotinoids, honeybees, and other (to include; residues other than those in bees, method
133 paper, review, not peer-reviewed, not about bees or pesticides). Those excluded for being about

134 honeybees, neonicotinoids and biopesticides were identified and included for the high level
135 quantification. Although every effort was made to capture the number of articles that would
136 have been included had they been in English (see Nuñez and Amano (2021) on how
137 monolingual searches can limit and bias results), there will be articles where both the abstract
138 and full-text were not in English that will have been missed.

139 2.3 Database and data extraction

140 Data was extracted from all articles that made it through the full-screen stage (S1) (i.e., detailed
141 data extraction). In addition to the bibliographic data (including Title, Author, Publication
142 Year), the following data were extracted from each article into a main database (S3):
143 Methodological approach, Geographic distribution; Bee taxa; Insecticide type and Effect type.
144 (See S2 Table 1 for full list of database headings and their definitions). Database categories
145 were initially tested, and then revised throughout the data extraction process where needed
146 while ensuring consistency with those already processed (see Pickering and Byrne (2014)
147 Figure 1, steps 6-10).

148 2.4 Data manipulation

149 The number of articles in each variable of relevance to our research questions was then
150 extracted from our database. Some articles investigated multiple categories per variable, in
151 which case these articles were counted more than once. One study was conducted in more than
152 one geographic location, in this case, the weighting was distributed accordingly e.g. split into
153 0.5 and 0.5. Some extracted data required subsequent categorisation before data could be
154 analysed. For example, (i) Active Ingredients were assigned to substance groups using the
155 Pesticide Properties Database (Lewis et al., 2016) and further refined to insecticide class based
156 on definitions given by ALS (2013) and expert advice (personal communication). Although
157 there is some debate as to how closely related the substance class sulfoximine are to
158 neonicotinoids, for the purpose of this review we classed them separately. (ii) Effect type data
159 were assigned to one of five broad effect type categories (Behaviour; Reproduction/Biomass;
160 Mortality; Physiological, sensory or morphological; and Other effects), and (iii) Synergistic
161 effect types were further categorised into broad synergistic effect groups. All categorizations
162 are visible in the database (S3). All data handling was conducted in R version 4.1.0 (R Core
163 Team, 2021). We used ggplot2 for the heatmaps (Wickham, 2016), rworldmap (South, 2011) for
164 the maps, plotrix (Lemon, 2006) for the pie charts and colorspace (Zeileis et al., 2019) to define
165 RGB colours.

166 **3 Results**

167 The initial database search yielded 9,643 articles (Supplementary material, S1). After duplicate
168 removal, a total of 4,772 articles were screened by title and abstract. The bulk of these articles
169 were excluded as they did not meet our inclusion criteria for detailed data extraction (although
170 data were extracted from those on honeybees and neonicotinoids for high level quantification).
171 The remaining 309 articles were then full-text screened and half of these subsequently excluded
172 with reasons (e.g. biopesticides) (see S1). By the end of the screening process, 138 primary
173 research articles met our inclusion criteria for detailed data extraction as primary research on
174 non-honeybees and (synthetic) non-neonicotinoid insecticides.

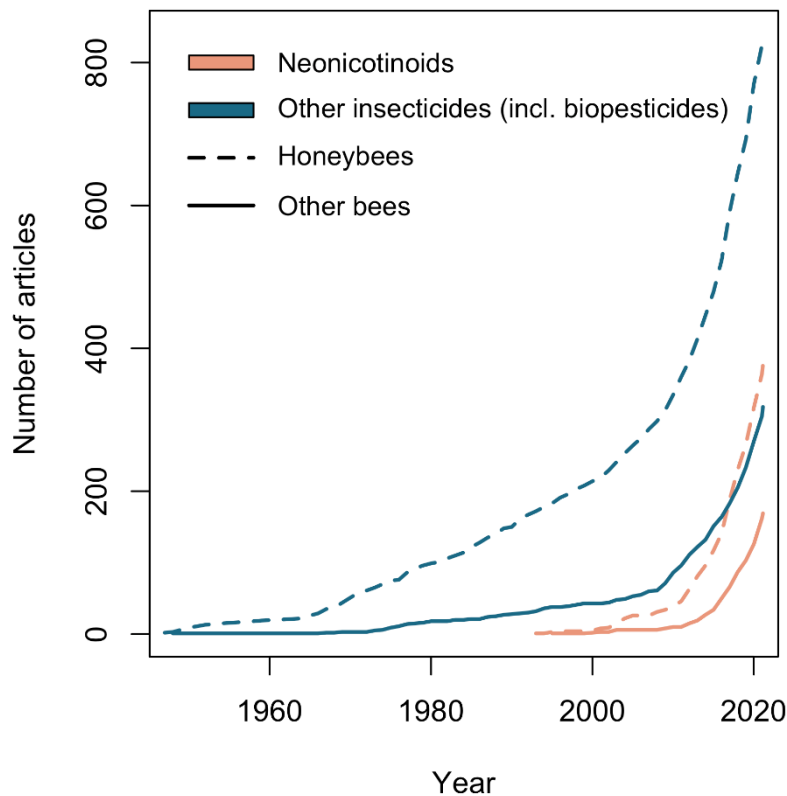
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176 **High level quantification**

177 **3.1 Comparison with honeybees, and neonicotinoid insecticides**

178 The research on bees and insecticides identified as part of our high level quantification showed
179 a strong bias towards the study of honeybees. During screening we identified 1,538 articles
180 about bees (honeybees and other bees) and insecticides (neonicotinoids and other insecticides,
181 including biopesticides), 80% of which were on honeybees (n = 1,215). Similarly, of the 1,132
182 articles looking at bees and non-neonicotinoid insecticides (Figure 1, blue lines), 74% were on
183 honeybees (n = 839 articles). Bee and insecticide publications began c. 1950s, but those about
184 honeybees increased at a much more rapid rate than for other bees (increase in the 1970s), with
185 research on both groups of taxa appearing to accelerate around 2010.

186 Articles on neonicotinoid insecticides were also prominent in the literature screened (Figure 1,
187 pink lines). Beginning in the early 1990s, for both honeybees and other bees, the number of
188 published articles about neonicotinoids increased at a much higher rate than articles about other
189 insecticides. Overall, articles looking at neonicotinoid insecticides made up a third (34%) of
190 the search results on all bees and all insecticides (n = 515) (Figure 1). A closer look at these
191 results specifically for non-honeybees shows that neonicotinoids are by far the most studied
192 insecticide class (Figure 2) overall, with almost three times as many studies (186 compared to
193 59) than pyrethroids which are the second most studied class.



