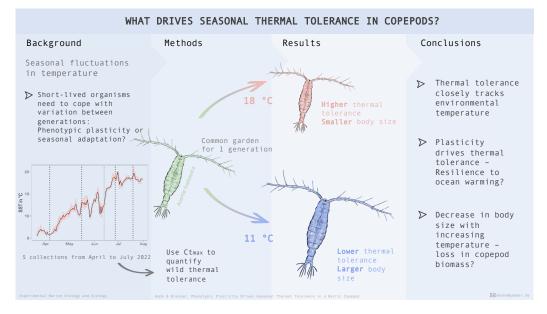
¹ Graphical Abstract

² Phenotypic Plasticity Drives Seasonal Thermal Tolerance in a Baltic

³ Copepod

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₅ Highlights

⁶ Phenotypic Plasticity Drives Seasonal Thermal Tolerance in a Baltic

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- Acartia hudsonica shows strong seasonality in thermal tolerance.
- The observed seasonal differences in CT_{max} are driven by phenotypic plasticity not adaptation.
- Body size in A. hudsonica is negatively correlated to environmental
 and developmental temperature.
- This is the first record of A. hudsonica in the Baltic Sea known to the authors.

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Phenotypic Plasticity Drives Seasonal Thermal Tolerance in a Baltic Copepod

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19 Abstract

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Seasonal changes in environmental conditions require substantial physiologi-20 cal responses for population persistence. Phenotypic plasticity is a common 21 mechanism to tolerate these changes, but for organisms with short gener-22 ation times rapid adaptation may also be a contributing factor. Here, we 23 aimed to disentangle the impacts of adaptation from phenotypic plasticity 24 on thermal tolerance of the calanoid copepod Acartia hudsonica collected 25 throughout spring and summer of a single year. We used a common garden 26 (11 °C and 18 °C) design to determine the relative impacts of plasticity ver-27 sus adaptation. Acartia hudsonica were collected from five time points across 28 the season and thermal tolerance was determined using critical thermal max-29 imum (CT_{max}) followed by additional measurements after one generation of 30 common garden. As sea surface temperature increased through the season, 31 field collected individuals showed corresponding increases in thermal toler-32 ance but decreases in body size. Despite different thermal tolerances of wild 33 collections, common garden animals did not differ in CT_{max} within thermal 34 treatments. Instead, there was evidence of phenotypic plasticity where higher 35 temperatures were tolerated by the 18 °C versus the 11 °C treatment animals 36

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across all collections. Acclimation also had significant effects on body size, 37 with higher temperatures resulting in smaller individuals, consistent with 38 the temperature size rule. Therefore, the differences in thermal tolerance 39 and body size observed in field collected A. hudsonica were likely driven by 40 plasticity rather than adaptation. However, the observed decrease in body 41 size suggests that nutrient availability and ecosystem functioning could be 42 impacted if temperatures consistently increase with no change in copepod 43 abundance. This is the first record of A. hudsonica in the Baltic Sea known 44 to the authors.

⁴⁶ Keywords: CT_{max}, Seasonality, Acartia

47 1. Introduction

Environmental variation is ubiquitous across habitats and organisms are 48 able to respond to and tolerate this variation in multiple ways. When vari-49 ation is both predictable and experienced within the lifespan of an individ-50 ual, it is expected that plasticity will evolve (Pereira et al., 2017; Bitter 51 et al., 2021). Conversely, if environmental variation is unpredictable or at 52 timescales longer than generation time, plasticity is unlikely to evolve but 53 selection should act with each environmental fluctuation. In this latter case, 54 the resulting changes in selective pressure through time can lead to fluctuat-55 ing selection, which can not only drive rapid adaptation but also contribute 56 to the maintenance of genetic diversity in populations (Bergland et al., 2014). 57 One of the major drivers of fluctuating selection in the wild is seasonal 58 change. In this scenario, the environmental changes that occur within a 59

⁶⁰ year result in divergent selective pressures at different temporal periods. For

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example, summer months may favor warm tolerant genotypes while cooler 61 spring or fall temperatures may favor genotypes that have higher perfor-62 mance at low temperature. While these changes are relatively consistent 63 on a vearly basis, the temporal scale is beyond many organisms' lifespan, 64 which can lead to adaptation to different periods within the fluctuating sea-65 sonal environmental change. This phenomenon has been observed in diverse 66 species and ecosystems, including marine copepods (Hairston and Dillon, 67 1990), Lonchopterid flies (Niklasson et al., 2004), dandelion (Vavrek et al., 68 1996), two-spotted ladybugs (Brakefield, 1985), and swallows (Brown et al., 69 2013), among others (Siepielski et al., 2009; Bell, 2010). There is also evi-70 dence for fluctuating selection at the genomic level: in *Drosophila* fruit flies 71 genome-wide allele frequencies consistently and repeatedly shift between sea-72 sons due to selection, helping to maintain genetic variation within popula-73 tions (Johnson et al., 2023). Thus, adaptive responses to seasonal change 74 may be a common phenomenon across taxa with short generation times. 75

In addition to contributing to the maintenance of genetic variation within 76 populations, the mechanisms underlying rapid seasonal adaptation can help 77 shed light on how populations may respond to ongoing anthropogenic driven 78 global change. For example, many species experience yearly temperature 79 changes with an amplitude greater than those predicted under global warm-80 ing (Bujan et al., 2020) and across terrestrial and aquatic ectotherms, popula-81 tions routinely achieve increased thermal tolerance in warmer seasons (Hop-82 kin et al., 2006; Bujan et al., 2020). This suggests that mechanisms enabling 83 seasonal responses, such as plasticity and adaptation, might similarly drive 84 resilience to global warming. 85

