Impact of the 2004 tsunami on the Vulnerable Nicobar megapode *Megapodius nicobariensis*

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Abstract The small, isolated populations of the Nicobar megapode Megapodius nicobariensis, currently categorized as Vulnerable on the IUCN Red List, on the Nicobar islands in the Indian Ocean have declined by c. 70% since 1994. The 2004 tsunami is believed to be the major cause of this decline. The populations of megapodes on the islands of Megapode and Trax were wiped out by the tsunami. I estimate from a 2006 survey that 395-790 breeding pairs of the Nicobar megapode now survive on the coasts of the various islands compared to 2,318-4,056 pairs in 1994. The tsunami also adversely influenced nest-site selection and mound-nest ecology: > 50% of mounds were found closer to the shore in 2006 than in 1994, probably because of loss of suitable habitat, and they may become inundated with seawater during high tides. Most of the mound-nests found were constructed after the tsunami and were significantly smaller in volume than those present in 1994. Restoration of suitable habitat is critical for the long-term viability of the Nicobar megapode. However, post-tsunami impacts such as the creation of large-scale plantations in coastal areas, which are encroaching upon megapode habitat, may be severe. Because of the decline in numbers and habitat destruction and hunting the Nicobar megapode probably now qualifies for categorization as Endangered on the IUCN Red List.

Keywords Coastal biodiversity, habitat selection, isolated population, *Megapodius nicobariensis*, Nicobar islands, Nicobar megapode

Introduction

I solated populations of rare species may be disproportionately affected by both stochastic and deterministic factors such as hurricanes, fires, changes in habitat quality, inbreeding and genetic drift (Soulé, 1987; Burgman et al., 1993). To conserve such species there is a need to assess the impacts of such factors on the long-term survival of individual populations (Karen, 1998). An earthquake with a magnitude of 9.15, with its epicentre at 3.29° N and 95.94° E off the coast of Sumatra and with a focal depth of 30 km, occurred on 26 December 2004 at 06.29. It triggered a tsunami that, amongst other effects, engulfed the entire

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low-lying coastal forests of the Nicobar islands in the Bay of Bengal. The waters, which took several days to recede, devastated the people and wildlife of the Nicobar islands (Sankaran, 2005). This was the first recorded major natural catastrophe for the islands.

One species expected to have been adversely affected by the tsunami is the Nicobar megapode Megapodius nicobariensis. The megapodes utilize external sources of heat to incubate their eggs (Jones et al., 1995). The Nicobar megapode, a mound nesting bird, is endemic to the Nicobar Islands, separated from its nearest congener by a distance of > 1,500 km (Lister, 1911; Olson, 1980). There are two subspecies: M. n. nicobariensis on the Nancowry group of Islands north of the Sombrero channel and *M. n. abbotti* on the Great Nicobar group of islands south of the channel (Hume & Marshall, 1878; Abdulali, 1964; Ali & Ripley, 1983; Fig. 1). The species builds three types of mounds: A, built in open areas away from trees; B, built against a living tree base; C, built on or around a dead log or stump (Dekker, 1992; Sankaran, 1995b; Sankaran & Sivakumar, 1999; Sivakumar, 2000) and the sources of heat that create suitable incubation conditions vary with mound type, location and dimensions (Jones, 1987; Bowman et al., 1994; Palmer et al., 2000; Sinclair, 2002; Sinclair et al., 2002; Sivakumar & Sankaran, 2003).

Historically the Nicobar megapode occurred on all Nicobar Islands (Hume, 1874; Kloss, 1903; Dekker, 1992; Sankaran, 1995a,b) except Car Nicobar (Butler, 1899), Chaura (Abdulali, 1967) and Bati Malv (Sankaran, 1995a). There are a few records of megapodes from the Andaman group of islands (Hume, 1874; Butler, 1899; Sewell, 1922) and from the Coco Islands further north (Kloss, 1903; Abdulali, 1964). None of the records from the Andaman group are of recent origin and the species is now believed to be absent there (Sankaran, 1995a,b). In 1994 the population of M. n. abbotti was estimated to be 3,400-6,000 and the number of active mounds 849 (Sankaran, 1995a). The population of breeding adults of M. n. nicobariensis was estimated to be 1,200-2,100 and the number of active mounds c. 300 (Sankaran, 1995a). Habitat loss, hunting and egg collection are considered the major threats to the megapodes (Jones et al., 1995; Sivakumar & Sankaran, 2003).