194

195 Figure 1 The number of articles on honeybee [dashed line] and other (non-honeybee) bee [solid line]
196 research for neonicotinoids [orange lines] and other (non-neonicotinoid) insecticides [blue lines] over time.

197

198 **Detailed data extraction (non-neonicotinoids and non-honeybees)**

199 3.2 *Insecticide type*

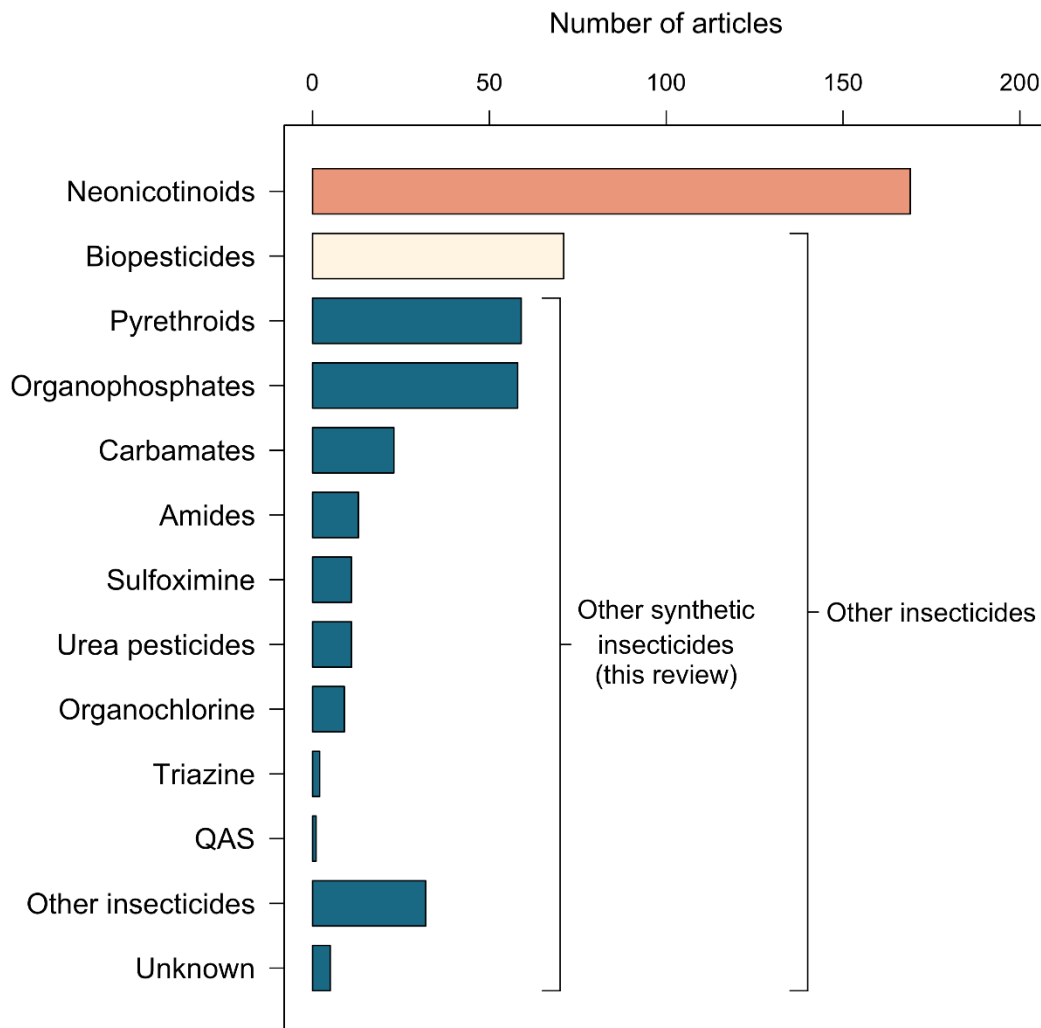
200 Of all the research included in the detailed data extraction for this review (on non-honeybees
201 and non-neonicotinoids), the most studied insecticide classes were pyrethroids (included in
202 43% of articles, n = 59), organophosphates (42%, n = 58) and carbamates (17%, n = 23). The
203 remaining insecticide classes together were included in 49% of the articles (n = 67) (Figure 2).
204 Most articles looked at more than one insecticide, which is why their sum is >100%. Although
205 articles on biopesticides were excluded from further data analysis, we note that 22% (n= 71
206 articles) of the articles about other (non-neonicotinoid) insecticides (Figure 2) assessed
207 biopesticide(s).

208 The type of insecticide and the amount it was researched, changed over time (S4 Figure 1).

209 Although the first ever study on bees using synthetic insecticides was recorded from the 1950s,

210 it was not until the 1980/90s that there was an increase in the diversity of insecticide types

211 researched, and not until the past decade that we see an increase in the quantity of articles being
212 published.



213

214 Figure 2 The number of articles for each insecticide class on non-honeybees; neonicotinoid insecticides
215 (orange bar), biopesticides (beige bar) and other (non-neonicotinoid) insecticides (blue bars). The blue bars
216 represent substance classes of synthetic non-neonicotinoid insecticides that were the focus of this systematic
217 review.

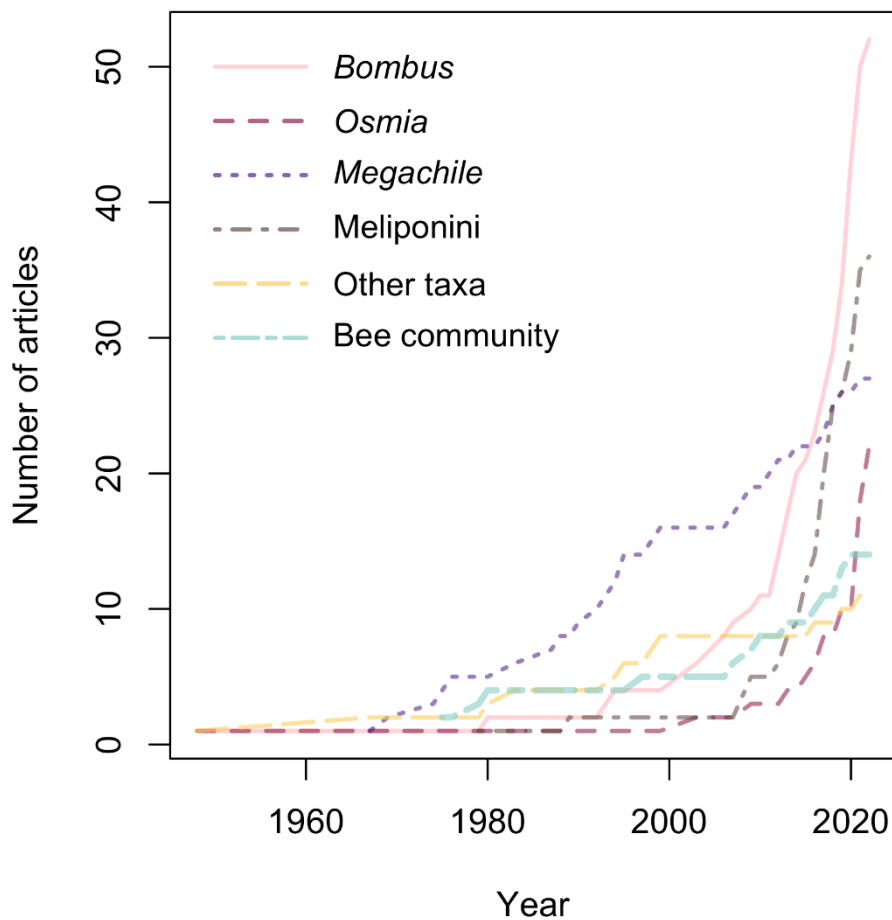
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219 3.3 Bee taxa

