# Aboveground biomass and diversity of woody climber in evergreen forest, southern Vietnam 

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#### Abstract

Woody climbers play an important role in overall plant diversity and carbon storage in a tropical forest. This study aimed at investigating woody climbers in tropical evergreen forests, in southern Vietnam. Forty-four survey plots of $2,500 \mathrm{~m}^{2}$ each were used for field data collection in four forest types experiencing different disturbances as rich (standing volume/SV>200 $\mathrm{m}^{3} / \mathrm{ha}$ ), medium ( $100<\mathrm{SV} \leq 200 \mathrm{~m}^{3} / \mathrm{ha}$, poor ( $50<\mathrm{SV} \leq 100 \mathrm{~m}^{3} / \mathrm{ha}$ ), and very poor ( $10<\mathrm{SV} \leq 50 \mathrm{~m}^{3} / \mathrm{ha}$ ) forests. Twenty-six woody climbers were recorded with 17 species in rich, 20 in medium, 10 in poor, and 13 in very poor forests. Two species that appeared in poor and very poor forests are missing in rich and medium forests. The diversity Shannon-Weiner index $\left(H^{\prime}\right)$ and evenness $(J)$ indices are in the order of reducing from rich forest ( $H^{\prime}=2.37, J=0.83$ ) to very poor forest ( $H^{\prime}=0.92, J=0.57$ ). Stem density and aboveground biomass (AGB) of woody climbers were significantly different among four forest types; 876 stems/ha in the rich forest, 1,246 stems/ha in the medium forest, $785 \mathrm{stems} / \mathrm{ha}$ in the poor forest, and $1,991 \mathrm{stems} / \mathrm{ha}$ in the very poor forest; AGB was $5,570 \mathrm{~kg} / \mathrm{ha}$ in the rich forest, $9,444 \mathrm{~kg} / \mathrm{ha}$ in the medium forest, $3,573 \mathrm{~kg} / \mathrm{ha}$ in the poor forest, and $20,560 \mathrm{~kg} / \mathrm{ha}$ in the very poor forest. It is concluded that previous disturbances significantly changed the diversity and structure of woody climbers by reducing diversity but increasing stem density and AGB in the forest with higher intensity of disturbance.


KEYWORDS: Carbon storage, Disturbance regime, Forest structure, Host-plant, Life-form.

## 1. INTRODUCTION

Woody climber (liana) is a prominent life-form and contributes to overall plant diversity occupying an important niche in the tropical forest ecosystem around the world [1-4]. These woody plants germinate at the ground, root permanently in the soil, then climb host trees to reach the forest canopy [5]. Woody climbers account for $10-45 \%$ of total woody stems, up to $35 \%$ species diversity [6], [7], and $10-30 \%$ total aboveground biomass (AGB) in tropical forests [8]. In a heavily disturbed natural forest, woody climbers can rapidly colonize the open habitats [9]. Woody climbers depend on trees for their growth and compete, hosts, both above and belowground for resources. Therefore, hosts with a heavy load of woody climbers often show decreased growth, increased mortality, and reduced biomass [10], [11]. However, woody climbers also provide a valuable food source for animals [6], increase community stability by physically linking trees together, and provide access for animals [3].

The abundance, diversity, and distribution of woody climbers are affected by several abiotic and biotic factors such as rainfall, seasonality, soil fertility, forest canopy structure, and disturbance regimes [12-14]. Their abundance and biomass are increasing across tropics as a result of forest fragmentation, natural and anthropogenic disturbances, and global environmental changes [3], [15-17]. These changes will consequently influence forest structure, composition, and carbon sequestration [18]. Studies on woody climber communities have been mostly neglected in Vietnam, because of their low commercial valuables compared to timber. However, they still contribute a significant role in biomass, carbon storage, plant diversity, and stability in the
forest ecosystem. Hence, this study aimed at investigating woody climber AGB, structure, and diversity in tropical evergreen forests, southern Vietnam.