The 1994 survey of the Nicobar megapode (Sankaran, 1995b) was the only one made before the tsunami. I therefore took this as a baseline for comparison with the status and distribution of the megapode in 2006, one nesting season after the tsunami. Post-tsunami restoration, especially of coastal habitat, is still required in the Nicobar

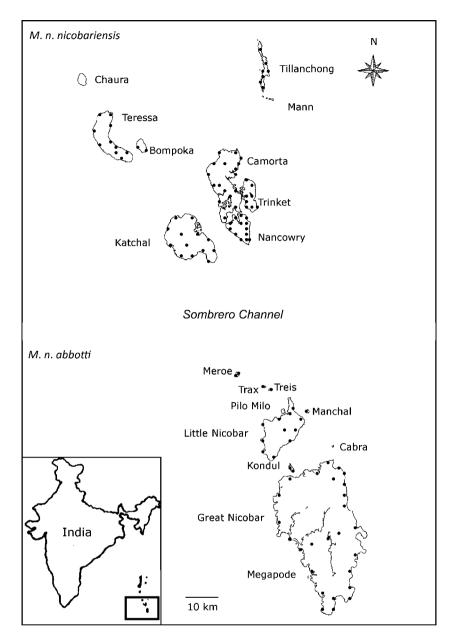


Fig. 1 Megapodius nicobariensis occurs as two subspecies: M. n. nicobariensis on the Nancowry island group north of the Sombrero channel and M. n. abbotti on the Great Nicobar island group south of the channel. The locations of the transects on each island are indicated as black circles. The inset indicates the location of the islands in the Bay of Bengal.

islands. However, any ill-planned restoration could adversely affect the coastal-living megapodes. I therefore studied the post-tsunami conservation status of the coastal populations of the Nicobar megapode to provide information to support appropriate conservation action.

Study area

The Nicobar Islands can be subdivided into three subgroups based on ornithological affinities (Sankaran, 1995a). To the south lies the Great Nicobar group consisting of two islands > 100 km², nine islets > 5 km² in area, and a few rocks. Among them, Great Nicobar, Little Nicobar, Kondul and Pilo Milo are inhabited, and Meroe, Treis, Trax, Menchal, Megapode, Cabra and Pigeon are uninhabited. Fifty-eight km north of the Great Nicobar group is the

Nancowry group (Fig. 1) comprising three islands >100 km², two of 36 and 67 km², three < 17 km², two small islets and a few rocks. Except islets, all other islands of the Nancowry group are inhabited. The northernmost subgroup comprises Batti Malv and Car Nicobar, which are 88 km north of the Nancowry group. Batti Malv is uninhabited and Car Nicobar has a population of > 19,000 people (Sankaran, 1995b).

Two groups of indigenous people inhabit the Nicobar Islands, along with settlers from the Indian mainland. The Shompen, who now number < 150, are a semi-nomadic hunter-gather tribe who inhabit the forests of the central uplands. The Nicobarese, who have several settlements along the coast, constitute the largest tribal group of > 27,000 people. The entire population of coastal-living Nicobarese, including settlers, were adversely affected by

the tsunami and these people still need to be provided with new housing and, in most cases, alternative livelihoods.

Methods

This study was carried out during February–May 2006. Because the Nicobar megapode is shy and cryptic and therefore difficult to see in the forest whereas nesting mounds are stationary, inanimate and represent breeding signs, the easiest way to estimate and monitor a megapode population is by counting the number of active mounds (Dekker, 1992; Sankaran, 1995b; Sivakumar & Sankaran, 2003). Surveys conducted in 1994 (Sankaran, 1995b) also utilized mound counts, thus facilitating a pre-tsunami comparison with the data collected in 2006.

The coastlines of the 15 islands on which the species was reported in 1994 were surveyed for mounds using a standardized survey protocol (Sankaran, 1995b). To estimate the total number of active mounds the coastline of each island was divided into suitable and unsuitable coastal habitat for mound building. Coastal habitat suitable for mound building has a sandy-loam substratum and littoral forests (Sivakumar, 2000). The extent of these two coastal habitat types was measured using satellite images (from 2006) and vegetation maps (Sankaran, 2005).

