



FEATURE ARTICLE

# Temporal variation of the small eukaryotic community in two freshwater lakes: emphasis on zoosporic fungi

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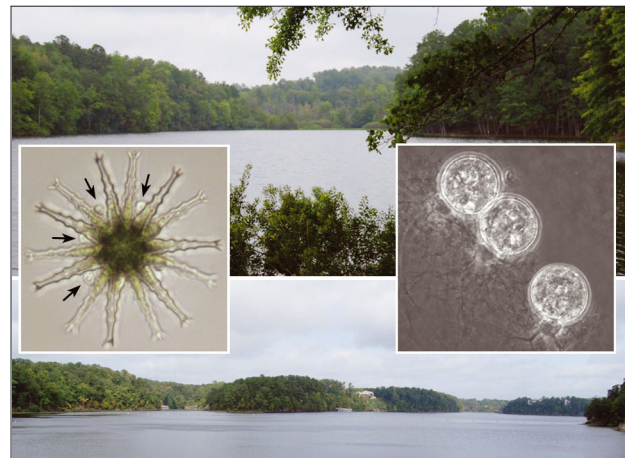
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**ABSTRACT:** Applications of molecular approaches to the study of microbial eukaryotic communities in freshwater lakes are transforming our understanding of these ecosystems. One of the most unexpected discoveries is that zoosporic fungi significantly dominate the planktonic fungal diversity. Although zoosporic fungi are now recognized as an important component of aquatic microbial food webs, our knowledge of their community structure and temporal variability remains poor. The objectives of our study were (1) to compare and describe the contribution of zoosporic fungi to the eukaryotic diversity in 2 lakes differing in their trophic status during the mixing and the stratified seasons and (2) to phylogenetically identify the recovered zoosporic fungal sequences. The small eukaryotes (0.6 to 8  $\mu\text{m}$ ) of the euphotic zone of the oligotrophic Lake Tuscaloosa and meso-eutrophic Lake Lurleen (Alabama, USA) were collected over 1 yr. Analyses of the 28S rDNA clone libraries showed that zoosporic fungi dominated the small planktonic fungal community and were more diverse in the meso-eutrophic lake and during the thermal stratification. Although the overall structure of the eukaryotic community was similar between the 2 lakes, at lower taxonomic levels, community composition differed. Analyses of the retrieved fungal sequences revealed that zoosporic fungi mostly affiliated with Rhizophydiales and Chytridiales or formed environmental clades. Although the phytoplanktonic community was also monitored, zoosporic fungal parasites were rarely observed on algae. These results provide new insights into the diversity and seasonality of the zoosporic fungal community in lake ecosystems.

**KEY WORDS:** Zoosporic fungi · Molecular diversity · Temporal variation · Freshwater lakes · Plankton

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The meso-eutrophic Lake Lurleen (top) and humic oligotrophic Lake Tuscaloosa (bottom) harbor a high diversity of planktonic zoosporic fungi, such as the saprobe *Rhizoclostridium aurantiacum* (right: young sporangia in culture) and the unidentified parasite (arrows) on the alga *Straustrum rotula* (left: individual in environmental sample).

Photos: MJ Powell, PM Letcher, E Lefèvre

## INTRODUCTION

Recent applications of molecular approaches to characterize eukaryotic communities have revealed that zoosporic fungi significantly contribute to the eukaryotic diversity in a range of ecosystems, such as high-elevation soils (Freeman et al. 2009), deep-sea hydrothermal ecosystems (Le Calvez et al. 2009, Nagahama et al. 2011), and streams (Nikolcheva & Bärlocher 2004, Seena et al. 2008). In freshwater lakes,

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where fungi in general have rarely been factored in as a component of the microbial food web (Wetzel 2001, Sigee 2005), a high diversity of zoosporic fungal sequences has been unexpectedly recovered in recent studies (Lefranc et al. 2005, Slapeta et al. 2005, Lefèvre et al. 2007, 2008, Lepère et al. 2008, Chen et al. 2008, Luo et al. 2011, Monchy et al. 2011).

The typical life cycle of a zoosporic fungus begins with the attachment of a free-swimming uniflagellate zoospore to decaying or living organic substrates. The encysted zoospore develops into a mature thallus (sporangium typically with rhizoids) obtaining its energy from the substrate. When the thallus reaches maturity, the sporangium releases zoospores (typically 3 to 8  $\mu\text{m}$  diameter) into the environment (Fuller & Jaworski 1987, Powell 1993). Multiple roles for zoosporic fungi in lake ecosystems have been suggested, including (1) parasites affecting phytoplanktonic successions and impacting primary production, (2) saprobes playing an important role in decomposition of recalcitrant organic material such as chitin and cellulose, and (3) prey, via the consumption of their zoospores by predators, transferring energy from primary producers and detritus to higher-trophic-level organisms (reviewed by Gleason et al. 2008). However, data supporting these assumptions are relatively scarce (van Donk & Ringelberg 1983, Kagami & Urabe 2002, Kagami et al. 2004, 2007a, Rasconi et al. 2009), and consequently little is known about the ecological importance and dynamics of zoosporic fungi in freshwater lakes.

In contrast, phylogenetics has advanced in the past decade with an impressive amount of molecular and ultrastructural data assembled for zoosporic fungi sampled globally (Letcher & Powell 2005a,b, Letcher et al. 2005, 2006, 2008a,b, James et al. 2006a,b, Mozley-Standridge et al. 2009, Simmons et al. 2009, Wakefield et al. 2010, Vélez et al. 2011). These analyses have profoundly improved our understanding of the phylogenetic relationships and diversity of zoosporic fungi. However, the majority of sequences recovered from lakes have not matched any described zoosporic fungi and do not affiliate with any known zoosporic fungal taxa. These results suggest that lakes harbor a high and largely unexplored zoosporic fungal diversity that perhaps represents novel phylogenetic lineages.

In this context, the present study had 2 main objectives: (1) from an ecological standpoint, to describe the community structure of freshwater zoosporic fungi in 2 lakes differing in their trophic status and the dynamics of the recovered diversity during 2 contrasted seasons of the year; and (2) from a phyloge-

netic standpoint, to explore the diversity of freshwater zoosporic fungi and accurately place the recovered sequences within the zoosporic fungal phylogeny. To address these objectives, an rDNA environmental survey was conducted on the small planktonic fraction (0.6 to 8  $\mu\text{m}$ ). The small planktonic fraction was selected in order to relate the detected diversity to similar previous studies also conducted on the small fraction, which harbors a relatively high diversity of zoosporic fungi (Lefranc et al. 2005, Lefèvre et al. 2007, 2008, Lepère et al. 2008) (see Table 2). Because of the exploratory aspect of our study for undescribed zoosporic fungi, we chose to use universal eukaryotic primers instead of more specific fungal primers designed from mostly terrestrial and non-zoosporic fungi (White et al. 1990, Borneman & Hartin 2000). In addition, although the main goal of the present study focused on the zoosporic fungal community, the use of universal eukaryotic primers provided an overview of the small eukaryotic community, therefore giving us the opportunity to place the recovered zoosporic fungal diversity in the context of the whole eukaryotic community. Finally, to address our second objective, we chose to target the ribosomal large subunit (LSU, i.e. 28S) gene, which has been shown to be more phylogenetically informative for zoosporic fungi than the more conserved small subunit (SSU, i.e. 18S) gene or the more variable ITS region (Letcher et al. 2008a, b,c). In addition, to characterize the fungal parasitism potentially associated with algae, the composition and dynamics of the phytoplanktonic community were microscopically determined.