## 2. MATERIAL AND METHOD

### 2.1 Study site

The study was conducted at Dong Nai Biosphere Reserve (DNBR), southern Vietnam (Figure 1A). The core zone of DNBR covers an area of 172,502 ha. The study area is characterized as a tropical monsoon climate. There are two distinct seasons with a rainy season during April-November and a dry season during DecemberMarch. The site has a mean annual rainfall of $2,450 \mathrm{~mm}$ mostly falling during August-September and a mean annual temperature of $25.4^{\circ} \mathrm{C}$. Acrisols and Luvisols dominate DNBR. Due to intense rainfall during the rainy season, many small areas in DNBR are periodically inundated for a long time of up to several months, which significantly influences the diversity and growth of plant communities [19].


Figure 1. Map of the study area (A) and plot layout (B)

Vegetation in the study site was described as lowland tropical forest dominated by dipterocarps [20], [21]. Most of the forests in DNBR were degraded by selective logging and war during 1961-1996 [22], [23]. Most stems of commercially valuable species (Anisoptera costata, Dipterocarpus alatus, Dipterocarpus dyeri, Heriteria cochinchinensis, Dipterocarpus costatus, Hopea odorata, Lagerstroemia calyculata, Shorea roxburghii, and Sindora cochinchinensis) with diameter at breast height larger than 40 cm were cut.

### 2.2 Data collection

Based on standing volume (SV), forests in the study site were classified to four types as a rich forest with SV $>200 \mathrm{~m}^{3} / \mathrm{ha}$, medium forest with $100<\mathrm{SV} \leq 200 \mathrm{~m}^{3} / \mathrm{ha}$, poor forest with $50<\mathrm{SV} \leq 100 \mathrm{~m}^{3} / \mathrm{ha}$, and very poor forest with $10<\mathrm{SV} \leq 50 \mathrm{~m}^{3} / \mathrm{ha}$. In each forest type, a number of plots were established for data collection (Table 1). The position for the setting plot was selected typically by using a forest cover map.

The main plot of $2.000 \mathrm{~m}^{2}(50 \mathrm{~m} \times 20 \mathrm{~m})$ was used (Figure 1B). In the main plot, all woody climbers with DBH (diameter at 1.3 m from the ground) $\geq 10 \mathrm{~cm}$ were identified to species level and measured for DBH. In the center of the main plot, a sub-first plot of $200 \mathrm{~m}^{2}(20 \mathrm{~m} \times 10 \mathrm{~m})$ was set up for identifying species and measuring DBH of all woody climbers with DBH $>5 \mathrm{~cm}$ and $<10 \mathrm{~cm}$. In the central of sub-first plot, a subsecond plot of $25 \mathrm{~m}^{2}(5 \mathrm{~m} \times 5 \mathrm{~m})$ was set up for identifying species and measuring DBH of all woody climbers with $\mathrm{DBH} \leq 5 \mathrm{~cm}$.

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### 2.3 Data analysis

### 2.3.1 Estimation

Species diversity was estimated by Shannon - Weiner index ( $H^{\prime}$ ) [24] and Evenness index (J) [25]. These indices were calculated by Eq. (1) and Eq. (2).

$$
\begin{align*}
& H^{\prime}=-\sum_{i=1}^{S} p_{i} \cdot \ln p_{i}  \tag{1}\\
& J=\frac{H^{\prime}}{\ln (s)} \tag{2}
\end{align*}
$$

where, $s$ is the number of species in the plot and $p_{i}$ is the proportion of individuals belonging to the $i^{\text {th }}$ species. Dry AGB of each woody climber [26] was estimated by Eq. (3).

$$
\mathrm{AGB}=\exp [-1.347+2.391 \cdot \operatorname{Ln}(\mathrm{DBH})] \quad(3)
$$

where, DBH is the diameter at 1.3 m from ground in centimeter and AGB is dry aboveground biomass in kilogram.

Stand parameters including diversity indices, AGB, mean DBH, basal area, stem density were separately calculated for each survey plot among four forest types.

