

1 **Diving into spring biodiversity: A natural heritage assessment of Australia's**  
2 **Great Artesian Basin discharge wetlands**

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11 **Abstract**

12 The Great Artesian Basin (GAB) is Australia's largest groundwater resource and feeds thousands of springs in the  
13 country's arid centre. GAB springs are globally important "time capsules" that support relictual species not found  
14 elsewhere on Earth, yet they are considerably threatened by ongoing water abstraction for pastoral, agricultural, and  
15 industrial practices. Biomonitoring and robust management plans are needed to prevent further extirpations of GAB-  
16 dependent taxa, but there remains a paucity of biodiversity and ecological knowledge in the literature. We conducted a  
17 comprehensive review of GAB spring biota for South Australia, the state containing the majority of springs. Almost 500  
18 taxa are recorded from GAB springs and surrounding wetlands. Invertebrates represent the largest proportion of GAB  
19 endemics and biota overall, yet they are highly threatened and have not received adequate conservation attention.  
20 Community composition differs considerably among spring clusters, highlighting their intrinsic value as refugia for both  
21 endemics and cosmopolitan taxa. Further, several potential biodiversity hotspots have been overlooked in the literature  
22 and the true extent of GAB spring biodiversity is likely far greater than currently known. Our findings underscore the  
23 importance of freshwater sources in arid regions and the need for fundamental research in the face of ongoing human  
24 impacts.

25 **Key words:** Great Artesian Basin, groundwater-dependent ecosystems, aquifers, springs, biodiversity, biogeography

## 26 1 Introduction

27 The Great Artesian Basin (GAB) is Australia's largest groundwater resource, spanning over one fifth of the  
28 continent's area or almost 2 million km<sup>2</sup> (Habermehl, 2020). As the Basin is a confined aquifer mostly sealed off from  
29 the Earth's surface by impermeable rock, water is only accessible via fissures or artificially drilled boreholes. At the  
30 outer margins of the GAB where the confining layer is thinner, water is often forced to the surface to form springs and  
31 associated wetlands. Thousands of such springs are found throughout Central Australia, with >5,000 individual spring  
32 vents (discrete discharge points of water) in South Australia (hereafter SA) (Arabana Aboriginal Corporation, 2021;  
33 Government of South Australia, 2023), >2,000 in Queensland (Queensland Department of Regional Development,  
34 Manufacturing and Water, 2023), and >400 in New South Wales (New South Wales Department of Planning, Industry  
35 and Environment, 2021). Springs fed by the GAB therefore represent a widespread and permanent potable water source  
36 that has supported humans, flora, and fauna in Australia's arid and semi-arid zones for thousands of years (Priestley et  
37 al., 2018). From an ecological and evolutionary standpoint, GAB-fed springs also support wetlands that represent  
38 "museums of biodiversity", supporting plant and animal species not found elsewhere on Earth (Murphy et al., 2015). As  
39 relicts of the continent's mesic past, species endemic to GAB springs can have exceptionally small distributions and it is  
40 not uncommon for taxa to be restricted to a single cluster of springs, also termed *ultra-short range endemics* (Harvey,  
41 1990; Gotch et al., 2008; King, 2009; Murphy et al., 2009, 2012; Guzik et al., 2012, 2019; King et al., 2014).

42  
43 The conservation and management of groundwater resources are generally inadequate worldwide (Famiglietti,  
44 2014). Springs fed by the Great Artesian Basin are no exception and are, generally speaking, considerably threatened by  
45 a range of industrial and pastoral practices (Mudd, 2000; Fairfax & Fensham, 2002; Lewis & Harris, 2020). Substantial  
46 reductions in artesian pressure (hereafter drawdown) and spring flow have occurred as a result of the sinking of over  
47 50,000 artificial boreholes and the direct abstraction of Basin water, both historically and in the present day (Mudd, 2000;  
48 Gotch et al., 2016; Great Artesian Basin Coordinating Committee, 2019; Beasley-Hall et al., 2023). Indeed, a complete  
49 cessation of flow has occurred for an estimated 800 springs across Australia (Andersen et al., 2016; Fensham et al.,  
50 2016). Concomitant declines in observations of endemic fauna have also been documented, many of which have occurred  
51 as a result of reduction in size and/or complete extinction of their corresponding spring habitat (Kinhill-Stearns, 1984;  
52 McLaren et al., 1986; Zeidler & Ponder, 1989; Kinhill, 1997; Fatchen, 2000; NSW National Parks and Wildlife Service,  
53 2002; Fensham & Fairfax, 2003; Kodric-Brown et al., 2007; Rossini et al., 2018). Initiatives such as the Great Artesian  
54 Basin Sustainability Initiative (hereafter GABSI), a basin-wide bore capping program to curb the uncontrollable flow of  
55 drilled bores, has "saved" over 250 gigalitres of water (Great Artesian Basin Coordinating Committee, 2019). The work  
56 of GABSI combined with the implementation of the Great Artesian Basin Strategic Management Plan, a longer-term  
57 framework related to GAB natural resource management, appears to have improved artesian pressure (Habermehl, 2020).  
58 However, net water flows in GAB springs nonetheless appear to be reducing across certain regions of the Basin, and  
59 understandings of the relationship between artesian pressure and spring flow remain poorly understood (Green & Berens,  
60 2013). While drawdown represents the major cause of spring extinction, grazing, trampling, and soiling of wetlands by  
61 livestock (Fatchen, 2000), overabundant invasive (Noack, 1994; Kerezy, 2015) and native (Davies et al., 2010; Lewis &  
62 Packer, 2020) species, climate change (Ordens et al., 2020), and tourist activity (Witjira National Park Co-management  
63 Board, 2022) pose threats to both spring health and the ecological communities dependent on springs themselves.

64

65 Wetlands supported by GAB springs have high biodiversity value and, correspondingly, their ecological  
66 communities are protected under the *Environment Protection and Biodiversity Conservation Act 1999* (Cth) (hereafter  
67 EPBC Act), a key piece of Australian federal environmental legislation. However, it is recognised that in many cases, the  
68 extinction of GAB spring endemic species represents a permanent loss of Australia's genetic diversity and unique  
69 evolutionary lineages that have persisted over millions of years in isolation (Murphy et al., 2015). Basic taxonomic and  
70 biological information is still largely unknown, inaccessible, or disparate, making it difficult to accurately assess and  
71 monitor all species present at springs in the face of rapid, human-driven change. This lack of taxonomic information is  
72 particularly the case for endemic spring invertebrates (Rossini, 2020). Similarly, the relationship between environmental  
73 characteristics of springs and biodiversity is also not well understood. Metrics commonly used by industry to measure the  
74 robustness of springs, such as increased spring flow rate and aquifer pressure, are associated with formation of wetland  
75 areas (White & Lewis, 2011; BHP, 2018, 2021; Fensham & Laffineur, 2022) and biodiversity is thought to increase with  
76 the number of active springs in a region (Harris, 1992; Rossini et al., 2018). Conversely, drawdown can lead to the  
77 extinction of springs and their corresponding biota (Kinhill, 1997: 199; Fensham et al., 2010). However, more precise  
78 interactions among spring characteristics are not known and these trends have not been observed for all springs (Green &  
79 Berens, 2013). As such, the vulnerability of GAB spring taxa therein is almost impossible to gauge based on spring traits  
80 alone, complicating conservation management and planning. Thus, there is a clear need to compile baseline information  
81 on these taxa in a centralised, accessible resource to facilitate future conservation work on this community.