## MATERIALS AND METHODS

### Study sites and sampling

Lake Tuscaloosa (LT) and Lake Lurleen (LL) are 2 freshwater reservoirs situated in the Black Warrior River basin, Alabama, USA (Ward et al. 2005). LT has a surface area of 24  $\text{km}^2$  and a maximum depth of 30 m, while LL has a surface area of 1  $\text{km}^2$  and a maximum depth of 10 m. Samplings and measurements were taken every 1 to 2 wk from December 2007 to November 2008 (LT) and from October 2008 to November 2009 (LL) at a central point in each lake (LT: 33° 17' 24.61" N, 87° 30' 41.43" W; LL: 33° 17' 27.65" N, 87° 30' 40.51" W). Temperature and dissolved oxygen vertical profiles were obtained using a YSI550A multiparameter probe (YSI). To describe the phytoplanktonic community and characterize the

fungal parasitism potentially associated with this community, phytoplanktonic samples were collected in triplicate using a vertical plankton tow (60  $\mu\text{m}$  mesh size) from the bottom of the euphotic zone (approximated using a Secchi disc) to the surface. In the laboratory, phytoplanktonic samples were processed following the protocol described by Wetzel & Likens (1990). Cells were observed using a Nikon Labophot-2 microscope and identified using identification keys from Smith (1950) and Wehr & Sheath (2003). Biovolumes were calculated using equations provided by Hillebrand et al. (1999), and a ratio of cellular organic carbon to cell volume of 0.1 was used to estimate phytoplanktonic biomass (Wetzel & Likens 1990). Water samples for molecular analysis and chlorophyll *a* measurements were collected in triplicate every meter from the bottom of the euphotic zone (varying from 2 to 5 m and 1 to 3 m for LT and LL, respectively, see Fig. 1) to the subsurface using a 1 l Van Dorn bottle. For each replicate, water from the different depths sampled was pre-filtered through a 150  $\mu\text{m}$  mesh size nylon filter and pooled into an 8 l carboy. In the laboratory, water samples for chlorophyll *a* analysis were filtered through glass fiber filters, and chlorophyll *a* was extracted using the acetone extraction method (Wetzel & Likens 1990). Trophic state indices (Carlson 1977) were calculated based on Secchi depth and chlorophyll *a* concentrations. For molecular analysis, 200 ml of water from 1 replicate was pre-filtered through 8  $\mu\text{m}$  pore-size carbonate filters by gravity, and cells of interest (<8  $\mu\text{m}$ ) were collected on 0.6  $\mu\text{m}$  pore-size polycarbonate filters using an electric vacuum pump. Filters were stored at  $-80^\circ\text{C}$  until DNA extraction.

### Molecular analysis

For each lake, 1 replicate of the samples collected during the summer thermal stratification (LT: 7 samples collected from June to September 2008, LL: 8 samples collected from June to September 2009) and the winter mixing (LT: 8 samples collected from January to March 2008, LL: 5 samples collected from December 2008 to February 2009) were processed for DNA extraction. For the cell lysis, filters were incubated with 573  $\mu\text{l}$  of TE (Tris-EDTA: 1 mM EDTA, 10 mM Tris-HCl), Proteinase K and RNase A (both at 0.1 mg  $\text{ml}^{-1}$  final concentration), and 0.5% SDS at  $37^\circ\text{C}$ . After 1 h incubation, 75  $\mu\text{l}$  of chloroform was added, the mixture was briefly vortexed and centrifuged at  $11\,000 \times g$  for 2 min, and the DNA-containing aqueous phase was collected. The DNA

purification was performed using the NucleoSpin Plant kit (Macherey-Nagel). DNA was quantified using a Nanodrop ND-1000 (NanoDrop Technologies), and its integrity (i.e. unsheared high molecular weight DNA) was checked on a 1% agarose gel. For each sample, a partial 28S (~900 base pairs) fragment was amplified using LROR (5'-ACC CGC TGA ACT TAA GC-3') and LR5 (5'-TCC TGA GGG AAA CTT CG-3') eukaryotic primers (Vilgalys & Hester 1990). PCR reactions were performed in 50  $\mu\text{l}$  containing distilled  $\text{H}_2\text{O}$ , 2.5 to 5 ng of DNA, 200  $\mu\text{mol l}^{-1}$  of each dNTP, 0.2  $\mu\text{mol l}^{-1}$  of each primer, 2.5 U of *Taq* DNA polymerase, and 1X NEB ThermoPol Buffer (New England Biolabs). The PCR program consisted of 2 min at  $94^\circ\text{C}$ , followed by 30 cycles of 1 min at  $94^\circ\text{C}$ , 1 min at  $50^\circ\text{C}$ , and 1 min at  $72^\circ\text{C}$ , and a final extension of 10 min at  $72^\circ\text{C}$  (MJ Research PTC-200 thermocycler, Bio-Rad). For each lake, PCR products from the same season were pooled together and purified using the NucleoSpin Extract II kit (Macherey-Nagel). Four clone libraries (LT summer thermal stratification, LT winter mixing, LL summer thermal stratification, and LL winter mixing) were constructed using the TOPO-TA cloning kit (Invitrogen). For each library, 200 clones were randomly selected and amplified using the primers LROR and LR5. A restriction fragment length polymorphism (RFLP) was performed directly on the PCR products of the positive clones (i. e., containing a ~900 base pair DNA insert) using the enzymes *MspI* and *AluI* (New England Biolabs). Enzymatic restriction was performed overnight at  $37^\circ\text{C}$ , and profiles were checked on a 2.5% NuSieve 3:1 agarose (Lonza) gel. Clones presenting the same restriction profile were grouped as an operational taxonomic unit (OTU). One representative of each OTU was sequenced by Macrogen USA. The program Analytic Rarefaction (<http://strata.uga.edu/software/index.html>) was used to construct species accumulation curves (Fig. S1 in the supplement at [www.int-res.com/articles/suppl/a067p091\\_supp.pdf](http://www.int-res.com/articles/suppl/a067p091_supp.pdf)), and SPADE (<http://chao.stat.nthu.edu.tw/softwareCE.html>) was used to estimate the coverage of our samples (Lee & Chao 1994) (see Table 1).

### Phylogenetic analysis

A Blast search ([http://blast.ncbi.nlm.nih.gov/Blast.cgi?CMD=Web&PAGE\\_TYPE=BlastHome](http://blast.ncbi.nlm.nih.gov/Blast.cgi?CMD=Web&PAGE_TYPE=BlastHome)) was performed to affiliate the recovered sequences. Zoospore fungal sequences (Table S1 in the supplement) were aligned using ClustalX (Thompson et al. 1997) and BioEdit (Hall 1998). Maximum parsimony (MP)

and Maximum likelihood (ML) phylogenetic trees were constructed. MP trees were constructed using PAUPRat (Sikes & Lewis 2001); a majority rule consensus tree was constructed, and branch support was determined as described by Letcher et al. (2004a). From ModelTest (v.3.7) (Posada & Crandall, 1998), the Akaike information criterion was used to determine the most appropriate model of DNA substitution. ML analyses were performed in GARLI v.0.951 (Zwickl 2006) ([www.bio.utexas.edu/faculty/antisense/garli/Garli.html](http://www.bio.utexas.edu/faculty/antisense/garli/Garli.html)). The run was repeated 10 times from random starting trees using the auto-terminate setting to determine the best ML tree. Of the 10 trees produced, 3 were equally likely. GARLI was also used to generate 100 ML nonparametric bootstrap replicates from which a majority-rule consensus tree was calculated in PAUP\* 4.0b10 (Swofford 2002). Trees were rooted with the choanoflagellate *Mono-siga ovata*.

### Zoosporic fungal isolation

Samples for zoosporic fungal isolation were collected in both lakes in February 2008 on 1 occasion using a vertical plankton tow (60  $\mu\text{m}$  mesh size). The particulate organic material collected was baited with sweet gum (*Liquidambar* sp.) pollen, onion skin, and shrimp exoskeleton (Fuller & Jaworski 1987). Isolated species were identified based on microscopic observation of the thallus and identification keys (Sparrow 1960). Cultures were grown and DNA extracted as described by Letcher et al. (2004a). Partial 28S rDNA was amplified and sequenced with the LROR and LR5 primers.