### 2.3.2 Statistical Analysis

The difference in each concerned parameter among four forest types was assessed by univariate analysis of variance (ANOVA) and post-hoc test. All analyses were conducted using SAS 9.2 at $p=0.05$ (SAS Institute Inc., Cary, NC, USA).

## 3. RESULTS

Twenty-six species belonging to 16 families were recorded in 44 survey plots. Of which, 17 species appeared in the rich forest, 20 species in the medium forest, 10 species in the poor forest, and 13 species in the very poor forest (Table 1). The most abundance species is Bauhinia cardinalis with 34 stems in very poor forest and 14 stems in the poor forest. Nine species appear in only one forest type, 6 species in two forest types, 3 species in three forest types, and 7 species appear in all four forest types. Two species (Tetrastigma quadrangulum and Acacia comosa) appearing in poor and very poor forests are missing in rich and medium forests. While 11 species appearing in rich and medium forests are missing in poor and very poor forests.

Table 1. The scientific name of woody climbers and their abundance found in 44 survey plots

| No | Scientific | Family | Stem number |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  | Rich | Medium | Poor | Very poor |
|  |  | Plot number | 17 | 14 | 9 | 4 |
| 1 | Abrus precatorius L | Fabaceae | 2 | 9 |  |  |
| 2 | Acacia comosa Gagnep. | Fabaceae |  |  |  | 4 |
| 3 | Acacia concinna (Willd.) DC | Fabaceae |  | 1 |  | 13 |
| 4 | Ancistrocladus tectorius (Lour.) Merr. | Ancistrocladaceae | 8 | 7 | 3 | 2 |
| 5 | Artabotrys intermedius Hassk. | Annonaceae | 5 | 8 |  |  |
| 6 | Bauhinia cardinalis Pierre ex Gagnep | Fabaceae | 4 | 8 | 14 | 34 |
| 7 | Caesalpinia minax Hance | Fabaceae |  |  |  | 8 |
| 8 | Calycopteris floribunda (Roxb.) Lamk. | Combretaceae | 1 | 8 |  | 1 |


| 9 | Combretum latifolium Blume | Combretaceae | 2 | 1 | 2 | 1 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 10 | Combretum tetralophum C. B. Clarke. | Combretaceae | 1 | 4 |  |  |
| 11 | Dalbergia Candenatensis (Dennst.) Prain | Fabaceae | 7 | 3 |  | 1 |
| 12 | Dalbergia curtisii Prain | Fabaceae | 3 | 2 |  |  |
| 13 | Entada pursaetha DC. | Fabaceae |  | 2 | 4 |  |
| 14 | Erycibe cochinchinensis Gagnep. | Convolvulaceae | 1 |  |  |  |
| 15 | Erycibe elliptilimba Merr. \& Chun | Convolvulaceae | 1 | 3 |  |  |
| 16 | Gnetum macrostachyum Hook.f. | Gnetaceae | 4 | 3 | 1 | 1 |
| 17 | Melodorum fruticosum Lour. | Annonaceae |  | 5 |  |  |
| 18 | Phytocrene oblonga Wall. | Icacinaceae |  | 1 |  |  |
| 19 | Sageretia theezans (L.) Brongn. | Rhamnaceae |  | 1 |  |  |
| 20 | Salacia chinensis L | Celastraceae |  | 2 |  |  |
| 21 | Sargentodoxa cuneata (Oliv.) Rehd. \& Wils | Sargentodoxaceae | 2 | 6 | 3 | 5 |
| 22 | Sphenodesme pentandra (Roxb.) Jack | Verbenaceae | 21 | 6 | 6 | 3 |
| 23 | Strychnos ignatii Bergius | Loganiaceae | 5 | 4 | 1 | 1 |
| 24 | Tetrastigma quadrangulum Gagnep. \& Craib. | Vitaceae |  |  | 1 |  |
| 25 | Uncaria acida (Hunt.) Roxb. | Rubiaceae | 2 |  |  |  |
| 26 |  | Annonaceae |  |  |  |  |
| 26 | Uvaria microcarpa Champ. ex Benth. |  | 5 |  | 1 | 1 |
|  | Species total |  | 17 | 20 | 10 | 13 |