82  
83 Data are crucial to decision-making capabilities in the context of biodiversity assessment, monitoring, and  
84 developing an understanding of ecosystem health (Vaughan et al., 2001; Stall et al., 2019). For industry partners,  
85 policymakers, and not-for-profit stakeholders, effective data sharing is essential for future progress and predictions in  
86 operations, products, and services (e.g., in the geosciences, (Hanson et al., 2017)). Guiding principles for scientific data  
87 management and stewardship have been developed by a global consortium of stakeholders to ensure digital assets are  
88 *findable, accessible, interoperable, and reusable* (FAIR) (Wilkinson et al., 2016) to promote the open sharing and reuse  
89 of research data, promote reproducibility, increase transparency, and ultimately accelerate scientific discovery. FAIR  
90 practices recommend data storage on open-access platforms, using persistent identifiers and built-in metadata in standard  
91 formats and vocabularies widely adopted by the research community. Currently, data on GAB springs, particularly  
92 biodiversity and hydrogeological information, has been collected and maintained in a sequestered and diffuse manner  
93 (38,39,44). No single centralised and public database or inventory that collates data for all fauna, flora, and funga—a  
94 term seeking to more accurately encapsulate biodiversity in conservation policy frameworks (Kuhar et al., 2018)—  
95 associated with the GAB springs exists. The closest resource is a publicly accessible dataset established by the  
96 Queensland Government which combines information on metrics such as water quality, chemistry, flow rate,  
97 biodiversity, and spring condition (Queensland Government, 2018) (absent for SA and New South Wales). The  
98 development of a biotic inventory has been identified as an urgent action to recover the community of native species  
99 dependent on GAB springs (Fensham et al., 2010). Such data can help to inform future management plans and early  
100 warning systems that detect changes in these ecosystems of high biodiversity value (Vaughan et al., 2001; Brack et al.,  
101 2015; Obura et al., 2019) and so improving their accessibility is a clear priority.

102  
103 The first published sampling efforts began in the 1980s, which largely concerned small areas of the GAB and/or  
104 specific taxa, such as filamentous algae and the crustaceans (Kinhill-Stearns, 1984; Symon, 1984; Badman, 1985;

105 Greenslade, 1985; McLaren et al., 1985; Mitchell, 1985; Thompson, 1985; Mollemans, 1989; Skinner, 1989; Zeidler,  
106 1989). The majority of these checklists concern SA. In 2010, Fensham (Fensham et al., 2010) undertook one of the first  
107 robust scoping exercises of spring endemics and documented 30 taxa associated with SA GAB springs. Rossini *et al.*  
108 (Rossini et al., 2018) subsequently increased this number to 52 across 154 SA spring groups in a synthesis on GAB  
109 spring endemics. These checklists have been supplemented by genetic studies of the taxa found in SA GAB springs such  
110 as crustaceans (Murphy et al., 2009, 2013, 2015; Guzik et al., 2012, 2019; Guzik & Murphy, 2013; Stringer et al., 2019),  
111 snails (Ponder et al., 1995; Murphy et al., 2012), beetles (DeBoo et al., 2019), spiders (Gotch et al., 2008), and microbes  
112 (Byers et al., 1998). Collating fundamental biodiversity information such as this and ensuring the data are accessible and  
113 usable is a first step towards future science-based decision making (i.e., FAIR). Here, we collated all published  
114 information on the occurrence extents of fauna, flora, and fungi associated with GAB springs in SA. This state was  
115 selected for its high number of GAB springs (>60% of Australia's springs) and the comparatively well-studied nature of  
116 their biodiversity from taxonomic and phylogenetic standpoints (Ponder et al., 1989; King, 2009; Guzik et al., 2012,  
117 2019; Guzik & Murphy, 2013; Murphy et al., 2013, 2015; King et al., 2014), allowing for a more accurate reflection of  
118 the true number of taxa in these ecosystems. We also collated metadata associated with the conservation status, extent of  
119 occurrence, and taxonomic status of spring biota, and used these to calculate biodiversity metrics for springs, in keeping  
120 with existing checklists (Fensham & Price, 2004; Rossini et al., 2018). Further, we explored connectivity of these  
121 communities, allowing us to put forward preliminary conservation priorities for GAB-fed wetlands.

## 122 2 Materials and Methods

### 123 2.1 Literature review and database construction

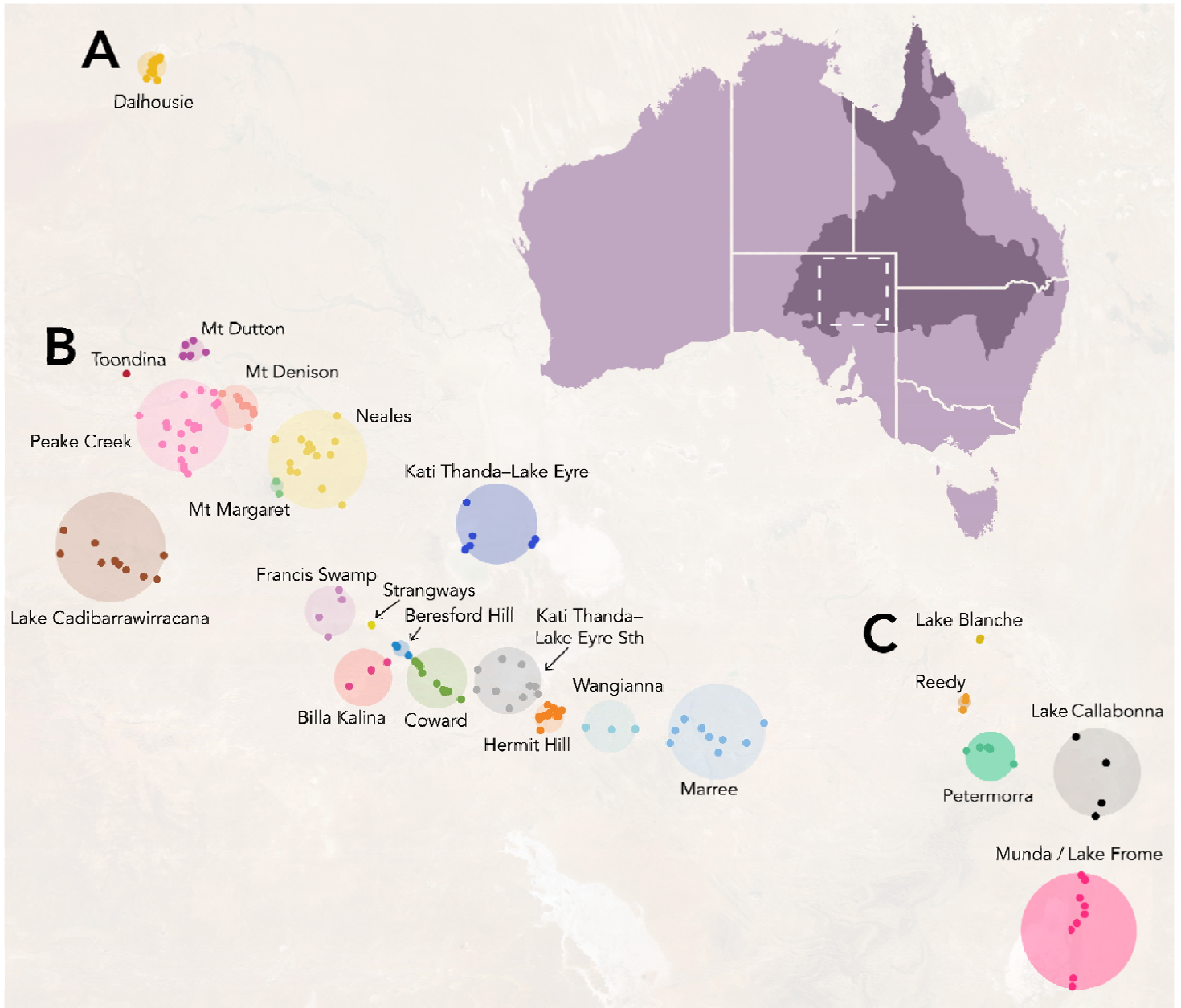
124 A non-systematic review of all available information relevant to the biodiversity of GAB springs in SA was undertaken  
125 in the present study. A formal systematic searching strategy was not possible as a large proportion of information on  
126 springs is present in unpublished or grey literature such as government reports, internal publications from mining  
127 companies, and museum records. We therefore relied heavily on repositories such as the Government of South  
128 Australia's WaterConnect portal (South Australian Department for Environment and Water, 2023) and The Atlas of  
129 Living Australia (Belbin et al., 2021). We were specifically interested in data indicating the presence or absence of  
130 species in or around spring wetlands as well as their occurrence extents, if available, to gauge where taxa occurred and  
131 their degree of endemism. The scope of this review spanned the three spring supergroups in SA: Kati Thanda–Lake  
132 Eyre, Munda / Lake Frome, and Dalhousie (Figure 1). **Supergroups** are clusters of **spring complexes**, which themselves  
133 contain **spring groups** composed of individual **springs**. Spring complexes share similar geomorphological characteristics  
134 and water chemistry, whereas groups are clusters of springs sourcing from the same fault of structure (Lewis et al., 2013).  
135 Individual springs are composed of permanent wetland vegetation with at least one **vent**, a discrete discharge point of  
136 water (Gotch, 2013). Within these three supergroups, we focused on 23 spring complexes containing 170 spring groups  
137 (Supplementary File S1, available publicly via FigShare: <https://doi.org/10.25909/24457105>). Whilst the number of these  
138 spring clusters, as well as their naming conventions, have previously not been standardised (Gotch et al., 2016), we  
139 selected these locations due to their widespread use in the literature and Australian Government publications (Lewis &  
140 White, 2013; Murphy et al., 2013; Rossini et al., 2018; Department of the Environment, 2022).