### Sequences accession numbers

Nucleotide sequences obtained in the present study were deposited in GenBank under accession numbers JN049527 to JN049529 and JN049533 to JN049557.

## RESULTS

### Physicochemical and biological characteristics of the study sites

LT and LL exhibited the same pattern in regards to their seasonal stratification regime, with complete mixing of the water column occurring in winter–

spring and thermal stratification in summer–fall (Fig. 1). Temperature profiles (Fig. 1) showed that both lakes were monomictic and holomictic. Dissolved oxygen profiles and chlorophyll *a* measurements, however, showed remarkable differences between the 2 lakes. While LT displayed a metalimnetic oxygen minimum during summer stratification, LL displayed a metalimnetic oxygen maximum. Chlorophyll *a* concentrations in LL ( $9.33 \pm 0.24 \mu\text{g l}^{-1}$ ) annually averaged ~6-fold higher those in LT ( $1.55 \pm 0.24 \mu\text{g l}^{-1}$ ) (Fig. S2 in the supplement). The trophic state indices (Carlson 1977) (Fig. S3 in the supplement) demonstrated that LT was oligotrophic and LL was meso-eutrophic (i.e. between mesotrophic and eutrophic). Phytoplankton richness extended across cyanobacteria and 5 groups of photosynthetic eukaryotes (Fig. 2). In total, 59 species were identified in LT and 79 in LL (Table S2 in the supplement). Only 15 algal species were common to both lakes. The annual average phytoplanktonic biomass was  $2.8 \text{ mg C l}^{-1}$  in LT and  $125 \text{ mg C l}^{-1}$  in LL. In LT during the mixing season, 2 species of centric diatoms, *Urosolenia* sp. and *Cyclotella* sp., dominated the phytoplanktonic community, accounting for 92% of the total biomass. In LL during the mixing season, *Synura uvella* represented 51% of the phytoplanktonic biomass. During the stratified season, 2 species of dinophytes, *Peridinium* sp. and *Gymnodinium* sp., and 2 species of filamentous cyanobacteria, *Planktothrix* sp. and *Planktolyngbya limnetica*, represented 90% of the phytoplanktonic biomass. In our microscopic examination of the phytoplankton in LT and LL, fungal sporangia were occasionally observed on several algal species (*Staurastrum rotula*, *Synedra acus*, *Cyclotella* sp., *Synura uvella*, *Oocystis lacustris*, *Botryococcus braunii*, and *Arthrodesmus* sp.) (Fig. 3, Table S2). However, the observation of infected algal cells was infrequent during the time of the present study. In addition, because of the lack of distinctive morphological features displayed by zoosporic fungi (Powell 1993, Kagami et al. 2012), microscopic observation of the thallus did not allow us to identify the observed organisms at the species level.

### Small eukaryotic diversity

In our study, 4 clone libraries were constructed. A total of 455 positive clones were analyzed, yielding 119 unique OTUs (Table 1). Although none of the accumulation curves generated reached an asymptote (Fig. S1 in the supplement), coverage estimates (between 73.2 and 94.7%) (Table 1) indicated that

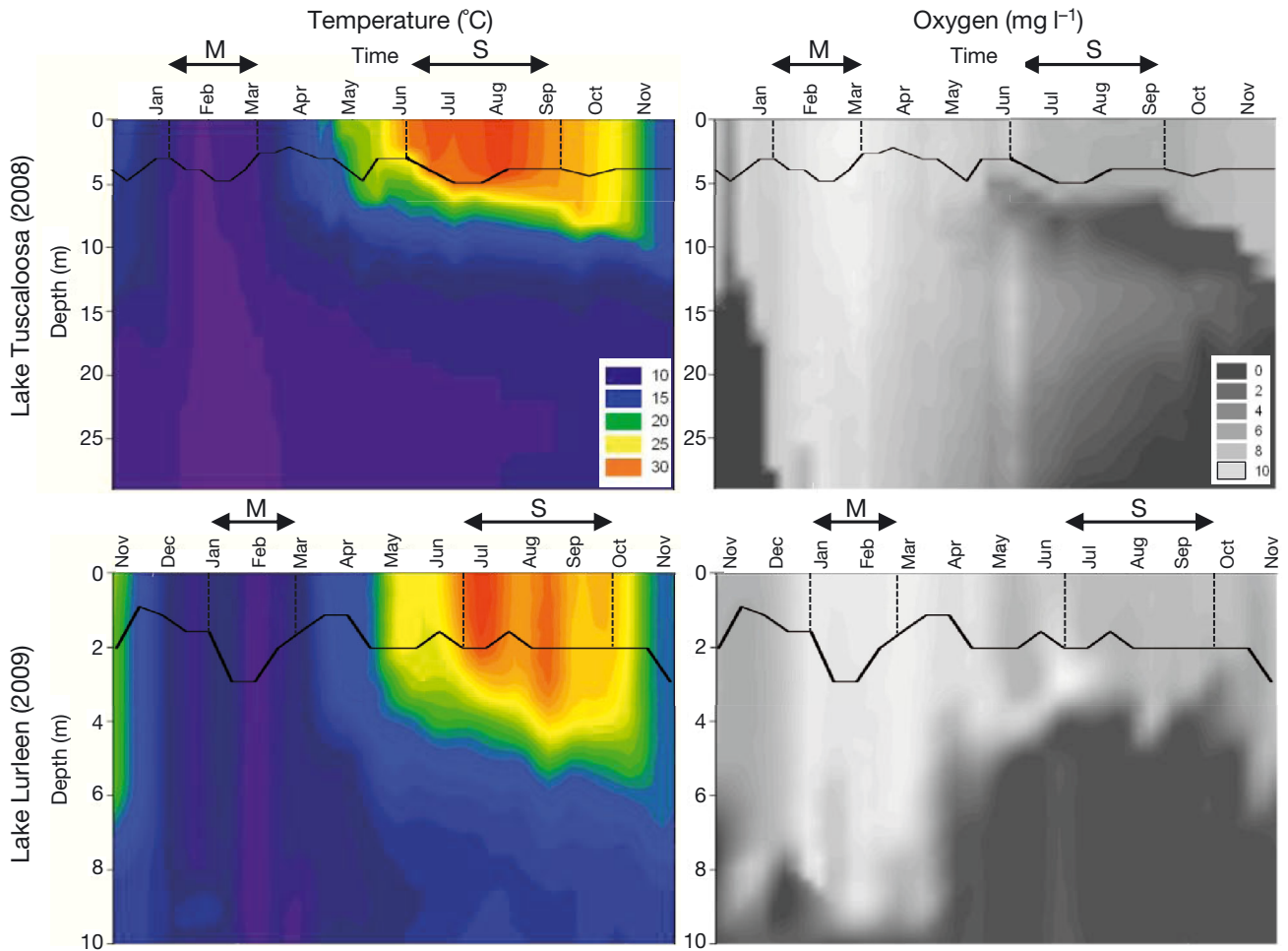


Fig. 1. Temperature (left) and oxygen concentration (right) isopleths for Lake Tuscaloosa (from December 2007 to November 2008, upper) and Lake Lurleen (from November 2008 to November 2009, lower). Secchi depth is represented by a black line. Sampling depths and seasons (M: mixing season, S: stratified season) covered by our molecular survey are delimited by vertical dashed black lines and horizontal arrows, respectively

the eukaryotic community was relatively well sampled. Among the 119 sequences generated in the present study, only 11 had  $\geq 97\%$  identity to identified organisms in GenBank. The BLAST of the 119 sequences indicated that a majority (74 to 79%) affil-

iated with 4 phylogenetic groups: Fungi, alveolates, stramenopiles, and cryptophytes, which were found in both lakes during both seasons (Fig. 4). The remaining sequences (i.e. 'other groups'; Fig. 4) were related to Cercozoa, Katablepharida, Telonema, Cen-

