

Measure the difference in efficiency in waste disposal and collection services from the EU targets in the Campania municipalities.

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1 **Measure the difference in efficiency in waste disposal and collection services from**
2 **the EU targets in the Campania municipalities.**

3
4 **Abstract**

5
6 The study analyses the economic and environmental performance of the 353 municipalities in the region
7 of Campania in the waste disposal and collection services. This study consists of three steps, firstly
8 municipal performance in the management of waste services from a linear economy point view is
9 assessed. Secondly, a circular economy paradigm is considered and jointly the economic (minimization
10 of management costs) and environmental (waste minimization) performance of municipalities
11 management is measured; two different DEA models are employed using information provided by the
12 Institute for Environmental Protection and Research for the year 2016. Third, in order to rank the most
13 virtuous municipalities in the environmental dimension (toward a circular economy paradigm), the study
14 defines a measure of the efficiency deviation from environmental sustainability. The results show a
15 cluster of municipalities in the metropolitan area of Naples and Caserta with a worse performance in the
16 environmental dimension but with a good performance in the economic dimension. The succession of
17 national and regional regulations has accentuated the uncertainty in the executive process and in the
18 management of the waste cycle, creating a regulatory vacuum. Local governments should act on citizen
19 motivations, promoting awareness on environmental issues, and should implement time-saving
20 collection methods.

21
22 ***JEL: D61; H3; H76; I38; Q24; Q53; Q56; R14;***

23
24 ***Key words:*** *environmental performance, economic performance, environmental sustainability,*
25 *Environmental economics, waste services, DEA, Campania municipalities, separate waste collection*
26 *(SWC), Legislative Decree 152/2006.*

27
28 ***Highlights:***

29 *1. Collecting recyclable materials is the basis for the processes of circular economy's recovery,*
30 *reuse, and recycling*

- 31 2. *We analyze the economic and environmental performance of the 353 municipalities of the*
32 *Campania region in waste disposal and collection services.*
- 33 3. *Two different input-oriented models are used using the input and output information provided by*
34 *the Institute for Environmental Protection and Research for the year 2016.*
- 35 4. *The analysis highlighted the importance of jointly assessing the economic and environmental*
36 *dimensions of the production process relating to the municipal solid waste management service*
37

38 **1. Introduction**

39

40 Significant concerns over the environmental impact of waste, primarily involving human health and
41 the environment, have emerged in European countries. There is a need for a more efficient use of
42 resources throughout their entire life cycle. In this context, a development strategy achieving zero waste
43 is highly relevant in the path towards a circular economy, since it allows to keep products at their highest
44 value for as long as possible through recycling (European Commission Report 2014; Malinauskaite, et
45 al., 2017). Collecting recyclable materials is the basis for the processes of circular economy's recovery,
46 reuse, and recycling. Over the last decades, waste management has become a pivotal element in European
47 environmental protection policies. The growing focus on municipal solid waste reduction, re-use and
48 recycling by national and local governments has prompted important reforms of municipal waste
49 policies.

50 Regulatory frameworks, both at European and national level, pose great emphasis on the prevention
51 and minimization of unsorted urban waste production and on the maximize the recovery of materials
52 through the separation collection. The recent European framework directives (UE 2018/850, 2018/851
53 and 2018/852) impose on the one hand the necessary misalignment between economic growth and the
54 increase in consumption and production waste, and on the other hand the optimization of waste
55 management systems for the purpose of maximizing the recovery of waste (recycling objective) and of
56 minimizing the production of unsorted waste (reduction landfills use objective). In particular, the circular
57 economy package sets targets for the recycling of municipal waste (reaching at least 55% by 2025, 60%
58 by 2030 and 65% by 2035), the reuse of different materials and the reduction of use of landfills (by 2035
59 a maximum of 10% of the total urban waste). However, it is not possible to achieve they, without taking
60 into account the effectiveness issue of waste management system at municipal level. These, among other
61 things, allow the achievement of both direct and indirect economic benefits, as the creation of new firms
62 and the increase in jobs (Agovino et al., 2019a).

63 In deep, the Directive 2008/98/EC established specific targets, in order to limit the production of
64 waste, using it as a resource, so as to create a “recycling society with a high level of efficiency”. The
65 Directive has been transposed in Italy by the Legislative Decree 205/2010, which set the targets for
66 separate waste collection, recovery and recycling, defining the responsibilities among the actors of the
67 national waste management system. However, it is not possible to achieve these targets, without taking
68 into account the effectiveness issue of waste management policies as well as the efficiency of disposal
69 and collection waste services. In Italy, the regulation of the waste management system passed through
70 three milestones: 1) in 1982, Decree of the President of the Republic (D.P.R.) n. 915/1982; 2) in 1997,
71 L.D. n. 22/1997; 3) in 2006, L.D. n. 152/2006. After the introduction of L.D. 152/2006, the level of
72 separate waste collection increased at a national level, but high regional differences persisted. Northern
73 regions seem to be more dynamic and reactive to the regulatory novelties, while Southern regions
74 achieved poorer and more heterogeneous results in terms of separate waste collection (Agovino et al.,
75 2021). Crociata et al. (2016) studied the separate collection rate across Italian provinces between 1999
76 and 2012. They highlighted the presence of two clusters of provinces: i) Northern Italian provinces with
77 values above the national average (good cluster); ii) Southern Italian provinces, including Campania,
78 featuring values below the average (bad cluster). These studies showed that improving separate waste
79 collection (SWC) achievements plays a pivotal role in narrowing the territorial gaps between Campania
80 and top regions in terms of environmental performance. However, achieving a high level of SWC
81 requires significant efforts, that can only be obtained through the joint action of citizens and institutions
82 (Agovino et al., 2018). There are regions where the planning framework is completely lacking or totally
83 inadequate; this is the case of Sicily, where municipal waste disposed of in landfills still accounts for
84 69% of the total waste produced, but also of Lazio and Campania, which cannot close the cycle within
85 the regional territory.