Stand parameters of all woody stems are shown in Table 2. For stems with DBH of $\geq 10 \mathrm{~cm}$, density is significantly different among four forest types, where the highest density was found in the very poor forest, reducing to the medium forest, and poor and rich forests. The difference of DBH was not different. While the differences in basal area and AGB were significant among four forest types. The largest basal area was found in the very poor forest $\left(0.76 \mathrm{~m}^{2} / \mathrm{ha}\right)$. While, the difference among the three other forest types was not significant, which is around $0.08-0.15 \mathrm{~m}^{2} / \mathrm{ha}$. The highest AGB of $7,797 \mathrm{~kg} / \mathrm{ha}$ was found in the very poor forest, while in three other forest types AGB was much lower, ranging $728-1,339 \mathrm{~kg} / \mathrm{ha}$. Both $H^{\prime}$ and $J$ diversity indices, and species number were in order of reducing from rich forest ( $H^{\prime}=2.04, J=0.89$ ) to very poor forest ( $H^{\prime}=0.092, J=0.57$ ).

For stems with DBH $5 \mathrm{~cm}>$ and $<10 \mathrm{~cm}$, the differences of stem density, basal area, and AGB were also significantly different among the four forest types (Table 2). The highest density ( 350 stems/ha), AGB ( 9,303 $\mathrm{kg} / \mathrm{ha}$ ), and basal area ( $1.30 \mathrm{~m}^{2} / \mathrm{ha}$ ) were found in very poor forest and the lowest density ( $66 \mathrm{stems} / \mathrm{ha}$ ), AGB $(1,602 \mathrm{~kg} / \mathrm{ha})$, and basal area $\left(0.23 \mathrm{~m}^{2} / \mathrm{ha}\right)$ was found in poor forest. Similar to $\geq 10 \mathrm{~cm}$ DBH stems, both $H^{\prime}$ and $J$ diversity indices, and species number were in order of reducing from rich forest ( $H^{\prime}=2.37, J=0.90$ ) to very poor forest ( $H^{\prime}=1.35, J=0.84$ ).

For stems with $\mathrm{DBH} \leq 5 \mathrm{~cm}$, the difference was found only in stem density, which was highest in the very poor forest ( $1,600 \mathrm{stems} / \mathrm{ha}$ ), reducing to the medium forest ( $1,085 \mathrm{stems} / \mathrm{ha}$ ), and to rich ( $779 \mathrm{stems} / \mathrm{ha}$ ) and poor (711 stems/ha) forests (Table 2). $H^{\prime}$ diversity index was in the order of reducing from rich forest ( $H^{\prime}=$ 2.25 ) to very poor forest ( $H^{\prime}=2.07$ ), but $J$ diversity index ( $0.83-0.96$ ) was more or less similar in all four forest types, while species number was in the order of reducing from rich forest ( 15 species) to very poor forest (9 species).

The total density of woody climbers was 876 stems in the rich forest, 1,246 stems in the medium forest, 785 stems in the poor forest, and 1,991 stems in the very poor forest. While, total AGB was $5,570 \mathrm{~kg} / \mathrm{ha}$ in rich forest, $9,444 \mathrm{~kg} / \mathrm{ha}$ in medium forest, $3,573 \mathrm{~kg} / \mathrm{ha}$ in poor forest, and $20,560 \mathrm{~kg} / \mathrm{ha}$ in very poor forest.