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173 **Fig. 1:** Springs fed by the Great Artesian Basin (GAB) in South Australia. Springs are classified hierarchically as  
174 supergroups, the broadest classification (A, Dalhousie; B, Kati Thanda–Lake Eyre; C, Munda / Lake Frome); complexes  
175 (labelled coloured circles); groups (coloured points); springs; and finally vents, discrete water discharge points. The  
176 approximate area of the GAB is shown in dark purple and the location of the SA GAB springs is indicated by the dashed  
177 square.

178

179

180 We retrieved Aboriginal names for spring groups from a key anthropological study (Hercus & Sutton, 1985) to  
181 ensure locations were referred to by a dual naming convention whenever possible. The validity of these names were

182 confirmed by The Arabana Aboriginal Corporation in a personal communication to the authors. In South Australia, the  
183 GAB springs can be found on the country of the Antakarinja, Arabana, Dhirari, Dieri, Karangura, Kokatha, Kuyani,  
184 Ngamini, Pirlatapa, Southern Aranda, Southern Arrente, Thirrari, Wangkangurru, and Yandruwandha peoples (Hercus &  
185 Sutton, 1985). The springs are the focus of cultural activities and oral histories for many of these peoples and almost all  
186 spring groups are documented as having mythological significance, historical significance, or both (Hercus & Sutton,  
187 1985; Nursey-Bray & Arabana Aboriginal Corporation, 2015; Nursey-Bray et al., 2020). For locations with dual names,  
188 but lacking “official” formatting (e.g., not used by federal/state government legislation or publications), we followed the  
189 Australian Government’s style manual (Australian Government, 2023). While Kati Thanda–Lake Eyre has an official  
190 name in government legislation, the formatting of the dual naming of Munda / Lake Frome is not standardised and we  
191 default to the style manual mentioned above in this case. We chose not to standardise names across springs, groups, or  
192 complexes as they refer to specific locations with distinct associated mythologies and histories (Hercus & Sutton, 1985).  
193 As a result, we do not refer to e.g. the Coward complex as Pitha Kalti-kalti / Coward (corresponding to a spring group),  
194 nor have we generalised the names of individual springs to the group level (e.g., Ngalpangkardanha and Pangki  
195 Warrunha refer to different locations within the Strangways group, but not the entire group itself). Alternative names and  
196 identifying codes for all locations were also recorded where present, e.g. for spring groups with several English or  
197 Aboriginal names or spelling variants. Records below the spring group level, where present, were standardised to the  
198 group level due to the rarity of such records and the inconsistency of spring vent naming conventions. As GAB spring  
199 fauna are often morphologically cryptic yet genetically distinct, we recorded species as separate taxa if explicitly  
200 indicated by their species authorities based on genetic data following Rossini *et al.* (Rossini et al., 2018).

201

202 To supplement presence/absence records of spring taxa, we also gathered metadata related to taxonomic status,  
203 conservation status, common names, synonyms, and endemism where available. Taxonomic status was coded into three  
204 categories: taxa for which there is a corresponding formal taxonomic description (“described”), taxa awaiting taxonomic  
205 description (“undescribed”), or those for which species-level occurrence records were not available (e.g., a family-only  
206 record; “unidentified”). For taxa with species-level identifications, conservation status information was retrieved from the  
207 Species Profile and Threats Database to capture listings under the IUCN Red List of Threatened Species and Australian  
208 federal and state environmental legislation (Department of Climate Change, Energy, the Environment and Water, 2023).  
209 Finally, for each endemic taxon we noted whether it occupied only one spring group, complex, or supergroup. With the  
210 exception of undescribed species with well-established occurrence records, endemism was not recorded for taxa without  
211 a species-level identification.

## 212 **2.2 Biodiversity metrics**

213 For each spring group, we first transformed presence/absence records to weight them by whether they were at the  
214 resolution of the spring group level (hereafter confident records) or those which referred to a taxon’s presence within a  
215 certain complex, but did not supply spring group information (hereafter coarse records). To assess the inclusion of  
216 uncertain occurrence information, we calculated metrics for our entire dataset and a subset of the data only considering  
217 confident records. As the dataset included only presence/absence records, we were limited in our choice of biodiversity  
218 metrics and focused on species richness and endemism. We calculated species richness values (hereafter taxon richness)  
219 by the number of putative taxa in each spring. Spring groups were ranked by the degree of endemism of their biota using  
220 a modification of the scoring system developed by Fensham and Price (Fensham & Price, 2004). Originally applied to

221 GAB flora in spring complexes in Queensland, the ranking has since been expanded to fauna across Australia (Rossini et  
222 al., 2018) and relies on the number of populations corresponding to the most widespread taxon in a spring dataset. The  
223 desert goby *Chlamydogobius eremius* (Zietz, 1896) is the most widespread SA GAB spring endemic, occurring across 24  
224 known groups (Gotch et al., 2016; Rossini et al., 2018); we used this value as a proxy for the taxon's number of  
225 populations, although we acknowledge this may be an underestimate as spring groups do not necessarily share permanent  
226 wetlands. Rankings for each endemic taxon were first calculated by dividing 24 by the number of groups the taxon  
227 occurred in such that *C. eremius* (24/24) would receive the lowest ranking due to its comparably widespread distribution.  
228 As all endemic taxa assessed here do not occur beyond their respective supergroups, each taxon was then scored by  
229 whether it was further restricted to a single spring complex (+1) or group (+2). These scores, hereafter *endemcity*  
230 *rankings*, were summed for each spring group. We visualised these metrics using QGIS (QGIS Association, 2023).  
231 Spring groups were first mapped using latitude and longitude information corresponding to vents retrieved from the  
232 Government of South Australia's WaterConnect portal (Government of South Australia, 2023); for groups containing  
233 multiple vents, centroids were calculated to approximate their location. Circles corresponding to the above metrics per  
234 group were then scaled using the Flannery method and an exponent method of 0.57. To assess differences in community  
235 composition among spring groups, we calculated pairwise Jaccard distances using a binary matrix of presence/absence  
236 records using the *proxy* package in R v.4.3.0 (Meyer & Buchta, 2022; R Core Team, 2023). Principal components  
237 analysis was performed using the native R *stats* package and visualised using *ggplot2* (Wickham, 2016; R Core Team,  
238 2023). Springs without occurrence records were excluded from the analysis, as were taxa known to occur in SA GAB  
239 springs generally but without specific location information. Finally, we produced rarefaction curves using the R package  
240 iNEXT (Hsieh & Chao, 2022) from our entire dataset and a subset of the data considering only confident records.  
241 Rarefaction was performed using the rarefaction and extrapolation models for species richness ( $q = 0$ ), 95% confidence  
242 intervals, and 100 replications.

