Invasive traits of freshwater fish database (ITOFF)

Abstract

AIM: Species invasions are a major driver of global biodiversity loss, but only a minority of invasions are successful. Evidence suggests that invasive success is linked to life-history traits. Yet, data on invasive success and species' traits remain fragmented across multiple sources. Here we present the Invasive Traits of Freshwater Fish (ITOFF) database, an interdisciplinary framework that integrates multiple datasets to elucidate the role of life-history traits in shaping invasive success. ITOFF allows seamless access to invasive species data and fosters collaborative actions through knowledge sharing. ITOFF is supported by an innovative web-application that makes complex relationships between invasive and native species accessible to a broad audience. The scientific contribution of ITOFF is illustrated by examining the role of life-history traits and phylogeny in invasion success.

LOCATION: Global.

METHODS: Generalized linear models were used to test the contribution of generation time, trophic level, longevity, and temperature range to invasive success. Through divisive cluster analysis we investigate the role of multiple traits in determining invasive success. Finally, we construct phylogenetic trees to investigate the role of evolutionary history in the invasion process.

RESULTS: ITOFF unifies data for 1917 freshwater fish species representative of invasive species, those species they endanger, and species impacted by invasives but not considered endangered. Invasive species are generally characterized by greater temperature ranges, but are indistinguishable from impacted, endangered, and critically endangered species for the remaining life-history traits. Further, we show that invasive species are generally not distinct from impacted or endangered species when considering multiple traits or phylogeny.

MAIN CONCLUSIONS: ITOFF provides an accessible platform for the improved forecasting of species invasions. ITOFF data shows that classical predictions of life-history traits determining invasive success do not hold amongst freshwater fish species. Forecasting of invasive species must therefore shift towards a wholistic approach encompassing the species and the environment.

1 INTRODUCTION

Invasive species are a leading threat to global biodiversity (Reid et al., 2019; WWF, 2020). 2 3 Further, the economic and social costs of invasive species are severe (Andersen et al., 4 2004; Pimentel, Zuniga and Morrison, 2004; Hoffmann and Broadhurst, 2016). Nevertheless, 5 measures can be implemented to mitigate the negative impacts of invasion (Olson, 2006). 6 For example, recreational and commercial fisheries have been used to control invasive 7 lionfish (*Pterois miles*) populations (Ulman et al., 2021). However, the control of invasive 8 species can be expensive (Heikkila and Peltola, 2004) and the impacts of invasion are often 9 irreversible (Andersen et al., 2004; Heikkila and Peltola, 2004; Keith and Spring, 2013). Prevention is, therefore, the optimal management strategy (Andersen et al., 2004). As 10 invasive propensity differs amongst species (Manchester and Bullock, 2000; Sakai et al., 11 2001), species with greater invasive propensity must be identified for the improved efficacy 12 of preventative management (Sakai et al., 2001). Life-history traits, such as fecundity and 13 14 age of sexual maturation, are generally acknowledged to be critical determinants of invasive 15 success (Holway and Suarez, 1999; Deacon, Ramnarine and Magurran, 2011; Chapple, 16 Simmonds and Wong, 2012). However, few comparative studies have identified traits that promote invasive success in freshwater fish species at the global or major geographical 17 region scale (Bernery, 2022). To address this gap, we present the Invasive Traits of 18 19 Freshwater Fish (ITOFF) database. The ITOFF database provides easy access to research-20 ready data concerning the life-history traits and impacts of invasive freshwater fish species at the global scale. 21

Life-history traits that promote invasive success ('invasive traits') are predicted to be shared
by a significant proportion of successful invasive species (Van Kleunen, Weber and Fischer,
2010). The comparison of trait values between invasive and 'non-invasive' species allows
the identification of invasive traits (Capellini *et al.*, 2015). However, identifying 'non-invasive'

26 species has proved challenging because it is extremely difficult to predict if a currently non-27 invasive species will become invasive in the future (Garcia-Berthou, 2007). For mammalian 28 species, comparing successful and unsuccessful invaders enables the identification of traits 29 critical to invasive success (Capellini et al., 2015). This is possible because of the large 30 number of documented, failed mammalian invasions (Capellini et al., 2015). However, 31 similar analyses are not appropriate for freshwater fish as introductions are not commonly 32 reported in sufficient detail (Garcia-Berthou, 2007). To circumvent this issue, data can be 33 collated on traits of species that are under threat of extinction from invasive species, 34 because endangered species are unlikely to become invasive themselves. Further, we can incorporate trait data for species known to be impacted by invasives but not considered to be 35 at risk of extinction (here onwards referred to as impacted species). This would be an 36 appropriate control group. Data on the life-history traits of invasive species, those species 37 38 they endanger, and impacted species have already been recorded (ISSG, 2000; Froese and Pauly, 2022; IUCN, 2022). Nevertheless, these data are stored in fragmented datasets 39 managed by different bodies. Moreover, in their current format, data are not easily 40 accessible for research purposes. The unified and public provision of life-history trait data for 41 42 invasive species in an immediately accessible and re-usable format for research is urgently 43 needed for the improved, predictive understanding of species invasions.

The ITOFF database creates the unique opportunity to address critical questions related to invasion biology by collating publicly available data on the life-history traits of the most common invasive, critically endangered and impacted freshwater fish species (freely available at <u>https://itoff-dataset.wp.st-andrews.ac.uk</u>). The accessibility of ITOFF is further supported by a web application (available at: <u>https://itoff-dataset.wp.st-</u> <u>andrews.ac.uk/webapplication/</u>) that facilitates the visualization of links between invasive, impacted, endangered and critically endangered species. Networks can be constructed

51 based on the user's focal species and climate region of choice. To our knowledge, this is the

52 first visualisation of the network of impacts associated with invasive species. These networks

reveal the complexity of impacts associated with invasive species. The harmonious
incorporation of invasive, impacted, endangered, and critically endangered freshwater fish
species in one dataset provides a novel approach to the study of invasion biology, while
creating a powerful yet simple tool for future research.

57 To illustrate the scientific significance and accessibility of ITOFF, we investigate differences 58 in life-history traits often predicted to facilitate species invasions among invasive, critically 59 endangered, endangered, and impacted species for temperate, subtropical, and tropical 60 regions. We focus on four traits shown to be critical to invasive success (Pimm, 1991; Crowder 61 and Snyder, 2010; Zerebecki and Sorte, 2011; Deacon and Magurran, 2016; Rosenthal et al., 2021). Namely, generation time, trophic level, longevity, and temperature range. We expected 62 invasive species to have shorter generation times and reduced longevity to facilitate fast 63 64 population growth (Pimm, 1991) and rapid adaptation (Rosenthal et al., 2021). Further, we expected invasive species to have wider temperature ranges, as this increases the likelihood 65 of survival across a range of environments (Zerebecki and Sorte, 2011). Finally, we expected 66 invasive species to be characterised by intermediate trophic levels (i.e. generalists). 67 68 Generalist species have broader resource use which increases competitive advantage (Crowder and Snyder, 2010) and facilitates behavioural flexibility that in turn promotes survival 69 under novel conditions (Deacon and Magurran, 2016). 70

71 There is the possibility that invasive success is not shaped by individual traits but rather by a 72 combination of several. For example, invasive gammarid amphipods cannot be identified from individual traits but can be predicted using ecological profiles that combine multiple 73 characteristics (Grabowski, Bacela and Konopacka, 2007). We therefore test the hypothesis 74 75 that invasive propensity is not tied to individual traits but arises from a combination of multiple 76 traits. As phylogenetic history influences trait similarity, we additionally utilise the detailed taxonomic information provided by the ITOFF database, to untangle the role of adaptation, 77 chance, and evolutionary history in the invasion process. 78

79 **METHODS**

80 The ITOFF Database

A dedicated worksheet was created for the collection of all data. Data fields include group (i.e., invasive, endangered, or impacted), taxonomy, climate region (divided broadly into temperate, tropical, subtropical, polar, and boreal), IUCN status (a detailed measure of extinction risk), impacts on other species, and life-history traits predicted to be critical to invasive success (Table A1 Appendix).