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Copepods provide an ideal model to understand population level re-86 sponses to fluctuating selection across seasons. These organisms are short-87 lived with generation times of a few weeks leading to multiple generations 88 per year, each subjected to a unique thermal regime. Previous work on ther-89 mal tolerance of copepods has found contributions of both plasticity and 90 adaptation. For instance, rearing temperature drives plastic responses and 91 significantly influences thermal tolerance (González, 1974), egg production 92 (Holste and Peck, 2006) and growth (Sasaki and Dam, 2020). Temperature 93 also affects adult body size with warmer temperatures leading to smaller 94 individuals (Viitasalo et al., 1995; Sasaki et al., 2019), consistent with the 95 temperature-size rule (Atkinson, 1994). Conversely, there is ample evidence 96 for local adaptation to temperature (Lonsdale and Levinton, 1985; Pereira 97 et al., 2017; Karlsson and Winder, 2020) as well experimental evolution stud-98 ies showing adaptive responses to high temperature after only a few gener-90 ations (Dam et al., 2021; Brennan et al., 2022b). Finally, planktonic cope-100 pods are an integral part of marine food webs and function as an essential 101 link between primary production and higher trophic levels (Turner, 2004; 102 Dzierzbicka-Głowacka et al., 2019). Therefore, our understanding of drivers 103 of copepod responses to temperature change has important implications for 104 the resilience of marine ecosystems as a whole. 105

In this study, we focus on the calanoid copepod Acartia hudsonica (Pinhey, 1926). Acartia copepods are among the most-studied copepod genera, in part due to their world-wide distribution and high abundance, making them foundational to marine and coastal ecosystems (Walter and Boxshall, 2023). In the Baltic Sea, this group is one of the dominant zooplankton and

critical to local ecosystems (Diekmann et al., 2012; Dzierzbicka-Głowacka 111 et al., 2019). Acartia hudsonica, specifically, is a cold adapted species that is 112 generally abundant in winter and spring months out-competing more warm 113 adapted congeners at low temperatures. The species can tolerate a broad 114 temperature range from at least 4 to 18 °C (Sullivan and McManus, 1986). 115 For populations native to the Eastern United States, A. hudsonica produces 116 resting eggs at temperatures > 16 °C and abundances strongly decline in 117 summer (Sullivan and McManus, 1986). 118

Here, we seek to disentangle the impacts of adaptation from plasticity in 119 thermal tolerance of A. hudsonica collected throughout spring and summer of 120 2022. We hypothesized that the thermal tolerance of wild collected individ-121 uals would closely follow the environmental temperature. Further, because 122 developmental temperature strongly impacts copepod body size (Horne et al., 123 2016), we predicted that body size would decrease as temperature increased. 124 We used common garden conditions at two different temperatures to deter-125 mine if the observed thermal tolerance and body size shifts between wild 126 individuals were driven by adaptation or plasticity. Together, these results 127 help to reveal the underlying mechanisms driving seasonal thermal tolerance 128 of A. hudsonica and provide insight into how this species may respond to 129 warming conditions in the future. 130

¹³¹ 2. Material & Methods

132 2.1. Sampling and cultures

All samples were collected by 100 µm WP2 net on board the research vessel *Polarfuchs* in Kiel Fjord (54°19'50"N 10°09'20"E). Live samples were

stored at collection temperature until processing. The experiment included
five sampling dates (referred to as collections) from April 2022 to July 2022
that spanned SST from 5.81 °C to 19.16 °C (Table 1, Hiebenthal et al. unpublished data).

Table 1: Overview of sampling dates and correspondingSST averages.

Collection name	Collection date	Daily mean	2-week mean
Collection 1	06 Apr 2022	$5.81 \pm 0.05 \ ^{\circ}\text{C}$	$6.36 \pm 0.47 \ ^{\circ}\mathrm{C}$
Collection 2	16 May 2022	$12.66 \pm 0.53 \ ^{\circ}\mathrm{C}$	11.44 ± 0.84 °C
Collection 3	13 Jun 2022	$12.37 \pm 0.33 \ ^{\circ}\mathrm{C}$	12.86 ± 1.52 °C
Collection 4	27 Jun 2022	$18.35 \pm 0.13 \ ^{\circ}\text{C}$	$16.55 \pm 1.83 \ ^{\circ}\mathrm{C}$
Collection 5	19 Jul 2022	19.16 ± 0.79 °C	18.11 ± 0.68 °C

For all sampling dates, approximately 440 adult animals were sorted and 139 split into two 6 L culture buckets with air supply and held at their collection 140 temperature. Over the following two days, CT_{max} assays were run on the 141 wild-caught copepods. After the initial assays, the cultures were moved to 142 a cold (11 °C) or warm (18 °C) culture room. No major fluctuations in 143 temperature occurred throughout the experiment. All cultures were kept on 144 a 12:12 light regime at a common salinity of 15 and were allowed to reproduce, 145 with water changes approximately every 7 days. Feeding was ad libitum with 146 Rhodomonas sp. and Isochrysis qalbana (Holste and Peck, 2006; Ismar et al., 147 2008; Mahjoub et al., 2014). 148

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After the initial culture establishment at treatment temperatures (2-4

days), the cultures were filtered through a nested 200 µm and 50 µm mesh 150 sieve. Adults were retained on the 200 µm mesh and kept for further cultur-151 ing. Offspring, eggs and nauplii, of the parental generation were retained on 152 the 50 µm mesh and placed in a new culture bucket to start the F1 gener-153 ation. Development was monitored to catch the onset of maturation. Once 154 the F1 generation reached adulthood, CT_{max} assays were repeated. Collec-155 tion 3 collapsed before reaching F1. Therefore, this collection is excluded 156 from further analysis. 157