86 A significant part of current literature has focused on waste management issues (collection and
87 disposal process) in order to evaluate the efficiency of the process and the effectiveness of the
88 implemented policies and/or study the relationships between policy makers guidelines and citizen action
89 (Barr et al., 2001; Barr et al., 2005; McDonald and Oates, 2003; Tonglet et al., 2004; Timlett and
90 Williams, 2009; Lombardo 2009; Marques and Simões, 2009; Koushki et al., 2004; Agovino et al.,
91 2018). The negative impact of poor management of the waste process can have both environmental and
92 economic effects (Distaso, 2012). Most of the current literature has focused on the issue of waste
93 management in Campania addressing the complexity of the underlying causes of a waste emergency in
94 the region (D’Alisa and Kallis 2016, Garofalo et al., 2019).

95 In the light of all the above, Campania is an interesting case study since, despite the declaration of
96 emergency status from year 1994 to 2009 year ('waste crisis'). Numerous papers have been written on
97 the topic of waste in Campania (Armiero and D'Alisa, 2012; Triassi et al., 2015; De Biase, 2009;
98 Membretti, 2015; D'Alisa and Kallis, 2016; Agovino et al., 2019b), addressing the complexity of the
99 factors underlying the waste emergency that recently affected the region. A mix of technical-
100 administrative errors and political, industrial and only partially identified criminal interests may be held
101 responsible for the poor environmental performance of the area (see Distaso, 2012; Armiero and D'Alisa,
102 2012; D'Alisa and Kallis, 2016). In 2009, during the waste emergency, the rate of separated waste in
103 Campania was 29.3 %, far below the 48 % achieved in Northern regions and about one half of the 57.8
104 % recorded in Trentino-South Tyrol, the best performing region (Ispra, 2014). In the space of 22 years,
105 in the territories belonging to two provinces (i.e. Naples and Caserta), about 10 million tons of waste of
106 any type and degree of danger have been buried (Legambiente, 2013). These constitute only a part of the
107 waste buried in the Campania territories. This is confirmed by the discovery of the largest illegal landfill
108 in Europe. The landfill has an area of 25 hectares, and it is estimated that about one million cubic meters
109 of waste are buried in it; over 50% of the waste was found to be "special hazardous waste".

110 This region has gone from one of the worst Italian regions in terms of separate waste collection to one
111 of the most virtuous in southern Italy, as shown by Agovino et al., 2020. However, it is often at the local
112 level (especially at the municipal level) that the national laws are operationalized, and further waste
113 management policies and schemes are adopted. It is for this latter reason that the study investigates the
114 performance of the waste services process in Italian municipalities after about ten years on from the
115 'waste crisis'. These facts certainly constitute strong arguments for the need for careful monitoring of
116 the performance of local waste management systems on the various EU targets mentioned above.

117 This study consists of three steps. Firstly, the performance of Italian municipalities in the collection
118 and disposal process of solid urban waste considering only an economic dimension is assessed. In this
119 case, the more efficient local government units will employ, on equal output levels, a lower amount of
120 expenditure to dispose waste. Secondly, following the suggestions made by Sarra et al. 2017, the joint
121 economic and environmental performance of the municipalities in the waste services process is
122 evaluated. In this case, the municipalities with the best performance will use, on equal levels of 'good'
123 output, both a lower amount of expenditure to dispose waste and a lower amount of bad output which
124 increases the risk of landfill use (economic and environmental dimension of efficiency). The landfill is
125 in fact the last stage of the waste cycle. It contains all the materials that are no longer usable
126 (undifferentiated) and that must be permanently eliminated. To this end, two different input-oriented

127 models, the Banker, Charnes and Cooper Data Envelopment Analysis models are employed using input
128 and output information provided by ISPRA (The Institute for Environmental Protection and Research)
129 for the year 2016. Third, in order to rank the most virtuous municipalities in the environmental
130 dimension, the study defines a measure of the efficiency deviation (gap) from environmental
131 sustainability (a measure of the efficiency deviation from EU targets). This gap may occur when the
132 environmental dimension is neglected in the production process to a greater economic efficiency:
133 expenditure minimization, while environmental damage occurs.

134 The remainder of this study is as follows. Section 2 presents the empirical strategy, the case study and
135 the dataset. Section 3 reports the results and discusses them. Section 5 presents some conclusions.

136

137 **2. Material and methods**

138

139 This section presents the methods and the dataset used to investigate the waste service process
140 performance of the Campania municipalities.

141

142 *2.1 Methods*

143

144 The study employs Data Envelopment Analysis (DEA) to evaluate municipal's efficiency of waste
145 services.

146 In the municipal solid waste sector especially, DEA analysis has been performed to evaluate particular
147 services provided by municipalities, like waste collection (Worthington et al., 2001; Benito-López et al.,
148 2011, Simões et al., 2013), to estimate the performance of different regions in recycling (Lozano et al.,
149 2004, Chang et al., 2013, Yeh et al., 2016, Crociata and Mattoscio, 2016), to optimally select sites for
150 locating SW facilities (Khadivi and Ghomi, 2012). Most of those studies aim to evaluate the cost
151 efficiency (Huang et al., 2011, Rogge and De Jaeger, 2013, Spallini et al., 2016, Amaral et al., 2022) or
152 the operational efficiency (Ichinose et al., 2013, García-Sánchez, 2008) of municipal solid waste services.
153 Recent studies have focused on the importance of inter-municipal cooperation the efficiency of waste
154 management (Guerrini et al., 2017, Máñez et al., 2016, Pérez-López et al., 2018, Sarra et al., 2017,
155 Amaral al., 2022).