220 Of all the non-honeybees studied, the most widely represented bee taxon was *Bombus* (in 38%
221 of articles, $n = 52$), followed by *Megachile* ($n= 27$) and *Osmia* ($n =22$). Bee genera
222 *Tetragonisca*, *Scaptotrigona*, *Partamona*, *Trigona*, *Plebeia*, and *Melipona* were represented
223 much less so (between 2-12 articles each). All six genera fall under the tribe Meliponini and
224 combined are the second most represented bee taxon ($n= 36$). Eleven articles had assessed

225 other taxa (see S2 Table 2 for full list of species), and 14 had assessed effects on the bee
226 community (Figure 3).

227 Among the bee groups, the research interest for *Megachile* has been steady since the 1960s,
228 with around five published articles per decade. In recent years, there has been an increasing
229 focus on *Bombus*, *Osmia* and Meliponini (Figure 3). Of particular note is Meliponini,
230 exhibiting the steepest rate of increase during the past decade.



231

232 Figure 3 **The number of articles over time broken down by bee taxa.** Each colour/line type (pink, red, purple,
233 brown, yellow and green) represents a different bee taxa; *Bombus*, *Osmia*, *Megachile*, Meliponini, other taxa and
234 bee community. The six genera within the tribe Meliponini (*Partamona*, *Trigona*, *Tetragonisca*, *Scaptotrigona*,
235 *Plebeia*, and *Melipona*) were not individually included in order to improve the interpretability of the figure.

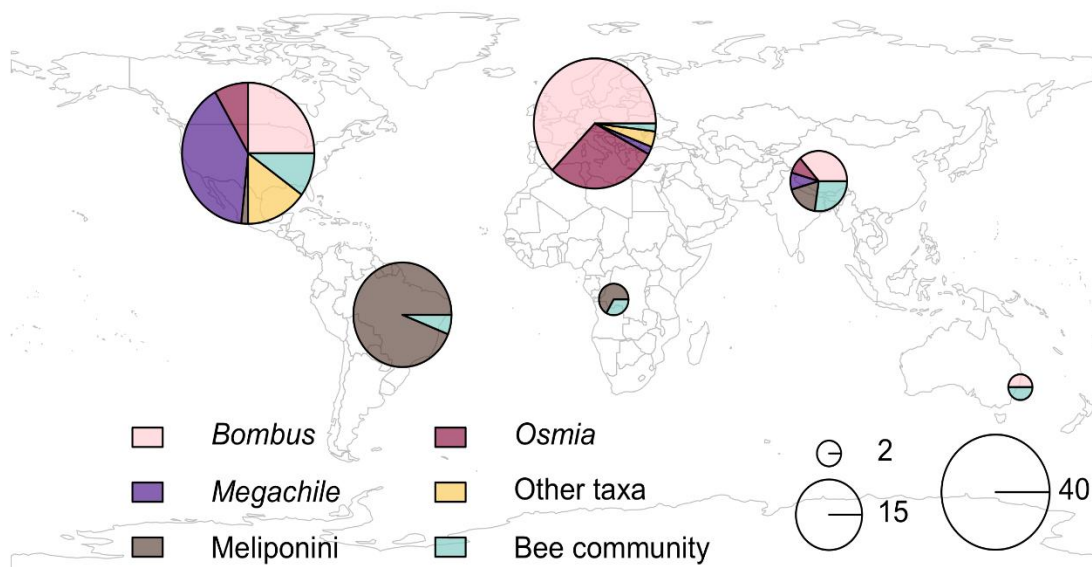
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237 3.4 Geographic distribution

238 We identified articles on non-honeybees and synthetic non-neonicotinoid insecticides from a
239 total of 22 countries. The country with the highest number of articles was Brazil (n = 33),
240 followed by the United States (n = 30), Canada and United Kingdom (n = 16 each). The

241 remaining countries accounted for less than 30% of the total articles included in this review;
242 Belgium and Italy ($n = 7$), Poland, Germany, China, France ($n \leq 5$ each), Australia,
243 Czechoslovakia, Estonia, Mexico, New Zealand, Nigeria, Pakistan, Sri Lanka, and Zimbabwe
244 ($n = 1$ each). The continent with the highest number of articles was North America ($n = 48$),
245 followed by Europe ($n = 43$), South America ($n = 33$), Asia ($n = 9$), Africa ($n = 2$) and Oceania
246 ($n = 2$; S4 Figure 2). There is a mismatch between some of these trends and pesticide usage
247 (tonnes) globally (S4, Figure 3). Where pesticide usage (tonnes) in order of highest to lowest
248 are; Asia (406k), South America (92k), Europe (70k), North America (69k), Africa (29k) and
249 Oceania (15k).

250 The bee groups studied differed among continents; for example Meliponini dominated research
251 from South America and Africa, while *Bombus* dominated research from Europe and
252 *Megachile* was the most studied bee group in North America (Figure 4). For the three
253 dominating continents (North America, Europe and South America), which together accounted
254 for 91% of the articles, we found that 92% of the South American articles, 76% of the European
255 articles and 54% of the North American articles had been published after 2011.



256

257 Figure 4 **Geographic distribution of articles with breakdown of bee taxa.** The area of the circles is proportional
258 to the number of articles, and the colours indicate the proportion of articles from each continent that focuses on
259 each of the bee taxa (*Bombus*, *Osmia*, *Megachile*, Meliponini, other) or wild bee community.