158 2.2. Temperature assays

 CT_{max} was used as a proxy for thermal tolerance. CT_{max} is defined as 159 the temperature at which locomotion is affected in a way that prevents the 160 individual to move away from harmful conditions, eventually resulting in 161 death (Cowles and Bogert, 1944). In this study, CT_{max} was the temperature 162 at which the individuals showed no visible response to a stimulus (details 163 below). For each collection and treatment the CT_{max} of twenty males and 164 twenty females was quantified. Individuals were sorted under temperature-165 controlled conditions and placed in 12 ml glass tubes filled with 5 ml of filtered 166 seawater. Ten individuals, five males and five females were simultaneously 167 run per trial. The starting temperature of the experimental tank matched 168 the wild collection or culturing temperature. To minimize bias, an assistant 169 placed the individuals in random order in the experimental tank, leaving the 170 experimenter blinded to the animals' sex until after the experiment. Heating 171 was monitored with a thermometer (PCE-HPT1, PCE instruments, Men-172 schede, Germany) placed into an additional glass tube filled with filtered 173 seawater. The animals were given a 30-minute acclimation period before 174

heating was started. With a 300 W and 500 W heater, heating was con-175 sistent at ~ 0.2 °C/min (Fig. S.1). Once 22.5 °C was reached, the 300 W 176 heater was removed, slowing the ramping temperature to ~ 0.1 °C/min to 177 facilitate monitoring the individuals. Throughout the experiment, the exper-178 imenter was blinded to the exact temperature. Animals were continuously 179 monitored and when movement ceased, gentle pipetting was used to trigger 180 a reaction. If still no movements occurred, the corresponding temperature 181 was considered CT_{max} . 182

After the experiment, the sex for all animals was confirmed and the indi-183 viduals were preserved in 95% ethanol and later photographed using a Nikon 184 imaging microscope and Nikon imaging software (NIS-Elements v. 5.20.00). 185 All images were obtained using the same magnification to ensure consistency. 186 From the pictures, the prosome length was measured for each copepod using 187 ImageJ (Schneider et al., 2012). Three measurements were obtained per in-188 dividual and averaged in the later analysis. While formalin is usually used to 189 preserve plankton samples for length analysis (Connolly et al., 2017; Aguilera 190 et al., 2020), previous work on zooplankton has shown no effect of ethanol 191 preservation on body size (Black and Dodson, 2003) and any size effect would 192 be consistent across the experiment. Further, ethanol preservation allows for 193 downstream genetic analyses, which is essential when dealing with cryptic 194 copepod species. 195

196 2.3. Genotyping

To confirm species identity, 4-15 copepods per collection were genotyped using the mtCOI region (Table S.1). For the DNA extraction, ethanolpreserved copepods were rinsed with ultra-clean water and rehydrated for

one hour. Individual copepods were transferred into 100 µl of 5% Chelex
solution, incubated for 20 minutes at 95 °C in a water bath, then centrifuged
for 5 minutes at 8000 rpm. PCR reactions were conducted in 20 µl volume
with 9.8 µl ultra clean water, 2 µl dNTPs, 2 µl Buffer, 2 µl LCO1490 forward primer or HCO2198 reverse primer (5mM, Folmer et al. (1994)), 0.2 µl
DreamTaq DNA polymerase (Thermo Fisher Scientific Inc., Massachusetts,
US), and 2 µl of DNA.

Amplification conditions were: 3 min of denaturation at 94 °C followed 207 by 33 cycles of 45 s denaturation at 94 °C, 45 s annealing at 48 °C, 60 s 208 extension at 72 °C and a final extension at 72 °C for 7 min. Samples were 209 sequenced on a Sanger sequencing platform by Eurofins Deutschland (Ger-210 many). Sequences were checked and aligned to generate consensus sequences 211 using CodonCodeAligner (CODONCODE, 2010). The Acartia genus consists 212 of multiple cryptic species making species identification difficult. Therefore, 213 we followed Figueroa et al. (2020) and used a bayesian approach to determine 214 species and clade for all samples. Using known samples from Figueroa et al. 215 (2020), we aligned all raw reads with MUSCLE (Edgar, 2004), converted 216 to NEXUS format in R using APE (Paradis and Schliep, 2019), and used 217 MrBayes to build a phylogenetic tree with Acartia dana as the outgroup. 218 Tree plots were made using *qqtree* (Yu et al., 2017). Since thermal tolerance 219 measurements showed outliers in collections 4 and 5, 38 experimental outliers 220 and additional trial animals from those collections were genotyped. 221

222 2.4. Data analysis and statistics

Data manipulation, visualization, and statistics were conducted in R version 4.2.2 (R Core Team, 2022).

To understand the factors influencing CT_{max} and length we used gener-225 alized linear models. Model selection was done by comparing model fit and 226 Akaike's information criterion (AIC) as well as considering biological rele-227 vance. The models included main effects of collection number (1-5), treat-228 ment (wild, warm, cold), sex, length, and the interaction of treatment and 229 collection. Random effects of vial number, tank used for the trial, and time 230 of day did not improve model fit and were therefore excluded from the final 231 models. A similar model was used to determine the factors influencing cope-232 pod length. Here, we included main effects of developmental temperature 233 and sex. Pairwise post-hoc testing was done comparing the model means 234 using the *emmeans* package (Lenth, 2023). 235

236 3. Results

The sampling days spanned SST values from 5.81 °C to 19.16 °C (Fig. 1A; 237 see Table 1 for all values and standard deviations). The thermal tolerance of 238 wild-caught individuals mirrored the environmental temperature (Fig. 1B). 230 CT_{max} values were lowest in Collection 1 (25.9 °C ±1.4 °C), where the tem-240 perature over the two weeks prior to sampling averaged 6.36 °C ± 0.47 °C. 241 CT_{max} then increased throughout the collections (Col-2: 27.9 °C ±0.7 °C; 242 Col-3: 27.9 °C ± 1.0 °C; Col-4: 29.1 °C ± 0.6 °C), reaching its maximum at 243 Collection 5 (29.8 °C ± 2.0 °C), where the two-week SST average was 18.11 244 °C ± 0.68 °C (Spearman's $\rho = 0.758$, p < 0.001). The prosome length of wild 245 individuals showed an opposite trend, with size decreasing as SST increased 246 (Fig 1C, Spearman's $\rho = -0.572$, p < 0.001). Two week averages were chosen 247 to characterize SST as this is the approximate the amount of time it takes 248