86 A list of non-native freshwater fish species was generated using data from the Global 87 Invasive Species Database (GISD: ISSG, 2000), the U.S. Geological Survey's 88 Nonindigenous Aquatic Species (NAS) database (USGS, 2022), CABI's Invasive Species Compendium (ISC; CABI, 2022), the R interface to FishBase (Froese and Pauly, 2022), and 89 90 data provided by the Global Register of Introduced and Invasive Species (ISSG, 2021). The 91 IUCN Red List of Threatened Species (2022) was used to generate three lists: (1) critically 92 endangered and (2) endangered species for which invasive species are considered a threat, and (3) least concern species that are known to be impacted by invasives. Finally, through 93 94 the screening of invasive species' profiles, provided in text format by online databases (ISSG, 2000; CABI, 2022; USGS, 2022; Froese and Pauly, 2022), a list of species known to 95 96 be impacted by invasive species was generated. Cross-referencing this list with the IUCN Red List allowed for the identification of impacted species considered to be of 'least concern' 97 in terms of extinction risk; these species were included in the impacted group of the dataset. 98 99 Scientific names for all species were updated to be in accordance with FishBase (Froese 100 and Pauly, 2022). Further, the database was curated to identify and omit cryptic duplicate 101 entries.

102 All species in the ITOFF database are categorised as non-native (n = 979), endangered (n =103 299), critically endangered (n = 213), or impacted (n = 381). Non-native species are further 104 categorised as invasive, established, extirpated, failed, or reported. Categorisation of nonnative species within ITOFF reflects the policy of governments, NGOs, and international 105 106 agencies by defining species as invasive only if there is an associated negative impact 107 (Reimanek et al., 2002). Species are considered reported if known to have been introduced 108 to a novel ecosystem, but survival beyond introduction is unknown. Reported is the default 109 categorisation for all non-native species. Non-native species that have successfully founded 110 one or more self-sustaining populations are considered 'established'. Non-native species that have never survived to establishment are considered 'failed'. Established species that were 111 later eradicated across their non-native range by human management are considered 112 'extirpated'. Finally, species are categorised as 'invasive' if there are known (or a reported risk 113 114 of) negative economic, ecological or social impacts associated with the species within its nonnative range. It should be noted that of the 535 extant invasive and established species in the 115 ITOFF database, for which IUCN status is reported, only 7 % are endangered or critically 116 endangered. 117

The ITOFF dataset reports IUCN status and native/non-native status as separate variables. 118 Accordingly, critically endangered, endangered, and impacted species with known non-119 native populations have been identified and recategorized as non-native within the ITOFF 120 121 database. Further, impacted species reported by the IUCN Red List of Threatened Species to have undergone large-scale population declines or localised extinctions due to invasive 122 species were recategorized as 'locally threatened'. This was necessary to ensure that 123 124 impacted species in the dataset are representative of native species that are tolerant of 125 invaders.

All data were collected by a small team of researchers (N = 11). All researchers followed the same systematic methodology for data collection as outlined below. For all species, online 128 databases, namely FishBase (Froese and Pauly, 2022) and the IUCN Red List of Threatened 129 Species (IUCN, 2022), were screened for data on relevant traits. FishBase was further used 130 for its unique estimate of trophic level, and the life-history tool was used to produce appropriate 131 estimates for key traits (age of sexual maturity, generation time, and longevity) when data 132 were otherwise unavailable. Note that estimates for trait values were only included if data were available for the species or the species' family. Estimated and observed values for the same 133 trait are recorded in separate columns within the ITOFF database for clarity. For all non-native 134 135 species, data on traits were additionally collected from species profiles provided by the GISD 136 (ISSG, 2000), ISC (CABI, 2022), and the NAS database (USGS, 2022). Means of introduction was extracted from non-native species' profiles and the R interface to FishBase (USGS, 2022; 137 CABI, 2022; Froese and Pauly, 2022). Links between species (e.g., predation of an invader 138 on one or more native species) were recorded from species profiles provided by the IUCN 139 140 Red List of Threatened Species, and above-mentioned databases for non-native species. Links between species were only included if reported to the species level (i.e., reports of an 141 impact by an invasive species on a taxonomic rank higher than the species level are not 142 included in the ITOFF database). 143

For all traits, if different data sources provided contrasting values, the mean value was used to populate the dataset. Detailed taxonomic information for all species was extracted from the GISD (ISSG, 2000), NAS database (USGS, 2022), and the IUCN Red List of Threatened Species (IUCN, 2021).

To validate the data, hybrids and species of taxonomic uncertainty (i.e., disputed subspecies or inconsistent and unclear nomenclature used in literature) were identified and omitted from the species pool. Conditional formatting rules, created via the Excel Visual Basic for Applications (VBA) tool, were used to control for data entry errors. Namely, entries not matching specified entry formats were visually highlighted, easily recognised, and corrected.

153 Checks were then made in R (R Core Team, 2021) using base functions to ensure consistent 154 formatting was used throughout and to control for duplicates and typos.

155 Linking life-history traits and invasive success

We used the ITOFF database to examine differences in life-history traits amongst invasive, critically endangered, endangered, and impacted species in temperate, subtropical, and tropical regions. Polar and boreal regions were excluded from the investigation because of limited data availability. The four traits of interest, generation time, longevity, temperature range, and trophic level, have been systematically reported as key traits for invasive success (Pimm, 1991; Crowder and Snyder, 2010; Zerebecki and Sorte, 2011; Deacon and Magurran, 2016; Rosenthal *et al.*, 2021).

Values of life-history traits were compared between the four treatment groups using four 163 164 generalized linear models (one per trait). Both observed and estimated trait values as reported 165 within ITOFF were included in the analyses. Each full model included trait value as the 166 response variable and species' status (i.e. invasive, impacted, endangered) as a fixed factor. Separate analyses were conducted for each climate region. Species that occupy multiple 167 climate regions were represented in multiple counts. Locally threatened species were 168 excluded from the analyses as they are representative of native species that have been 169 170 severely impacted by invasive species within part or all of their sympatric range. Error distributions were chosen based on data type and distribution of the response variable. 171

172 Multiple traits in determining invasive success

To investigate the role of multiple traits in determining invasive success a hierarchical cluster analysis, divisive analysis using the diana function of the Cluster package in R v 4.2.0. (Maechler *et al.*, 2022), based on age of sexual maturity, trophic level, and preferred temperature range was performed. We utilised a random subset of 75 tropical species included in the ITOFF database (25 invasive, 25 endangered, and 25 impacted). A subset was

used because data on the three traits of interest were often not reported collectively for all species, and data availability differed between treatment groups. To investigate the relationship between adaptation and evolutionary history in the context of invasive species, we constructed a taxonomic tree utilising the species pool used for the cluster analysis.

182 **RESULTS**

183 **ITOFF Database**

Data availability and sample size differ amongst species categories within ITOFF. Limited data availability reflects knowledge gaps which can be mitigated with the ITOFF database. For example, of all non-native species in the ITOFF database, only 31 are recorded in polar climatic zones. Further, relatively few species are reported as failed invaders, and data availability within this group is limited.

The ITOFF web application, clearly shows that endangered and impacted species are affected by multiple invasive species (Figure 1). Further, invasive species are shown to have impacts on other invasive species. Information on the type of impact (i.e., competition, predation), and profiles for the species selected are provided when using the web application directly.

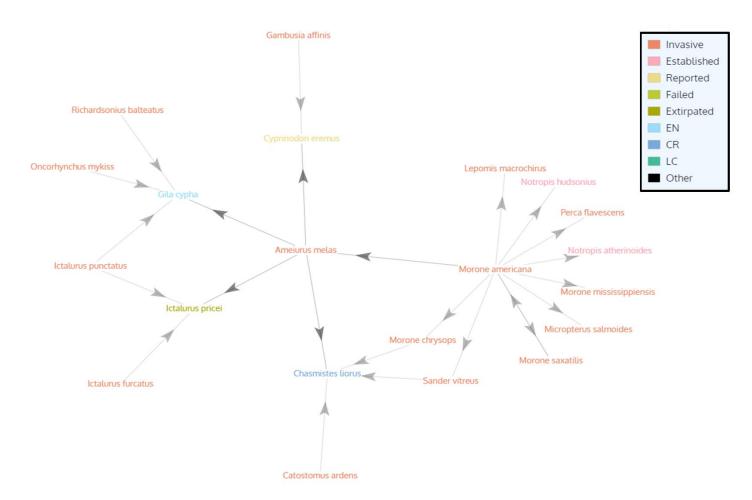


Figure 1. Network showing the direction of impacts of invasive species, as taken from the ITOFF web application. The network is illustrated with *Ameiurus melas* selected as the focal species. The degree of separation between species is set to two, distribution of the species across climate regions is not specified. The web application is illustrated using a colourblind-friendly palette.