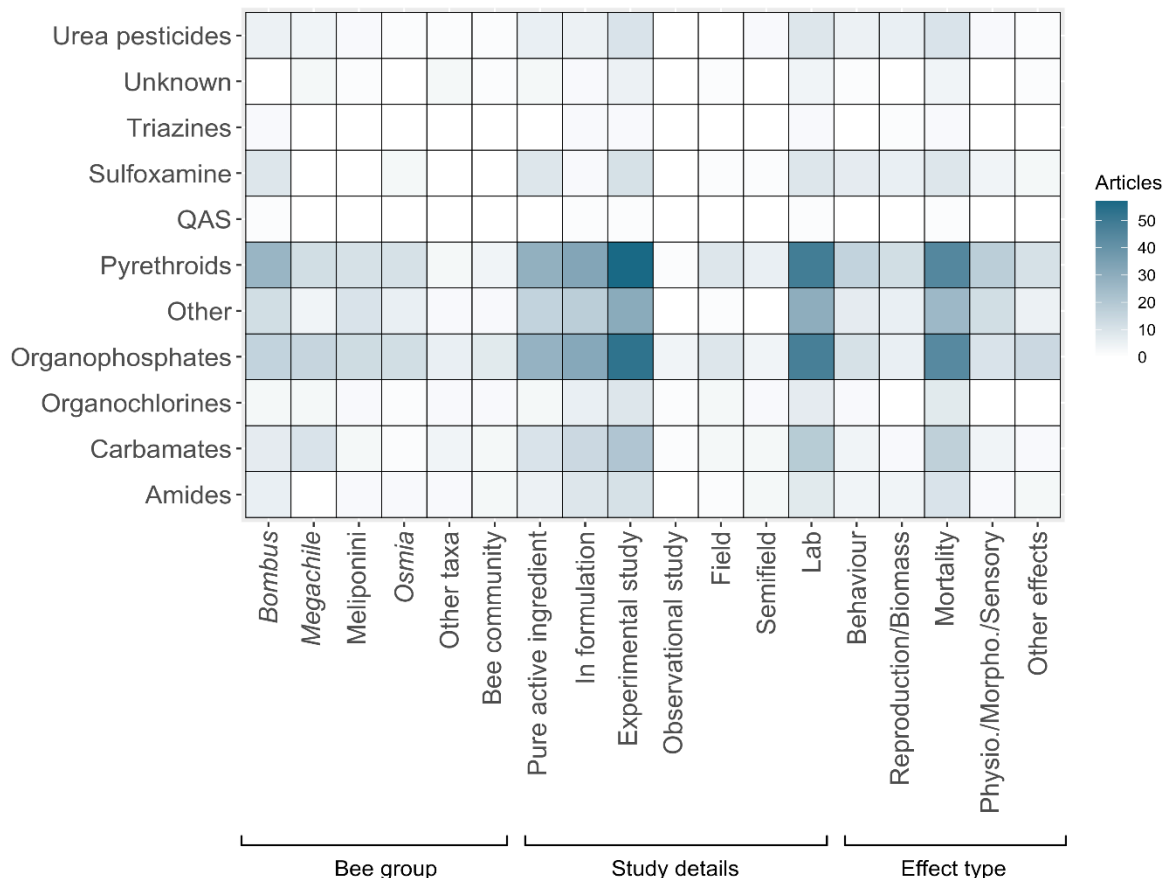
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261

262 3.5 Methodological approaches

263 The most studied effect type was mortality (79% of articles, n = 109) followed by those looking
 264 at behaviour; such as foraging, nesting, etc (33% of articles n = 46), sensory, morphological or
 265 physiological effects (31% of articles n = 43) and effects that were about reproduction/biomass;
 266 such as offspring/worker/male/queen production, sex ratio, etc (25% of articles, n = 34). Any
 267 remaining other effect types such as pollination services, genomic, species richness, etc. were
 268 categorised as ‘other’ (17% of articles, n = 24) (Figure 5, S3 heatmap raw data). All of the
 269 these sub-lethal effect types once combined are close to but still less than the cumulative
 270 number of articles on mortality effects (S4 Figure 4).

271 Ninety-six percent of articles (n=132) were experimental (rather than observational) articles,
 272 carried out in the laboratory (82%, n = 113), rather than field or semi-field experiments. The
 273 use of both, insecticide active ingredients and formulations when designing experiments was
 274 similar overall (Figure 5, S3 heatmap raw data).



275

276 Figure 5 Heatmap showing the range of non-honeybee taxa, article details and effect type on the x-axis for
 277 the different types of synthetic insecticides on the y-axis. (colour scale on right indicated the number of studies
 278 e.g. the darker the colour the more studies)

279

280 *3.5.1 Synergistic and Cocktail effects*

281 A number of articles investigated synergistic or cocktail effects overall. Nine percent of the
282 articles (n =12) assessed the effect of a synthetic non-neonicotinoid insecticide together with
283 another non-pesticide stressor (synergistic effects). A total of five such effects were recorded;
284 the most studied stressors were parasites and substances that are combined with insecticides to
285 increase their intended effect (synergist component e.g. piperonyl butoxide; S4 Figure 5) (n =
286 3 respectively). Followed by adjuvant, diet and other (n = 2 respectively). All except one article
287 were experimental lab based studies looking at the effect on mortality. These all took place
288 across two countries, the UK and USA.

289 Twenty-four articles (17%) assessed cocktail effects between a synthetic non-neonicotinoid
290 insecticide and other pesticide(s). The most studied cocktail effects were with neonicotinoids
291 (n = 15), followed by fungicides (n = 8) (S4 Figure 6). Over half the articles were experimental
292 lab based studies (63%), with the majority looking at the effect on mortality (71%). These
293 studies took place across three continents.

294

295

296 **4 Discussion**

297 There has been increasing interest in the environmental impacts of pesticides, and in particular
298 the area of bees and pesticides has received a lot of public and research attention. Here we
299 show that there has been a focus on research on honeybees and the neonicotinoid class of
300 insecticides (for non-honeybees). However, despite an increase in research in this area over
301 time, there are still knowledge gaps, in particular around the use of synthetic non-neonicotinoid
302 insecticides, non-honeybees and sublethal effects.