156 Unlike parametric techniques, DEA does not require a detailed description of the production process
157 and allows to assess the relative efficiency of homogeneous operating units (DMUs). The efficiency
158 scores are obtained assuming the *input-orientation* in the (collection and disposal) solid urban waste

159 services production process manage by the municipalities (our DMUs). The orientation to input assumes
 160 that DMUs have more control over the inputs (i.e. costs) than over the outputs (i.e. waste). Municipalities
 161 performance is assessed for two different models i.e., Model I and Model II respectively. In particular,
 162 Model I computes municipalities performance considering only the economic dimension of production
 163 process. In this case, DMUs with a better *performance* (more efficient) will employ, on equal output
 164 levels (total disposed waste), a lower amount of expenditure to collect and dispose waste.

165 Model II, following Sarra et al. 2017, jointly assesses the economic and environmental performance
 166 of municipalities in the urban waste services process. In this case, the municipalities with the best
 167 performance will use, on equal levels of *good output* (environmental benefit), both a lower amount of
 168 expenditure to dispose waste and a lower amount of *bad output* which damages the environment
 169 (*economic and environmental dimension of efficiency*). The waste from Sorted Waste (SW, good output)
 170 and Unsorted Waste (UW, bad output) are treated differently when municipal performances are evaluated
 171 from an *environmental dimension*. Collecting recyclable materials is the basis for the processes of
 172 recovery, reuse, and recycling, while the amount of unsorted waste cannot be reused and it must be
 173 disposed of in either landfills or incinerators, and this damaging the environment. Environmental
 174 performance of the production process increases when UW to be treated is reduced. There are different
 175 literature approaches that account for bad output (see Halkos and Petrou, 2019), Model II, in light of the
 176 analysis aims, treats UW as a classical input (indirect approaches): UW is an undesirable output which
 177 public agents have to minimize. Indeed, considering UW among the inputs allows to consider an *input-*
 178 *oriented model* where the maximum reduction refers not only to expenditure (costs) but also to
 179 undesirable waste as a deterrent factor for environmental sustainability.

180 Both Models (II and I) analyse the efficiency of DMUs by the *input-oriented* Banker, Charnes and
 181 Cooper (BCC-I) DEA model. The following linear programming (LP) problem under the BCC-I model
 182 is solved:

$$\begin{aligned}
 &183 \\
 &184 \quad h_o = \min_{\theta_o, \gamma} \theta_o \\
 &185 \quad \text{s.t.} \\
 &186 \quad \theta_o x_{oki} - \sum_i^n \gamma_i x_{ki} \geq 0 \quad k = 1, \dots, K \\
 &187 \quad \sum_i^n \gamma_i y_{mi} \geq y_{omi} \quad m = 1, \dots, M
 \end{aligned}$$

$$\sum_i^n \gamma_i = 1 \quad \text{convexity constraint}$$

$$\theta_{oi}, \gamma \geq 0$$

188

189

190

191 where y_{mi} is the amount of the m -th output to DMU i , x_{ki} is the amount of the k -th input to DMU i and
 192 γ are the weights of DMU i . The value of θ_o obtained is the ES for the i -th municipalities and is bounded
 193 between 0 and 1. According to the Farrell (1957) definition, when $\theta=1$ the DMU is technically efficient
 194 (i.e. a point on the frontier), while inefficient DMUs will score less than one.

195 In summary, the empirical analysis is divided into three steps.

196 The first two steps evaluate the performance of the waste management service considering two different
 197 dimensions and employing two different BCC-I DEA models, which assumed Variable Returns to Scale,
 198 under *input-orientation* to calculate the ES (θ_o) of DMUs:

199 - Model I of the economic dimension: The production process involves as a factor of production one
 200 input – i.e., the total annual costs per capita of managing the urban hygiene service - to produce the final
 201 output (disposed waste).

202 - Model II of the environmental and economic dimension: The production process involves as a factor
 203 of production one input – i.e., the total annual costs per capita of managing the urban hygiene service –
 204 to produce two outputs: a good output (environmental benefit) and a bad output (environmental damage).
 205 Both input and undesirable outputs are the values that need to be minimized. The model incorporates
 206 environmental damage as an input, which needs to be decreased to improve the performance of a DMU.
 207 The following constrain is added to the BCC-I DEA models considered: $\theta_o d_{osi} - \sum_i^n \gamma_i d_{si} \geq 0 \quad s =$
 208 $1, \dots, S$; where d_{si} is the amount of the s -th bad output to DMU i .

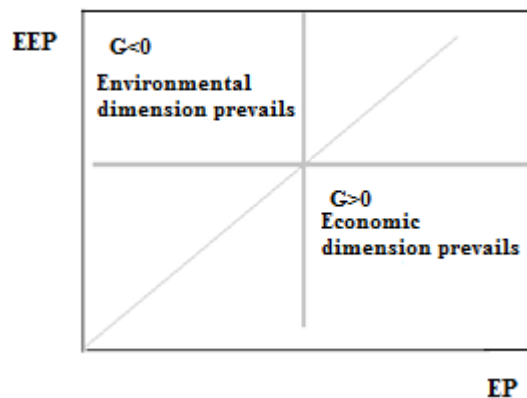
209 In the third step of analysis, in order to operate a ranking of the most virtuous municipalities in the
 210 environmental dimension, we define a measure of the efficiency deviation from environmental
 211 sustainability, labelled G. G is given by the difference between the Model I and Model II efficiency
 212 scores. This gap occurs when the environmental dimension is neglected in favor of greater economic
 213 efficiency (expenditure reduction while environmental damage occurs).