Table 1. Results of the RFLP analysis for the 4 libraries constructed from Lake Tuscaloosa and Lake Lurleen. Dissolved oxygen and temperature averaged over the 2 sampling seasons for each lake are indicated. Sample coverage was calculated using the program SPADE (<http://chao.stat.nthu.edu.tw/softwareCE.html>)

Lake	Trophic status	Thermal condition	Oxygen ( $\text{mg l}^{-1}$ )	Temperature ( $^{\circ}\text{C}$ )	No. of OTUs/ positive clones	No. of unique OTUs	OTUs shared between seasons	Sample coverage (%)
Lake Tuscaloosa	Oligotrophic	Mixing	10.1	10.8	29 / 109	25	4	87
		Stratified	6.8	28.4	38 / 102	30		86.3
Lake Lurleen	Meso-eutrophic	Mixing	11	9.8	22 / 132	19	1	94.7
		Stratified	7.6	28.8	52 / 112	45		73.2

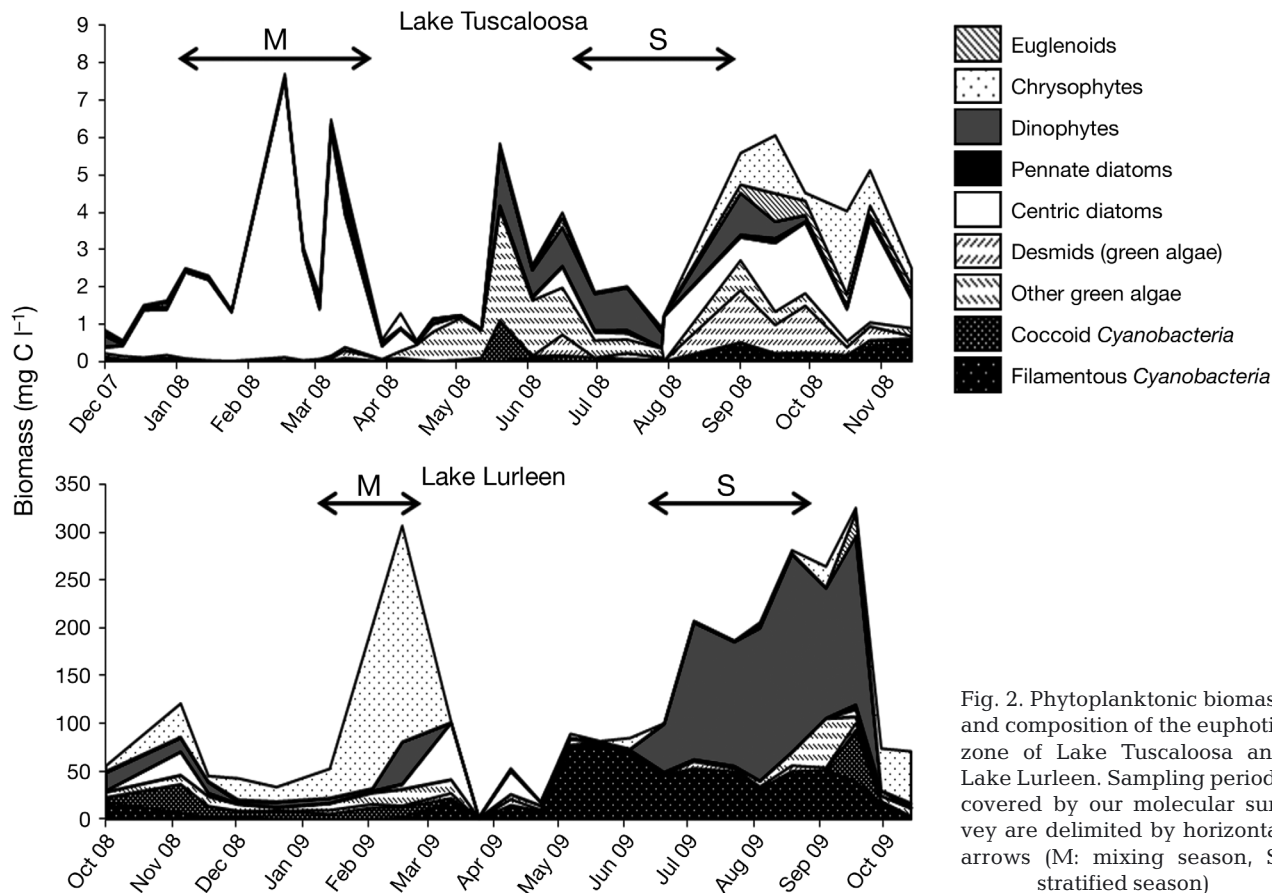
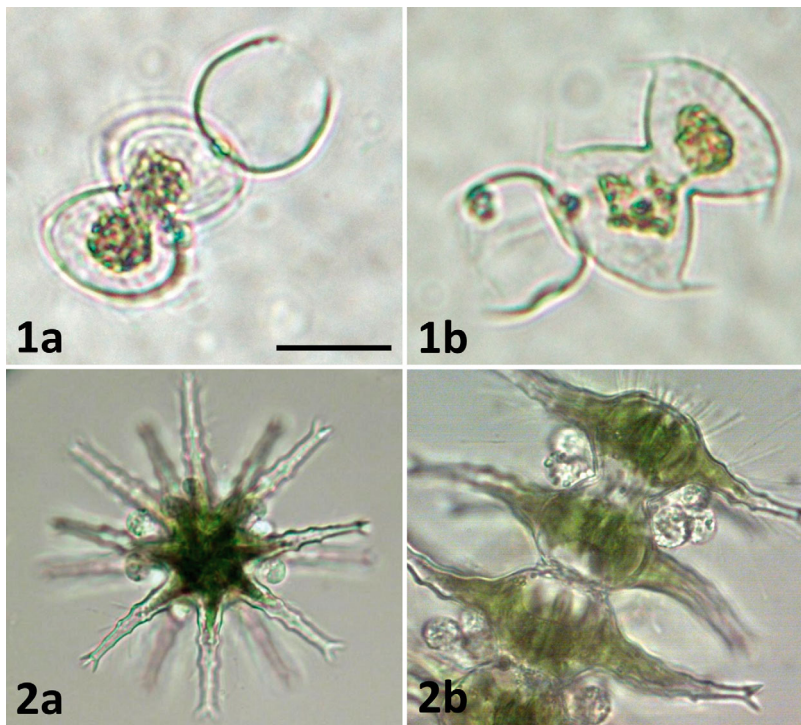


Fig. 2. Phytoplanktonic biomass and composition of the euphotic zone of Lake Tuscaloosa and Lake Lurleen. Sampling periods covered by our molecular survey are delimited by horizontal arrows (M: mixing season, S: stratified season)



trohelizoa, Euglenozoa, Ichthyospora, Glaucophyta, and Chlorophyta and were each represented in a relatively low proportion ( $\leq 2\%$ ). In each library, a majority of the sequences recovered (72 to 84%) belonged to taxa whose primary mode of nutrition was heterotrophy (Fig. 4). Only 7 (6%) of the 119 unique OTUs were found in both lakes. Few OTUs were recovered from both the mixing and stratified periods, with only 4 in LT and only 1 in LL. Overall, the number of unique OTUs recovered was

Fig. 3. Micrographs of algae bearing fungal sporangia observed in Lake Tuscaloosa: (1a,b) *Arthrodesmus* sp. bearing an empty fungal sporangium (observed in September 2008), (2a,b) *Staurastrum rotula* bearing several developing fungal sporangia (observed in November 2008). Scale bar = 25  $\mu\text{m}$  for 1a,b & 2a and 50  $\mu\text{m}$  for 2b