197 Linking life-history traits and invasive success

Invasive species had greater preferred temperature ranges than impacted species across temperate (13.23 \pm 1.07, p = 0.005), subtropical (12.2 \pm 1.06, p = 0.001), and tropical (8.38 \pm 1.05, p = 0.001) climate regions. Endangered and critically endangered species did not differ significantly from impacted species for preferred temperature range across any of the three climate regions. For all other traits, observed differences between invasive, endangered, critically endangered, and impacted species did not match predictions (Table 1, Figure 2). Further, there was no consistent pattern in trait values between climate zones. The full model outputs are provided in Table A2 (Appendix).

In temperate zones, invasive, endangered, and critically endangered species did not differ from impacted species in terms of mean trophic level (Table 1, Figure 2). Contrary to our predictions, invasive, and critically endangered species had significantly greater generation times than impacted species. Further, invasive, endangered, and critically endangered species all had greater longevity than impacted species (Table 1, Figure 2).

Similar to temperate zones, invasive, endangered, and critically endangered species had greater generation times and longevity than impacted species in subtropical zones (Table 1, Figure 2). Additionally, neither invasive nor critically endangered species differed significantly from impacted species for trophic level (Table 1, Figure 2). Conversely, endangered species had significantly lower trophic level values than impacted species.

In tropical zones, invasive species had significantly greater generation times, longevity, and preferred temperature ranges than impacted species but did not differ significantly for trophic level (Table 1, Figure 2). In contrast, endangered species did not differ significantly from impacted for any of the four traits. Contrary to our predictions, critically endangered species had greater trophic level scores and reduced longevity relative to impacted species (Table 1, Figure 2). Critically endangered species did not differ from impacted species for generation time or temperature range in tropical zones.

Table 1. General direction of predicted and observed differences between critically endangered,

endangered, and invasive fish species relative to species impacted by invasives but considered least

concern in terms of extinction risk (impacted; control group). Green arrows indicate significantly
 greater trait values for that group, red arrows indicate significantly lower trait values, while a dash

indicates no significant differences. Data are split between three major climate regions (Temperate,

Subtropical, and Tropical). Asterisks denote statistical significance (* $p \le 0.05$, ** $p \le 0.01$, *** $p \le$

229 0.001). N represents the total number of species included from the corresponding climate region. n

indicates the number of species for which data are available for the corresponding trait.

Climate Region: To (N = 744)	emperate						
Trait Critically endangered Endangered					Invasive		
	Predicted	Observed	Predicted	Observed	Predicted	Observed	
Generation time (n = 567)		1 ***			₽	1 **	
Trophic level (n = 725)	-						
Longevity (n = 596)		1 ***		1 **	➡	1 ***	
Preferred temperature range (n = 309)	₽		₽			^ **	

Climate Region: Subtropical (N = 800)						
Trait	Critically e	ndangered	Endangere	d	Invasive	
	Predicted	Observed	Predicted	Observed	Predicted	Observed
Generation time (n = 518)		1 ***			➡	***
Trophic level (n = 779)	-			*		
Longevity (n = 545)				1 ***	➡	1 ***
Preferred temperature range (n = 342)	₽		₽			

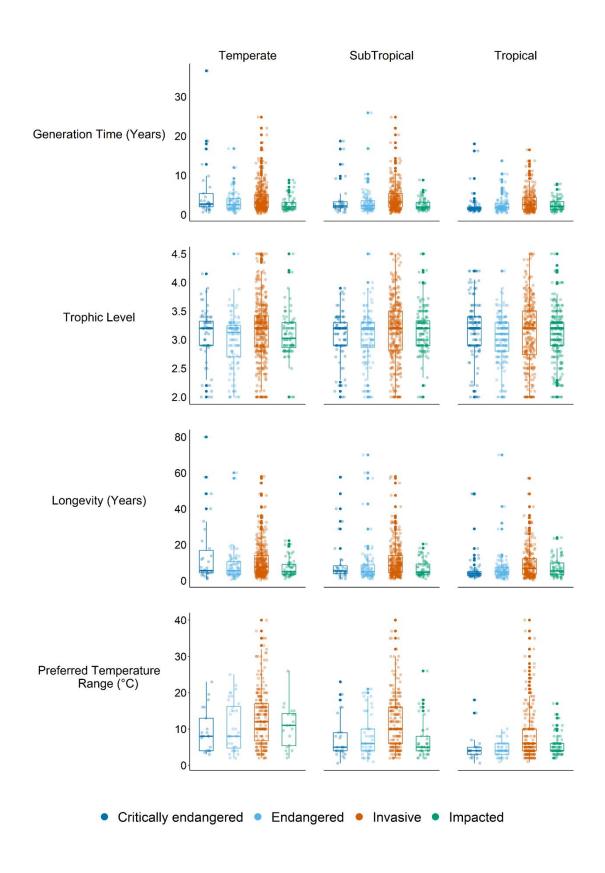
Climate Region: T (N = 974)	ropical					
Trait	Critically e	ndangered	Endangere	d	Invasive	
	Predicted	Observed	Predicted	Observed	Predicted	Observed
Generation time (n = 519)					₽	1 ***
Trophic level (n = 935)	₽		₽			
Longevity (n = 531)		**			-	1 ***
Preferred temperature range (n = 423)	₽		₽			1 ***

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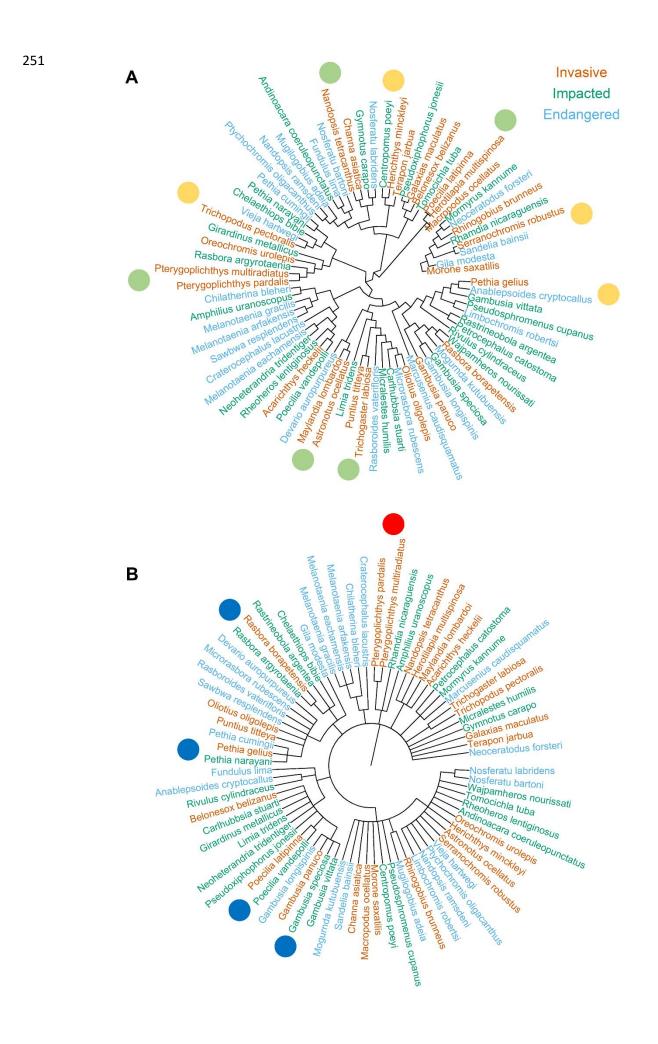


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Figure 2. Box and scatter plots showing observed values for generation time (years) (n = 1604), trophic level (n = 2439), longevity (years) (n = 1672), and preferred temperature range (°C) (n = 1074) for invasive species, those species they endanger, critically endanger, and species impacted by invasives but not considered to be at risk of extinction (impacted; control group). Species that occupy multiple climate regions are represented in multiple counts.