303 *4.1 Comparison with honeybees, and neonicotinoid insecticides*

304 Honeybees are important crop pollinators globally, with large, often domesticated, populations
305 in Europe and Africa, and elsewhere outside their native range. Honeybees have been used as
306 a model species to represent bees in pesticide risk assessment and regulatory testing (OECD).
307 It is therefore unsurprising that we show the majority of bee and insecticide research has
308 focussed on honeybees, with all other bee species less well represented. This trend is mirrored
309 in other reviews of various pesticide groups and types (Abati et al., 2021; Cullen et al., 2019;
310 Lundin et al., 2015; Tosi et al., 2022). However, there are more than 20,000 bee species in the

311 world. Many of these taxa differ from honeybees in ways that influence insecticide-related
312 risks. For example; differing life histories, sociality, nesting behaviour, foraging range and
313 floral preference can impact the risk of exposure (Knapp et al., 2023); Willis Chan et al.
314 (2019), and genetic factors and body size can moderate the bees' sensitivity to insecticides
315 (Devillers et al., 2003; Hayward et al., 2019), so that insecticide risk estimates for non-
316 honeybees cannot be accurately extrapolated from honeybees (Cresswell et al., 2012; Rundlöf
317 et al., 2015; Woodcock et al., 2017). Thus, there is a clear need for more research to understand
318 impacts of insecticides on wider, non-honeybee diversity. Honeybee research has been
319 increasing gradually over time, while non-honeybee research has increased a lot in rate since
320 the mid-2000s. This suggests that this pattern is already changing, which may be related to
321 mounting evidence that the sensitivity to pesticides differs among bee species in combination
322 with increasing awareness regarding the importance of wild bees as crop pollinators.

323 Neonicotinoids were the most studied class of insecticides for non-honeybees, despite their
324 relatively short history compared to organophosphates, carbamates and organochlorines. This
325 is due to a very strong increase in neonicotinoid research from around 2010, which was around
326 two decades after the first neonicotinoids had been released on the market and likely reflects
327 their importance in agriculture, in combination with increasing awareness that they may pose
328 a risk to bees. A huge diversity of synthetic insecticides have been developed for pest control,
329 of which the neonicotinoids are only one class. Given the wide global usage of neonicotinoids
330 and reduction in the use of this class specifically in some regions (e.g. the EU) we need to
331 understand more about non-neonicotinoids and bees.

332 **4.2 Insecticide type**

333 It is not surprising that most articles on non-honeybees and non-neonicotinoids have
334 investigated impacts of pyrethroids, organophosphates and carbamates, given the wide global
335 usage of these classes, some of which currently are now more dominant than the neonicotinoids
336 (Maggi et al., 2019). It is possible that the fact that these substances now are more used than
337 the neonicotinoids, will change the pattern, so that neonicotinoid research is less dominant in
338 a near future.

339 The diversity of insecticide compounds researched in this context has increased over time, with
340 some classes such as the sulfoximines appearing in research relatively recently (since 2018),
341 reflecting the development of new chemistries and products for the insecticidal market. The
342 emergence of such novel insecticides is arguably a necessity, in response to continued

343 insecticide resistance and negative sub-lethal impacts on beneficial insects. However, the
344 benefit of replacing harmful substances with novel compounds has been questioned as they
345 sometimes have similar sub-lethal impacts (Siviter and Muth, 2020). Although not covered in
346 great detail, this review highlights that there is non-honeybee research taking place into the
347 effects of naturally derived substances e.g. biopesticides.

348 **4.3 Bee taxa**

349 Although we see a diversity of non-honeybees being studied in the context of insecticides, there
350 are interesting trends in the species studied. The bumblebees (*Bombus sp.*) are most widely
351 represented in the literature for non-honeybees and non-neonicotinoids, with a visible increase
352 in rate of research since 2010. These species are dominant pollinators in the Northern
353 Hemisphere, and some species have been domesticated since the 1990s. This has made them a
354 good candidate for research, as they are commercially available in many countries. The two
355 genera of solitary bee *Megachile* and *Osmia* are also well represented in the literature (with the
356 latter genus increasing rapidly from 2010 and the former group increasing steadily since the
357 1970s), again this is most likely to due to their importance as crop pollinators as well as the
358 availability of domesticated species. Interestingly, there has been a rapid increase in research
359 on stingless bees Meliponini and non-neonicotinoid insecticides, especially since the mid-
360 2000s. This tribe of bees are especially important pollinators in tropical regions. Most of this
361 research comes from Brazil and might reflect increased research interest or capacity in this
362 country, or a reaction to changed pesticide policies in Brazil (e.g. Braga et al., 2020). As the
363 response to pesticide exposure can differ in magnitude and effect type among taxa (as
364 mentioned above), it is encouraging to see pesticide risk assessments are beginning to move
365 beyond focusing solely on honeybees to also include *Bombus* and *Osmia* species (OECD).
366 However, there are at least 20,000 documented bee species in the world, and it has been
367 estimated that more than 10% of the regional species pool visits crop flowers (Kleijn et al.
368 2015), suggesting that several hundreds of bee species can be exposed to insecticides. We
369 identified articles on 42 non-honeybee species and 15 studies on wild bee communities from
370 nine countries, showing that most species are still under-researched or are not researched at all
371 in terms of impacts of insecticides. Given that scientific research is often an important pre-
372 cursor to making decisions around the development and implementation of risk assessment
373 protocols, the lack of research on a range of non-honeybee taxa suggests that we simply lack
374 the knowledge to make informed decisions about many bee species in a regulatory and risk
375 assessment context, and highlights a key knowledge gap for future work.

376 **4.4 Geographic distribution**

377 As with much research on bees and pesticides (c. Cullen et al. (2019); Lundin et al. (2015)),
378 the majority of research on non-honeybees and non-neonicotinoid insecticides has come from
379 North America and Europe. This may reflect capacity and interest in bee-related research, but
380 also policy requirements. In contrast to previous reviews about bees and pesticides, we identify
381 South America as one of the top producers of articles. We suggest two reasons for this. First,
382 as we show, honeybee research is dominating the field at a global level, probably because they
383 are easy to access and important for food production. However, in South America, stingless
384 bees are often used instead of honeybees, suggesting that our scope resulted in the inclusion of
385 a larger proportion of the bee and insecticide research from South America than from North
386 America and Europe. Second, it is possible that the very recent increase in articles from South
387 America was captured in this, but not in other reviews, only because our literature search was
388 more recent. The high and rapidly increasing number of articles from South America is solely
389 driven by Brazil and likely reflects an ongoing debate regarding the insufficient pesticide
390 regulations in the country (see e.g. Braga et al., 2020).

391 Both insecticide usage patterns and bee communities differ substantially across the globe and
392 extrapolation of results from South America, North America and Europe to other parts of the
393 world might therefore not be possible. In addition, most insect pollinated crops are grown
394 outside Europe and North America (Gallai et al., 2009), showing the importance of
395 understanding interactions between insecticides and bees globally. Knowledge gaps from
396 places like Asia and Africa, which have high levels of undiscovered bee diversity (Orr et al.,
397 2021) and these as well as Oceania produce large amounts of insect pollinated crops (USDA,
398 2022a; USDA, 2022b; USDA, 2022c), are of concern. Equally, relatively low number of
399 research from Asia is of concern given that it accounts for 62% of the global pesticide usage
400 (FAOSTAT 2022, see S4 Figure 3).