## 243 3 Results

### 244 3.1 Biodiversity of the SA GAB

245 The database we compiled based on our comprehensive literature review captured 3,463 occurrence records  
246 corresponding to 495 putatively distinct taxa (De Deckker, 1979; Greenslade, 1985; Thompson, 1985; Sluys, 1986;  
247 Sokol, 1987; Ling et al., 1989; Mollemans, 1989; Ponder et al., 1989, 1995, 1996; Zeidler & Ponder, 1989; Harvey,  
248 1990, 1998; Zeidler, 1991, 1997; Fatchen & Fatchen, 1993; Morton et al., 1995; Ponder, 1995, 1996a, 1996b, 1996c,  
249 1996d; Clark et al., 2003; Perkins, 2005; Framenau et al., 2006; Furler & Willing, 2006; Kodric-Brown et al., 2007; Page  
250 et al., 2007; Gotch et al., 2008, 2016; Clark, 2009; King, 2009; Murphy et al., 2009, 2013, 2015; Fensham et al., 2010;  
251 Unmack & Dowling, 2010; Guzik et al., 2012, 2019; King et al., 2014; Rossini et al., 2018; DeBoo et al., 2019; Stringer  
252 et al., 2019; Belbin et al., 2021; Department of the Environment, 2022; Witjira National Park Co-management Board,  
253 2022; Froese & Pauly, 2023). Of these records, 2,300 were considered coarse—i.e., corresponding to the supergroup or  
254 complex level, but not informing the presence of taxa at a specific spring group. Invertebrates were by far the most  
255 speciose group in the dataset (42%, 206 taxa) followed by vascular plants (22%, 111 taxa), vertebrates (21%, 102 taxa),  
256 algae (14%, 68 taxa), non-vascular plants (1%, 5 taxa), and fungi (0.4%, 2 taxa) (Figure 2). Sixty-five taxa (13%) are  
257 known only from SA GAB springs, almost all of which are invertebrates. Despite this, invertebrates are also the most  
258 poorly known in the system from a taxonomic standpoint. Just over one-third of invertebrate taxa in the database have



259 formal taxonomic names, whereas the remainder of the fauna are either undescribed (15%) or have an unknown  
260 taxonomic status due to a lack of species-level occurrence records (49%). For other taxonomic groups, these values range  
261 from 50–100% (described), 0–1.5% (undescribed), and 0–50% (unidentified). Further, apart from the Gastropoda no  
262 invertebrate taxon has had its conservation status assessed at the global, federal, or state level (Figure 2). Fifty-eight of  
263 the 170 spring groups assessed in this study had no corresponding occurrence records in the literature. A rarefaction  
264 curve derived from the dataset suggests the artesian wetlands of SA as a whole have not been adequately surveyed, and if  
265 additional locations were examined with equal sampling intensity, dozens of additional taxa would likely be documented  
266 (Figure 3).

267

268 In addition to occurrence records in the dataset, we also identified several records of potential local extinctions  
269 in the literature (Table 1). In all, we retrieved absence records corresponding to one isopod taxon, which may represent  
270 multiple species (*Phreatomerus latipes* (Chilton, 1922) Central, North, South haplotypes) (Kinhill-Stearns, 1984; Kinhill,  
271 1997; Fensham et al., 2010; Guzik et al., 2012), one ostracod (*Ngarawa dirga* De Deckker, 1979) (McLaren et al., 1985;  
272 Kinhill, 1997; Fensham et al., 2010), 15 snail taxa (*Fonsochlela accepta*, members of *F. aquatica*, *F. billakalina*, *F.*  
273 *variabilis*, *Trochidrobia punicea*, *T. smithii* species complexes [all Ponder, Hershler & Jenkins, 1989], *Sinumelon*  
274 *pedasum* Iredale, 1937) (Zeidler & Ponder, 1989; Ponder et al., 1995; Fensham et al., 2010; Rossini et al., 2018;  
275 Department of the Environment, 2022), and four fishes (the Dalhousie goby *Chlamydogobius gloveri* Larson, 1995, Lake  
276 Eyre hardyhead *Craterocephalus eyresii* (Steindachner, 1883), spangled perch *Leiopotherapon unicolor* (Günther, 1859),  
277 Dalhousie gudgeon *Mogurnda thermophila* Allen & Jenkins, 1999) (Zeidler & Ponder, 1989; Kodric-Brown et al., 2007;  
278 Gotch et al., 2016; Rossini et al., 2018; Department of the Environment, 2022; Froese & Pauly, 2023). To the best of our  
279 knowledge, all of these records corresponded to local extinctions as opposed to complete species extinctions (i.e.,  
280 relevant taxa were present in at least one additional location). These extirpations have occurred across 13 spring groups  
281 in the Dalhousie and Kati Thanda–Lake Eyre supergroups (Table 1).

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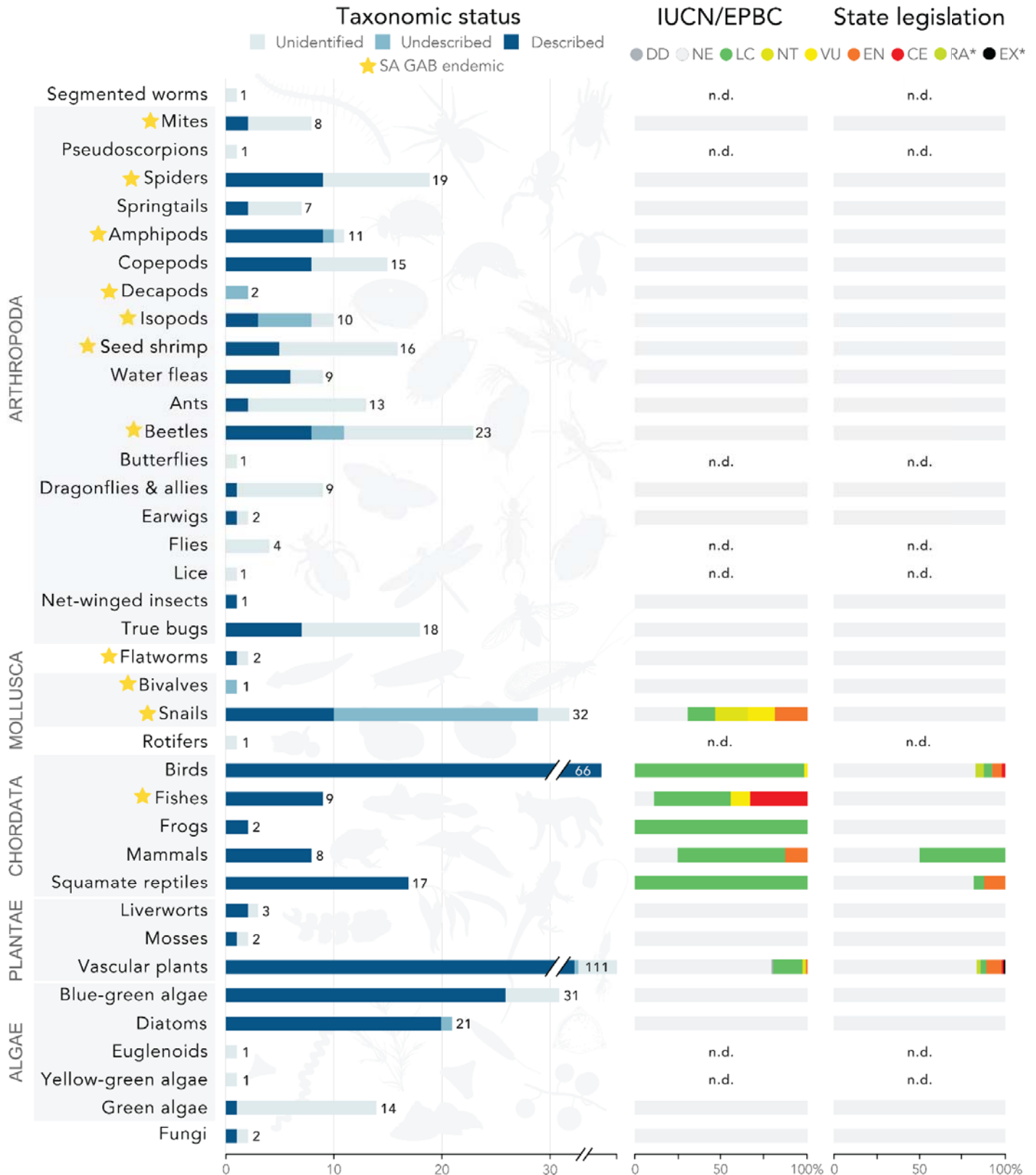
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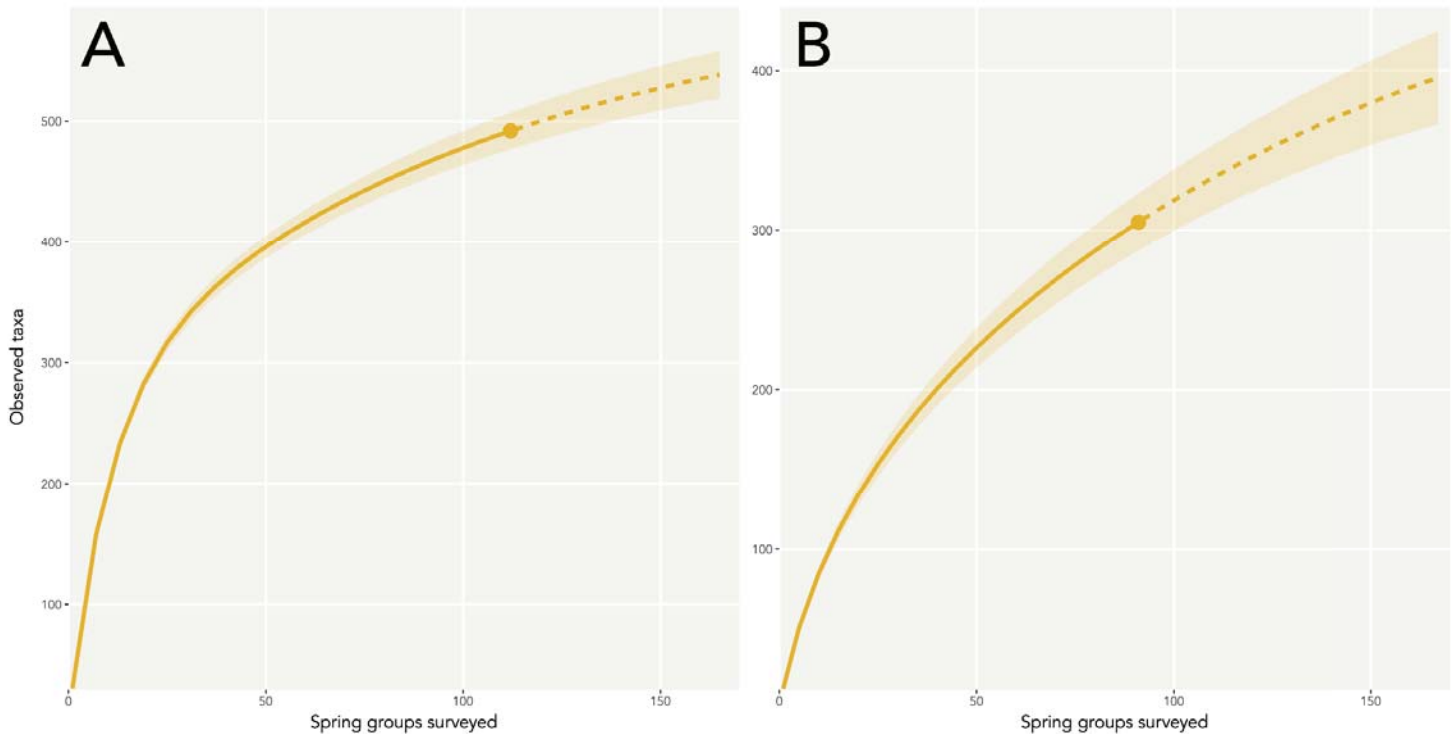
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335 **Figure 2:** Fauna, flora, and fungi associated with SA GAB spring wetlands, with major organismal groups indicated by  
 336 grey boxes. Groups lacking species-level records have an unknown conservation status and are indicated with n.d. (no  
 337 data). \* Rare (RA) and extinct (EX) conservation statuses here are specific to SA/TAS and NSW legislation,  
 338 respectively, and only refer to populations/taxa within those states. Silhouette credits Maxime Dahiriel, Armelle Ansart,  
 339 Mathieu Pélissié, and Lafage via PhyloPic.