241 Multiple traits in determining invasive success

242 Following a divisive cluster analysis based on three traits, some clusters were found to be composed exclusively by invasive species suggesting that combined traits may be linked to 243 invasive success (Figure 3A, green circles). However, other clusters contained invasive, 244 impacted and/or endangered species (Figure 3A, yellow circles). Accordingly, multiple 245 invasive species have greater similarity to impacted and endangered species as opposed to 246 247 other invasives. Following phylogenetic analyses of the same species, we identified genera that contain a mix of invasive, impacted and/or endangered species (Figure 3B, blue circles). 248 However, a genera (Pterygoplichthys) containing only invasive species was also identified 249 (Figure 3B, red circle). 250



252 Figure 3. (A) Hierarchical cluster (using the Divisive ANAlysis [DIANA] clustering algorithm) based on 253 trait values for age of sexual maturity, trophic level, and temperature range for a random subset of 75 254 tropical species in the ITOFF database (25 invasive species, 25 species endangered due to the spread 255 of invasive species, and 25 species impacted by invasives but not endangered [impacted]). Green circles indicate clusters composed of exclusively invasive species. Yellow circles indicate clusters 256 containing invasive, impacted and/or endangered species. (B) Taxonomic tree constructed for the same 257 258 species pool used in (A). Red circles indicate genera containing only invasive species. Blue circles 259 indicate genera containing invasive, impacted and/or endangered species.

260

261 **DISCUSSION**

ITOFF offers a unique and accessible platform to comprehensively address key questions in 262 263 invasion biology. ITOFF data provides strong evidence that classical predictions for lifehistory traits facilitating invasive success do not hold for freshwater fish on the major climate 264 zone scale (Capellini et al., 2015; Chapple, Simmonds, and Wong, 2012; Crowder and 265 Snyder, 2010). While invasive freshwater fish species are characterized by greater preferred 266 267 temperature ranges, they cannot be reliably identified by generation time, trophic level, longevity, or multiple trait analyses. It is therefore essential that we shift the focus of invasive 268 species management towards a holistic approach that considers the species (Chapple, 269 Simmonds, and Wong, 2012), environment (Pyšek et al., 2020), and evolutionary history 270 271 (Mazzamuto et al., 2016).

ITOFF Database

The ITOFF database reveals key gaps in data availability amongst invasive species and 273 those species they impact. For example, there is data inequality amongst non-native 274 275 species. Specifically, trait data are unavailable for many failed invaders. Nonetheless, 276 comparisons between non-native groups (i.e. invasive, and failed invaders) will allow the 277 identification of traits that facilitate invasive success during one or more stages of the 278 invasion process (Capellini et al., 2015). Future research should target both successful and 279 unsuccessful invasive freshwater fish species. Additionally, by using ITOFF it was possible 280 to identify key data gaps concerning the distribution of invasive species across climate

regions. One of the major factors affecting the spread of non-native species is tropicalization
(Osland *et al.*, 2021). High latitude regions are therefore hot spots for the spread of nonnative species (Goldsmit *et al.*, 2020; Hughes *et al.*, 2020). However, data on freshwater
invasions in these regions are limited relative to tropical, subtropical and temperate regions.
It is therefore critical that invasive species research is distributed evenly across climate
zones.

Another unique addition provided by ITOFF to invasion biology research is the visualization of impacts associated with invasive species. This is an accessible tool with applications for academic research, citizen science, and education. Networks highlight the complexity of impacts associated with invasive species and can be used to guide invasive species management and identify target study species.

292 Linking life-history traits and invasive success

293 Collectively, the results of our trait analyses provide weak evidence to support the classical 294 predictions that life-history traits promote invasive success (Capellini et al., 2015; Chapple, 295 Simmonds, and Wong, 2012; Crowder and Snyder, 2010). The influence of environmental 296 variability and propagule pressure on invasive success may explain the failure to detect life-297 history traits that ubiquitously promote invasive success (Moyle and Light, 1996; Tabak, Webb, 298 and Miller, 2018). However, a key finding of this study is the greater preferred temperature ranges for invasive species across temperate, tropical, and subtropical climate zones. This 299 result confirms predictions that broader abiotic tolerances increase the likelihood of invasive 300 success by facilitating survival in a greater range of novel environments (Zerebecki and Sorte, 301 2011; Bates et al., 2013). Nonetheless, it should be noted that there is large overlap in trait 302 303 values amongst all treatment groups for all traits, including preferred temperature range (Figure 2). 304

305 *Multiple traits in determining invasive success*

The confounding results of the hierarchical cluster and phylogenetic analyses stress the complexity of the invasion process and suggest that both life-history traits and context are key in determining invasive success. These results prompt discussion into the role of ancestral traits in determining invasive success and highlight the need for targeted research of closely related species that differ in invasive status.

Overlap of trait values between invasive species and those species they impact limits the 311 successful identification of potential invasive species on the climatic zone or global scale. This 312 key result is in line with previous conceptual research which predicted that ubiquitous invasive 313 314 traits are unlikely to be identified (Moyle and Light, 1996). Universal invasive species management is likely to be less efficient than localized management strategies and we advise 315 316 that this be reflected in policy. Successful prediction of invasion on reduced spatial scales is 317 possible with detailed knowledge of the biotic and abiotic conditions of an ecosystem (Kolar and Lodge, 2002; Marchetti, Moyle and Levine, 2004; Vila-gispert, Alcaraz and Garcia-318 Berthou, 2005). However, local scale approaches to invasive species management are work 319 320 intensive. Nonetheless, as current ability to forecast aquatic species invasions on the large spatial scale remains limited, localised management remains the most viable approach (Moyle 321 and Light, 1996). 322

323 Future directions

The accurate prediction of future invasive species is a central component of governmental non-native species management. In the United Kingdom, 'The Great Britain Invasive Nonnative Species Strategy' (Department for Environment Food and Rural Affairs [DEFRA], 2015) is used as a strategic framework for current management (e.g. 'River basin management plans'; Environment Agency, 2022). Horizon scanning is the only predictive management measure named in the DEFRA framework, and is largely reliant on the use of 'biological and ecological criteria' to identify species with a high-risk of becoming invasive in the UK. The results of this study demonstrate that predictive frameworks must consider ecosystem characteristics, evolutionary history and contemporary evolution in combination with lifehistory traits for improved accuracy of invasive species forecasts. We argue that urgent research concerning the relationship of life-history traits, evolutionary history, rapid adaptation and invasive success is needed. ITOFF opens the opportunity to test the above avenues of research, which ultimately will further our understanding of the factors contributing to successful invasion in freshwater fish, a model taxon for the study of invasion biology.

338 Data Availability Statement

The ITOFF core dataset is available at https://doi.org/10.5281/zenodo.10135093

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REFERENCES

Amiel, J. J., Tingley, R. and Shine, R. (2011) 'Smart moves: Effects of relative Brain size on Establishment success of invasive amphibians and reptiles', *PLoS ONE*, 6(4), pp. 4–7. doi: 10.1371/journal.pone.0018277.

Andersen, M. C. et al. (2004) 'Risk assessment for invasive species', Risk Analysis, 24(4).

Bates, A. E. *et al.* (2013) 'Geographical range, heat tolerance and invasion success in aquatic species', *Proceedings of the Royal Society B: Biological Sciences*, 280(1772). doi: 10.1098/rspb.2013.1958.

Bernery, C., Bellard, C. A., Courchamp, F., Brosse, S., Gozlan, R., Jaric, I., Teletchea, F., and Leroy, B. (2022) 'Freshwater fish invasions: A comprehensive review', *Annual Review of Ecology, Evolution, and Systematics*, 53, pp. 427-456 doi: 10.1146/annurev-ecolsys-032522-015551

CABI (2022) *Invasive Species Compendium*. Wallingford, UK: CAB International. <u>www.cabi.org/isc</u>.

Capellini, I. *et al.* (2015) 'The role of life-history traits in mammalian invasion success', *Ecology Letters*, 18(10), pp. 1099–1107. doi: 10.1111/ele.12493.

Chapple, D. G., Simmonds, S. M. and Wong, B. B. M. (2012) 'Can behavioral and personality traits influence the success of unintentional species introductions?', *Trends in Ecology & Evolution*, 27(1), pp. 57–64. doi: 10.1016/j.tree.2011.09.010.