214 **Figure 2** shows the deviation box. The vertical axis measures the economic performance achieved
 215 without damage to the environment while the horizontal axis measures the performance of the
 216 municipalities in the waste management and disposal process in the absence of environmental
 217 compliance assessments. There are four possible outcomes. In the first quadrant, there will be the most

218 virtuous municipalities: the environmental performance exceeds the economic one ($G < 0$); The
 219 municipalities in the second quadrant will have a high environmental performance and a high economic
 220 performance. The municipalities above the bisector, however, will have an environmental performance
 221 higher than the economic one ($G < 0$); while for the municipalities below, the bisector economic
 222 dimension will prevail ($G > 0$). The municipalities of the third quadrant will be characterized by a low
 223 environmental and economic performance. The prevalence of one or the other dimension depends on the
 224 position with respect to the bisector. In the fourth quadrant G , it is greater than zero and the municipalities
 225 reach high levels of economic performance but with a positive gap of environmental inefficiency.

226

227 **Figure 2. Deviation from environmental efficiency Box**



228

229 *Note: environmental inefficiency gap (G) = Scores mod. I - Scores mod. II*

230

231

232 *2.2 The case study*

233

234 Campania is a predominantly hilly and mountainous region and the plain part is mainly on the coast.
 235 The population is concentrated in the plains and hills. The main characteristics of this area are: the very
 236 high population density (427 inhabitants/sq. Km) first among the Italian regions, the very high
 237 concentration of population in the metropolitan city of Naples (about 53% of the regional population).
 238 The degree of urbanization is very low in over half of the municipalities, while almost 60% of the
 239 population lives in those with a high degree of urbanization (about 15.5%). With regard to the
 240 administrative structure in the Region, there are mainly medium-small municipalities (84.5% below

241 15,000 inhabitants) where 31% of the population lives; in addition, there are few municipalities with over
242 50,000 inhabitants where 38% of the population lives.

243 The Region produces about 2.6 million tons of waste per year (above the national average), with a per
244 capita level of 450 kg/inhabitant/year. The organization and management of waste collection, disposal
245 and recovery services and the collection of related taxes are functions entrusted to the municipalities.

246 However, in order to avoid fragmentation in service performance and achieve an optimal operational
247 scale, art. 199 of ‘Consolidated Environmental Law’ provides that the Municipalities have to exercise
248 their urban waste management competences in a collective form, on the basis of Optimal Territorial
249 Areas (ATO). Regional law n.14 of 2016 defined a structure divided into seven ATOs, four of which
250 correspond to the provincial (NUTS-3) territories (Avellino, Benevento, Caserta and Salerno) and three
251 within the territory of the Metropolitan City of Naples. Each ATO can be divided into homogeneous
252 areas called Sub-District Areas (SAD) within which the municipalities can regulate their relations
253 through agreements. With regard to the size and characteristics of the ATOs, it should be noted that those
254 of a provincial dimension, despite having a territorial extension and a number of municipalities greater
255 than the sub-provincial ones (in the metropolitan cities of Naples), nevertheless have a lower population.
256 Until the full operation of the local authorities, the municipalities can manage and decides how to contract
257 their services. For this reason, the analysis considers municipalities as DMUs.

258 The urban waste management system can be divided into various heterogeneous phases with their
259 own characteristics: collection and transport, treatment, energy recovery/recycling of materials/disposal.
260 The phase of waste collection and transport is characterized by the prevalence of the labor over capital
261 input and, therefore, by a limited incidence of *sunk costs*. The treatment phase is characterized by the
262 prevalence of the capital on labor input, by the high initial investment costs (sunk cost) and the long
263 amortization periods. The disposal phase is the last phase of the waste cycle. This is since the waste
264 management hierarchy set by the European Commission establishes that the use of the landfill should
265 only affect materials that cannot otherwise be exploited. The waste remaining after the treatment phase
266 is sent to the landfill.

267 **Table 1** shows the descriptive statistics of the variables used in the analysis. The data are provided by
268 ISPRA (The Institute for Environmental Protection and Research) and refers to 353 municipalities
269 (NUTS-4) in the Campania region for the year 2016.

270 The total cost (TC) of the municipal solid waste management service is on average 161.395 euros per
271 inhabitants and includes the following cost items, as set by the provisions of Presidential Decree 158/99

272 “Regulation containing rules for the development of the standardized method for defining the tariff for
 273 the urban waste management service urban”:

274 - *Operating costs of services regarding unsorted municipal waste (UWC)* is on average 60.580euros
 275 per inhabitants and includes four components: street sweeping and washing costs; collection and
 276 transport costs; treatment and disposal costs; other costs, inherent to the management of unsorted
 277 municipal waste, not included in the previous items.

278 - *Costs of managing separated waste collection cycle (SWC)* is on average 58.361 euros per
 279 inhabitants and includes: costs of separated collection of individual materials; treatment and recycling
 280 costs, net of proceeds from the sale of materials and recovered energy and CONAI contributions.

281 - *Common costs and capital costs (KC)* that include the following expenditures: administrative costs
 282 of the assessment, collection and litigation; general management costs; depreciation of machinery for
 283 collection, sweeping, collection containers, financial depreciation; return on capital (R).

284 The average per capita cost for the management of unsorted waste (on average 60.580 euros per
 285 inhabitant) is higher than the average per capita cost of services linked exclusively to separate waste
 286 collection (on average 58.361 euros per inhabitant).

287 SC (on average 225.546 kg per inhabitant) and WS (on average 143.663 kg per inhabitants) denote,
 288 respectively, urban waste separately collected and unsorted urban waste collected variables. Art. 183 of
 289 Legislative degree 152/06 (paragraph f) provides a definition of separate collection. It is defined as:
 290 “collection which aims to: i) group urban waste into homogeneous categories; ii) group packaging waste
 291 materials separately from other waste. Finally, separate collection must be performed according to
 292 cheapness, efficiency, transparency and efficiency criteria”.