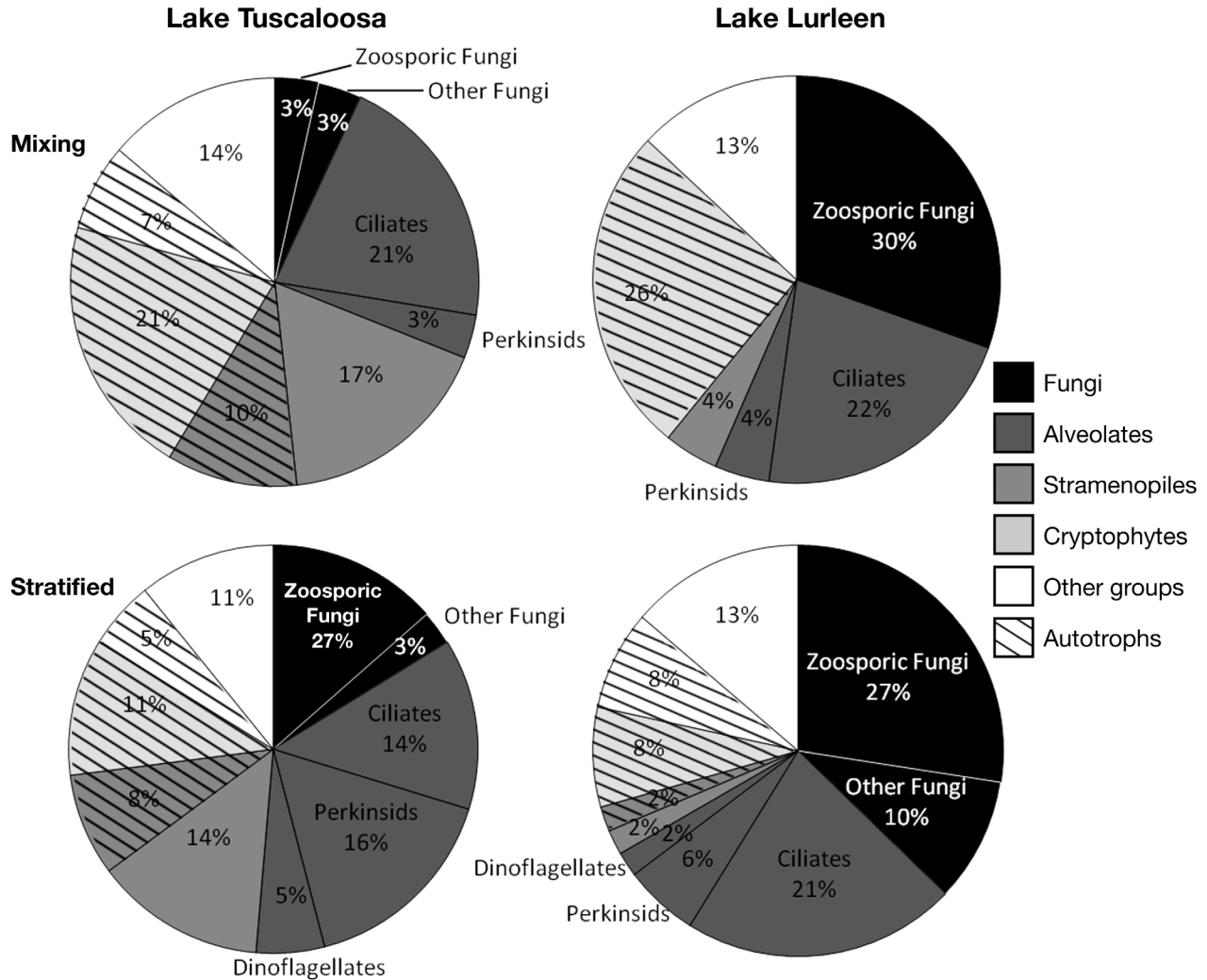


Fig. 4. Relative abundances of OTUs recovered from the 4 clone libraries constructed in the present study. OTUs were grouped into main phylogenetic groups based on BLAST search of the partial LSU rRNA gene

slightly higher in LL (73 OTUs for 244 clones analyzed) than in LT (63 OTUs for 211 clones analyzed). When only seasons were considered, the number of OTUs recovered during the stratified seasons was 1.7-fold greater than during the mixing season (i.e. 90 OTUs of 214 clones analyzed during the stratified season and 51 OTUs of 241 clones analyzed during the mixing season) (Table 1).

#### Zoosporic fungal diversity and phylogeny

A total of 31 unique fungal sequences were recovered from the clone libraries (25 zoosporic fungi and 6 other fungi). Only 3 of these sequences were similar ( $\geq 97\%$ ) to previously described and sequenced fungal species (1 Basidiomycetes *Oxyporus* sp. and 2

Chytridiomycetes, *Chytriomycetes hyalinus* MP053 and *Kappamyces* sp. JEL356). Most of the fungi recovered in our study (24) were detected in the eutrophic LL, where fungi dominated the eukaryotic diversity for both seasons (Fig. 4). Although fungi represented only 6% of the small eukaryotic diversity during the mixing season in LT, they were the second most represented group after alveolates during the stratified season (Fig. 4). The environmental LTME6I sequence, which affiliated with the aquatic saprobe zoosporic fungus *Coralloidiomyces digitatus* (order Rhizophydiales) (Letcher et al. 2008c), was the only sequence detected during both mixing and stratified seasons. The majority of environmental fungal sequences (83%) belonged to zoosporic fungi.

The phylogenetic placement of the 28S rDNA zoosporic fungal sequences is shown in Fig. 5. Maxi-