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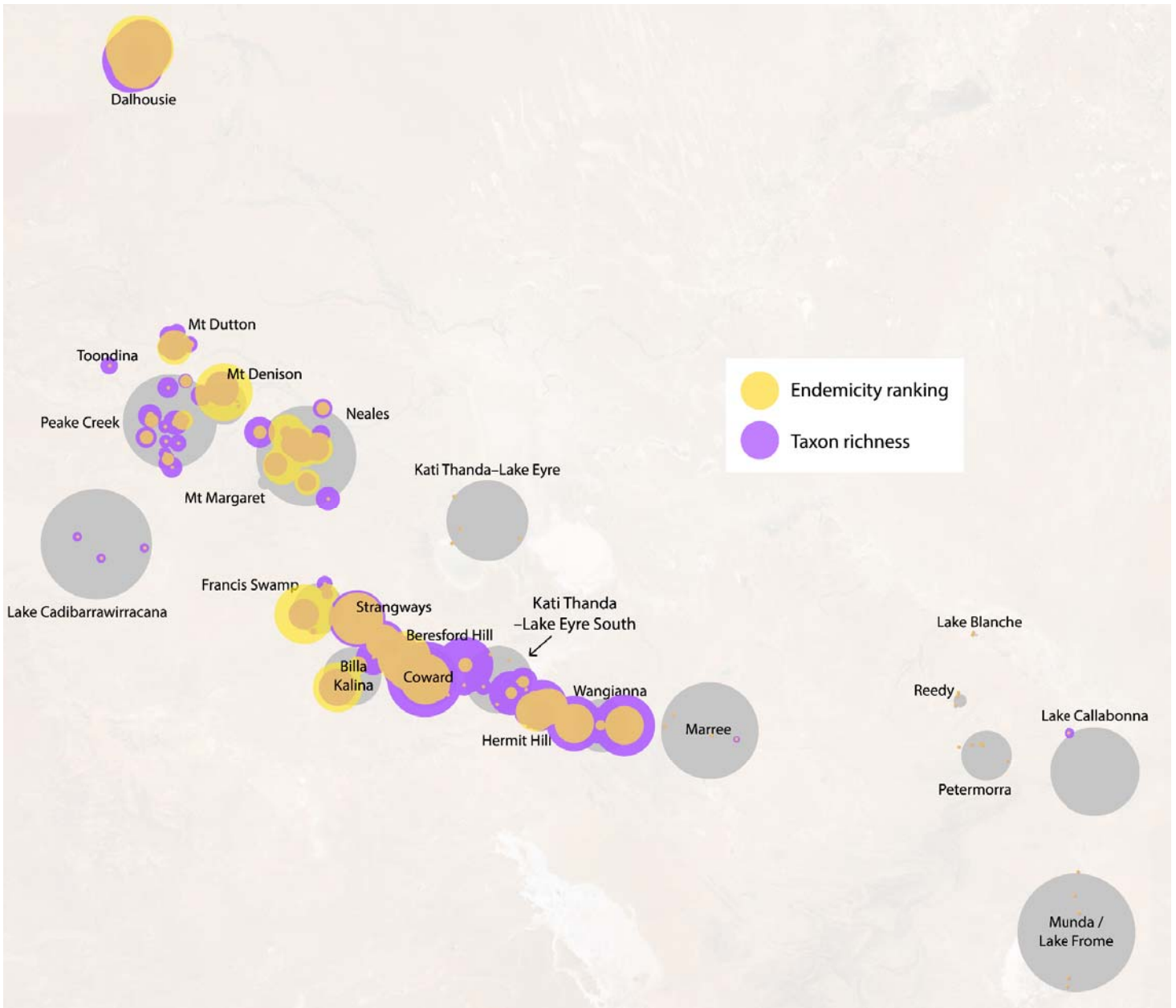
361 **Fig. 3:** Rarefaction curves predicting an increase in taxon richness if additional spring groups were surveyed, calculated  
362 from all occurrence records in the dataset (A) and a subset of the dataset containing only confident records (B). As the  
363 curve does not flatten at the maximum number of known spring groups in the state (170), an increase in sampling effort  
364 *within* springs (as opposed to the sole sampling of additional springs) may also capture a more comprehensive  
365 representation of these communities.  
366