Crowder, D. W. and Snyder, W. E. (2010) 'Eating their way to the top? Mechanisms underlying the success of invasive insect generalist predators', *Biological Invasions*, 12(9), pp. 2857–2876. doi: 10.1007/s10530-010-9733-8.

Deacon, A. E., Ramnarine, I. W. and Magurran, A. E. (2011) 'How Reproductive Ecology Contributes to the Spread of a Globally Invasive Fish', *PLoS ONE*, 6(9), pp. 1–7. doi: 10.1371/journal.pone.0024416. Drake, J. M. (2007) 'Parental investment and fecundity, but not brain size, are associated with establishment success in introduced fishes', *Functional Ecology*, 21(5), pp. 963–968. doi: 10.1111/j.1365-2435.2007.01318.x.

Garcia-Berthou, E. (2007) 'The characteristics of invasive fishes: what has been learned so far?', *Journal of Fish Biology*, 71, pp. 33–55. doi: 10.1111/j.1095-8649.2007.01668.x.

Froese, R. and Pauly, D. Editors. (2022) FishBase, World Wide Web electronic publication. www.fishbase.org

Goldsmit, J., McKindsey, C. W., Schlegel, R. W., Stewart, D. B., Archambault, P., and Howland, K. L. (2020) 'What and where? Predicting invasion hotspots in the Arctic marine realm', *Global Change Biol*ogy, 26, pp. 4752–4771. doi:

https://doi.org/10.1111/gcb.15159

Grabowski, M., Bacela, K. and Konopacka, A. (2007) 'How to be an invasive gammarid (Amphipoda: Gammaroidea) – comparison of life-history traits', *Hydrobiologia*, 590, pp. 75–84. doi: 10.1007/s10750-007-0759-6.

Heikkila, J. and Peltola, J. (2004) 'Analysis of the Colorado potato beetle protection system in Finland', *Agricultural Economics*, 31, pp. 343–352. doi: 10.1016/j.agecon.2004.09.019.

Hoffmann, B. D. and Broadhurst, L. M. (2016) 'The economic cost of managing invasive species in Australia', *NeoBiota*, 31, pp. 1–18. doi: 10.3897/neobiota.31.6960.

Holway, D. A. and Suarez, A. V (1999) 'Animal behavior: an essential component of invasion biology', *Tree*, 14(8), pp. 328–330. doi: https://doi.org/10.1016/S0169-5347(99)01636-5.

Hughes, KA, Pescott, OL, Peyton, J, Adriaens, T., Cottier-Cook, E. J., Key, G., Rabitsch,

W., Tricarico, E., Barnes, D. K. A., Baxter, N., Belchier, M., Blake, D. Convey, P.,

Dawson, W., Frohlich, D., Gardiner, L. M., González-Moreno, P., James, R.,

Malumphy, C., Martin, S., Martinou, A. F., Minchin, D., Monaco, A., Moore, N., Morley, S.

A., Ross, K., Shanklin, J., Turvey, K., Vaughan, D., Vaux, A. G. C., Werenkraut, V.,

Winfield, I. J., Roy, H. E. (2020) 'Invasive non-native species likely to threaten biodiversity

and ecosystems in the Antarctic Peninsula region', Global Change Biology, 26, pp. 2702-

2716. doi: https://doi.org/10.1111/gcb.14938

Jessop, A., Michalopoulou, A., Coonan, C., Mazzei, L., Sutherland O'Brien, E., Brady, G.,

Davison, C., Gourlay, W., Henderson, E., Lornie, A., McCloskey, E., Ramsay, H., Wilson, S.,

Shimadzu, H., and Barbosa, M. (2023). Invasive traits of freshwater fish database (ITOFF)

(Version 5) [Data set] Zenodo. https://doi.org/10.5281/zenodo.10135093

Keith, J. M. and Spring, D. (2013) 'Agent-based Bayesian approach to monitoring the

progress of invasive species eradication programs', PNAS, 110(33), pp. 1-6. doi:

10.1073/pnas.1216146110/-

/DCSupplemental.www.pnas.org/cgi/doi/10.1073/pnas.1216146110.

Invasive Species Specialist Group (ISSG) (2000) *Global Invasive Species Database* (GISD), http://www.iucngisd.org/gisd/

Invasive Species Specialist Group (ISSG) (2021) *Aquatic Species Extraction*, dataset, unpublished.

IUCN (2022) *The IUCN Red List of Threatened Species*, Version 2022-2. https://www.iucnredlist.org.

Manchester, S. J. and Bullock, J. M. (2000) 'The impacts of non-native species on UK biodiversity and the effectiveness of control', *Journal of Applied Ecology*, 37(5), pp. 845–864. Marchetti, M. P., Moyle, P. B. and Levine, R. (2004) 'Invasive species profiling? Exploring the characteristics of non-native fishes across invasion stages in California', *Freshwater Biology*, 49(5), pp. 646–661. doi: 10.1111/j.1365-2427.2004.01202.x.

Mazzamuto, M.V., Galimberti, A., Cremonesi, G., Pisanu, B., Chapuis, J.-L., Stuyck, J., Amori, G., Su, H., Aloise, G., Preatoni, D.G., Wauters, L.A., Casiraghi, M. and Martinoli, A. (2016), 'Preventing species invasion: A role for integrative taxonomy?', *Integrative Zoology*, 11, pp. 214-228. doi: https://doi.org/10.1111/1749-4877.12185 Olson, L. J. (2006) 'The economics of terrestrial invasive species: a review of the literature', *Agricultural and Resource Economics Review*, 35(1), pp. 178–194.

Osland, M. J., Stevens, P. W., Lamont, M. M., Brusca, R. C., Hart, K. M., Waddle, J. H., Langtimm, C. A., Williams, C. M., Keim, B. D., Terando, A. J., Reyier, E. A., Marshall, K. E., Loik, M. E., Boucek, R. E., Lewis, A. B. and Seminoff, J. A. (2021), 'Tropicalization of temperate ecosystems in North America: The northward range expansion of tropical organisms in response to warming winter temperatures', Global Change Biology, 27, pp. 3009-3034. doi: https://doi.org/10.1111/gcb.15563

Pimentel, D., Zuniga, R. and Morrison, D. (2004) 'Update on the environmental and economic costs associated with alien-invasive species in the United States', *Ecological Economics*, 52, pp. 273–288. doi: 10.1016/j.ecolecon.2004.10.002.

Pimm, S. (1991) The balance of nature? Chicago: University of Chicago Press.

Pyšek, P., Hulme, P.E., Simberloff, D., Bacher, S., Blackburn, T.M., Carlton, J.T., Dawson,

W., Essl, F., Foxcroft, L.C., Genovesi, P., Jeschke, J.M., Kühn, I., Liebhold, A.M., Mandrak,

N.E., Meyerson, L.A., Pauchard, A., Pergl, J., Roy, H.E., Seebens, H., van Kleunen, M., Vilà, M., Wingfield, M.J. and Richardson, D.M. (2020), 'Scientists' warning on invasive alien species'. *Biol Rev*, 95, pp. 1511-1534. doi: https://doi.org/10.1111/brv.12627

Reid, A. J. *et al.* (2019) 'Emerging threats and persistent conservation challenges for freshwater biodiversity', *Biological Reviews*, 94(3), pp. 849–873. doi: 10.1111/brv.12480. Rejmánek, M. *et al.* (2002) 'Biological invasions: politics and the discontinuity of ecological terminology', *Bulletin of the Ecological Society of America*, 83(2), pp. 131–133. doi: 10.2307/20168698.

Rosenthal, W. C. *et al.* (2021) 'Invasion and rapid adaptation of guppies (*Poecilia reticulata*) across the Hawaiian Archipelago', *Evolutionary Applications*, 14(7), pp. 1747–1761. doi: 10.1111/eva.13236.

Sakai, A. K. et al. (2001) 'The population biology of invasive species', Annual Review of

Ecology and Systematics, 32, pp. 305-332.

Sol, D. *et al.* (2008) 'Brain size predicts the success of mammal species introduced into novel environments', *American Naturalist*, 172(S1), pp. 63–71. doi: 10.1086/588304.

Sol, D. *et al.* (2012) 'Unraveling the Life history of successful invaders', *Science*, 337(6094), pp. 580–583.