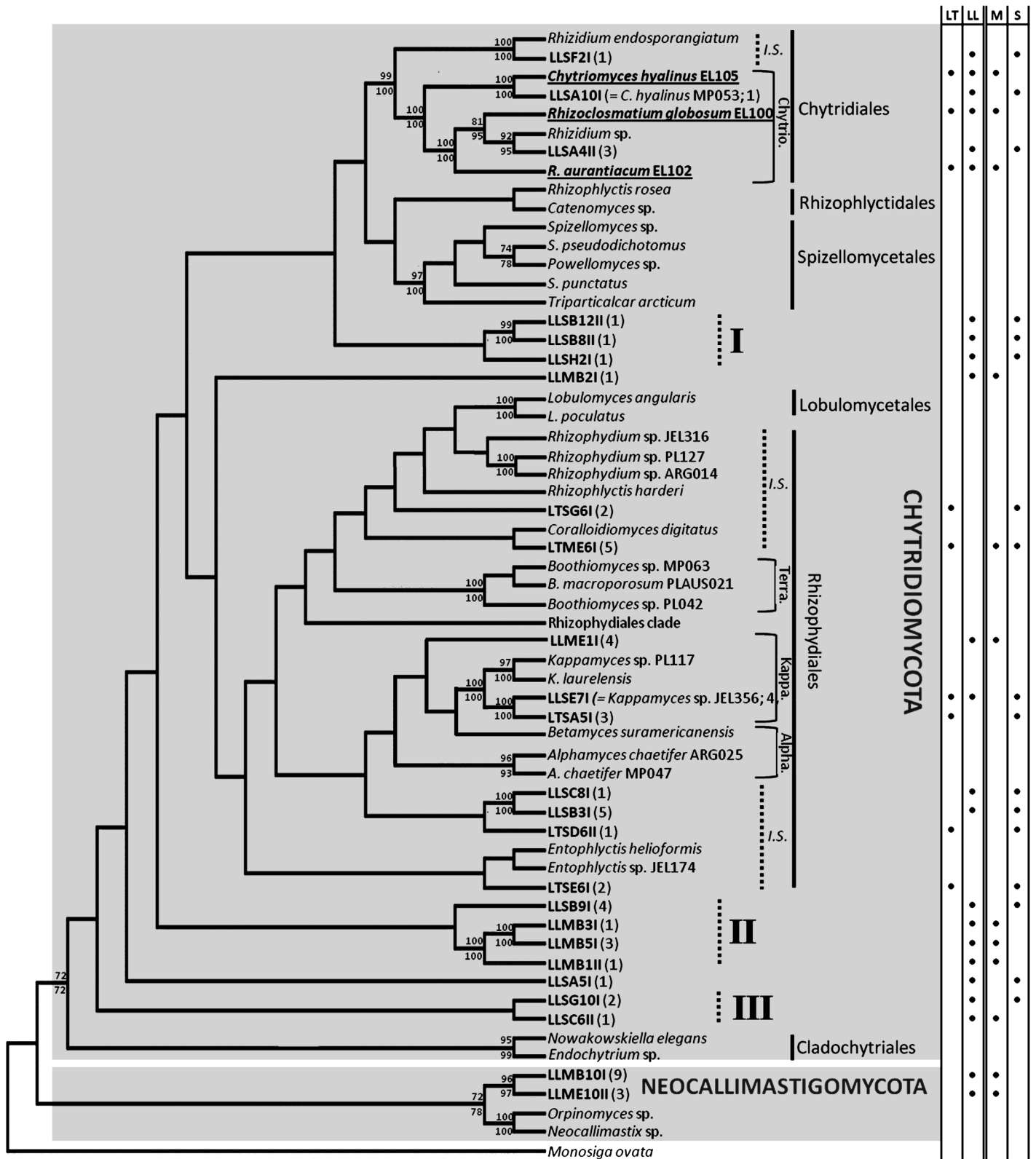


Fig. 5. Maximum likelihood phylogenetic tree of the basal fungi (i.e. zoosporic fungi) including a total of 60 partial 28S rRNA sequences (28 recovered from the present study, in **bold**, and 32 obtained from GenBank, Table S1 in the supplement). Support values ( $\geq 70\%$ ) above and below the nodes are bootstrap values derived from maximum likelihood and maximum parsimony analyses, respectively. Taxa that were cultured are underlined. For each environmental sequence, the number of clones retrieved in our clone libraries is indicated in parentheses. Novel environmental clades are indicated by the Roman numerals I, II, and III. Hierarchical classification is given to the right of the tree. Occurrence of sequences in Lake Tuscaloosa (LT) and Lake Lurleen (LL), during the mixing (M) and the stratified (S) seasons are indicated (•) adjacent to the tree. Chytrio.: Chytriomycetaceae, Terra.: Terramycetaceae, Kappa.: Kappamycetaceae, Alpha.: Alphamycetaceae, I.S.: incertae sedis



imum parsimony ( $L = 2478$  steps) and maximum likelihood ( $-\ln L = 15\,978.78$ , using the GTR + I +  $\Gamma$  model of evolution) phylogenies were identical, with similar support values at major nodes; thus, only the ML tree is shown (Fig. 5). Within the phylogeny, strong support ( $>70\%$ ) occurred primarily at terminal branches as well as at the divergence of Chytridiomycota from Neocallimastigomycota (72%). The remainder of the backbone was not strongly supported. However, distinct and recognizable monophyletic groups in the inferred phylogeny are congruent with major clades in Chytridiomycota and Neocallimastigomycota (James et al. 2006a,b) that have been previously delineated, including Spizellomycetales (Barr 1980), Rhizophydiales (Letcher et al. 2006), Rhizophlyctidales (Letcher et al. 2008b), Cladochytriales (Mozley-Standridge et al. 2009), Lobulomycetales (Simmons et al. 2009), and Chytridiales (Vélez et al. 2011). Half

of the zoosporic fungal sequences obtained grouped within previously delineated orders of Chytridiomycota, (i.e. Chytridiales and Rhizophydiales). The other half formed clades that did not include any known or described zoosporic fungi. Three species of zoosporic fungi belonging to the order Chytridiales (Vélez et al. 2011) were isolated from both LT and LL (Fig. 6). Based on microscopic observation of the thallus (Fig. 6) and 28S rDNA partial sequences, 2 isolates were identified as *Chytriomyces hyalinus* (EL105) and *Rhizoclostratium globosum* (EL100) (Fig. 6). The third isolate was identified as *Rhizoclostratium aurantiacum* (EL102) according to the morphological characteristics of its thallus (Sparrow 1960) (Fig. 6). Only 5 of the 28 zoosporic fungi cultured (EL100, EL102, and EL105) or detected (LLSA10I = *Chytriomyces hyalinus* and LLSE7I = *Kappamyces* sp.) in our study had been previously

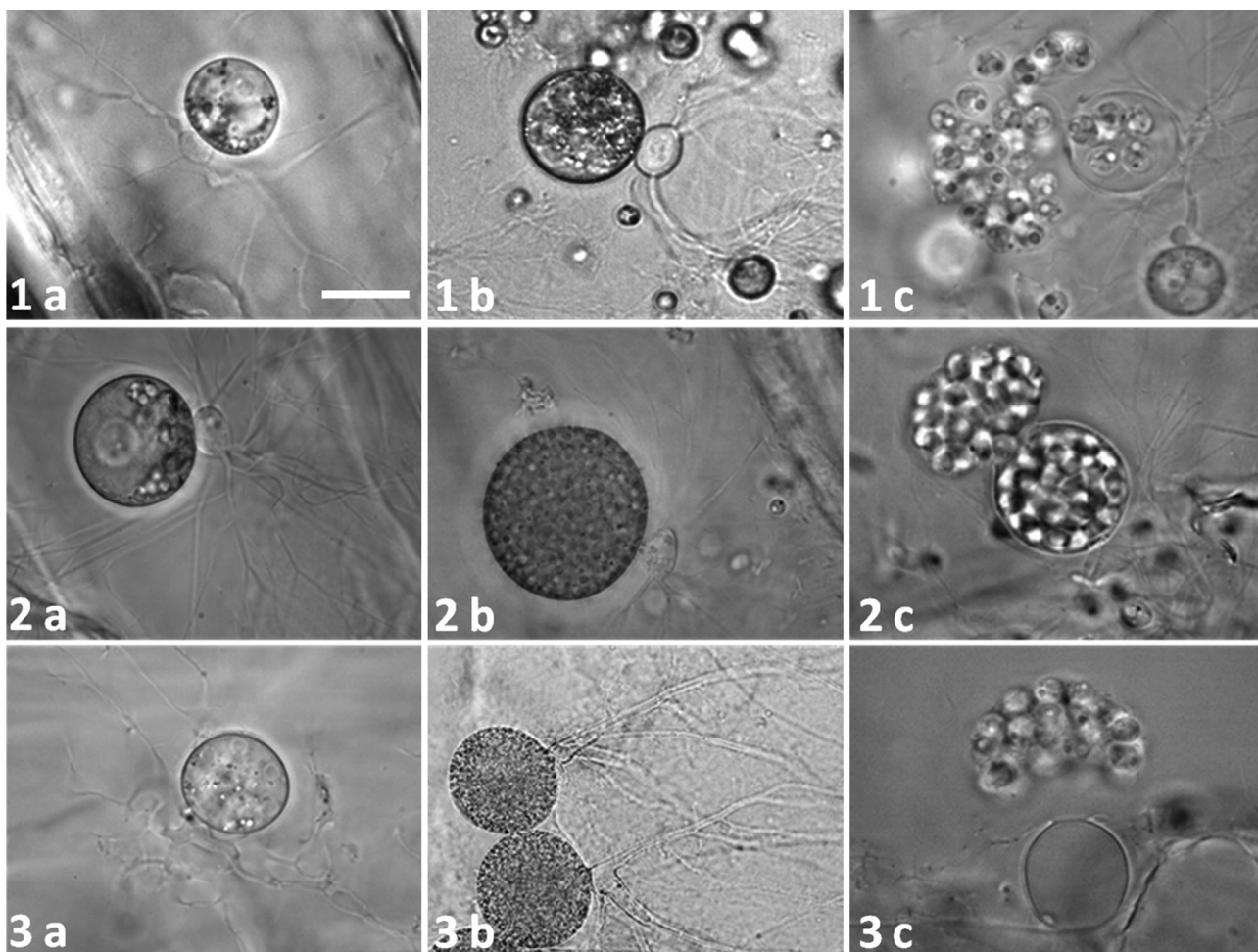


Fig. 6. Morphology of the 3 zoosporic fungi isolated from Lake Tuscaloosa and Lake Lurleen: (1) *Rhizoclostratium globosum* EL100, (2) *Rhizoclostratium aurantiacum* EL102, and (3) *Chytriomyces hyalinus* EL105 (a) germlings, (b) developing thalli, and (c) zoospore discharge. Scale bar in 1a = 20  $\mu\text{m}$  for all micrographs

isolated and described (Fig. 5). All of these except for LLSA10I were detected in both lakes.