Variable width transects of 10-600 m were used to count all the mounds present within a surveyed area (the low lying coastal littoral forests are of varying width). Transect length and distance between transects was determined by island size but was uniform for an island. The mean length of a transect was 2 km but in some islands, the entire coast was surveyed because the total coastline was < 2 km. The census was carried out with seven observers walking at 20-m intervals parallel to the shore; for transects > 140 m wide we walked the transect more than once to cover the entire width. The interior forests of Great Nicobar, Little Nicobar, Camorta, Katchal and Teressa islands were surveyed, by seven observers, with 1-km transects of 140 m width and 1 km long. The total number of active and abandoned mounds, mound size, green canopy cover over mound, and the distance between high tide mark and mounds were recorded.

Active mounds were identified by signs of recent digging by megapodes or by whether the soil was compact and hard, with vegetation growth (i.e. an abandoned mound) or loose and easily penetrable with a stick (i.e. an active mound; Sankaran, 1995b).

Because the distribution of mounds was not uniform (Sankaran, 1995b) the suitable and unsuitable segments of coastal areas were surveyed separately. A total of 328 km of coastal habitat was identified as suitable, of which 157.5 km were sampled in 80 transects (68 of 2 km, 10 < 2 km, and two > 2 km). Of the 358.8 km of unsuitable coastal habitat, 77.9 km were surveyed in 39 transects (29 of 2 km, five

> 2 km and five < 2 km). On most islands mound distribution was not uniform and therefore mound abundance, D, on an island was estimated using the following formula:

$$D = \left(\frac{N}{S_a}\right) H_a,$$

where N = total number of mounds found in S_a , a = habitat type (suitable or unsuitable), S = total length of coastal area sampled in habitat a and H = total length of coastal area available for habitat a. Standard error of the mean number of mounds present per km of coast of an island was estimated when there were > 2 transects on an island (Table 2).

Up to 20% of the megapodes are estimated to live in the interior forests of the islands (Dekker, 1992), which comprise 72% of the total area of the islands. Because of the difficulty in working in such forests survey intensity was lower, with 11 transects on Great Nicobar, four each on Little Nicobar and Camorta, three on Katchal, and two each on Teressa and Nancowry islands.

Basal circumference, height and diameter of mounds were measured. As mounds are uneven in shape and have a cone-like appearance, volume was calculated approximately using the equation for the volume of a cone: $1/3\pi r^2 h$, where r is the radius and h the height (Sivakumar & Sankaran, 2003).

More than one pair of megapodes may use a mound. Sankaran (1995a) used two pairs per mound for the lower limit and 3.5 pairs as the upper limit but Sivakumar (2000) observed that a mean of 2.5 pairs used a mound. Most of the mounds that I observed were < 1 year old (based on their size and the fact that most mounds were washed away by the tsunami) and too small to accommodate more than two pairs (Sivakumar & Sankaran, 2003). Thus, for the population estimate I used one and two pairs as the lower and upper limits, respectively.

Results

In 2006 I found the Nicobar megapode on all but two islands (Trax and Megapode) from where it had been reported in 1994 (Sankaran, 1995a). Comparing the distribution of mounds from the shore inland there were significant differences between 1994 (Sankaran, 1995a) and 2006 (Wilcoxon signed ranks test, Z = -2.934, P < 0.005; Fig. 2). In 2006 > 90% of mound nests were within 30 m of the shore and, of these, c. 16% of mounds were within 5 m of the shore.

I found significantly fewer active mound nests than reported in 1994 (paired sample t-test, t = 2.061, df = 14, P < 0.05). Based on an estimated total of 394 active mounds (Table 1), I estimate that 394–788 breeding pairs of Nicobar

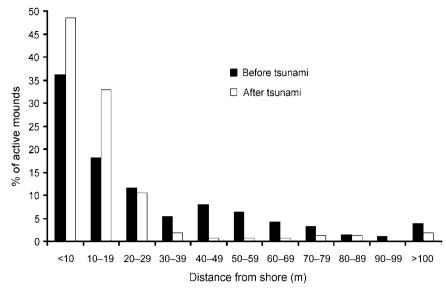


Fig. 2 Distance of active Nicobar megapode mound nests from the shore in 1994 (before the tsunami) and in 2006 (after the tsunami).

megapode now occur on the coastal habitat of the islands, which is \leq 70% of the numbers reported in 1994.