Suarez, A. V. and Tsutsui, N. D. (2008) 'The evolutionary consequences of biological invasions', *Molecular Ecology*, 17(1), pp. 351–360. doi: 10.1111/j.1365-294X.2007.03456.x.

Tabak, M. A., Webb, C. T. and Miller, R. S. (2018) 'Propagule size and structure, life history, and environmental conditions affect establishment success of an invasive species' *Scientific Reports*, 8. doi: https://doi.org/10.1038/s41598-018-28654-w

Ulman, A. *et al.* (2021) 'Low pufferfish and lionfish predation in their native and invaded ranges suggests human control mechanisms may be necessary to control their Mediterranean abundances', *Frontiers in Marine Science*, 8, pp. 1–13. doi: 10.3389/fmars.2021.670413.

U.S. Geological Survey (USGS) (2022) *Nonindigenous Aquatic Species Database*, Gainesville, FL. http://nas.er.usgs.gov

Vall-Llosera, M. and Sol, D. (2009) 'A global risk assessment for the success of bird introductions', *Journal of Applied Ecology*, 46(4), pp. 787–795. doi: 10.1111/j.1365-2664.2009.01674.x.

Van Kleunen, M., Weber, E. and Fischer, M. (2010) 'A meta-analysis of trait differences between invasive and non-invasive plant species', *Ecology Letters*, 13(2), pp. 235–245. doi: 10.1111/j.1461-0248.2009.01418.x.

Vila-gispert, A., Alcaraz, C. and Garcia-Berthou, E. (2005) 'Life-history traits of invasive fish in small Mediterranean streams', *Biological Invasions*, 7(1), pp. 107–116. doi: 10.1007/s10530-004-9640-y.

Winker, H. et al. (2011) 'Life history and population dynamics of invasive common carp,

Cyprinus carpio, within a large turbid African impoundment.', Marine and Freshwater

Research, 62, pp. 1270–1280. doi: https://doi.org/10.1071/MF11054.

Wong, J. W. Y., Meunier, J. and Kölliker, M. (2013) 'The evolution of parental care in insects:

The roles of ecology, life history and the social environment', Ecological Entomology, 38(2),

pp. 123–137. doi: 10.1111/een.12000.

Wourms, J. P. and Lombardi, J. (1992) 'Reflections on the evolution of piscine viviparity',

American Zoologist, 32(2), pp. 276–293. doi: 10.1093/icb/32.2.276.

WWF (2020) Living Planet Report 2020 - Bending the curve of biodiversity loss, Living Planet Report 2020.

Zerebecki, R. A. and Sorte, C. J. B. (2011) 'Temperature tolerance and stress proteins as mechanisms of invasive species success', *PLoS ONE*, 6(4). doi:

10.1371/journal.pone.0014806.

APPENDIX

ITOFF usage notes

The core ITOFF dataset can be downloaded from the Zenodo public repository (https://doi.org/10.5281/zenodo.10135093) and should be referenced by citing the present paper. Exemplary R code and a document describing the variables included in the core ITOFF dataset are also available via the Zenodo public repository. Additionally, a dynamic, continuously updated, version of the ITOFF database can be downloaded from the ITOFF project website (https://itoff-dataset.wp.st-andrews.ac.uk). Data are organised such that the dataset is easy to use and serves as a readily available tool for invasion biology research. The web application can be accessed via the ITOFF project website (https://itoff-dataset.wp.st-andrews.ac.uk/webapplication/). Guidance on how to use the web application is also provided on the project website.

The primary intended use of the dataset is the identification of shared invasive traits amongst freshwater fish. Traits can be identified through broad-scale comparisons made between invasive, critically endangered, endangered, and impacted species. Once identified, these traits will improve current ability to forecast species invasions and highlight 'high-risk' species with great invasive potential. The same data offer insights into traits that may increase a native species' susceptibility to invasion and can therefore be used to inform policy, conservation, and invasive species management. Future studies can address differences in invasive traits between climate regions. This research is of great importance given the current knowledge gap concerning how climate change will impact the outcome of future species invasions. Additionally, using the provided taxonomic information, future studies can untangle the role of adaptation, chance, and evolutionary history in the invasion process (Figure 1A, Appendix). Further, cluster analyses can investigate similarity across multiple traits for invasive, endangered, and impacted species (Figure 1B, Appendix). Cluster analyses will be important to invasive trait identification as invasive propensity is not always tied to individual traits but arises from a combination of multiple traits (Grabowski, Bacela and Konopacka, 2007).

Trait	Description	Link to invasive success	References
Reproductive guild (included as mode of reproduction, mode of fertilization, and extent of parental care)	Species are categorised as non- guarders, guarders, and bearers. Bearers include livebearers, mouthbrooders, or any species that utilises a brood pouch. Fish that actively defend externally laid eggs are guarders. Non-guarders include egg scatterers and any species for which neither parent actively defends eggs. Additional information is provided in a	Greater parental care promotes colonisation success, especially in harsh environments. Further, establishment success was correlated with greater parental investment in Californian fishes and a global study using 14 fish species. Further, live birth is linked to increased offspring survival, colonisation success, and is advantageous when mating	(Wourms and Lombardi, 1992; Marchetti, Moyle and Levine, 2004; Drake, 2007; Deacon, Ramnarine, and Magurran, 2011; Wong, Meunier and Kölliker, 2013)
	include egg scatterers and any species for which neither parent actively defends eggs. Additional	species. Further, live birth is linked to increased offspring survival, colonisation success, and is	Wong, Meunier and Kölliker,

Table A1. Descriptions of the life-history traits included in the ITOFF database and evidence for their link to invasive success.

Absolute fecundity	brooding, nesting, internal live- bearing). Total number of eggs in an average female (for egg-layers), or the average, total number of offspring produced over the entire lifetime of livebearers.	population sizes are limited following introduction). Greater reproductive output is positively associated with mammalian invasion success on the global scale	(Capellini <i>et al.,</i> 2015)
Brood size	The number of offspring produced, or the number of eggs laid at one discrete time	Production of larger broods reduces the risk of extinction for small populations and is linked to non-native mammalian establishment success	(Pimm, 1991; Capellini <i>et al.,</i> 2015)
Age of sexual maturity	The average age (years) at which an individual is able to successfully reproduce. Provided as a mean value across sex in the ITOFF database	Early maturity is linked to rapid population growth, establishment successs, and improved resistance to demographic stochasticity and anthropogenic management strategies in the invasive common carp, <i>Cyprinus carpio</i> . Conversely, non-native fish species of Catalan streams are characterised by late maturity relative to native species	(Vila-gispert, Alcaraz and Garcia-Berthou, 2005; Winker <i>et</i> <i>al.</i> , 2011)
Generation time	The average age (years) of parents at the time their first offspring are born	Short generation times promote rapid population growth, colonisation success, and facilitates rapid adaptation	(Rosenthal <i>et al.,</i> 2021)
Trophic level and feeding type	Trophic level is a value estimated by FishBase to show the 'rank of a species in a food web'. Defined as 1 + the mean trophic level of the focal species' food items (Froese and Pauly, 2022). Under feeding type, fish are categorised by mode of feeding (e.g. grazer or predator).	Generalist diets allow for behavioural flexibility in novel environments and increase the competitive ability of non-native species	(Crowder and Snyder, 2010)

Relative brain weight	Given as brain weight (mg) divided by body weight (g).	Larger brains may facilitate greater behavioural flexibility and promote survival in novel environments. Brain size is correlated with invasive success in birds, mammals, amphibians, and reptiles	(Sol <i>et al.</i> , 2008; Vall-Llosera and Sol, 2009; Amiel, Tingley and Shine, 2011)
Longevity (included as maximum age in years, maximum weight, and maximum length)	The maximum age, weight, or length that an individual of each species can reach. Reported as the mean value of the approximate age, weight, or length of the oldest known individuals amongst populations.	Though life-history strategies are generally acknowledged to be an important factor in determining invasive success, there is debate as to whether or not invasive species are more often characterised by fast or slow life-histories. Fast life- histories are predicted to promote rapid population growth and establishment success. However, slow life-histories can also increase establishment success in a wider range of environmental conditions by prioritising survival and delaying reproduction	(Sol <i>et al.,</i> 2012)
Temperature range (included as both preferred and tolerable)	Preferred temperature range: the range of temperatures over which the species can survive indefinitely. Reported in ITOFF as the range of water temperatures experienced throughout the species range. Tolerable temperature range: the range of temperatures over which individuals of the species can survive brief exposure	Wider abiotic tolerances increase the likelihood of survival within a wider range of novel environments	(Zerebecki and Sorte, 2011; Bates <i>et al.</i> , 2013)

Table A2. Summary of the results for generalized linear models investigating differences in life-history trait values between invasive species, species impacted by invasive species but not facing an elevated risk of extinction (impacted), endangered species that are threatened by invasive species, and critically endangered species that are threatened by invasive species. Results are shown for four life-history traits shown to be critical to invasive success. Data are split between major climate regions. Species present in multiple climate regions are represented in multiple counts. N represents the total number of species included for the corresponding climate region. n indicates the number of species for which data are available for the corresponding trait. Overall model adjusted R² is reported for models with Gaussian error distributions, McFadden's pseudo-R² is reported for models with Poisson error distributions.