The environmental clades I, II, and III as well as the lineages LLMB2I and LLSA5I were only found in LL. The remaining environmental sequences from LL grouped within Chytridiales, Rhizophydiales, and Neocallimastigomycota, whereas all environmental sequences from LT grouped within Rhizophydiales (Fig. 5). Within the Rhizophydiales, only 3 sequences (LLME1I, LTSA5I, and LLSE7I) fell into a known family, Kappamycetaceae (Letcher et al. 2008a). The other sequences grouped with phylogenetically unresolved taxa (*incertae sedis*). Within the order Chytridiales, however, all detected sequences except for LLSF2I (Fig 5) fell within the family Chytriomycetaceae (Vélez et al. 2011).

## DISCUSSION

### Study sites

LT and LL are located in the same geographic area and exhibited similar seasonal stratification regimes (Fig. 1). However, based on chlorophyll *a* concentration and Secchi depth measurements, LT and LL were characterized as oligotrophic and meso-eutrophic, respectively (Fig. S3 in the supplement at [www.int-res.com/articles/suppl/a067p091\\_supp.pdf](http://www.int-res.com/articles/suppl/a067p091_supp.pdf)). The metalimnic oxygen minimum observed in LT, resulting from the activity of heterotrophic organisms, and maximum in LL (Fig. 1), resulting from autotrophic organisms, indicated that these 2 lake systems fundamentally differed in their functioning (Wetzel 2001). In addition, the dark color of LT water, a result of a high concentration of humic compounds in this lake (Wetzel 2001), likely prevents the production of a high algal biomass, supporting the development of heterotrophic microorganisms. These observations suggest that the 2 lakes differ in their carbon source and microbial food web structure and, thus, constituted ideal contrasting sites for comparing microbial communities.

### Overview of the small eukaryotic diversity

Our study is the first to target the LSU gene in lakes. Previous similar studies used the SSU gene (Lefranc et al. 2005, Richards et al. 2005, Slapeta et al. 2005, Lefèvre et al. 2007, 2008, Chen et al. 2008, Lepère et al. 2008, Genitsaris et al. 2009, Luo et al. 2011, Monchy et al. 2011). Nevertheless, some of the

general results obtained in our study are in accordance with the findings of previous studies: (1) the most represented taxa in our genetic libraries were fungi, alveolates, stramenopiles and cryptophytes, (2) heterotrophs were phylogenetically the most diverse among the small eukaryotic community, (3) Cryptophyta was the most diverse taxon among the small phototrophic eukaryotes, and (4) only few taxa were found in common in the 2 lakes studied. These results suggest that the overall structure of the small eukaryotic community is similar in a wide range of freshwater lakes. However, at a lower phylogenetic level, the structure of the eukaryotic community seems to be determined by the local characteristics of a lake.

A link between trophic level and the small eukaryotic community structure in lakes has been proposed (Lefranc et al. 2005, Chen et al. 2008, Lepère et al. 2008). Results from these studies suggested that the diversity of the small eukaryotic community tended to be lower in more productive waters. In the present study, the meso-eutrophic LL displayed a slightly higher diversity than the oligotrophic LT. However, this result might only reflect the fact that fewer clones were analyzed in LT libraries (211) compared to LL libraries (244). Therefore, the small eukaryotic community is possibly not more diversified in the meso-eutrophic LL compared to the oligotrophic LT.

Although our sampling effort was slightly lower during the stratification (i.e. 102 and 112 clones analyzed for LT and LL, respectively) than during the mixing season (i.e. 109 and 132 clones analyzed for LT and LL, respectively), the number of recovered OTUs was higher during the stratified season. However, more intensive sampling is necessary to establish a clear relationship between eukaryotic diversity and thermal stratification.

### Zoosporic fungal diversity and phylogeny

With the exception of 1 study (Richards et al. 2005), zoosporic fungi have been systematically detected in lakes (Lefranc et al. 2005, Lefèvre et al. 2007, 2008, Lepère et al. 2008, Genitsaris et al. 2009, Luo et al. 2011, Monchy et al. 2011). Because some previous studies comparing aquatic systems differing in their trophic status found that zoosporic fungal diversity was higher in oligotrophic waters (Lefranc et al. 2005, Chen et al. 2008, Monchy et al. 2011), the oligotrophic LT was expected to harbor a higher zoosporic fungal diversity than the meso-eutrophic LL. However, the zoosporic fungal diver-

sity was generally higher in the more eutrophic lake (Fig. 4). Table 2 summarizes environmental eukaryotic molecular surveys conducted in several lakes differing in trophic status and in which zoosporic fungal sequences were detected. Although comparison among these studies is difficult because experimental designs (season, depth, and size fraction sampled), methodologies, and data analyses (sampling effort, PCR primers, and sequence percentage similarity cut-offs) differed (Table 2), no obvious relationship between zoosporic fungal diversity and trophic status could be suggested. In contrast to the present study, which used integrated samples (i.e. several sampling points in time were pooled), previous environmental molecular surveys (Table 2) only analyzed 1 or 2 samples isolated in time collected during various seasons of the year. Thus, although data from numerous studies are available (Table 2), it might be too early to establish a clear relationship between trophic status and zoosporic fungal community structure.

Previous studies have found that the zoosporic fungal community in lakes was either affiliated with members of the Rhizophydiales and Chytridiales or formed environmental clades (i.e. with no cultured or described species) (Lefèvre et al. 2007, 2008, Lepère et al. 2008, Monchy et al. 2011). In the design of the present study, we hypothesized that the systematic recovery of environmental clades of zoosporic fungi in previous surveys might have been related to the lack of representative fungal sequences in the constructed phylogeny. The LSU gene has been widely used in the taxonomy of lower fungi (i.e. zoosporic fungi) (Tanabe et al. 2005, Letcher et al. 2005, 2006, 2008a,b,c, James et al. 2006a,b, Hibbett et al. 2007, Mozley-Standridge et al. 2009, Simmons et al. 2009, Wakefield et al. 2010). Thus, because of the greater availability of LSU sequences, we expected our retrieved sequences to primarily fall into described taxa of zoosporic fungi. Surprisingly, our results were similar to what was obtained in previous studies, with half of our sequences affiliated with described zoosporic fungal taxa and the other half forming environmental clades. This suggests that the environmental clades detected in previous studies were not an artifact due to the lack of fungal sequences used as references in the phylogeny but actual environmental clades for which zoosporic fungi have not yet been described. Thus, our study confirms that lakes are undersampled for zoosporic fungal diversity and that these ecosystems harbor a highly unexplored zoosporic fungal diversity mainly within the order Chytridiales and Rhizo-

phydiales, known to contain both saprobes and algal parasites (Letcher et al. 2008a, Vélez et al. 2011). Interestingly, 2 environmental sequences (LLMB10I and LLMBE10II) affiliated to Neocallimastigomycota, which are known to be essentially anaerobic (Liggenstoffer et al. 2010), were detected. Only 1 previous similar study conducted by Slapeta et al. (2005) in a shallow freshwater pond also detected 1 environmental sequence related to this phylum. Although these sequences were detected in the plankton, they could actually originate from anaerobic sediments.