Table 2. Stand parameters of the woody stem in four different forest types

| DBH | Parameter | Forest type |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | Rich | Medium | Poor | Very poor |
| $\underset{\text { cm }}{\geq 10}$ | Density (stems/ha) | $6 \pm 2^{\text {a }}$ | $11 \pm 2^{\text {b }}$ | $8 \pm 5^{\text {ab }}$ | $41 \pm 18^{\text {c }}$ |
|  | DBH $\pm$ SE (cm) | $12.7 \pm 0.7$ | $12.7 \pm 0.5$ | $11.7 \pm 0.3$ | $14.7 \pm 0.5$ |
|  | Basal area ( $\mathrm{m}^{2} / \mathrm{ha}$ ) | $0.08 \pm 0.02^{\text {a }}$ | $0.15 \pm 0.03^{\text {b }}$ | $0.10 \pm 0.06{ }^{\text {ab }}$ | $0.76 \pm 0.31^{\text {c }}$ |
|  | Species number | 10 | 14 | 6 | 4 |
|  | $H^{\prime}$ | 2.04 | 2.34 | 1.41 | 0.92 |
|  | $J$ | 0.89 | 0.85 | 0.73 | 0.57 |
|  | AGB (kg/ha) | $728.0 \pm 210.9^{\text {a }}$ | $1,339.6 \pm 336.9^{\text {b }}$ | $883.7 \pm 569.8^{\text {a }}$ | 7,797.5 $\pm 3,186.2^{\text {c }}$ |
| $\begin{gathered} 5 \mathrm{~cm} \\ >\text { and } \\ <10 \\ \mathrm{~cm} \end{gathered}$ | Density (stems/ha) | $91 \pm 23^{\text {a }}$ | $150 \pm 37^{\text {b }}$ | $66 \pm 49^{\text {c }}$ | $350 \pm 192^{\text {d }}$ |
|  | DBH $\pm$ SE (cm) | $6.3 \pm 0.3$ | $6.4 \pm 0.2$ | $7.2 \pm 0.8$ | $6.7 \pm 0.1$ |
|  | $\begin{aligned} & \text { Basal area } \\ & \left(\mathrm{m}^{2} / \mathrm{ha}\right) \end{aligned}$ | $0.32 \pm 0.09^{\text {a }}$ | $0.49 \pm 0.12^{\text {b }}$ | $0.23 \pm 0.15^{\text {a }}$ | $1.30 \pm 0.70^{\text {c }}$ |
|  | Species number | 14 | 16 | 4 | 5 |
|  | $H^{\prime}$ | 2.37 | 2.57 | 1.13 | 1.35 |
|  | $J$ | 0.90 | 0.93 | 0.70 | 0.84 |
|  | AGB (kg/ha) | $2,261.1 \pm 658.3^{\text {a }}$ | $3,435.5 \pm 846.2^{\text {b }}$ | $1,602.9 \pm 1,068.8^{\text {a }}$ | 9,303.2 $\pm 5,001.4^{\text {c }}$ |
| $\begin{aligned} & \leq 5 \\ & \mathrm{~cm} \end{aligned}$ | Density (stems/ha) | $779 \pm 266^{\text {a }}$ | $1,085 \pm 279^{\text {b }}$ | $711 \pm 218^{\text {a }}$ | $1,600 \pm 432^{\text {c }}$ |
|  | DBH $\pm$ SE (cm) | $2.3 \pm 0.4$ | $2.8 \pm 0.2$ | $1.7 \pm 0.3$ | $2.4 \pm 0.1$ |
|  | Basal area ( $\mathrm{m}^{2} / \mathrm{ha}$ ) | $0.50 \pm 0.19$ | $0.85 \pm 0.30$ | $0.24 \pm 0.08$ | $0.74 \pm 0.21$ |
|  | Species number | 15 | 11 | 7 | 9 |
|  | H' | 2.25 | 2.30 | 1.82 | 2.07 |
|  | $J$ | 0.83 | 0.96 | 0.94 | 0.94 |
|  | AGB (kg/ha) | $2,581.1 \pm 1,022.3$ | 4,669.2 $\pm 1,739.8$ | 1,086.8 $\pm 385.6$ | 3,462.1 $\pm 976.2$ |

Different letters $\mathrm{a}, \mathrm{b}, \mathrm{c}, \mathrm{d}$, in a row indicate significant differences of means at $p=0.05$. AGB is dry aboveground biomass, DBH is stem diameter at 1.3 m from the ground, $H^{\prime}$ is Shannon - Weiner index, $J$ is Evenness index.