### 367 3.2 Biodiversity metrics and community composition

368  
369 Taxon richness was more heavily impacted by the exclusion of coarse occurrence records as most data in this category  
370 are not spring endemics, whereas endemicity rankings were essentially unchanged between spring locations. In other  
371 words, complexes may be taxon-rich without containing a large number of endemics (e.g., groups within the Francis  
372 Swamp complex; Figure 4) and vice versa (e.g., Beresford Hill). Groups with high levels of richness values generally  
373 corresponded to the Coward, Dalhousie, Billa Kalina, and Hermit Hill complexes and those containing large numbers of  
374 isolated endemics included Dalhousie, Francis Swamp, Mount Denison, and Coward (Figure 4; Supplementary File S2,  
375 available publicly via FigShare: <https://doi.org/10.25909/24457120>).  
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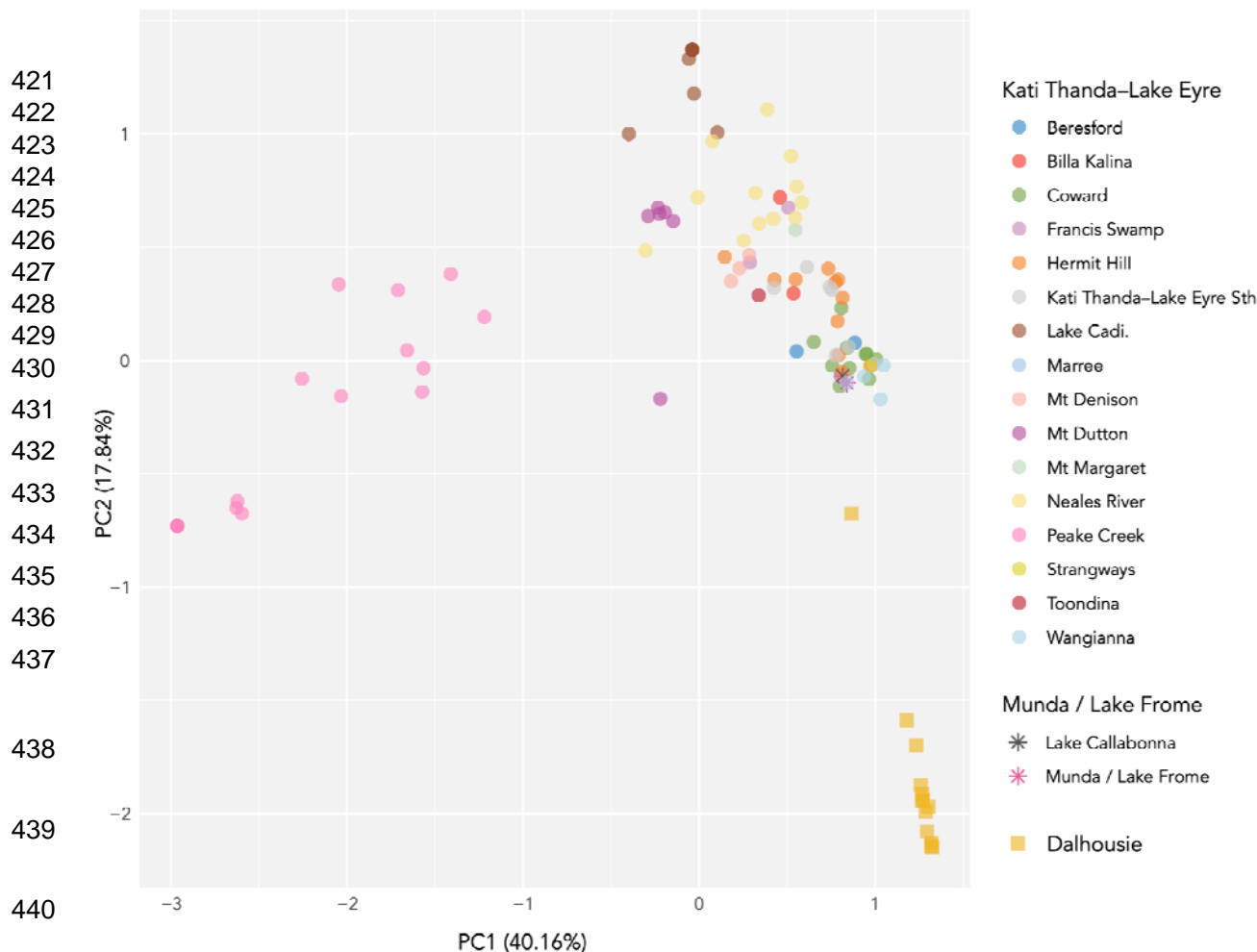
377 While there is some degree of overlap regarding taxon composition of spring complexes, particularly in the Kati  
378 Thanda–Lake Eyre region (Figure 5), locations with similar biodiversity metrics do not necessarily support the same  
379 biotic communities. The Peake Creek, Lake Cadibarrawirracanna, Neales River, and Mount Dutton complexes of the  
380 Kati Thanda–Lake Eyre supergroup are generally non-overlapping and groups from within the Dalhousie  
381 complex/supergroup are easily differentiated from other complexes (Figure 5). In contrast, groups from the Munda / Lake  
382 Frome complex are nested within those from the Kati Thanda–Lake Eyre supergroup. Groups belonging to the same  
383 complex are also generally more similar to one another than those in different complexes. This apparent lack of  
384 connectivity is perhaps unsurprising given the aridity of the surrounding landscape. Overall, spring groups within the  
385 Dalhousie supergroup/complex are the most distinct in the dataset in that they support taxa not known to be associated  
386 with other spring complexes, including highly isolated endemics (Figures 4, 5).  
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413 **Fig. 4:** Rankings of endemicity and biodiversity (i.e., considering non-endemic biota) in SA GAB spring groups  
414 calculated from occurrence records in the literature. Data shown here were calculated using all occurrence records in the  
415 dataset. Metrics calculated from only confident records are available in Supplementary File S2. We use the term “taxon  
416 richness” over “species richness” here as several taxa in the dataset are putative species based on morphological and/or  
417 molecular divergences per their respective taxonomic authorities (Greenslade, 1985; Murphy et al., 2009; Guzik et al.,  
418 2012, 2019; DeBoo et al., 2019; Stringer et al., 2019).

419  
420



441 **Fig. 5:** Principal components analysis (PCA) showing composition of faunal, floral, and fungal communities reliant on  
442 South Australian GAB springs. Both coarse and confident occurrence records in the dataset were used to calculate  
443 Jaccard distances between spring groups. Spring groups lacking occurrence records are not shown here. Groups are  
444 coded by their respective spring complex (colour) and supergroup (shape). Axis labels refer to the amount of variability  
445 explained by each principal component (PC).

#### 446 **4 Discussion**

447 Here, we present a robust literature review of the natural heritage value of the artesian wetlands of Australia, focusing on  
448 SA GAB wetlands as the state contains the majority of springs throughout the continent's arid zone. The data we have  
449 collated highlight several key trends: 1) invertebrates are a poorly-known component of GAB biodiversity; 2) the  
450 composition of GAB biota differs considerably by location; and 3) the true extent of GAB biodiversity is far greater than  
451 is currently appreciated in the literature. We discuss each of these findings below, as well as conservation implications  
452 for GAB springs more broadly and how a centralised biodiversity resource for the GAB could be implemented in future.

#### 453 **4.1 Biodiversity of South Australia's GAB springs**

454 We recorded almost 500 taxa associated with GAB springs and their surrounding wetlands. Invertebrates (namely insects,  
455 crustaceans, and arachnids) represented the largest proportion of this biota and the largest group of endemics (Figure 2).  
456 Invertebrates are poorly documented from biodiversity, ecological, and conservation standpoints relative to the remainder  
457 of GAB biota, and are also overrepresented in extinction records (Table 1). These groups may therefore be the most at

458 risk of decline due to difficulties in devising management strategies. In contrast, while vertebrates were the third-most  
459 speciose group associated with the system (N = 102), almost all were widespread species not of conservation concern  
460 (Fensham & Price, 2004) (Figure 2). We note here that vertebrates such as waterbirds are nonetheless dependent on the  
461 springs as breeding grounds (Badman, 1985), however, and may make more extensive use of these habitats in non-flood  
462 years and/or as “stepping stone” habitats to sustain migrations to nearby Kati Thanda–Lake Eyre. For example, the little  
463 grassbird (*Poodytes gramineus* (Gould, 1845)) and clamorous reed warbler (*Acrocephalus stentoreus* (Hemprich &  
464 Ehrenberg, 1833)) are not restricted to GAB springs but have an obligate relationship with stands of the common reed  
465 *Phragmites australis* (Cav.) Trin. ex Steud. in GAB wetlands in that region (Read, 1997).

466

467         Based on all available information on the occurrence extents of fauna, flora, and funga associated with GAB  
468 springs in SA, the composition of these biota is far from uniform. It has long been recognised that certain locations in the  
469 SA GAB springs are more speciose than others, but this has only been formally quantified with respect to endemic taxa  
470 and without consideration of the distinct spring groups within the Dalhousie complex (Rossini et al., 2018). Here we have  
471 showed that the bulk of endemic biodiversity at Dalhousie is contributed by the Irrwanjira / Main Pool group, and groups  
472 within the Francis Swamp, Mount Denison, and Coward complexes, which in fact house a more speciose and isolated  
473 endemic biota than initially recognised based on published records (Supplementary File S1). Further, taxon-rich locations  
474 are not necessarily rich in endemics (Figure 4). Overall, these data indicate the disturbance or extinction of any given  
475 spring group could represent an enormous loss of biodiversity when both endemic and non-endemic taxa are considered  
476 together. The biodiversity of the South Australian GAB has not been adequately surveyed (Figure 3) and over two-thirds  
477 of records we collated were coarse, meaning they lacked occurrence information at the spring group level. While records  
478 are helpful in distinguishing spring group locations and suggest considerable structure in taxa occupying different spring  
479 complexes and supergroups (Figure 5), they nonetheless highlight a lack of detailed past surveys and the sparseness of  
480 published information on GAB biodiversity.