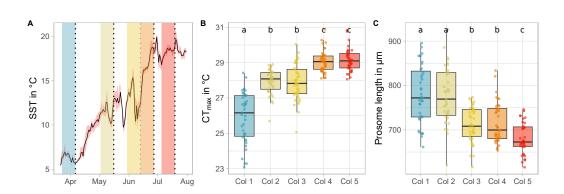


Figure 1: Thermal tolerance follows the seasonal changes in temperature. (A) Sea surface temperature at the collection site where sampling points are indicated by dotted lines and shaded boxes indicate the two-week period prior to sampling. (B) Critical thermal maxima and (C) mean prosome length for wild collected individuals. Boxplot colors correspond to sampling date and compact letters show the results from post-hoc tests where shared letters indicate no significant difference between measurements.

for copepods to mature from egg to adult and can therefore be considered developmental temperature; the two-week average showed a similar trend to daily mean temperatures (Fig. S.2).

Following the common garden, there was a significant effect of treatment, 252 collection and the interaction between treatment and collection on CT_{max} 253 (Fig. 2A, p < 0.001 for all, see Table S.3) for detailed report). While CT_{max} 254 of the wild individuals varied depending on the SST around sampling, the 255 thermal tolerance measurements within the cold and warm treatment were 256 similar across collections. The weighted model means for all warm treatments 257 did not significantly differ from each other (col-1: 28.7 °C ± 0.5 °C, col-2: 28.6 258 °C ±0.5 °C, col-4: 29.0 °C ±0.6 °C, p \geq 0.578) and were comparable to the 259 wild collections 4 and 5 (29.1 °C ± 0.6 °C and 29.2 °C ± 0.7 °C respectively, p 260

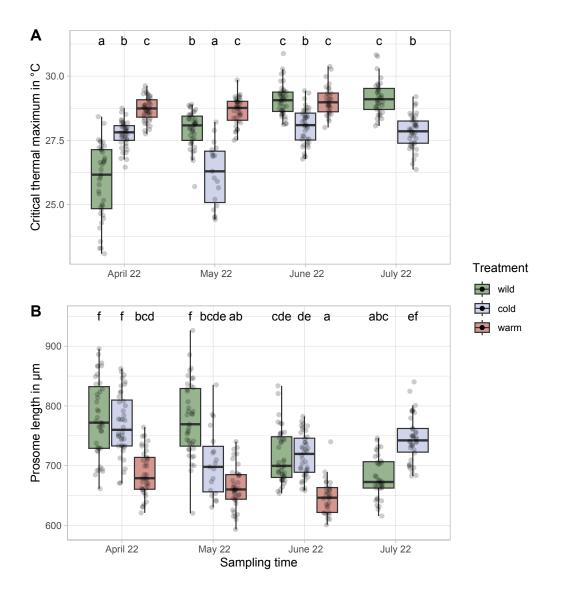


Figure 2: Phenotyping results following common garden for (A) critical thermal maxima and (B) prosome length. Colors of boxes correspond to treatment where "wild" are field collected animals, "cold" are F1 animals at 11°C, "warm" are F1 animals at 18°C. Compact letters show the results from post-hoc tests where shared letters indicate no significant difference between measurements.

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> 0.153) where the mean temperature resembled the warm treatment (warm 261 treatment 18 °C, wild SST 16.55 °C ± 1.83 °C, and 18.11 °C ± 0.68 °C respec-262 tively). For the cold treatment, weighted means did not differ significantly 263 (col-1: 27.7 °C ± 0.5 °C, col-4: 28.0 °C ± 0.7 °C, col-5: 27.8 °C ± 0.6 °C, p > 264 (0.789) and were comparable to the thermal tolerance of the wild collection 265 2 (27.9 °C ± 0.7 °C, p ≥ 0.992) where the developmental temperature was 266 similar (cold treatment 11 °C, wild SST 11.44 °C ± 0.84 °C). The exception 267 was the cold treatment of collection 2, which was significantly lower than 268 the other cold collections (p < 0.001). Here, CT_{max} was comparable to the 269 mean of the wild individuals within collection 1 (col-1: 25.9 °C \pm 1.4 °C, col2: 270 26.2 ± 1.2 °C, p = 0.965) despite the treatment temperatures differing by ~ 271 4 °C. The effect of treatment, collection on prosome length and the inter-272 action of both terms were significant (Fig. 2B, p < 0.001 for all, see Table 273 S.3). However, post-hoc testing revealed that there were no clear similari-274 ties within treatments, suggesting a more complicated connection between 275 prosome length, seasonality and common garden temperature. 276

There was a strong plastic effect of developmental temperature on thermal 277 performance and prosome length (p < 0.001, Fig. S.4). The reaction norms 278 for treatment effect on CT_{max} showed a positive effect of treatment temper-279 ature (Fig. 3A). There was a significant interaction between collections that 280 was driven by the outlier col-2 and when col-2 was removed there was no 281 interaction (with col-2: p , 0.001, without col-2: p = 0.897). Prosome length 282 was negatively affected by treatment temperature (Fig. 3B). Again, a signif-283 icant interaction was driven by col-2, that was not present when removing 284 the outlier collection (with col-2: p = 0.019, without col-2: p = 0.372). The 285

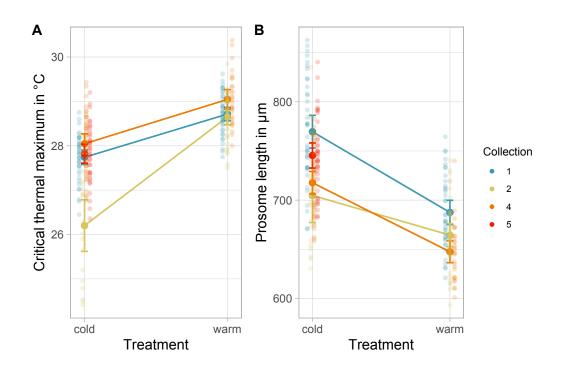


Figure 3: Reaction norms for treatment effect on (A) Critical thermal maximum; (B) Prosome length; bold points indicate mean values, error bars indicate 0.95 % confidence interval.

similar slopes of reaction norms and the absence of an interaction betweentreatment and collection support the presence of plasticity.