In poor and very poor forests, the highest stem number belonged to the DBH class of $<2.5 \mathrm{~cm}$. While in rich and medium forests the highest stem number belonged to the DBH class of 2.5-5.0 cm (Figure 2). Poor and very poor forests had the exponential shape of stems/DBH distribution. While rich and medium forests had inverted J shape of stems/DBH distribution. Generally, stems of woody climbers with DBH $>7.0 \mathrm{~cm}$ were rarely found in all four forest types.


Figure 2. Stem distribution of woody climbers by DBH classes in four forest types

Rich and medium forests had a similar pattern of AGB/DBH distribution (Figure 3) of inverted J shape with a peak at 2.5-5.0 cm DBH class. While poor forest had two peaks of AGB at $<2.5 \mathrm{~cm}$ DBH class and 5.0-7.5 cm DBH class. Much difference of AGB/DBH distribution was found in the very poor forest with three peaks at $<2.5 \mathrm{~cm}$ DBH class, $5.0-7.5 \mathrm{~cm}$ DBH class, and $>20 \mathrm{~cm}$ DBH class. Generally, AGB of woody climbers mostly focused on DBH of $<10.0 \mathrm{~cm}$.


Figure 3. AGB distribution of woody climbers by DBH classes in four forest types

## 4. DISCUSSION

To promote regeneration and growth of seedlings of commercially valuable species, silvicultural practices [27] were applied mostly in the poor forest as many seedlings were available three. While, it was not allowed to apply rich and medium forests as they are fully protected and fewer seedlings of valuable species were available in the very poor forest, leading to no silvicultural practice application. Such differences in
silvicultural application lead to a difference in stem density and AGB among four forest types as the lowest ones were found in the poor forest ( $785 \mathrm{stems} / \mathrm{ha}$ and AGB of $3,573 \mathrm{~kg} / \mathrm{ha}$ ). The lower species number was found in poor and very poor forests compared to much higher ones found in rich and medium forests (Table 1). This could be explained by the difference in living environment as shading level/sunlight penetration [9], [12], [13]. Canopy cover in the poor and the very poor forest was low, the canopy structure was disturbed leading to many open areas, where shade-intolerant and shade-tolerant species cannot survive [28].

In this study, highly disturbed forests as of very poor forest ( $1,991 \mathrm{stems} / \mathrm{ha;} 20,560 \mathrm{~kg} / \mathrm{ha}$ ) had much higher stem density and AGB compared to the less disturbed forest as of rich ( 876 stems $/ \mathrm{ha}$; $5,570 \mathrm{~kg} / \mathrm{ha}$ ) and medium ( $1,246 \mathrm{stems} / \mathrm{ha} ; 9,444 \mathrm{~kg} / \mathrm{ha}$ ) forests. A similar phenomenon was found in other studies [28], [29]. This could be explained by the fact that disturbed forests provide more favorable conditions than less disturbed forests to promote woody climber success [9], [12]. The decrease of diversity indices from rich forests to very poor forest was found in this study (Table 2). A similar pattern was also found in other researches [12], [29]. This indicated that disturbance and change of micro-environment (e.g. light) play a central role in the diversity of woody climbers as most species are shade-tolerant plants. While micro-environment change after disturbance promotes light-demanding woody climbers such as Tetrastigma quadrangulum and Acacia comosa in the present study (Table 1) leading to their high abundance in the poor and very poor forest while missing in rich and medium forests.