Climate Region: Temp (N = 744)	erate	-		
Trait	Impacted (Reference group)	Critically endangered	Endangered	Invasive
Generation time (n = 567, McFadden's R ² = 0.02)	2.71 ± 1.08 (z = 12.92, p < 0.001)	5.73 ± 1.11 (z = 7.06, p < 0.001)	3.24 ± 1.1 (z = 1.84, p = 0.07)	4.06 ± 1.08 (z = 4.97, p < 0.001)
Trophic level (n = 725, R ² = 0.02)	3.08 ± 0.05 (t = 60.18, p < 0.001)	3.09 ± 0.08 (t = 0.05, p = 0.95)	2.97 ± 0.07 (t = -1.65, p = 0.1)	3.16 ± 0.06 (t = 1.42, p = 0.16)
Longevity (n = 596, McFadden's R ² = 0.03)	7.16 ± 1.05 (z = 41.8, p < 0.001)	15.18 ± 1.07 (z = 11.66, p < 0.001)	8.38 ± 1.06 (z = 2.63, p = 0.009)	10.56 ± 1.05 (z = 7.87, p < 0.001)
Preferred Temperature range (n = 309, McFadden's R ² = 0.02)	10.85 ± 1.07 (z = 35.12, p < 0.001)	9.62 ± 1.1 (z = -1.23, p = 0.22)	10.61 ± 1.09 (z = -0.26, p = 0.79)	13.23 ± 1.07 (z = 2.83, p = 0.005)

Climate Region: Subtr	Climate Region: Subtropical				
(N = 800)					
Trait	Impacted	Critically	Endangered	Invasive	
	(Reference group)	endangered			
Generation time	2.59 ± 1.08	3.85 ± 1.12	3.14 ± 1.1	4.2 ± 1.08	
(n = 518, McFadden's	(z = 12.62, p <	(z = 3.45, p < 0.001)	(z = 2.05, p =	(z = 6.03, p <	
$R^2 = 0.02)$	0.001)		0.04)	0.001)	
Trophic level	3.16 ± 0.04	3.09 ± 0.06	3.03 ± 0.06	3.16 ± 0.05	
(n = 779, R ² = 0.01)	(t = 82.26, p	(t = -1.23, p = 0.22)	(t = -2.39, p =	(t = 0.18, p =	
	<0.001)		0.02)	0.86)	
Longevity	6.84 ± 1.05	10.63 ± 1.07	8.38 ± 1.06	10.71 ± 1.05	
(n = 545, McFadden's	(z = 41.46, p <	(z = 6.34, p < 0.001)	(z = 3.53, p <	(z = 9.11, p <	
$R^2 = 0.02)$	0.001)		0.001)	0.001)	
Preferred	6.98 ± 1.06	7.57 ± 1.09	7.9 ± 1.08	12.2 ± 1.06	
Temperature range (n	(z = 33.65, p <	(z = 0.91, p = 0.36)	(z = 1.63, p = 0.1)	(z = 9.18, p <	
= 342, McFadden's R ² =	0.001)			0.001)	
0.06)					

Climate Region: Tropie (N = 974)	cal			
Trait	Impacted (Reference group)	Critically endangered	Endangered	Invasive
Generation time (n = 519, McFadden's $R^2 = 0.02$)	2.49 ± 1.08 (z = 12.49, p < 0.001)	2.21 ± 1.11 (z = -1.15, p = 0.25)	2.54 ± 1.1 (z = 0.19, p = 0.85)	3.35 ± 1.08 (z = 3.65, p < 0.001)
Trophic level (n = 935, R ² = 0.01)	3.09 ± 0.03 (t = 99.95, p < 0.001)	3.21 ± 0.05 (t = 2.29, p = 0.02)	3.02 ± 0.05 (t = -1.6, p = 0.11)	3.11 ± 0.04 (t = 0.38, p = 0.71)
Longevity (n = 531, McFadden's R ² = 0.03)	7.16 ± 1.04 (z = 45.57, p < 0.001)	5.93 ± 1.07 (z = -2.96, p = 0.003)	6.71 ± 1.06 (z = -1.18, p = 0.24)	9.52 ± 1.05 (z = 5.98, p < 0.001)
Preferred Temperature range (n = 423, McFadden's R ² = 0.06)	5.26 ± 1.04 (z = 38.66, p < 0.001)	4.79 ± 1.11 (z = -0.91, p = 0.36)	4.55 ± 1.08 (z = -1.86, p = 0.06)	8.38 ± 1.05 (z = 9.63, p < 0.001)

ITOFF Supplementary R code. R code for basic analysis of the ITOFF database. This script will be provided in unison with the ITOFF database following publication of the ITOFF data paper.

This code is for exploration, preliminary analyses, and visualisation of the ITOFF core database ### This code should be used as a guide and adapted to your needs

First you will need to download the ITOFF core dataset. Data will be available at: ### https://itoff-dataset.wp.st-andrews.ac.uk/ and via the Zenodo public repository (in .csv format)

Load packages: library(dplyr) library(ggplot2) library(ggpubr) ### You may need to install these packages using install.packages("...") ### if you have not done so previously

Load the database using the file pathway as saved on your device: data <- read.csv("...", header=T, sep=",")</pre>

Ensuring the dataset has loaded correctly head(data) str(data)

Specifying factor and continuous variables
data\$Polar <-as.factor(data\$Polar)
data\$Boreal <-as.factor(data\$Boreal)
data\$Temperate <-as.factor(data\$Temperate)
data\$SubTropical <-as.factor(data\$SubTropical)
data\$Tropical <-as.factor(data\$Tropical)
data\$ITOFF.Group <-as.factor(data\$ITOFF.Group)
data\$System <-as.factor(data\$System)
data\$Status <- as.factor(data\$Status)</pre>

data\$IUCN_Status <-as.factor(data\$IUCN_Status)</pre> data\$ReproMode<-as.factor(data\$ReproMode) data\$ReproductiveGuild <-as.factor(data\$ReproductiveGuild) data\$RepGuild1<-as.factor(data\$RepGuild1)</pre> data\$RepGuild2<-as.factor(data\$RepGuild2)</pre> data\$MeanFecundity <- as.numeric(data\$MeanFecundity)</pre> data\$MinFecundity <- as.numeric(data\$MinFecundity)</pre> data\$MaxFecundity <- as.numeric(data\$MaxFecundity)</pre> data\$BroodSize <- as.numeric(data\$BroodSize)</pre> data\$AgeMaturity Years Estimate<- as.numeric(data\$AgeMaturity Years Estimate) data\$MeanAgeMaturity_Years<- as.numeric(data\$MeanAgeMaturity_Years) data\$GenerationTime Years Estimate<- as.numeric(data\$GenerationTime Years Estimate) data\$TrophicLevel_Estimate<- as.numeric(data\$TrophicLevel_Estimate) data\$TrophicLevel<- as.numeric(data\$TrophicLevel)</pre> data\$FeedingType <-as.factor(data\$FeedingType)</pre> data\$Relative BrainWeight<- as.numeric(data\$Relative BrainWeight) data\$Longevity Years Estimate<- as.numeric(data\$Longevity Years Estimate) data\$Tmax<- as.numeric(data\$Tmax)</pre> data\$Wmax<- as.numeric(data\$Wmax)</pre> data\$Lmax<- as.numeric(data\$Lmax)</pre> data\$TempPref_min<- as.numeric(data\$TempPref_min)</pre> data\$TempPref_max<- as.numeric(data\$TempPref_max)</pre> data\$TempRange<- as.numeric(data\$TempRange) data\$TolerableTemp_min<- as.numeric(data\$TolerableTemp_min)</pre> data\$TolerableTemp_max<- as.numeric(data\$TolerableTemp_max)</pre> data\$Temp_Range_tolerable<- as.numeric(data\$Temp_Range_tolerable)</pre>

###------

To see all variables currently included in the database: names(data)

To see the pool of species:

species <- unique(data\$Species)</pre>

species

###-----

Extracting data - To create a new, derivative dataframe you can subset

the data by categorical variables. For example, to create a dataframe comprised of only temperate

species:

temperate <- subset(data, Temperate == "1")</pre>

head(temperate)

To export this dataframe as a new csv file [ADD YOUR OWN CUSTOM PATH] write.csv(temperate, "ITOFF_Temperate_species.csv")

###-----

Data visualisation. Example 1.