401 **4.5 Methodological approaches**

402 Across all bee groups and insecticide classes included, there were consistent patterns in the
403 type of insecticide effects studied. Mortality was the most widely studied effect type for all
404 non-honeybee groups, with less focus on sub-lethal effects such as on behaviour or
405 reproduction. This may be unsurprising, as the vast majority of pesticide risk assessment to
406 date has focused on measures of mortality, such as LD₅₀. However, through the large number
407 of studies on neonicotinoids, there is increasing recognition that insecticides can have a wide
408 variety of sub-lethal effects (Desneux et al., 2007), such as on reproduction (Rundlöf et al.,

409 2015), learning ability (Stanley et al., 2015b; Williamson and Wright, 2013) and even delivery
410 of pollination services (Herbertsson et al., 2022; Stanley et al., 2015a) which may have longer
411 term implications for bee populations (Woodcock et al., 2016), ecosystems and crop
412 production. Although this review shows that a number of sub-lethal effects have been
413 investigated with increasing interest over the last decade (S4 Figure 4), crucially we note that
414 few articles assessed pollination services (n = 1), genomic effects (n = 2), and learning ability
415 (n = 1) and none assessed navigation (n = 0). Within the EU there is increased provision for
416 including sub-lethal effects in pesticide risk assessments for bees; however, the lack of research
417 in this area for non-honeybees means that for many taxa the scientific knowledge on sub-lethal
418 effects to inform risk assessment development and focus is lacking and represents another key
419 knowledge gap.

420 Most research on non-honeybees and non-neonicotinoid insecticides has taken place in
421 laboratory settings, and this pattern is consistent across all bee groups and insecticide classes
422 studied. Lab research allows much stronger control of experimental conditions but has been
423 criticised for not being field-realistic where, for example, it does not reveal how bee fitness is
424 affected in complex environments (Mommaerts et al., 2010). A combination of lab, semi-field
425 and field articles can build a clearer picture of impacts of pesticides from hazard to exposure
426 and ultimately risk. As such, this review highlights the need for more research on non-
427 neonicotinoid insecticides and non-honeybees in semi-field and field settings.

428 *4.5.1 Synergistic and cocktail effects*

429 External stressors can modulate how bees respond to pesticide exposure, but only a small
430 proportion of the articles included in this review assessed insecticides in combination with
431 another stressor (synergistic), or other pesticides (cocktail). Evidence for cocktail effects or
432 synergistic effects from multiple stressors is to date primarily from research on honeybees
433 (Siviter et al., 2021), where for example nutrition (Tong et al., 2019) or pathogens (Grassl et
434 al., 2018) can affect how bees respond to pesticide use. These interactions and context specific
435 impacts of pesticides are particularly important to understand as they also provide potential
436 avenues for reducing or mitigating effects of insecticides. For example, if impacts of a pesticide
437 are reduced when bees have access to more diverse forage and nutrition in a landscape (as seen
438 in Wintermantel et al. (2022)), then this could be implemented as a measure through agri-
439 environmental management to mitigate pesticide effects (Rundlöf et al., 2022).

440 This added complexity to assessing risk is important and valuable especially in the context of
441 modelling future impacts of insecticide use on non-target organisms. There is much scope and
442 urgency for expanding research on insecticides and their effects on non-honeybees, in the
443 context of global change and the reality that insecticides are applied in combination with other
444 insecticides, fungicides and herbicides. Meta-analyses show that multiple stressors (in the form
445 of synergistic and cocktail effects) can have an accumulative negative effect on bees (Siviter
446 et al., 2021; Tosi et al., 2022). Even less investigated is the area of co-formulants, but recent
447 evidence shows that fungicide co-formulants can have adverse effects on bumblebees (Straw
448 and Brown, 2021), underlining the need for more research in this field. This review highlights
449 that there is a substantial knowledge gap for the effect of multiple stressors on non-honeybees
450 exposed to synthetic non-neonicotinoid insecticides.

451

452

453 **5 Conclusions and future direction**

454 This review quantifies the extent of research articles on synthetic insecticides and bees. Despite
455 the high diversity in bee species globally with different life history traits that can modulate
456 their susceptibility to insecticides, and the wide variety of insecticide compounds used, we
457 confirm a bias towards research on honeybees and neonicotinoids. When focussing on the
458 literature on non-honeybees and non-neonicotinoids, the findings highlight the need for
459 expanding research on a diversity of bee taxa, in a variety of geographic regions (particularly
460 Asia, Africa and Oceania), and methodological approaches used. In particular, the focus on
461 *Bombus*, *Osmia*, *Megachile* and Meliponini means that understanding of other bee taxa is
462 sparse, while the focus on mortality indicates that knowledge of sub-lethal effects is
463 substantially behind. Both of these demonstrate that the scientific underpinning to support
464 recent developments in including non-honeybees and sub-lethal effects in assessing risk of
465 pesticides (e.g., within the EU) is not comprehensive, and requires more focus to best inform
466 policy, risk assessment and bee conservation. Given the growing recognition of the value of
467 pollinating insects to global food security and the increasing demand for sustainable solutions
468 to crop protection we suggest that research on insecticides and bees also investigate combined
469 effects on bees such as other pesticides and/or other pressures such as climate change.

470

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474

475

476 **Supporting Information**

477 S1 ROSES flow diagram for systematic review (PDF)

478 S2 Table 1 Database headings and definitions and Table 2 Species list (PDF)

479 S3 Systematic review raw data(base) and Heatmap raw data (XLS)

480 S4 Figure 1 Heatmap number articles of non-honeybee per insecticide class over time; Figure
481 2 Geographic distribution of articles based on country; Figure 3 World map of insecticide use
482 (tonnes) in 2019; Figure 4 Effect type investigated over time; Figure 5 Number of synergistic
483 effects articles; Figure 6 Number of cocktail effect articles (PDF)

484 R script available on request.

485

486

487 **References**

- 488 CADIMA. <https://www.cadima.info/>. Julius Kühn-Institut, Quedlinburg, Germany, 2017.
489 Abati, R., et al., 2021. Bees and pesticides: the research impact and scientometrics relations.
490 *Environmental Science and Pollution Research*. 28, 32282-32298.
491 ALS, Overview of pesticide classes . www.alsglobal.eu (accessed: 4 July 2022). ALS Europe,
492 2013.
493 Beketov, M. A., et al., 2013. Pesticides reduce regional biodiversity of stream invertebrates.
494 *Proceedings of the National Academy of Sciences*. 110, 11039-11043.
495 Botías, C., et al., 2015. Neonicotinoid Residues in Wildflowers, a Potential Route of Chronic
496 Exposure for Bees. *Environ Sci Technol*. 49, 12731-40.
497 Braga, A. R. C., et al., 2020. Global health risks from pesticide use in Brazil. *Nature Food*. 1,
498 312-314.
499 Casado, J., et al., 2019. Screening of pesticides and veterinary drugs in small streams in the
500 European Union by liquid chromatography high resolution mass spectrometry. *Science*
501 *of The Total Environment*. 670, 1204-1225.
502 Cresswell, J. E., et al., 2012. Differential sensitivity of honey bees and bumble bees to a dietary
503 insecticide (imidacloprid). *Zoology (Jena)*. 115, 365-71.
504 Cullen, M. G., et al., 2019. Fungicides, herbicides and bees: A systematic review of existing
505 research and methods. *PLoS One*. 14, e0225743.