Rich and medium forests had inverted $\mathbf{J}$ shape of stems/DBH distribution, which was also found in other less disturbed sites [29]. This indicated that micro-environment is not suitable for many species to germinate and recruit small stems in rich and medium forests. While such condition is suitable for many light-demanding species in poor and very poor forest and therefore stem density of woody climbers was very high in the smallest DBH class ( $<2.5 \mathrm{~cm}$ DBH class), leading to the exponential shape of stems/DBH distribution (Figure 3 and Figure 4).

## 5. CONCLUSIONS

Previous human disturbances have significantly impacted stem density, species diversity, and aboveground biomass of woody climbers in evergreen tropical forests, southern Vietnam. Very poor forest with high intensity of human disturbances as selective logging led to high density and AGB of woody climbers but low diversity. Meanwhile, silvicultural practice to promote regeneration and growth of commercially valuable species has led to low density, AGB, and diversity of woody climbers in the poor forest. Less disturbed forests as of rich and medium forests had lower density and AGB of woody climbers but higher diversity both in terms of species number and diversity indices. Therefore, maintaining woody climbers in the tropical forest could promote plant conservation, while still plays an important role in carbon sequestration.

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## 7. REFERENCES

[1] Gentry, H.A., \& Dodson, C. (1987) Contribution of non-tree to species richness of a tropical rain forest. Biotropica, 9, 149-155.
[2] Burnham, R.J. (2002). Dominance, diversity and distribution of lianas in Yasuní, Ecuador: who is on top? Journal of Tropical Ecology, 18, 845-864.
[3] Schnitzer, S.A., \& Bongers, F. (2022). The ecology of lianas and their role in forests. Trends in Ecology Evolution, 2002, 17:223-230.
[4] Mascaros, J., Schnitzer, S.A., \& Carsonc, W.P. (2004). Liana diversity, abundance, and mortality in a tropical wet forest in Costa Rica. Forest Ecology and Management, 190, 3-14.
[5] Gerwing, J.J. (2004). Life history diversity among six species of canopy lianas in an old-growth forest of the eastern Brazilian Amazon. Forest Ecology and Management, 190, 57-72.
[6] Nabe-Nielsen, J. (2001). Diversity and distribution of lianas in a neotropical rain forest, Yasun National Park, Ecuador. Journal of Tropical Ecology, 17, 1-19.
[7] DeWalt, S.J., Schnitzer, S.A., Alves, L.F., Bongers, F., Burnham, R.J., Cai, Z., Carson, W.P., Chave, J., Chuyong, G.B., Costa, F.R., \& Ewango, C.E. (2015). Biogeographical patterns of liana abundance and diversity. In: Schnitzer, S.A., Bongers, F., Burnham, R.J., \& Putz, F.E. (eds.) Ecology of lianas. John Wiley \& Sons, New York. pp. 131-146.
[8] DeWalt, S.J., \& Chave. (2004). Structure and biomass of four lowland neotropical forests. Biotropica, 36, 7-19.
[9] Laurance, W.F., Perez-Salicrup, D., Delamonica, P., Fearnside, P.M., D'Angelo, S., Jerozolinski, A., Pohl, L., \& Lovejoy, T.E. (2001). Rain forest fragmentation and the structure of Amazonian liana communities. Ecology, 82, 105-116.
[10] Putz, F.E. (1984). The natural history of lianas on Barro Colorado Island, Panama. Ecology, 65, 17131724.
[11] Kainer, K.A., Wadt, L.H.O., Sliva, D.A.P., \& Capanu, M. (2006). Liana loads and their association with Bertholletia excelsa fruit and nut production, diameter growth and crown attributes. Journal of Tropical Ecology, 22, 147-154.
[12] DeWalt, S.J., Schnitzer, S.A., \& Denslow, J.S. (2000). Density and diversity of lianas along a chronosequence in a central Panamanian lowland forest. Journal of Tropical Ecology, 16, 1-19.
[13] Toledo, M. (2010). Neotropical lowland forests along environmental gradients. Ph.D. Thesis, Wageningen University.
[14] Vivek, P., \& Parthasarathy, N. (2014). Liana community and functional trait analysis in tropical dry evergreen forest of India. Journal of Plant Ecology, 8, 501-512.
[15] Putz, F.E. (1983). Liana biomass and leaf area of a "tierrafirme" forest in the Rio Negro basin, Venezuela. Biotropica, 15, 185-189.
[16] Wright, I.J., Reich, P.B., Westoby, M., Ackerly, D.D., Baruch, Z., Bongers, F., Cavender-Bares, J., Chapin, T., Cornelissen, J.H., Diemer, M., Flexas, J., Garnier, E., Groom, P.K., Gulias, J., Hikosaka, K., Lamont, B.B., Lee, T., Lee, W., Lusk. C., Midgley, J.J., Navas, M.L., Niinemets, U., Oleksyn, J., Osada, N., Poorter, H., Poot, P., Prior, L., Pyankov, V.I., Roumet, C., Thomas, S.C., Tjoelker, M.Y., \& Wright, J. (2005).