Furthermore, there was an increase of CT_{max} with increasing developmen-288 tal temperature (Spearman's $\rho = 0.718$, p < 0.001). Conversely, prosome 289 length was negatively correlated with developmental temperature, where 290 higher temperatures during development led to smaller individuals (Spear-291 man's $\rho = -0.566$, p < 0.001). These effects were present in both male and 292 female animals, however, females had significantly higher thermal tolerance 293 and larger body size compared to males (CT_{max} females: 28.4 °C ±1.21 °C, 294 males: 27.7 °C ± 1.21 °C, length females: 744 μ m $\pm 69 \mu$ m, males: 694 μ m 295

²⁹⁶ $\pm 44 \ \mu m, p < 0.001 \ for \ both).$

In addition to the correlation of developmental temperature, there was 297 a negative correlation between prosome length and CT_{max} in wild animals 298 where smaller individuals had significantly higher thermal tolerance (p <299 0.001, Pearson' r = -0.229, Fig. 4A). While this result was likely driven by 300 the aforementioned relation of developmental temperature and length, the 301 negative trend was also present, though weaker, when looking only at the 302 warm treatment common gardened animals (p = 0.038, Fig. 4B); no effect 303 was observed within cold common gardened treatments (p = 0.108, Fig. 4B). 304

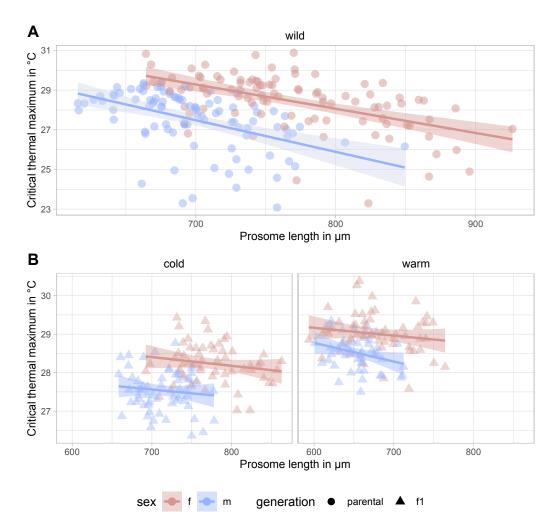


Figure 4: Correlation between prosome length and CT_{max} in A. hudsonica, (A) Parental generation, length has a significant effect on CT_{max} (p < 0.001, R² = 0.35) (B) F1 generation; cold treatment: no significant effect of length on CT_{max} (p = 0.108, R² = 0.323); warm treatment: a significant effect of length on CT_{max} (p = 0.038, R² = 0.242). Linear regression per sex with 95% confidence interval.

305 4. Discussion

Seasonal fluctuations have strong physiological effects on organisms occu-306 pying these variable conditions. We hypothesized that the thermal tolerance 307 of copepods would closely mirror their developmental temperature and that 308 both plastic and genetic mechanisms would contribute to the physiological 300 change observed. As predicted, the CT_{max} of A. hudsonica increased in 310 parallel with environmental temperature. However, under common garden 311 conditions collections showed similar levels of plasticity and converged on 312 common thermal tolerances and body sizes, indicating that phenotypic dif-313 ferences between collection times were driven by plasticity with no evidence 314 for rapid adaptation. Together, these results indicate that A. hudsonica has 315 substantial phenotypic plasticity to rapidly acclimate to large changes in 316 external temperature. 317

318 4.1. Seasonal variation in thermal tolerance

For both marine and terrestrial ectotherms, the ability to rapidly ac-319 climate to changes in environmental temperature is common and adaptive 320 (Gunderson and Stillman, 2015). In coastal marine organisms, particularly 321 those from temperate environments, the presence of thermal plasticity is 322 essential as shallow waters tend have high variance in their temperatures, re-323 quiring rapid phenotypic responses (Reusch, 2014). This is particularly true 324 for intertidal copepods which can experience daily temperature changes of 325 nearly 10°C (Leong et al., 2017) and therefore have high thermal tolerance 326 and plasticity (Healy et al., 2019). Further, across copepods there is near 327 universal presence of thermal plasticity that is dependent on environmental 328

temperature (Sasaki and Dam, 2021). For Acartia copepods specifically, A.
tonsa and A. hudsonica from the east coast of the United States are typically
plastic in their thermal tolerance (Sasaki and Dam, 2020) and laboratory
studies on A. tonsa show large effects of acclimation on thermal tolerance
(Sunar and Kir, 2021). Thus, plasticity plays an important role in enabling
most marine copepods to tolerate environmental temperature fluctuations.