293

294 **Table 1. Summary statistics**

Variables	Minimum	Mean	Maximum	Standard	
				Deviation	Median
<i>Inputs</i>					
UWC	1.73	60.580	277.07	36.547	54
SWC	0.95	58.361	207.76	35.155	56.74
KC	0.01	42.45	277.39	43.073	30.72
TC	34.66	161.395	522.23	64.71314	150.12
<i>Outputs</i>					
SW	6.508	225.546	609.248	83.11064	213.9122
UW	4.798	143.663	505.99	78.25892	123.1645

TW	137.581	369.429	1008.18	117.3605	354.9546
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Source: Elaboration of ISPRA 2016 data

Notes: all costs are expressed in euros per inhabitant; UW and SW are expressed in kg per inhabitant.

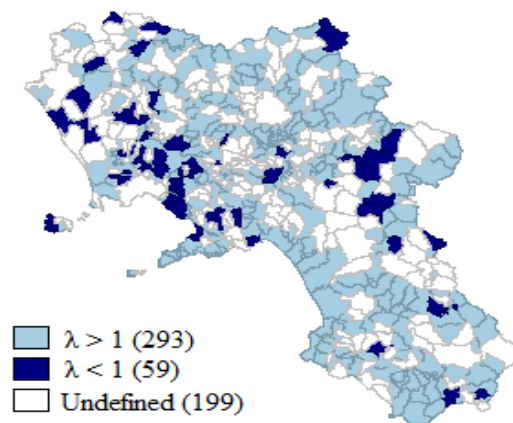
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296
297

298 In order to have additional information on the levels of UW and SW in each municipality, we define
299 a measure of environmental tightness in the management's urban waste services as the ratio of SW to
300 UW, λ labelled. This tightness may affect the performance of the production process and, therefore, the
301 results of efficiency deviation from the environmental sustainability i.e., to have or not to have a gap in
302 terms of environmental efficiency. This is because separate collection is a precondition for recycling (as
303 set by the European and Italian normative frameworks). On the contrary, the quantity of unsorted waste
304 (which cannot be recycled) ends up in landfills. If $\lambda < 1$ the amount of separate collection per capita is
305 less than the amount of unsorted waste per capita and it means that the process is characterized by a high
306 environmental tightness (high risk of environmental unsustainability). For municipalities with a value of
307 λ less than one, the probability of being placed in the third or fourth quadrant of *deviation from*
308 *environmental efficiency Box (Figure 2)* is high. Lower levels of λ make it more difficult to achieve
309 environmental sustainability by local government and implies negative externalities. Conversely, if λ is
310 above the unit, the municipal waste management should be performer better in terms of environmental
311 sustainability.

312 **Figure 3** shows that only 56 municipalities have values of λ less than one i.e., the amount of per capita
313 separate collection is less than the amount of unsorted waste per capita. These municipalities are mostly
314 concentrated in the Neapolitan and Caserta area and only a few municipalities are in the Avellino,
315 Salerno, Benevento area.

316

Figure 3. Box map of environmental tightness



317
318

Source: Elaboration of ISPRA 2016 data

319

320 3. Results and discussions

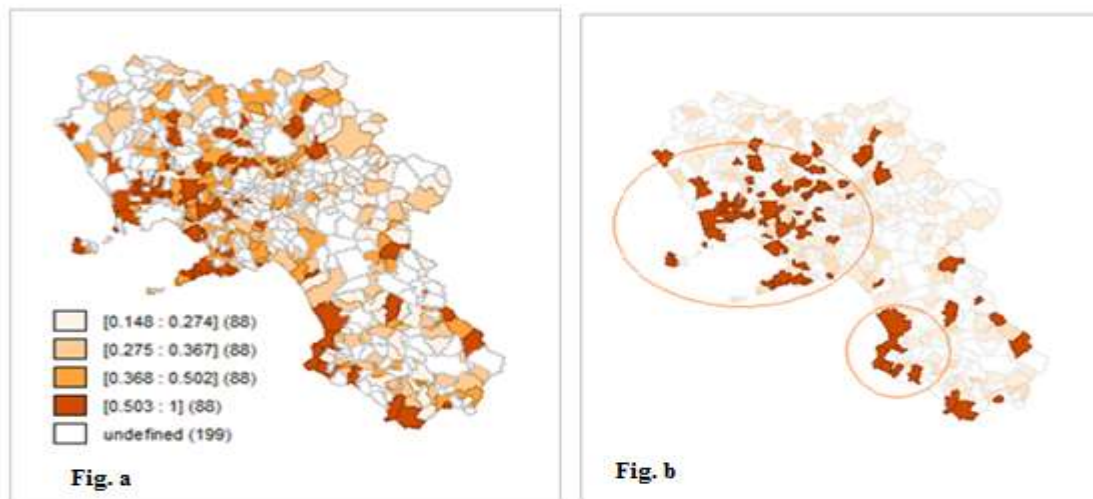
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322 **Figure 4** shows the efficiency scores of municipalities in the production process of the waste
323 collection and disposal services from Model I. The box map (**Figure 4a**) shows the distribution of the
324 performance considering only the economic dimension. The map features heterogeneity among the
325 municipalities. However, a small cluster of more efficient municipalities emerges in the area between the
326 Neapolitan and Caserta areas and a small cluster of municipalities on the Salerno coast (with the scores
327 from 0.5 to 1, **Figure 4b**). The municipalities of these latter areas, on equal output levels (disposed
328 waste), use a smaller amount of expenditure to dispose of it. However, this performance analysis does
329 not consider any damage caused to the environment due to environmental unsustainability of the waste
330 services production process. The results of the performance analysis that includes not only the economic
331 dimension but also the environmental one are presented in **Figure 5**.