Late twentieth-century patterns and trends in the climate of tropical forest regions. In: Malhi, Y., \& Phillips, O. (Eds.). Tropical forests and global atmospheric change. Oxford University Press, Oxford.
[17] Schnitzer, S.A., \& Bongers, F. (2011). Increasing liana abundance and biomass in tropical forests: emerging patterns and putative mechanisms. Ecology Letters, 14, 397-406.
[18] Ingwell, L.L., Wright, S.J., Becklund, K.K., Hubbell, S.P., Schnitzer, S.A. (2010). The impact of lianas on 10 years of tree growth and mortality on Barro Colorado Island, Panama. Journal of Ecology, 98, 879-884.
[19] Toledo, M., Pena-Claros, M., Bongers, F., Alarcon, A., Balcazar, J., Chuvina, J., Leano, C., Licona, J.C.C., Poorter ,L. (2012). Distribution patterns of tropical woody species in response to climatic and edaphic gradients. Journal of Ecology, 100, 253-263.
[20] Thai, V.T. (1963). Eco-genesis and classification of forest vegetation of Vietnam (From ecosystem perspective). Science and Techniques Publishing House, Ha Noi, Vietnam.
[21] Hung, D.V., \& Potokin, A.F. (2019). Diversity of Plant Species Composition and Forest Vegetation Cover of Dong Nai Culture and Nature Reserve, Vietnam. In IOP Conference Series: Earth and Environmental Science, 316, p. 012009.
[22] Stellman, J.M., Stellman, S.D., Christian, R., Weber, T., \& Tomasallo, C. (2003). The extent and patterns of usage of Agent Orange and other herbicides in Vietnam. Nature, 422, 681-687.
[23] Millet, J., Pascal, J.P., \& Kiet, L.C. (2010). Effects of disturbance over 60 years on a lowland forest in southern Vietnam. Journal of Tropical Forest Science, 22, 237-246.
[24] Shannon, C.E. (1948). A mathematical theory of communication. Bell System Technology Journal, 27:379-423.
[25] Pielou, E.C. (1966). The measurement of diversity in different types of biological collections. Journal of Theoretical Biology, 13, 131-144.
[26] Hozumi, K., Yoda, K., Kokawa, S., \& Kira, T. (1969). Production ecology of tropical rain forests in southwestern Cambodia. I. Plant biomass. Nature \& life in Southeast Asia, 6, 1-49.
[27] Ha, T.T., DoabJohn, C., Trinh, N.B., Zimmer, H.C., Tran, L.D., \& Nicholsa, D. (2019). Recovery of tropical moist deciduous dipterocarp forest in Southern Vietnam. Forest Ecology and Management, 433, 184204.
[28] Schnitzer, S.A., \& Carson, W.P. (2010). Lianas suppress tree regeneration and diversity in treefall gaps. Ecology Letters, 13, 849-857.
[29] Anbarashan, M., \& Parthasarathy, N. (2013). Diversity and ecology of lianas in tropical dry evergreen forests on the Coromandel Coast of India under various disturbance regimes. Flora, 208, 22-32.

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