Suitable coastal habitat for *M. n. nicobariensis* on the Nancowry island group (Fig. 1) shrank after the tsunami and only 37% of the coastal habitat is now available for

mound building, with an estimated 97 active mounds and thus 97–194 breeding pairs. I did not find any active nest mounds in unsuitable coastal habitat (coconut plantations, mangroves, human habitation and cliffs). The islands of Tillanchang and Trinket, despite their relatively small size,

Table 1 Summary of the survey of the two subspecies of Nicobar megapode *Megapodius nicobariensis* carried out on 14 islands (Fig. 1) in 2006 in suitable and unsuitable habitat for mound building (see text for definitions), with total length of the coast of each island, length of coast surveyed and number of transects used, number of active mounds found, estimated total number of active mounds (extrapolated to the whole coast, see text for explanation), and estimated number of active mounds per km (with SE; suitable habitat only).

	Suitable coastal habitat						Unsuitable coastal habitat			
Island	Length (km)	Length surveyed, km (no. of transects)	Active mounds	Estimated total active mounds	Active mounds ± SE (km ⁻¹)	Length (km)	Length surveyed, km (no. of transects)	Active mounds	Estimated total active mounds	
M. n. abbotti (Great Nicobar island group)										
Great Nicobar		42.5 (20)	64	195.8	1.46 ± 0.24	83	12 (6)	1	7	
Kondul	1	1 (1)	1	1.0		6.5	2 (1)	0	0	
Little Nicobar	55	17.5 (9)	25	78.6	1.46 ± 0.39	23	6 (2)	1	4	
Megapode	0	0	0	0		0	0	0		
Menchal	1	0.5 (1)	3	6.0		2.3	1 (1)	0	0	
Meroe	2	2 (1)	2	2.0		3.3	1 (1)	0	0	
Pilo Milo	0	1.5 (1)	0	0		3	3 (1)	0	0	
Trax	0.1	0.1 (1)	0	0		1.2	1.2 (1)	0	0	
Treis	2	2 (1)	3	3.0		0.7	0.7 (1)	0	0	
Subtotal	191	67 (35)	98	286	155 ± 0.67	123	27 (14)	2	11	
M. n. nicobariensis (Nancowry island group)										
Nancowry	17	16 (8)	7	7.4	0.44 ± 0.20	27.3	10 (5)	0	0	
Katchal	30	14 (7)	4	8.6	0.29 ± 0.18	48	12 (6)	0	0	
Camorta	35	21 (11)	4	6.7	0.23 ± 0.10	77.5	12 (6)	0	0	
Tillanchang	15	9 (5)	16	26.7	1.80 ± 0.12	27	6 (2)	0	0	
Trinket	15	11.5 (6)	20	26.1	1.75 ± 0.51	15	4 (2)	0	0	
Teressa	20	16 (7)	7	8.8	0.46 ± 0.22	33.3	6 (3)	0	0	
Bampoka	5	2 (1)	5	12.5		7.8	1 (1)	0	0	
Subtotal	137	90.5 (45)	63	96.8	1.05 ± 0.35	236	51 (25)	0	0	
Total	328	157.5 (80)	161	383	1.32 ± 0.39	358.8	77.9 (39)	2	11	

Table 2 Estimated total number of active mounds and breeding pairs of the two subspecies of Nicobar megapode in 1994 (Sankaran, 1995b) and 2006 (this study; number of active mounds is a sum of the estimated total number of active mounds in suitable and unsuitable habitat from Table 1).

Island	Estimated no. of active mounds in 2006	Estimated no. of active mounds in 1994	Estimated no. of breeding pairs in 2006 (range)	Estimated no. of breeding pairs in 1994 (range)
M. n. abbotti (G	reat Nicobar island group)			
Great Nicobar	203	515	203-406	1,030-1,803
Kondul	1	11	1–2	22-39
Little Nicobar	82	311	82-164	622-1,089
Megapode	0	2	0	4–7
Menchal	6	2	6–12	4–7
Meroe	2	1	2–4	2–4
Pilo Milo	0	0	0	0
Trax	0	3	0	6–11
Treis	3	4	3–6	8-14
Subtotal	297	849	297-594	1,698-2,972
M. n. nicobarien	isis (Nancowry island group	o)		
Nancowry	7	60	7–14	120-210
Katchal	9	69	9–18	138-242
Camorta	7	20	7–14	40-70
Tillanchang	27	10	27-54	20-35
Trinket	26	8	26-56	16-28
Teressa	9	119	9–18	238-417
Bampoka	13	26	13-26	52-91
Subtotal	98	312	98-196	624-1,092
Total	394	1,161	395–790	2,322-4,064

now hold nearly 50% of the population of this subspecies. The megapode populations on the larger islands of Camorta, Katchal, Teressa and Nancowry was estimated to be c. 63 breeding pairs, 88% less than estimated in 1994 (Table 2).