332

333

Figure 4. Box map of Economic efficiency scores



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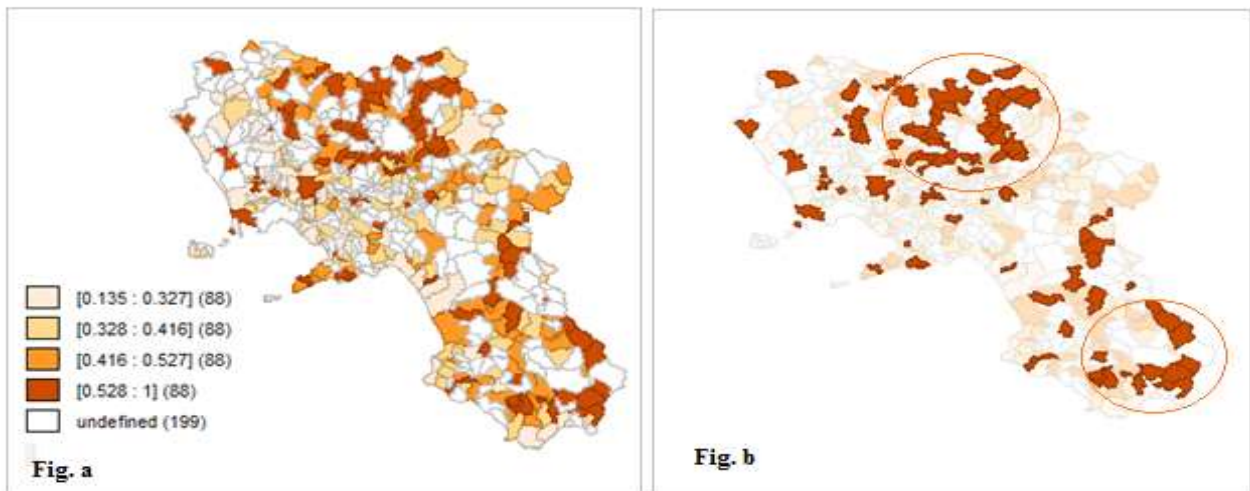
Source: Elaboration of ISPRA 2016 data

337 The box map (**Figure 5a**) shows the distribution of the efficiency scores taking into account
338 environmental sustainability i.e., when local governments are efficient from economic perspective but
339 without damaging the environment. The municipalities with a high score are those that in the production
340 process of waste collection and disposal services use fewer resources and minimize the quantity of

341 unsorted disposed waste: minimization of the spending, while reducing environment damage. **Figure 5b**
342 shows two small clusters of municipalities with a better performance located in the Benevento area and
343 in the Salerno area (with the scores in the rage 0.5-1).

344

345 **Figure 5. Box map of Economic and environmental efficiency scores**



346

347 *Source: Elaboration of ISPRA 2016 data*

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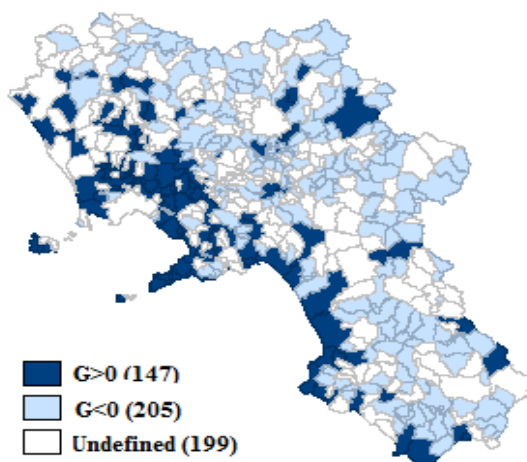
349 As introduced in Section 2.1, in order to operate a ranking of the most virtuous municipalities in the
350 environmental dimension, we define a measure of the deviation (or gap) of efficiency from environmental
351 sustainability, G . It is given by the difference between the Model I and Model II efficiency scores. This
352 gap occurs when the environmental dimension is neglected in favor of greater economic efficiency: cost
353 savings but with damaging the environment while producing and disposing of a high quantity of unsorted
354 waste. If the deviation from environmental sustainability, G , is negative, the municipality has a good
355 economic and environmental performance (205 municipalities in light blue, fig. 5); if G is positive (147
356 municipalities in blue, fig. 5) the municipality is efficient from an economic point of view but to the
357 detriment of environmental sustainability which would require a reduction of undesirable output. Figure
358 5 shows the distribution of G in Italian municipalities. **Figure 6** shows two clusters of municipalities that
359 have G values greater than zero: one in the Neapolitan area (that also includes some municipalities in the
360 Caserta area) and another along the Salerno coast. **Figure 7** presents the deviation box in order to have
361 more information on the placement of the municipalities in the box. The municipalities in the first
362 quadrant are the ‘most virtuous’, while the municipalities classified as the ‘worst’ from an environmental

363 point of view are located in the fourth quadrant. The latter, although they have a high economic
364 performance, present a high environmental unsustainability from the production process (such as
365 Frattamaggiore; Villaricca; Gricignano di Aversa Lacco Ameno; etc..).