- 506 David, A., et al., 2016. Widespread contamination of wildflower and bee-collected pollen with
507 complex mixtures of neonicotinoids and fungicides commonly applied to crops.
508 *Environ Int.* 88, 169-178.
- 509 Desneux, N., et al., 2007. The sublethal effects of pesticides on beneficial arthropods. *Annu*
510 *Rev Entomol.* 52, 81-106.
- 511 Devillers, J., et al., 2003. Comparative toxicity and hazards of pesticides to Apis and non-Apis
512 bees. A chemometrical study. *SAR QSAR Environ Res.* 14, 389-403.
- 513 EC, Commission Implementing Regulation (EU) No 485/2013 of 24 May 2013 amending
514 Implementing Regulation (EU) No 540/2011, as regards the conditions of approval of
515 the active substances clothianidin, thiamethoxam and imidacloprid, and prohibiting the
516 use and sale of seeds treated with plant protection products containing those active
517 substances. 2013, pp. 12-26.
- 518 EC, Commission Implementing Regulation (EU) 2018/783 of 29 May 2018 amending
519 Implementing Regulation (EU) No 540/2011 as regards the conditions of approval of
520 the active substance imidacloprid (Text with EEA relevance.). 2018a, pp. 31-34.
- 521 EC, Commission Implementing Regulation (EU) 2018/784 of 29 May 2018 amending
522 Implementing Regulation (EU) No 540/2011 as regards the conditions of approval of
523 the active substance clothianidin (Text with EEA relevance.). 2018b, pp. 35-39.
- 524 EC, Commission Implementing Regulation (EU) 2018/785 of 29 May 2018 amending
525 Implementing Regulation (EU) No 540/2011 as regards the conditions of approval of
526 the active substance thiamethoxam (Text with EEA relevance.). 2018c, pp. 40-44.
- 527 Elbert, A., et al., 2008. Applied aspects of neonicotinoid uses in crop protection. *Pest*
528 *Management Science: formerly Pesticide Science.* 64, 1099-1105.
- 529 FAO, Pesticides use. Global, regional and country trends 1990–2018. . FAOSTAT Analytical
530 Brief 16, Rome, 2021.
- 531 Gallai, N., et al., 2009. Economic valuation of the vulnerability of world agriculture confronted
532 with pollinator decline. *Ecological Economics.* 68, 810-821.
- 533 Garibaldi, L. A., et al., 2013. Wild pollinators enhance fruit set of crops regardless of honey
534 bee abundance. *science.* 339, 1608-1611.
- 535 Godfray, H. C., et al., 2015. A restatement of recent advances in the natural science evidence
536 base concerning neonicotinoid insecticides and insect pollinators. *Proc Biol Sci.* 282,
537 20151821.
- 538 Goulson, D., 2013. An overview of the environmental risks posed by neonicotinoid
539 insecticides. *Journal of Applied Ecology.* 50, 977-987.
- 540 Grassl, J., et al., 2018. Synergistic effects of pathogen and pesticide exposure on honey bee
541 (*Apis mellifera*) survival and immunity. *Journal of Invertebrate Pathology.* 159, 78-86.
- 542 Haddaway, N. R., et al., 2018. ROSES Reporting standards for Systematic Evidence
543 Syntheses: pro forma, flow-diagram and descriptive summary of the plan and conduct
544 of environmental systematic reviews and systematic maps. *Environmental Evidence.* 7,
545 1-8.
- 546 Hayward, A., et al., 2019. The leafcutter bee, *Megachile rotundata*, is more sensitive to N-
547 cyanoamidine neonicotinoid and butenolide insecticides than other managed bees. *Nat*
548 *Ecol Evol.* 3, 1521-1524.
- 549 Herbertsson, L., et al., 2022. Seed-coating of rapeseed (*Brassica napus*) with the neonicotinoid
550 clothianidin affects behaviour of red mason bees (*Osmia bicornis*) and pollination of
551 strawberry flowers (*Fragaria× ananassa*). *PloS one.* 17, e0273851.
- 552 Jeschke, P., et al., 2011. Overview of the status and global strategy for neonicotinoids. *Journal*
553 *of agricultural and food chemistry.* 59, 2897-2908.
- 554 Kleijn, D., et al., 2015. Delivery of crop pollination services is an insufficient argument for
555 wild pollinator conservation. *Nature communications.* 6, 1-9.