481

482         The observed variability in the dataset may also be associated with sampling bias. Spring groups with the  
483 highest endemicity rankings, such as those within the Dalhousie and Coward complexes, are also those consistently  
484 associated with high sampling effort in previous surveys (Badman, 1985; Greenslade, 1985; Mitchell, 1985; Sokol, 1987;  
485 Zeidler & Ponder, 1989; Noack, 1994; Ponder et al., 1995; Kovac, 2003). The Kati Thanda–Lake Eyre and Dalhousie  
486 supergroups corresponded to over 2,000 occurrence records, whereas observations within the Munda / Lake Frome  
487 supergroup were restricted to a single, coarse occurrence record of the copepod *Microcyclops dengizicus* (Lepeschkin,  
488 1900) (Zeidler & Ponder, 1989). Munda / Lake Frome has clearly not been adequately sampled, although this does not  
489 necessarily imply the location lacks endemic taxa yet to be characterised (McLaren et al., 1985; Rossini, 2020). Indeed,  
490 widespread taxonomic groups expected to be present in springs, such as the Hymenoptera and other cosmopolitan insect  
491 groups, were also poorly represented in the dataset due to a presumed bias in sampling methods and a lack of past  
492 taxonomic expertise (ADA, pers. obs.). As such, we caution against the derivation of conservation priorities from  
493 locations with high measures of metrics such as species richness alone, but trends are nonetheless evident in this review  
494 regarding spring locations and taxonomic groups that have been undersampled.

495

## 496 4.2 Conservation implications for South Australia's artesian springs

497 Populations of at least 13 taxa are thought to have become locally extinct in the SA GAB springs (Table 1). The bulk of  
498 these extirpations have occurred in spring groups that have ceased to flow as a result of drawdown, namely Venable,  
499 Palura Pintjanha / Priscilla, Marrinha / Hergott, Margaret, and Manda-wardunha / Mundowdna (Fensham et al., 2010).  
500 The Venable and Palura Pintjanha / Priscilla spring groups became extinct around 1990 following predictions by industry  
501 that mining activities would lead to a partial, if not complete, reduction in artesian flow at certain GAB springs (Kinhill-  
502 Stearns, 1982, 1983). Marrinha / Hergott ceased to flow in the mid-1980s as water was withdrawn to supply the nearby  
503 town of Marree (McLaren et al., 1985: 198). Other extirpations have been attributed to human modification of springs,  
504 the presence of overabundant invasive and native species, and potentially insufficient sampling effort. The apparent local  
505 extinction of the fish species, Lake Eyre hardyhead (*Craterocephalus eyresii*) at Toondina and Old Nilpinna may have  
506 been caused by spring excavations and competition from the invasive mosquitofish *Gambusia holbrooki* Girard, 1859,  
507 respectively (Gotch et al., 2016). Extirpations of other fish species at Dalhousie have been attributed to the overgrowth of  
508 the native reed *Phragmites* as a result of decreased grazing pressure following the exclusion of livestock (Kodric-Brown  
509 et al., 2007).

510

511 It is important to establish conservation priorities given the potential for further local, if not species-wide,  
512 extinctions in the SA GAB springs. Here, we ranked spring groups by the number of endemic species supported by each  
513 spring group and their degree of isolation. High-ranking locations mirror those proposed as conservation priorities in past  
514 studies. For example, McLaren *et al.* (McLaren et al., 1985) produced an inventory of fauna, flora, and funga to set out  
515 conservation priorities based on species diversity, the presence of rare species, “naturalness” (i.e., extent of interference  
516 by humans/livestock) and perceived vulnerability to degradation. Spring groups were then ranked as being of low,  
517 medium, or high conservation priority. High-ranking spring groups in this assessment include those found at Dalhousie  
518 and in the Coward and Mt. Denison complexes of the Kati Thanda–Lake Eyre supergroup. Further, Rossini *et al.*  
519 (Rossini et al., 2018) used a ranking system developed by Fensham and Price (Fensham & Price, 2004) (modified here to  
520 produce the endemicity rankings in the dataset) to conclude that groups within the Dalhousie, Strangways, Francis  
521 Swamp, Billa Kalina, and Mt Denison complexes were of high conservation significance relative to other springs. The  
522 highest-ranking spring groups by endemicity in the current study presented here corresponded to all six of the above  
523 complexes, often down to the spring group level (Supplementary File S2). We stress that these trends do not suggest  
524 certain spring groups are more important than others or that insignificant springs exist, as has been implied in past  
525 environmental impact statements (Kinhill-Stearns, 1983; Keane, 1997). This is especially the case for locations that could  
526 be perceived as having low biological importance, such as the extinct Papu-ngaljuru / Primrose spring group, but have  
527 outstanding cultural significance to Aboriginal peoples due to associated mythologies and its use as a major occupation  
528 site (Hercus & Sutton, 1985). Instead, our findings suggest that certain spring groups harbour high numbers of endemics  
529 relative to other locations (but remembering that insufficient sampling effort has occurred) and require targeted  
530 conservation efforts to preserve these short-range taxa.

531

532 In contrast to GABSI, which sought to increase artesian pressure Basin-wide and address the threat of  
533 drawdown, several conservation programs are currently in place which aim to directly improve the condition of SA GAB  
534 spring wetlands (Harris, 1992). Preliminary evidence suggests several of these practices have had a positive impact on  
535 SA GAB spring endemics. Springs are threatened by the presence of large-bodied and hard-hooved livestock which can

536 graze on or trample wetland vegetation, foul the water through their faeces, and potentially reduce invertebrate  
537 biodiversity if stocking levels are high (Hutchinson & King, 1980; Kovac & Mackay, 2009; Fensham et al., 2010; Gotch  
538 et al., 2016). Fencing has been successful in restoring endemic biodiversity of SA GAB springs in some cases, e.g., the  
539 endemic salt pipewort *Eriocaulon carsonii* subsp. *carsonii* F. Muell. in the Hermit Hill complex (Fatchen & Fatchen,  
540 1993) and aquatic invertebrates at springs that were heavily damaged by stock (Kinhill-Stearns, 1984). It is worth noting  
541 that the Dalhousie supergroup and portions of the Hermit Hill complex are also within National Parks, further reducing  
542 stocking pressures without the use of fencing (Harris, 1992). However, the complete exclusion of stock can also lead to  
543 the overproliferation of the native reed *Phragmites*, which occurs as a natural component of SA GAB ecological  
544 communities but when present at unnaturally high densities can reduce endemic floristic diversity (Davies et al., 2010;  
545 Lewis & Packer, 2020) and increase transpiration rates, reducing available habitat for aquatic animals (Fensham et al.,  
546 2010). Best practices for controlling *Phragmites* and preserving GAB spring invertebrates may involve allowing stock  
547 only to graze at certain periods during the year, as burning does not seem to be an appropriate tool to control its growth  
548 (Lamb et al., 2002). Invasive species such as the date palm (*Phoenix dactylifera* L.) and mosquitofish (*Gambusia*  
549 *holbrooki*), among others, are also of concern for the conservation of GAB spring endemics as they can reduce  
550 invertebrate diversity (Holmquist et al., 2011) and threaten native fauna through competition and predation (Gotch et al.,  
551 2016). Complete removal of date palms has now occurred at Dalhousie, which used to house a large population of the  
552 species (Fensham et al., 2010).