While the Baltic Sea experiences only small wind-driven and irregular 335 tides, the variation in water temperature at our collection site is neverthe-336 less high, ranging from 5.65 °C to 20.70 °C during the study period. This 337 high and relatively predictable variation likely favors the evolution of plas-338 ticity observed in the population (Bitter et al., 2021). Indeed, previous work 339 has shown that Acartia tonsa from less variable low latitude thermal envi-340 ronments harbor lower levels of phenotypic plasticity than those from more 341 variable high latitude sites (Sasaki and Dam, 2020), consistent with the lati-342 tudinal hypothesis of plasticity (Janzen, 1967; Ghalambor et al., 2006). Sim-343 ilarly, Sasaki and Dam (2020) found that seasonal variation in thermal LD50 344 of A. hudsonica from the east coast of North America was driven by plasticity 345 rather than adaptation. Therefore, the plasticity that A. hudsonica harbors 346 to respond to changing temperature across the season is likely adaptive in 347 this environment and is present across multiple populations. 348

The lack of seasonal adaptation is in contrast to the evidence of this phenomenon in other systems including both terrestrial *D. melanogaster* (Johnson et al., 2023) and the sister species to *A. hudsonica*, *A. tonsa* (Sasaki and Dam, 2020). For *A. tonsa*, Sasaki and Dam (2020) found that collections from different time points differed in their plasticity. It is unclear exactly

what drove the seasonal patterns in A. tonsa, but it may be due to the pres-354 ence of cryptic lineages emerging or surviving at different temperatures. The 355 comparison to *D. melanogaster* is also interesting as terrestrial organisms 356 typically have a better ability than marine organisms to buffer their thermal 357 environment via behavioral mechanisms, known as the Bogert effect (Bogert, 358 1949). Acartia copepods have enormous populations sizes that are regularly 359 in the hundreds of individuals per cubic meter of water (Möllmann, 2002). 360 Therefore, the efficiency of selection may be similar between Acartia and 361 Drosophila. Given this, one might expect similar signals of seasonal adapta-362 tion in our system relative to *Drosophila*; we observed no evidence to support 363 this expectation. 364

There are a number of possible explanations for the lack of seasonal adap-365 tation in our study. First, A. hudsonica may have sufficient plasticity to 366 respond to the seasonal thermal environment. Given the Bogert effect, envi-367 ronmental temperature is directly experienced to a greater degree in marine 368 systems and selection for plasticity may be strong and result in a highly 360 flexible phenotype. Alternatively, there may be no heritable genetic varia-370 tion for thermal tolerance in this population. However, populations of A. 371 hudsonica from North America can rapidly evolve to elevated temperatures 372 under laboratory conditions (deMayo et al., in press) and there is evidence 373 for local adaptation to temperature along latitudinal gradients in the sister 374 species, Acartia tonsa (Sasaki and Dam, 2019). Therefore, it is likely that 375 heritable variation in thermal tolerance is also present in this species and 376 population. An alternative explanation is that the phenotypes of focus were 377 not under selection or sensitive enough to capture any adaptive responses. 378

The temperatures at the collection site did not approach the critical ther-379 mal limit measured in the lab (maximum SST during sampling: 20.70 °C). 380 Therefore, it is unlikely that CT_{max} was directly under selection. It is possi-381 ble that an alternative phenotype would show a seasonal adaptive response 382 as has been observed in other systems (Hairston and Dillon, 1990; de Ville-383 mereuil et al., 2020). Finally, much of the evidence for seasonal adaptation 384 in D. melanogaster has been found at the genomic level (Johnson et al., 385 2023). Given this, our populations may similarly be experiencing fluctuating 386 selection that would be detectable using genomic approaches. 387

While we observed high levels of plasticity for CT_{max} , under future tem-388 perature conditions it is unlikely that most ectotherms have sufficient plastic-389 ity to respond to temperature changes without adaptive responses (Gunder-390 son and Stillman, 2015). Indeed, DeMayo et al showed that under warming 391 conditions, plasticity alone is insufficient to maintain high population fitness 392 in A. hudsonica (deMayo et al., in press). However, the species can rapidly 393 adapt after just four generations to recover fitness levels. Therefore, it is 394 likely that both plasticity and adaptation will be required to tolerate future 395 environmental conditions and more work is needed to understand the relative 396 contribution of each to overall resilience. 397

398 4.2. Plasticity in body size and potential potential impacts of warming

As temperatures increased, body size decreased in *A. hudsonica* (Fig. 1), a common phenomenon across ectotherms known as the temperature-size rule (Angilletta and Dunham, 2003; Rubalcaba and Olalla-Tárraga, 2020). This concept applies to copepods (Escribano and McLaren, 1992; Viitasalo et al., 1995), including those from tropical (Ortega-Mayagoitia et al., 2018)

and temperate environments (Riccardi and Mariotto, 2000). However, other 404 factors such as phytoplankton density may affect the body size of individ-405 uals in the wild (Deevey, 1964, 1966), though this would not have affected 406 the common garden animals in our study. The reduction in body size in 407 response to increasing temperature may be driven by the disproportionate 408 increase in respiration and metabolism relative to ingestion and assimilation 409 of nutrients (Lehman, 1988), leading to lower overall energy available for 410 growth and therefore a smaller body size. Further, there may be a trade-off 411 that favors smaller individuals at high temperatures; reproductive efficiency, 412 the ratio of egg production and respiration, is maximized at smaller body 413 sizes and therefore may be adaptive in warmer temperatures. Similarly, at 414 higher temperatures oxygen availability (aerobic scope) may be decreased in 415 larger individuals relative to smaller individuals, favoring smaller body sizes 416 (Rubalcaba et al., 2020). Alternatively, in cold temperatures growth periods 417 may be prolonged while the growth rate remains relatively stable, leading to 418 larger individuals under cold conditions (Vidal, 1980). 419