M. n. abbotti occurs on all of the islands of the southern Great Nicobar group (Fig. 1) except for Pilo Milo, Megapode and Trax where the megapode now appears to be extinct (megapodes were not found on Pilo Milo in 1994). Of the 314-km coast 61% contains suitable habitat, and I found a total of 286 active mounds. In unsuitable

coastal habitat I found 11 active mounds. I estimate that 297–594 breeding pairs occur on the coastal habitat of these islands. I found the largest population of this subspecies on Great Nicobar, with an estimated 203–406 breeding pairs. The second largest population is on Little Nicobar. These two islands are the largest in this group and provide breeding habitats for 96% of the total population of the subspecies.

There was a significance difference in the composition of mound types between 1994 and 2006 ($\chi^2 = 35.4$, df = 2,

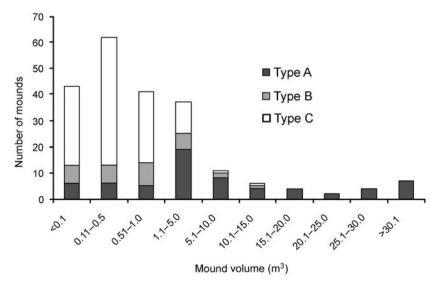


Fig. 3 Frequency distribution of mound nest volume of the three mound types (see text for details) of the Nicobar megapode.

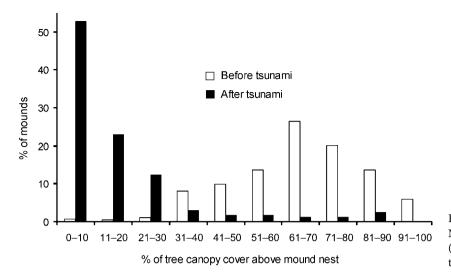


Fig. 4 Percentage of tree canopy cover above Nicobar megapode mound nests in 1994 (before the tsunami) and in 2006 (after the tsunami).

P < 0.001). In 2006 and 1994, respectively, 11 and 21% were Type A, 17 and 41% Type B, and 72 and 37% Type C. Mounds varied in volume (0.01–71.45 m³). Of the 217 observed mounds, the majority (84%) were < 5 m³ and 67% were < 1 m³. Less than 6% of mounds were > 20 m³ and all of these were type A. Because most of the mounds were new and constructed after the tsunami the mean volume (3.78 \pm SE 0.62 m³) was smaller than in 1994 (6.34 \pm SE 0.16 m³; Wilcoxon signed ranks test, Z = -2.497, P < 0.05). Type A mounds were generally of a greater range of volumes than Types B and C, which were generally small in size (Fig. 3). Green canopy cover above mounds was significantly less in 2006 than in 1994 (Wilcoxon signed ranks test, Z = -2.090, P < 0.05; Fig. 4).

Of the 26 transects surveyed in the interior of the islands, mounds were found only on three transects, one on the southern tip of Great Nicobar, and two mounds from two different locations on Camorta. These mounds were not included in our population estimates.

Discussion

The major reason for the decline of the Nicobar megapode since 1994 appears to be the tsunami, which washed away both suitable habitat and nest mounds. In addition, the tsunami appears to have been the cause of the extirpation of the megapode on the islands of Megapode and Trax. Megapode was fully submerged for 3 weeks after the tsunami and the coastal areas of Trax were fully inundated. The Nicobar megapode appears preferentially to build mound nests in coastal areas because of the sandy-loam substratum there (Sivakumar, 2000; Sivakumar & Sankaran, 2003). Compared to the previous survey (Sankaran, 1995b) the mounds were closer to the shore, possibly because the tsunami reduced the extent of suitable coastal habitat. These mounds, which were built near to the high tide mark, could be affected by abnormally high tides or waves. Maintaining

mound temperature at a constant rate is important for successful egg incubation (Sivakumar & Sankaran, 2003), and any influence of the nearby sea water on incubation temperature could adversely affect the hatching success of these mounds and lead to further population declines. I observed that human threats to the megapode, such as habitat degradation, hunting and egg collection, continue in the Nicobar islands.