553

554 As stated above, there is a large proportion of endemic taxa at several complexes within the SA GAB, namely  
555 those contained within the Dalhousie and Kati Thanda–Lake Eyre supergroups. The data presented here reiterates the  
556 need for targeted surveys in these locations to not only gather additional biodiversity data, but also to develop an  
557 understanding of population dynamics, habitat requirements, and accidental human-mediated translocations of fauna,  
558 flora, and funga per the national recovery plan (Fensham et al., 2010). The plan notably highlights the fact that the  
559 construction of a robust biodiversity inventory for GAB springs has been hindered by a lack of survey effort and  
560 taxonomic expertise. The utility of emerging environmental DNA techniques to capture a holistic picture of SA GAB  
561 spring biodiversity in a non-invasive manner would overcome several of these limitations (Vörös, 2017; West et al.,  
562 2020; Saccò et al., 2022; Beasley-Hall et al., 2023); in this regard, the database developed as part of this study would  
563 assist in the selection of fieldwork locations for a pilot study in this regard. Once taxa are established as occurring within  
564 the SA GAB, monitoring efforts should be undertaken on a regular basis to develop deeper understandings of ecological  
565 knowledge as opposed to presence/absence information.

566

567 Although taxa reliant on GAB-fed springs are protected as a single ecological community under the EPBC Act,  
568 species-level listings are also far overdue. This is particularly the case for invertebrates dependent on the GAB. The  
569 majority of GAB molluscs and crustaceans meet the criteria to be listed as Critically Endangered (Rossini, 2020), but  
570 almost 90% of SA GAB invertebrates recorded here lack conservation listing at a global, federal, or state level, the bulk  
571 of which are either new species awaiting formal description or those completely lacking species-level identifications.  
572 This large knowledge gap is only exacerbated by a lack of understanding of the relationship between spring  
573 characteristics and biodiversity metrics (Rossini et al., 2018; Fensham & Laffineur, 2022) and indeed, relationships  
574 among spring ecological and hydrogeological characteristics themselves (Mudd, 2000; Green & Berens, 2013; Love et al.  
575 (eds), 2013). We hope that resources like the database developed here will inform the prioritisation of certain spring



576 locations and taxonomic groups, ultimately leading to the listing of invertebrates and other spring endemics under  
577 relevant legislation on a per-species basis.

### 578 **4.3 Towards a FAIR and centralised database of GAB spring biodiversity**

579 This GAB database is a significant first step towards tailored conservation of spring biota and represents an important  
580 addition to understandings of spring biodiversity. There are at least 64 endemic taxa across 170 SA spring groups and an  
581 additional 366 non-endemics that occur in association with springs. Our findings present the most comprehensive  
582 collation of biodiversity data for the artesian springs of SA to date, both geographically and from a taxonomic standpoint.  
583 Data related to GAB spring biodiversity is often diffuse, inaccessible, and ultimately maintained in a decentralised  
584 fashion. Early surveys conducted on the springs in the 1980s were largely taxon-specific and conducted by museum  
585 employees or those working in industry, leading to their results either being published locally without digitisation or in  
586 the form of internal reports (Badman, 1985; Greenslade, 1985; McLaren et al., 1985; Mitchell, 1985; Thompson, 1985;  
587 Ling et al., 1989; Mollemans, 1989; Skinner, 1989; Zeidler, 1989). Currently, the only digitised database available to the  
588 general public is the Queensland Government's Springs Database (Queensland Government, 2018), which is freely  
589 available and also includes a Creative Commons Attribution 4.0 licence, a useability rating, and version information. In  
590 contrast, the existing data for SA are neither publicly available or centralised (Kinhill-Stearns, 1982, 1984; Kinhill,  
591 1997). Moreover, nomenclature used to describe SA GAB spring vents and groups has not been standardised, leading to  
592 inconsistencies in names and identifying codes among datasets that refer to identical locations. Similar issues are evident  
593 for hydrogeological information related to the GAB springs (PGBH pers. obs.). There is a clear need to progress beyond  
594 these "silos" and establish a database for the GAB springs nationwide that is not only publicly accessible, but also  
595 supports iterative refinement, updating of biodiversity information over time, and standardises nomenclature used to refer  
596 to springs, particularly spring groups. To this end, in the database published here we have standardised spring  
597 nomenclature across the Government of South Australia's WaterConnect database and identifiers in use by industry  
598 stakeholders.

## 599 **5 Conclusion**

600 Here, we have presented the most robust review of biota supported by the Great Artesian Basin in South Australia. This  
601 dataset is an important resource that we hope will facilitate future studies on SA GAB spring endemics, investigations  
602 into their population dynamics, basic biology, and taxonomy, and ultimately facilitate the listing of relevant taxa under  
603 state and federal environmental legislation. Such resources are essential given the multitude of threats currently facing  
604 springs and their biota, with potential extirpations of populations of at least 13 species in the SA GAB springs to date.  
605 We have also highlighted springs of particular conservation concern which may assist in determining future conservation  
606 priorities for this system. The dataset presented here stresses three major points about GAB spring biodiversity. Firstly,  
607 the majority of taxa reliant on these wetlands are invertebrates, and these animals are also the most poorly known and  
608 conserved. Secondly, ecological communities reliant on artesian springs in SA are largely non-overlapping, irrespective  
609 of whether only endemic taxa, or endemics as well as "incidental" species, are concerned. The extinction or considerable  
610 disturbance of any spring group is therefore likely to lead to a considerable loss of biodiversity and/or genetic diversity.  
611 Finally, the artesian springs of SA have not been adequately sampled by past survey efforts and a considerable proportion  
612 of taxa likely remain to be documented, particularly in understudied locations such as the Munda / Lake Frome

613 supergroup. There also remains a dearth of occurrence records for certain taxa for which taxonomic expertise has been  
614 lacking in the past, such as the wasps and allies (Greenslade, 1985). We recommend that datasets such as these are made  
615 publicly available with the capacity to be modified and updated on an ongoing basis, embodying the gold standard of  
616 digital asset storage.

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## 897 8 Statements and Declarations

### 898 8.1 Author contributions

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### 907 8.2 Competing interests

908 The authors feel it is important to disclose that authors PGBH and MTG, independent researchers affiliated with The  
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911 Research findings and conclusions expressed in this publication are based solely on the analysis of the data and the  
912 scientific merit of the paper.

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916 **9 Tables**

917 **Table 1:** Records of possible local extinctions of fauna in the SA GAB springs. To our knowledge there are no records of  
 918 floral extinctions in the SA GAB. ^ Taxon represents multiple genetically distinct clades which we consider putative  
 919 species, but which of these were present in the below location(s) prior to their apparent extirpation is unknown. KT–LE  
 920 Sth = Kati Thanda–Lake Eyre South. \* GAB spring endemic.

<b>Taxon</b>	<b>Broad grouping</b>	<b>Spring group(s)</b>	<b>Spring complex(es)</b>
<i>Phreatomerus latipes</i> <sup>^*</sup>	Isopod	Venable, Marrinha / Hergott	Hermit Hill, Marree
<i>Ngarawa dirga</i> <sup>*</sup>	Ostracod	Venable, Manda-wardunha / Mundowdna	Hermit Hill, Marree
<i>Fonscochlea accepta</i> <sup>*</sup>	Hydrobiid snail	Venable, Palura Pintjanha / Priscilla	Hermit Hill, KT–LE Sth
<i>Fonscochlea aquatica</i> <sup>^*</sup>	Hydrobiid snail	Margaret	Francis Swamp
<i>Fonscochlea billakalina</i> <sup>*</sup>	Hydrobiid snail	Margaret	Francis Swamp
<i>Fonscochlea variabilis</i> <sup>^*</sup>	Hydrobiid snail	Venable, Palura Pintjanha / Priscilla	Hermit Hill, KT–LE Sth
<i>Trochidrobia punicea</i> <sup>^*</sup>	Hydrobiid snail	Venable, Palura Pintjanha / Priscilla	Hermit Hill, KT–LE Sth
<i>Trochidrobia smithii</i> <sup>^*</sup>	Hydrobiid snail	Margaret	Francis Swamp
<i>Sinumelon pedasum</i>	Camaenid snail	Irrwanjira / Errawanyera	Dalhousie
<i>Chlamydogobius gloveri</i>	Fish	Irrwanjira / Errawanyera, Frog Dreaming, Kirki / Dalhousie Proper, Cadni Dreaming	Dalhousie
<i>Craterocephalus eyresii</i>	Fish	Old Nilpinna, Thuntinha / Toondina	Peake Creek, Toondina
<i>Leiopotherapon unicolor</i>	Fish	Idnjundura / Kingfisher, Ilpikwa	Dalhousie
<i>Mogurnda thermophila</i>	Fish	Frog Dreaming	Dalhousie

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