Regardless of the mechanism, as temperatures warm due to anthropogenic 420 causes, decreases in body size may affect ecosystem interactions. This is par-421 ticularly true in the Baltic Sea where the heating rate is around three times 422 higher than the ocean average due to its unique topography (Reusch et al., 423 2018; Szymczycha et al., 2019; Dutheil et al., 2022). A size reduction in 424 A. hudsonica, or other prey organisms, might impact higher trophic level 425 predators, for example by requiring the consumption of more individuals to 426 maintain the same amount of nutrient intake (Garzke et al., 2015). If abun-427 dance does not increase with decreasing body size, nutrient availability may 428

be reduced for consumers who will also require increased energy needs un-429 der higher temperature (Brown et al., 2004). Further, Garzke et al. (2015) 430 observed that large copepods from colder temperatures clear algae biomass 431 more efficiently than smaller individuals, exerting top-down control on phy-432 toplankton. This in turn suggests that smaller copepods are less efficient 433 grazers, with less control over the planktonic community. As copepods are 434 important grazers on large phytoplankton and microzooplankton (Sommer 435 et al., 2003; Armengol et al., 2017), less efficient grazing or a shift to different 436 prey size classes might have unforeseen cascading effects across ecosystems. 437

Finally, the correlation between body size and CT_{max} (Fig. 4) suggests 438 that body size itself may influence the thermal tolerance of an individual 439 with small body size being of advantage in warm environments. While the 440 effect was weak in comparison with the effects of developmental temperature 441 and the mechanistic link between body size and thermal tolerance remains 442 unknown, this relationship could potentially be used to predict an individ-443 ual's thermal tolerance. This may be of interest when analyzing historical 444 samples where body size measurements from the same location across years 445 could enable predictions about past thermal tolerance and environmental 446 temperatures. More work would be needed to develop these predictions. 447

448 4.3. Outliers and mixed species

The outliers in collection 2 under cold conditions, characterized by unexpectedly low thermal tolerance and small body sizes, were most likely the result of an impending culture collapse. This was potentially caused by poor food quality during that period due to ciliate and bacteria growing in the algae cultures used for feeding. The negative effects of ciliates on copepod

fitness are well described with effects ranging from decreased egg produc-454 tion (Burris and Dam, 2014) to increased adult mortality (Visse, 2007). The 455 species identity of the ciliates in this study could not be determined. How-456 ever, ciliate peak abundances correlated with culture collapse, and after the 457 establishment of more frequent water changes, the cultures improved. Since 458 the survival and fitness of the animals was clearly affected by external factors 459 unrelated to the experiment, length and thermal tolerance measurements for 460 the cold treatment in collection 2 should be treated with caution and were 461 therefore excluded from parts of the analysis. 462

In the wild collection 5, five individuals showed a CT_{max} above 33 °C, 463 which was well beyond the distribution of values for any other collection 464 (Fig. S.5). In the F1 generation, the warm treatment of collection 4 and 465 5 and the cold treatment of collection 5 also had similar high performing 466 individuals (11, 40, and 2 individuals, respectively) (Fig. S.5). These ex-467 treme outliers suggested a mixed species composition, which was confirmed 468 by genotyping individuals of each collection as well as high thermal outliers 460 (see Supplement *Mixed species*). High thermal outlier individuals were A. 470 tonsa without exception (Table S.1, Fig. S.6). The increased thermal toler-471 ance for A. tonsa relative to A. hudsonica is consistent with other studies. 472 On the east coast of North America, A. hudsonica dominates plankton com-473 munities early in the year when water temperatures are low and is replaced 474 by A. tonsa as temperatures increase (Borkman et al., 2018; Sullivan and Mc-475 Manus, 1986). This pattern appears to be similar for copepod communities 476 in Kiel Bight as well. Since A. tonsa was present only in later collections and 477 just as individual outliers, this paper could not determine how thermal tol-478

erance of A. tonsa changes within a season. However, we would hypothesize
that A. tonsa follows an overall similar trend than shown for A. hudsonica
with thermal tolerances shifted towards warmer temperatures. Additional
experiments would be required to test this hypothesis.

483 4.4. First record of A. hudsonica in the Baltic Sea

From our literature and database review, the molecular barcoding in this 484 study is the first record of A. hudsonica in the Baltic Sea. There are two 485 possible explanations for this novel species presence. First, A. hudsonica 486 may have recently invaded the Baltic Sea. In the North Sea, A. omori - a487 species that co-occurs with A. hudsonica off the Japanese coast (Ueda, 1987) 488 - was first described by Seuront (2005) in the early 2000. The successful 489 invasion of A. omori suggests that A. hudsonica could similarly have been 490 introduced to the North Sea. As the North Sea and Baltic Sea are connected, 491 it is possible that A. hudsonica then moved to the Baltic Sea. Alternatively, 492 there is ample shipping in the region and an independent local introduction 493 could have occurred. 494

Secondly, it cannot be ruled out that A. hudsonica has historically been 495 misidentified in the Baltic Sea. Copepods from the genus Acartia have a 496 record of mis-identifications in databases and are difficult to distinguish mor-497 phologically (Figueroa et al., 2020). Instead, phylogenetic methods are re-498 quired for accurate species identification. Before the 1970's, A. hudsonica 499 was a subspecies of A. clausi, but is now considered its own species (Brad-500 ford, 1976; Ueda, 1986). Given this, it is possible that A. hudsonica is native 501 to the Baltic Sea but has been, and still is, commonly identified as A. clausi. 502

503 5. Conclusions

Here we showed that Acartia hudsonica has high phenotypic plasticity 504 in response to changing temperature within a single year. We found that 505 thermal tolerance closely tracks environmental temperature, indicating that 506 A. hudsonica has capacity to tolerate increasing temperatures that fall within 507 the current range experienced in nature. However, the observed decrease in 508 body size suggests that nutrient availability and ecosystem functioning could 509 be impacted if temperatures consistently increase with no change in copepod 510 abundance. By focusing on the relative impacts of plasticity and adaptation 511 to population responses to temperature change we can begin to understand 512 the resilience populations and ecosystems to ongoing global change. 513

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522 Conflicts of interest

⁵²³ The authors declare that there is no conflict of interest to disclose.