Plot of log(Absolute Fecundity) for groups of invasive, endangered, and impacted species

summary(data\$Status)

Creating a new variable for log(Absolute Fecundity)

data\$logMeanFecundity<- log(data\$MeanFecundity)

str(data)

Creating a dataframe to plot from

invasive <- subset(data, Status=="Invasive")</pre>

EN <- subset(data, Status=="EN")

CR <- subset(data, Status=="CR")

LC <- subset(data, Status=="LC")

plottingdata<-rbind(invasive,EN,CR,LC)

For simplicity we will focus only on species found in SubTropical, Temperate
and Tropical climate zones:
plottingdata_temp <- subset(plottingdata , Temperate == "1")</pre>

plottingdata_subtrop <- subset(plottingdata , SubTropical == "1")
plottingdata_trop <- subset(plottingdata , Tropical == "1")</pre>

plottingdata <- rbind(plottingdata_temp, plottingdata_subtrop, plottingdata_trop)</pre>

Creating the plot

```
FecundityPlot <- ggboxplot(plottingdata, x = "Status", y = "logMeanFecundity",
```

```
color = "Status", palette = "viridis",
```

```
ylab = "log(Mean Fecundity)", xlab = "",
```

add = "jitter",

order = c("LC", "EN", "CR", "Invasive")) +

scale_x_discrete(labels=c('Impacted', 'Endangered',

```
'Critically Endangered', 'Invasive'))
```

FecundityPlot

Example 2. Plot of Generation Time (years) for groups of non-native, endangered, ### and impacted species

```
### Creating the plot
```

order = c("LC", "EN", "CR","Invasive")) +

scale_x_discrete(labels=c('Impacted', 'Endangered',

```
'Critically Endangered', 'Invasive'))
```

GenerationTimePlot

To show these plots simultaneously:

ggarrange(FecundityPlot, GenerationTimePlot,

```
ncol = 1, nrow = 2, common.legend=TRUE)
```

To also show differences by climate region: Subsetting by climate region
First need to create a new dataset with Climate Region in long format:
temper <- subset(plottingdata,Temperate=="1")
subtrop <- subset(plottingdata,SubTropical=="1")
trop <- subset(plottingdata,Tropical=="1")</pre>

temper\$ClimateRegion<- "Temperate" subtrop\$ClimateRegion<- "SubTropical" trop\$ClimateRegion<- "Tropical"

Merge

newplottingdata <- rbind(temper,subtrop,trop)</pre>

str(newplottingdata)

Plot for absolute fecundity as shown for each climate region

FecundityPlot2<- ggboxplot(newplottingdata, x = "Status", y = "logMeanFecundity",

color = "Status", palette = "viridis",

ylab = "log(Mean Fecundity)", xlab = "",

add = "jitter",

order = c("LC", "EN", "CR","Invasive")) +

scale_x_discrete(labels=c('Impacted', 'Endangered',

'Critically Endangered', 'Invasive'))

FecundityPlot2

FecundityPlot2 <- facet(FecundityPlot2, facet.by="ClimateRegion",ncol=3)

FecundityPlot2

###-----

CLUSTER AND TAXONOMIC ANALYSES

Load required packages [NOTE THESE MAY HAVE TO BE INSTALLED FIRST]:

library(ape) library(Rcpp) library(ggtree)

library(cluster)

Ensure variables are specified as factors: data\$Class<-as.factor(data\$Class) data\$Order<-as.factor(data\$Order) data\$Family<-as.factor(data\$Family) data\$Species<-as.factor(data\$Species) data\$Genus<-as.factor(data\$Genus) data\$Phylum<-as.factor(data\$Phylum) data\$Kingdom<-as.factor(data\$Kingdom)</pre>

HEIRARCHICAL CLUSTER ANALYSIS

We will base our cluster analysis on three traits: age of sexual maturity,

trophic level, and preferred temperature range

Creatinga a dataframe for age of sexual maturity, trophic level, and temperature range

data2 <- data.frame(data\$Species,data\$Kingdom,data\$Phylum,data\$Class,

data\$Order,data\$Family,data\$Genus,data\$System, data\$ITOFF.Group,data\$Status,data\$AgeMaturity_Years, data\$TrophicLevel,data\$TempRange)

str(data2) data2 <- na.omit(data2) str(data2)

names(data2)[names(data2) == "data.Species"] <- "Species"
names(data2)[names(data2) == "data.AgeMaturity_Years"] <- "AgeMaturity_Years"
names(data2)[names(data2) == "data.TrophicLevel"] <- "TrophicLevel"
names(data2)[names(data2) == "data.TempRange"] <- "TempRange"</pre>

names(data2)[names(data2) == "data.Kingdom"] <- "Kingdom" names(data2)[names(data2) == "data.Phylum"] <- "Phylum" names(data2)[names(data2) == "data.Class"] <- "Class" names(data2)[names(data2) == "data.Order"] <- "Order" names(data2)[names(data2) == "data.Family"] <- "Family" names(data2)[names(data2) == "data.Genus"] <- "Family" names(data2)[names(data2) == "data.System"] <- "System" names(data2)[names(data2) == "data.ITOFF.Group"] <- "ITOFF.Group" names(data2)[names(data2) == "data.Status"] <- "Status"

From this dataframe we will select a random subset of 90 species ### that we will perform our heirarchical cluster analysis on

Create list of species
ids <- unique(data2\$Species)
Randomly draw IDs from the species pool
draw <- ids %>% sample(90)

Create dataframe containing only the randomly selected species
DendrogramData <- data2[data2\$Species %in% draw,]</pre>

Here we will create a separate dataframes that only contains variables
of relevance to the cluster analysis

dropa <- c("Kingdom","Phylum","Class","Order","Family","Genus","System","ITOFF.Group", "Status") clustdata <- DendrogramData[,!(names(DendrogramData) %in% dropa)] str(clustdata)

We will also reformat our data so that it is appropriate for the ### requirements of the cluster package

rownames(clustdata) <- clustdata\$Species head(clustdata) rownames(clustdata) dropb <- c("Species") clustdata = clustdata[,!(names(clustdata) %in% dropb)] head(clustdata)

Perform divisive clustering using DIANA (DIvisive ANAlysis clustering algorithm)
Divisive clustering based on: Age at maturity, Trophic level, and Temp range
clustdata <- scale(clustdata)
HC <- diana(clustdata)
HC</pre>

ILLUSTRATING

ClustTree <- ggtree(HC,layout="circular", branch.length='none') + geom_tiplab() ClustTree

We can also construct a TAXONOMIC TREE using the same species pool

Specifying the variables on which we will base our tree frm <- ~Kingdom/Phylum/Class/Order/Family/Genus/Species

Constructing the tree using the 'DendrogramData' dataframe
phylotree <- as.phylo(frm, data = DendrogramData, collapse=FALSE)
phylotree\$edge.length <- rep(1, nrow(phylotree\$edge))
phylotree</pre>

Plotting the taxonomic tree
phylotree2 <- ggtree(phylotree,layout="circular") + geom_tiplab()
phylotree2</pre>

Please see https://guangchuangyu.github.io/ggtree-book/chapter-ggtree.html
For details on how to customise the aesthetics of dendrograms using
the ggtree package