A majority of the zoosporic fungi detected in our molecular survey were only found in 1 lake and were only represented by few clones in our libraries (Fig. 5). This suggests that the zoosporic fungi detected by the molecular survey are probably part of the 'rare biosphere' (Sogin et al. 2006), supporting the idea that molecular surveys tend to detect less abundant ('rare') species, while culture methods are selective for more common species. The isolation of only 3 zoosporic fungi with selective baiting techniques in the present study highlights the difficulty of adequately culturing zoosporic fungi from aquatic systems. Although zoosporic fungal isolation from the small planktonic fraction (0.6 to 8  $\mu\text{m}$ ) was attempted in the present study, only isolates from the fraction collected using the plankton tow (>60  $\mu\text{m}$ ) were obtained. Thus, since our sampling strategy differed between culturing and molecular approaches, comparison of the diversity obtained between the 2 methods cannot be made. In addition, as the baiting technique only selects for saprobes and facultative algal parasites (Letcher & Powell 2001, 2002, Letcher et al. 2004b, 2008a,b, Wakefield et al. 2010), zoosporic fungi that are obligate parasites on phytoplankton and known to be diverse and recurrent in lake ecosystems (Gromov et al. 1999, Ibelings et al. 2004, Kagami et al. 2007b, Rasconi et al. 2009, 2012, Sønstebo & Rohrlack 2011) remain undetected using culturing methods.

Previous studies have shown that fungal sporangia are commonly observed attached to planktonic algae (Kagami et al. 2007b, Rasconi et al. 2009, 2012). Although fungal epidemics of phytoplanktonic populations are a recurrent phenomenon in lakes (Canter & Lund 1951, van Donk & Ringelberg 1983, Kagami & Urabe 2002, Miki et al. 2011), the mechanisms underlying the dynamics of parasitic fungi on phytoplankton are not well understood. Rasconi et al. (2009, 2012) suggested that the community of algal parasitic fungi might be more related to the community composition of their host

Table 2. Summary of environmental molecular surveys conducted in lake ecosystems. Histograms in the last column indicate percentage of zoosporic fungi and other fungi among eukaryotes detected for each study. nd: not determined/no data

Lake	Trophic status	Thermal condition	Zone sampled	Oxygen (mg l <sup>-1</sup> )	T (°C)	Planktonic fraction (µm)	Primer used (gene)	No. of OTUs (% similarity cutoff)	Coverage (%)	Source	Zoosporic fungi (black) and other fungi (grey) (%)
Stechlin	Oligotrophic	Mixing	Euphotic	Saturation	20	>0.6	EukA/B (18S)	24 (97)	nd	Luo et al. (2011)	
Tuscaloosa	Oligotrophic	Mixing	Euphotic	9	10	0.6–8	LROR/LR5 (28S)	29 (97)	87	Present study	
		Stratification	Euphotic	6.8	28.4	0.6–8		38 (97)	86	Present study	
Godivelle	Oligotrophic	nd	Euphotic	8	14.9	0.2–5	1f/1520r (18S)	18 (99)	71	Lefranc et al. (2005)	
Pavin	Oligo-mesotrophic	nd	Euphotic	9.9	15	0.2–5		26 (99)	46	Lefranc et al. (2005)	
Pavin	Oligo-mesotrophic	Fall stratification	Mixolimnion	8.7	10.7	0.6–5	1f/1520r (18S)	46 (97)	90	Lefèvre et al. (2007)	
		Spring stratification	Mixolimnion	8	9.6	0.6–5		50 (97)	86	Lefèvre et al. (2008)	
Pavin	Oligo-mesotrophic	Stratification	Euphotic	nd	17.5	0.6–150	NS1/TTS4 (18S-ITS2)	46 (99)	nd	Monchy et al. (2011)	
Bourget	Mesotrophic	nd	Euphotic	nd	nd	0.2–5	1f/82f/1520r (18S)	120 (99)	91	Lepère et al. (2008)	
Lurleen	Meso-eutrophic	Mixing	Euphotic	9	10	0.6–8	LROR/LR5 (28S)	22 (97)	95	Present study	
		Stratification	Euphotic	7.6	28.8	0.6–8		52 (97)	73	Present study	
Aydat	Eutrophic	Stratification	Euphotic	nd	19.5	0.6–150	NS1/TTS4 (18S-ITS2)	63 (99)	nd	Monchy et al. (2011)	
Aydat	Eutrophic	nd	Euphotic	7.4	25.5	0.2–5	1f/1520r (18S)	12 (99)	89	Lefranc et al. (2005)	
Tailhu	Eutrophic	nd	Euphotic	5.16	nd	0.8–20	EukA/B (18S)	131 (98)	86	Chen et al. (2008)	
Cheuvreuse	nd	nd	nd	9.6	4	>0.2	42f/82f/1498r/1520r (18S)	15 (97)	nd	Slapeta et al. (2005)	
Koronia	Hyper-eutrophic	nd	Euphotic	7.8	22.6	>1.2	EukA/1633rE (18S)	11 (98)	38	Genitsaris et al. (2009)	

than the trophic status of the lake. Indeed, the higher algal biomass in LL, which represents a higher abundance of potential hosts for parasitic zoosporic fungi, could explain the higher diversity detected in LL. Similarly in LT, the more diverse algal community during the stratified season could also potentially explain the higher diversity of zoosporic fungal community during this season. However, although fungal sporangia were observed on several algal species in both lakes (Fig. 3), their occurrence was very low, suggesting that most of the zoosporic fungi detected in our clone libraries were saprobes. Therefore, it seems unlikely that the composition and dynamics of the phytoplanktonic community explains the temporal variation of the zoosporic fungal community in LT and LL. Quantitative methods have recently been developed to detect zoosporic fungi in aquatic ecosystems (Rasconi et al. 2009, Jobard et al. 2010, Lefèvre et al. 2010). However, more adapted and specific methods are needed for the study of the zoosporic fungal community. In the present study, LROR and LR5 primers, although widely used for zoosporic fungal taxonomy (James et al. 2006a,b, Letcher et al. 2008a,c, Mozley-Standridge et al. 2009, Wakefield et al. 2010), amplified various other eukaryotes. The availability of specific fungal primers represents a major limitation for the molecular detection of fungi in natural environments. Primers initially designed to target fungi, nu-SSU-0817-5'/nu-SSU-1196-3' (Borneman & Hartin 2000) and FF390/FR1 (Vainio & Hantula 2000), once used in molecular environmental surveys in lake ecosystems (Monchy et al. 2011 [nu-SSU-0817-5'/nu-SSU-1196-3'], Kagami et al. 2012 [FF390/FR1]), also target an important proportion of eukaryotes (81 to 91 % of sequences detected using nu-SSU-0817-5' and nu-SSU-1196-3' fungal primers were not fungal sequences) (Monchy et al. 2011). These methodological limitations stress the importance of developing new techniques adapted to the isolation and detection of zoosporic fungi from truly aquatic habitats. In addition, direct observation and identification of zoosporic fungi in natural samples is a very difficult task due to their small size and lack of distinctive morphological characteristics (Powell 1993, Lefèvre et al. 2007, 2008, Kagami et al. 2012). The molecular data generated from lake surveys will provide greater possibilities for the design of specific primers and the development of molecular techniques allowing quantification of zoosporic fungi in lakes, thus improving our understanding of their fundamental importance in pelagic microbial food webs.

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