524 Data accessibility

The data set and scripts can be found on https://github.com/HahnAlexandra/ Plasticity_Acartia_hudsonica. The individual sequences are uploaded on NCBI GenBank https://www.ncbi.nlm.nih.gov/genbank/, find individual accession numbers in Table S.1. SST data from Kiel Fjord will be made available on https://www.pangaea.de by our collaborator Claas Hiebenthal.

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Table 1: Overview over sampling dates and correspond-ing SST average.

Collection name	Collection date	Daily mean	2-week mean
Collection 1	06 Apr 2022	$5.81 \pm 0.05 \ ^{\circ}\text{C}$	$6.36 \pm 0.47 \ ^{\circ}\text{C}$
Collection 2	16 May 2022	$12.66 \pm 0.53 \ ^{\circ}\mathrm{C}$	11.44 ± 0.84 °C
Collection 3	13 Jun 2022	$12.37 \pm 0.33 \ ^{\circ}\mathrm{C}$	12.86 ± 1.52 °C
Collection 4	27 Jun 2022	$18.35 \pm 0.13 \ ^{\circ}\text{C}$	$16.55 \pm 1.83 \ ^{\circ}\mathrm{C}$
Collection 5	19 Jul 2022	19.16 ± 0.79 °C	18.11 ± 0.68 °C

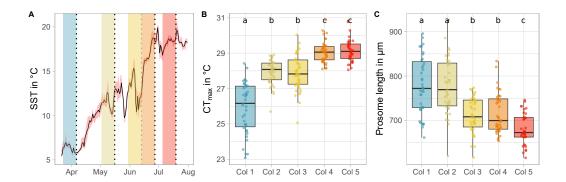


Figure 1: Thermal tolerance follows the seasonal changes in temperature. (A) Sea surface temperature at the collection site where sampling points are indicated by dotted lines and shaded boxes indicate the two-week period prior to sampling. (B) Critical thermal maxima and (C) mean prosome length for wild collected individuals. Boxplot colors correspond to sampling date and compact letters show the results from post-hoc tests where shared letters indicate no significant difference between measurements.

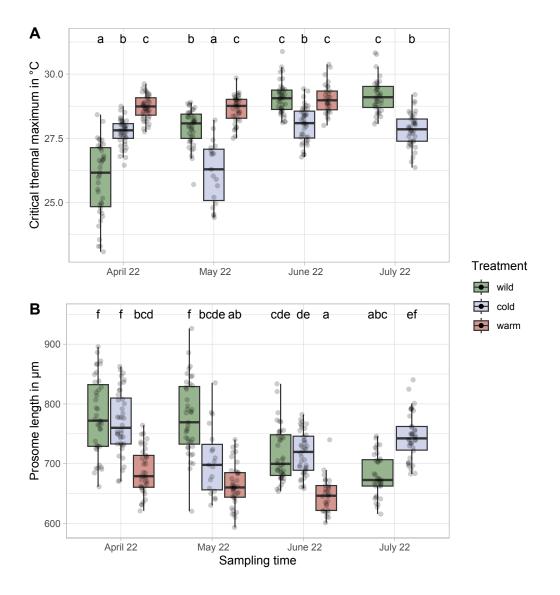


Figure 2: Phenotyping results following common garden for (A) critical thermal maxima and (B) prosome length. Colors of boxes correspond to treatment where "wild" are field collected animals, "cold" are F1 animals at 11°C, "warm" are F1 animals at 18°C. Compact letters show the results from post-hoc tests where shared letters indicate no significant difference between measurements.

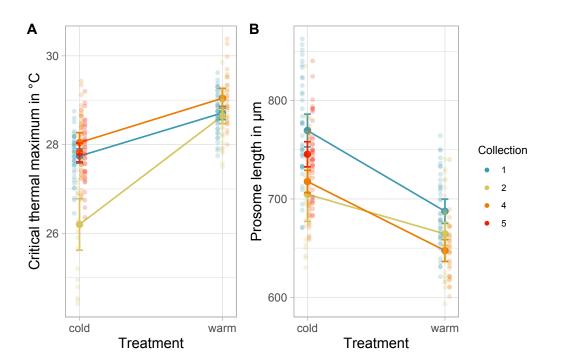


Figure 3: Reaction norms for treatment effect on (A) Critical thermal maximum; (B) Prosome length; bold points indicate mean values, error bars indicate 0.95 % confidence interval.

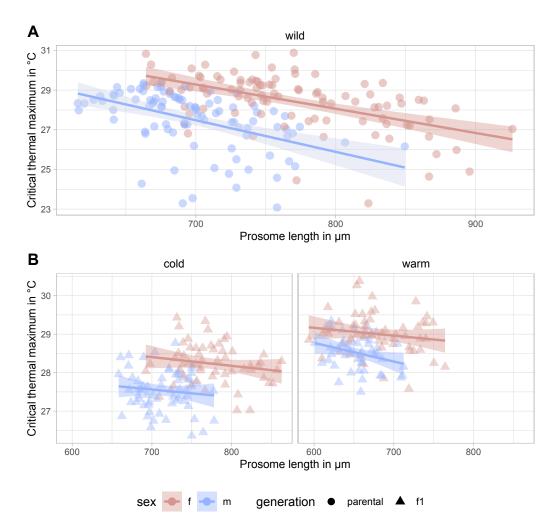


Figure 4: Correlation between prosome length and CT_{max} in A. hudsonica, (A) Parental generation, length has a significant effect on CT_{max} (p < 0.001, R² = 0.35) (B) F1 generation; cold treatment: no significant effect of length on CT_{max} (p = 0.108, R² = 0.323); warm treatment: a significant effect of length on CT_{max} (p = 0.038, R² = 0.242). Linear regression per sex with 95% confidence interval.