366

367 **Figure 6. Box map of deviation from environmental sustainability**

368

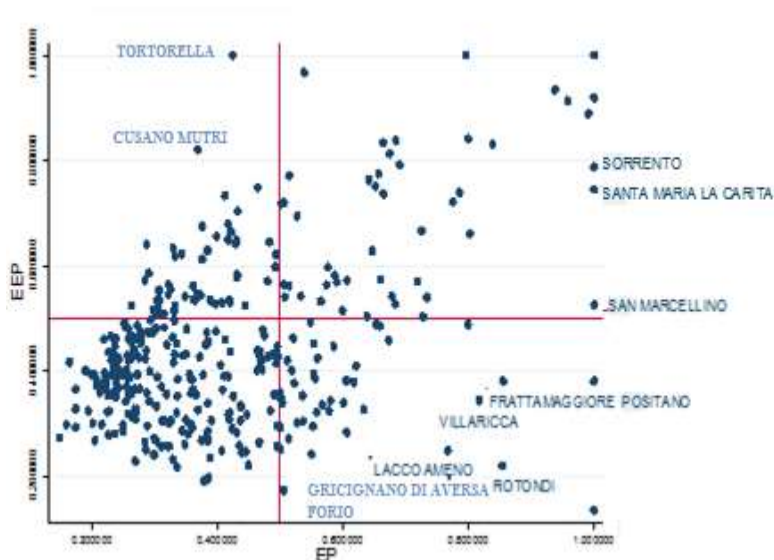


369

370 *Source: Elaboration of ISPRA 2016 data*

371

372 **Figure 7. Deviation box**



373

374 *Source: Elaboration of ISPRA 2016 data*

375 The environmental efficiency gap in these areas (especially in the cluster of the Caserta and Naples
376 area), other factors being equal, depend on the levels of unsorted waste that should be reduced to improve

377 environmental performance. As shown in **Figure 3**, these municipalities and especially those in blue in
378 the Neapolitan area have a waste services production process characterized by environmental tightness
379 (λ). On the other hand, these areas were also those most affected by the waste emergency in Campania.
380 The huge amounts of waste on the streets of Naples have gone around the world (see for example The
381 Economist, 2009; New York Times, 2007). There is a mix of public (technical and administrative) errors
382 combined with private interests (political, industrial and criminal) that can be partially identified: delays
383 in the design and preparation of suitable landfills; delays in the design and construction of incinerators
384 and composting plants for the organic fraction of waste; low levels of separate waste collection (Armiero
385 and D'Alisa, 2012; D'Alisa and Kallis, 2016). In addition, the poor waste management is linked to a low
386 institutional quality and to the Kapp theory of social cost (see also Berger, 2008; Cecere et al. 2014;
387 Garofalo et al., 2019; Agovino et. al., 2019c).

388 Achieving a high level of SWC requires significant efforts, which can only be achieved through the
389 joint action of citizens and institutions (Agovino et al., 2018). Missed objective in Campania during the
390 period of the waste crisis, due to the declaration of a state of emergency between 1994 and 2009 (Armiero
391 and Fava, 2016). During this period, the state was the only active agent in the waste management process,
392 excluding the active participation of citizens. Citizens organized themselves into associations and
393 protested, risking imprisonment to be heard by central or local administrations, which continued to
394 exclude them from decision-making processes (D'Alisa et al., 2010; Caggiano and De Rosa, 2015;
395 Armiero, 2014; Martinez-Alier et al., 2016; Lucchini and Membretti, 2016; Armiero and D'Alisa, 2012).
396 As a consequence of this state of emergency, a collaboration between citizens and local authorities has
397 been established in Campania, resulting in better waste management and an increase in the regional level
398 of SWC (Agovino et al., 2018). However, differences remain between the municipalities in the region,
399 with many areas ranking at the bottom of the distribution (Garofalo et al., 2019).

400 Among the factors that could explain the deviation highlighted in these areas, there is the cost to
401 dispose of waste in landfills, that continues to be very low (110 euros per ton of waste disposed) when
402 compared to the damage caused to the environment. A too low cost does not discourage the dumping of
403 waste, on the contrary, this makes it more convenient from an economic point of view (a lower disposal
404 cost). In Italy, 383 landfills are currently active where almost 20 million tons of urban and special waste
405 have been disposed (Ispra, 2017). The best (most virtuous) situations in terms of environmental
406 sustainability in waste management (Veneto, Piedmont and Sardinia region) are often present where the
407 cost of landfill disposal is higher. In the case of the Campania region, the low costs associated with the
408 disposal of municipal waste in landfills have on the one hand, discouraged the creation of sustainable

409 methods (of collection and disposal) from an environmental point of view, thus favoring the excessive
410 use of landfills. This is also due to the failure to start a waste cycle, which has implied that landfills have
411 become a central node in the disposal of waste in Campania rather than an accessory element.

412 The sequence of national and regional regulations, that have shifted tasks between the authorities
413 (region, provinces and municipalities), has further accentuated the uncertainty of both the executive and
414 management of the waste cycle process. This has generated both confusion in the preparation of local
415 administrative interventions as well as a loss of citizens trust in the public institutions. This uncertainty
416 of competence has led to a regulatory gap on the central issues relating to the starting of an integrated
417 cycle, creating the conditions, on the one hand, for the exploitation of public functions for clientelist
418 purposes and, on the other, for the infiltration of organized crime. D'Alisa et al. (2010) argued that
419 Campania's waste problem cannot be analyzed as one of simple waste mismanagement. The latter spoke
420 of a "crisis of democracy" in the waste management process that generated conflict between citizens and
421 government. Integrated waste management is only achieved through the joint action of citizens and
422 institutions (central and local) (Agovino et al., 2018).

423 In the light of Directive 2008/98/EC and Italian L.D. 152/2006, evaluating public service provision
424 performance is a crucial element to improving any service's efficiency (Nogueira and Jorge, 2016).
425 Efficient management entails identifying and measuring the performance of the various parts operating
426 in a particular sector (Pires et al., 2018). The control and monitoring of these elements are vital for the
427 intelligent use of resources. In recent decades, numerous studies have attempted to analyze efficiency in
428 waste service provision (Romano et al., 2020). The type of management may vary with the region's
429 characteristics, so that to improve efficiency, the kind of management should be appropriate to the area
430 (Yang et al., 2018). Bosch et al., 2000, Dijkgraaf and Gradus, 2007, the competitive environment in
431 which the service is provided may be more critical than the dichotomy between public-private
432 management.