I found the largest population of *M. n. nicobariensis* on Tillanchang despite its smaller size, and the population there is approximately the same size as in 1994. Tillanchang is the only Wildlife Sanctuary within the range of this subspecies, although I observed evidence of megapode hunting there. Because the intensity of the tsunami waves was diminished by the islands of Camorta, Katchal and Chaura before reaching Tillanchang, the impact of the tsunami was less on this island. However, the tsunami caused a drastic change in the number of mounds on all other islands (Table 2). As the earthquake occurred south of the Great Nicobar island group the tsunami waves arrived there first, and this may explain the 65% decrease in the population of *M. n. abbotti* on this island group compared to 1994.

The sources of heat that create suitable incubation conditions within a mound nest vary with mound type, location and size (Bowman et al., 1994; Palmer et al., 2000; Sinclair, 2002; Sinclair et al., 2002; Sivakumar & Sankaran, 2003). Type A mounds are larger in size and normally used by > 3 pairs (Sivakumar & Sankaran, 2003) but the number of such mounds was greatly reduced by the tsunami. I saw a greater number of type C mounds, probably because of the availability of dead trees after the tsunami; type C mounds may become type A once the tree decomposes fully (Sivakumar & Sankaran, 2003). Types B and C are normally smaller in size and accommodate fewer birds (Sivakumar & Sankaran, 2003). The active mounds were mostly small in volume (Fig. 3), indicating that they were constructed after the tsunami.

The temperature generated through fermentation of vegetative material inside a mound probably aids incubation (Sivakumar & Sankaran, 2003) although ambient temperature may also have an influence (Jones, 1987; Sinclair, 2002; Sinclair et al., 2002). The reduced green canopy cover above mounds in 2006 compared to 1994 means that more direct sunlight is now falling on the mounds, which may therefore warm up more quickly (Sinclair, 2002; Sinclair et al., 2002). If temperatures become too hot this could have a detrimental effect on egg development.

The Nicobar megapode is included in Schedule I of the Indian Wildlife (Protection) Act, 1972 and is categorized as Vulnerable on the IUCN Red List (BirdLife International, 2008). However, as the population has decreased by c. 70% since 1994 and habitat destruction and hunting are still adversely affecting the species it probably qualifies for categorization as Endangered based on criteria (IUCN, 2001) A1acd, i.e. a reduction in population size (A), with an estimated reduction of \geq 70% over the last 10 years (1) based on direct observation (a), an index of abundance (b) and a decline in extent of occurrence (c).

Island ecosystems are known for their resilience because of the possibilities for restoration by recolonization (Gunderson, 2000; Elmqvist et al., 2003). In this context, habitat restoration will be crucial for the long-term viability of the Nicobar megapode populations. There are currently two National Parks and two Wildlife Sanctuaries within the range of the megapode. Both National Parks are on Great Nicobar. Although these Parks offer some protection to M. n. abbotti, large extents of potential coastal habitat, especially along the west coast, are outside protected areas. The Wildlife Sanctuary of Megapode Island was completely submerged by the tsunami and no megapodes survived. Tillanchang Wildlife Sanctuary, which includes the whole of the island, is the only protected area that protects M. n. nicobariensis. Overall < 5% of the range of this subspecies is protected.

The tsunami washed away most of the planted as well as wild coastal coconut *Cocos nucifera* and areca nut *Areca catechu* palms and re-establishment of these plantations is important for the future survival of the islands' people. However, in the absence of appropriate planning the ongoing planting may encroach upon the coastal habitat of the Nicobar megapode.

Following the tsunami hunting pressure on the megapode has increased. Although the Nicobarese attach traditional cultural values to megapodes (the species is a symbol of love because of its monogamous breeding behaviour), scarcity of animal protein has forced them to hunt the species. A conservation education programme addressing this issue needs to be initiated. The two aboriginal tribes, the Nicobarese and Shompens, are exempt from the Indian Wildlife (Protection) Act, 1972. Considering the changing lifestyle of these two peoples

(Math et al., 2006) either this immunity needs to be reviewed or the tribes need to be involved in the conservation and management of both the megapode and the islands' protected areas.

The Andaman & Nicobar Forest Department has already adopted some of the recommendations of this study, such as inter-departmental coordination in the re-establishment of coconut plantations so as to avoid plantation in prime megapode habitat. In addition, the Department has agreed, in principle, to initiate a programme to eradicate introduced cats and dogs, known to predate the megapode (Sivakumar, 2000), from the coastal areas of the Nicobar islands. The Department has also initiated a long-term monitoring programme of the habitat of the megapode.

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