- 556 Klein, A.-M., et al., 2007. Importance of pollinators in changing landscapes for world crops.
557 *Proceedings of the royal society B: biological sciences.* 274, 303-313.
- 558 Knapp, J. L., et al., 2023. Ecological traits interact with landscape context to determine bees'
559 pesticide risk. (accepted). *Nature ecology & evolution.*
- 560 Lemon, J., 2006. Plotrix: a package in the red light district of R. *R-news.* 6, 8-12.
- 561 Lewis, K. A., et al., 2016. An international database for pesticide risk assessments and
562 management. *Human and Ecological Risk Assessment: An International Journal.* 22,
563 1050-1064.
- 564 Lundin, O., et al., 2015. Neonicotinoid Insecticides and Their Impacts on Bees: A Systematic
565 Review of Research Approaches and Identification of Knowledge Gaps. *PLoS One.* 10,
566 e0136928.
- 567 Maggi, F., et al., 2019. PEST-CHEMGRIDS, global gridded maps of the top 20 crop-specific
568 pesticide application rates from 2015 to 2025. *Scientific data.* 6, 1-20.
- 569 Main, A. R., et al., 2020. Beyond neonicotinoids - Wild pollinators are exposed to a range of
570 pesticides while foraging in agroecosystems. *Sci Total Environ.* 742, 140436.
- 571 Michener, C., 2007. *The Bees of the World* Johns Hopkins University Press.
572 Baltimore.[Google Scholar].
- 573 Mommaerts, V., et al., 2010. Risk assessment for side-effects of neonicotinoids against
574 bumblebees with and without impairing foraging behavior. *Ecotoxicology.* 19, 207-15.
- 575 Nuñez, M. A., Amano, T., 2021. Monolingual searches can limit and bias results in global
576 literature reviews. *Nature Ecology & Evolution.* 5, 264-264.
- 577 OECD, OECD Work Related to Bees/Pollinators. Available from:
578 <https://www.oecd.org/chemicalsafety/testing/work-related-beepollinators.htm>.
579 Organisation for Economic Co-operation and Development
- 580 Ollerton, J., et al., 2011. How many flowering plants are pollinated by animals? *Oikos.* 120,
581 321-326.
- 582 Orr, M. C., et al., 2021. Global Patterns and Drivers of Bee Distribution. *Current Biology.* 31,
583 451-+.
- 584 Pickering, C., Byrne, J., 2014. The benefits of publishing systematic quantitative literature
585 reviews for PhD candidates and other early-career researchers. *Higher Education*
586 *Research & Development.* 33, 534-548.
- 587 Pisa, L. W., et al., 2015. Effects of neonicotinoids and fipronil on non-target invertebrates.
588 *Environ Sci Pollut Res Int.* 22, 68-102.
- 589 R Core Team, R: A language and environment for statistical computing. R Foundation for
590 Statistical Computing, Vienna, Austria, 2021.
- 591 Rundlöf, M., et al., 2015. Seed coating with a neonicotinoid insecticide negatively affects wild
592 bees. *Nature.* 521, 77-80.
- 593 Rundlöf, M., et al., 2022. Flower plantings support wild bee reproduction and may also mitigate
594 pesticide exposure effects. *Journal of Applied Ecology.* 59, 2117-2127.
- 595 Silva, V., et al., 2019. Pesticide residues in European agricultural soils—A hidden reality
596 unfolded. *Science of The Total Environment.* 653, 1532-1545.
- 597 Siviter, H., et al., 2021. Agrochemicals interact synergistically to increase bee mortality.
598 *Nature.* 596, 389-392.
- 599 Siviter, H., Muth, F., 2020. Do novel insecticides pose a threat to beneficial insects?
600 *Proceedings of the Royal Society B-Biological Sciences.* 287.
- 601 South, A., 2011. rworldmap: a new R package for mapping global data. *R Journal.* 3.
- 602 Stanley, D. A., et al., 2015a. Neonicotinoid pesticide exposure impairs crop pollination services
603 provided by bumblebees. *Nature.* 528, 548-50.

- 604 Stanley, D. A., et al., 2016. Investigating the impacts of field-realistic exposure to a
605 neonicotinoid pesticide on bumblebee foraging, homing ability and colony growth.
606 *Journal of Applied Ecology*. 53, 1440-1449.
- 607 Stanley, D. A., et al., 2015b. Bumblebee learning and memory is impaired by chronic exposure
608 to a neonicotinoid pesticide. *Sci Rep*. 5, 16508.
- 609 Straw, E. A., Brown, M. J., 2021. Co-formulant in a commercial fungicide product causes lethal
610 and sub-lethal effects in bumble bees. *Scientific reports*. 11, 1-10.
- 611 Tilman, D., et al., 2002. Agricultural sustainability and intensive production practices. *Nature*.
612 418, 671-677.
- 613 Tong, L., et al., 2019. Combined nutritional stress and a new systemic pesticide
614 (flupyradifurone, Sivanto®) reduce bee survival, food consumption, flight success, and
615 thermoregulation. *Chemosphere*. 237, 124408.
- 616 Tosi, S., et al., 2022. Lethal, sublethal, and combined effects of pesticides on bees: A meta-
617 analysis and new risk assessment tools. *Science of The Total Environment*. 156857.
- 618 USDA, Cotton World Production.
619 [https://ipad.fas.usda.gov/cropexplorer/cropview/commodityView.aspx?cropid=26310](https://ipad.fas.usda.gov/cropexplorer/cropview/commodityView.aspx?cropid=2631000)
620 [00](https://ipad.fas.usda.gov/cropexplorer/cropview/commodityView.aspx?cropid=2631000) accessed 19 January 2023. Vol. 19/01/2023. U.S. Department of Agriculture, 2022a.
- 621 USDA, Rapeseed World Production.
622 [https://ipad.fas.usda.gov/cropexplorer/cropview/commodityView.aspx?cropid=22260](https://ipad.fas.usda.gov/cropexplorer/cropview/commodityView.aspx?cropid=2226000)
623 [00](https://ipad.fas.usda.gov/cropexplorer/cropview/commodityView.aspx?cropid=2226000) accessed: 19 January 2023. U.S. Department of Agriculture, 2022b.
- 624 USDA, Sunflower World Production.
625 [https://ipad.fas.usda.gov/cropexplorer/cropview/commodityView.aspx?cropid=22240](https://ipad.fas.usda.gov/cropexplorer/cropview/commodityView.aspx?cropid=2224000)
626 [00](https://ipad.fas.usda.gov/cropexplorer/cropview/commodityView.aspx?cropid=2224000) accessed: 19 January 2023. U.S. Department of Agriculture, 2022c.
- 627 Wickham, H., Data analysis. *ggplot2*. Springer, 2016, pp. 189-201.
- 628 Williamson, S. M., Wright, G. A., 2013. Exposure to multiple cholinergic pesticides impairs
629 olfactory learning and memory in honeybees. *J Exp Biol*. 216, 1799-807.
- 630 Willis Chan, D. S., et al., 2019. Assessment of risk to hoary squash bees (*Peponapis pruinosa*)
631 and other ground-nesting bees from systemic insecticides in agricultural soil. *Scientific*
632 *Reports*. 9, 11870.
- 633 Winfree, R., et al., 2007. Native bees provide insurance against ongoing honey bee losses.
634 *Ecology letters*. 10, 1105-1113.
- 635 Wintermantel, D., et al., 2022. Flowering resources modulate the sensitivity of bumblebees to
636 a common fungicide. *Science of The Total Environment*. 829, 154450.
- 637 Woodcock, B. A., et al., 2017. Country-specific effects of neonicotinoid pesticides on honey
638 bees and wild bees. *Science*. 356, 1393-1395.
- 639 Woodcock, B. A., et al., 2016. Impacts of neonicotinoid use on long-term population changes
640 in wild bees in England. *Nat Commun*. 7, 12459.
- 641 Zeileis, A., et al., 2019. colorspace: A toolbox for manipulating and assessing colors and
642 palettes. arXiv preprint arXiv:1903.06490.