433 An important role to improve environmental performance and to achieve behavioral models consistent
434 with the objectives set by legislation is also played by well-designed incentive system. This should
435 guarantee various principle as the full cost recovery and the polluter pays' principle. For instance, a
436 quantity-based price system (related to volume, weight, or frequency as Pay-As-You-Throw schemes,
437 PAYT) acting as a Pigouvian tax could be establish a negative incentive approach (discouraging citizens
438 from producing waste) (Thøgersen 2003; Acuf and Kaffi, 2013; Buccioli et al., 2015). However, despite
439 the continuous normative changes, the financing of urban waste services, in Campania, are not directly
440 related to the services provided to each citizen (user-payer). If not well-designed, a system could not only

441 rise late payment phenomenon while endangering the fully costs recover (as in the flat-rate system, see
442 Bilitewski, 2008), but it could also not favorite the achievement of environmental sustainability goal
443 (waste reduction) undermining environmental performance. In particular, at national level, there are
444 24.43% of uncollected percentages, and there is an average rate of 43% in the South, 15% in the Center
445 and 19% in the North (CRIF Ratings, 2020). The late payment phenomenon also affects the timing of
446 payments from local authorities to companies to which the waste services are outsourced. The
447 outsourcing of services is mostly entrusted to in-house companies (71% of municipalities). They mainly
448 concern larger municipalities (such as Naples) and municipalities in the province of Avellino, where
449 integrated waste management is carried out by an in-house company (*Irpiniambiente SpA*).

450 Moreover, the level of citizen efforts required to sort their waste can depend on the type of program
451 operated by local institutions. If the local administration does not provide an easy separate collection
452 system (e.g. door-to-door collection), the level of effort required to recycle is high and only individuals
453 with intrinsic motivation and strong pro-environmental attitudes will be likely to recycle (e.g. taking up
454 valuable storage space) (Agovino et al., 2018). In addition to the benefits for the environment, high levels
455 of recycling can give tax benefits to local governments which are responsible for the waste services
456 production process. For example, in Veneto art. 39 of Regional Law 3 of 21 January 2000 introduced tax
457 benefits for municipalities that were able to achieve certain separate collection objectives. This benefit
458 resulted in a reduction in a tax for disposing of waste in landfills. Nationwide, law of 28 December 2015,
459 n. 221 (Related Environmental art.32 and 34 - entered into force on February 2, 2016), has made some
460 changes to Legislative Decree 152/2006 (art.205), providing for the application of a 20% increase of
461 Eco-Tax to municipalities that have not reached the legal minimum of sorted waste. Conversely, a
462 reduction in the tax is envisaged for the cases where the minimum level of separate waste collection
463 exceeds the target. Specifically, while on one hand, L.D. 152/2006 has favored the growth of SWC, on
464 the other hand it has increased the disparities among Italian regions. The main problem is related to the
465 fact that L.D. 152/2006 defined higher thresholds, without indicating the path to achieve them, leaving
466 this problem to the regions; moreover, L.D. 152/2006 does not take into account the economic,
467 infrastructural, institutional, social, and demographic disparities that characterize Italian regions. Only
468 recently L.D. 152/2006 was amended, with the definition of penalties (paragraph 3 of article 152) and
469 incentives (paragraph 3 bis of article 152). In particular, in the case of penalties, an additional 20% is
470 defined on the tribute for the delivery of waste to landfills for administrations that have not reached the
471 percentages set by L.D. 152/2006. These penalties have the objective of pushing up SWC rates in
472 Southern regions, characterised by a high use of landfills. Furthermore, L.D. 152/2006 does not consider

473 the historical dif- ferences existing between Italian regions and the possible causes that generate delays
474 in these regions. In this regard, the Green Book (AA.VV, 2018) shows that in Northern Italy 69% of the
475 waste generated is burned to produce energy, while in the South 62% ends up in landfills, due the low
476 cost of land; the latter result has the effect of reducing the percentage of SWC in the regions of Southern
477 Italy. In general, the legislation ends up penalising the regions of Southern Italy, which on average
478 continue to record low SWC percentages (38% against 64% of Northern Italy) and to favour the regions
479 of Northern Italy. In this way, the gap between Northern and Southern Italy will continue to widen.
480 National legislation must consider territorial differences and above all the socio- economic characteristics
481 of the regions (Agovino et al., 2021).

482

483 **4. Conclusion**

484

485 The study analyzed the economic and environmental performance of the 353 municipalities in
486 Campania (NUTS-4) in the waste disposal and collection services. The analysis highlighted the
487 importance of jointly assessing the economic and environmental dimensions of the production process
488 relating to the municipal solid waste management service. A production process that ensures the
489 expenditure minimization does not necessarily translate into a system that protects the environment. The
490 analysis shows how the municipalities with good economic performance are the less virtuous with a high
491 efficiency deviation from environmental sustainability. Numerous factors contribute to explaining these
492 results recorded in these areas: waste emergency; delay in starting an integrated waste cycle; a low
493 Ecological Tax; the need for a negative economic incentive system such as PAYT. The latter, in addition
494 to being able to achieve economic sustainability of the waste services process (cost recovery), can favor,
495 as a Pigouvian tax, an allocative efficiency in the waste production. It is also important to underline how
496 good environmental performance achieved through the separate collection of recyclable materials and
497 the reduction of unsorted waste (for the intermediate ‘recycling’ target and the ultimate goal of ‘zero
498 waste’) can also guarantee a good economic performance. Recycling allows to obtain a double efficiency:
499 *Economic efficiency*: waste recycling is the process of transforming waste into reusable materials; an
500 intelligent waste disposal system is a way to reduce energy consumption and industrial costs;
501 *Environmental efficiency*: recycling is the most advantageous alternative to conventional waste disposal
502 systems, which in addition to being no longer sufficient to dispose of the ever increasing load of waste
503 produced, have a non-negligible environmental impact.

504

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