

Benthos of the York River

David J. Gillett and Linda C. Schaffner

Virginia Institute of Marine Science
Gloucester Point, VA 23061 U.S.A.

ABSTRACT

Benthic organisms and their communities are key components of estuarine systems. We provide an overview of the biology and key ecological features of benthic communities of York River Estuary (YRE), which is the site of the Chesapeake Bay National Estuarine Research Reserve in Virginia (CBNERRVA). Major subtidal benthic habitats in YRE include soft mud and sand bottoms, with only limited distribution of submerged aquatic vegetation and oyster shell. Major taxonomic groups of macrofauna dominating muds and sands of YRE include annelids, molluscs and crustaceans; similar to those found in other temperate estuaries of the US Mid-Atlantic. Meiofaunal assemblages of YRE soft bottoms are dominated by nematodes and copepods. Species distribution patterns in YRE are strongly correlated with salinity and bottom type, while other factors such as eutrophication and hypoxia may be growing in importance. Much of the YRE benthos fails to meet the restoration goals set by the Chesapeake Bay Program. The poor condition of the benthos is expressed as low biomass and abundance and may be associated with degraded water quality, hypoxia and sediment disturbance processes. No comprehensive inventory of the benthic biota of the CBNERRS sites is available, which will make it difficult to assess future changes due to human impacts such as climate change or the introduction of exotic species. Given this paucity of data, a systemic cataloging of the benthic resources of the reserve sites and any potential invasive species is a much needed avenue of future research for CBNERRVA.

INTRODUCTION TO THE BENTHOS

The soft mud and sand habitats of the York River Estuary, as well as the interspersed patches of aquatic vegetation and oyster shell, support a wide variety of fauna and flora and are an important part of this productive coastal ecosystem. These bottom habitats and their resident organisms are called the benthos, derived from the Greek for "bottom of the sea." The animals comprising benthic communities, the zoobenthos¹, include almost every known phylum and exclusively encompass a number of them. For the purposes of this paper we have limited ourselves to a discussion of the benthic invertebrate residents and their communities of the York River Estuary. This is not to slight the countless numbers of bacteria, Archea, and protozoans that comprise the microbenthos, or the bottom-dwelling fish and crustaceans of the estuary, all of which are discussed in other papers in this issue.

Most benthic invertebrates are quite small and can be clearly distinguished only with the aid of magnification. They are classified into three major groups based on adult size. The smallest are the meiobenthos, which pass through a 500- μm mesh, but are retained on a 63- μm screen. Important taxa of meiobenthos include harpacticoid copepods, nematodes, ostracods and Foraminifera (see HIGGINS and THIEL, 1988). Macrobenthos are retained on a 500- μm mesh screen and are not readily identifiable without magnification. Annelid worms, bivalves, gastropods, crustaceans, tunicates, and insect

¹The generic terms *benthos* and *benthic*, which are used to describe the bottom realm, have also been variously used to describe any and all of the organisms, from bacteria and microalgae to seagrasses and demersal predators, that are associated with benthic habitats. Use of the term *zoobenthos* provides more clarity, but in practice is rarely used by benthic ecologists working in the U.S.

larvae are commonly encountered macrobenthos in estuaries. The largest size-based category, the megabenthos, can be identified without magnification because individuals are typically multiple centimeters in size. This group includes animals such as crabs, bivalves, gastropods, sponges, colonial entoprocts and hydrozoans. Benthic organisms may progress through different categories as they grow. Many animals classified as macrobenthos start off as meiobenthic juveniles and are known as "temporary meiobenthos."

Beyond size, the mobility of an animal (motile versus sessile) and how it associates with the sediment or hard substrate (infaunal versus epibenthic) are other common ways benthic organisms are classified. Epibenthic animals live on or just above the substrate. They may be firmly attached (sessile), relatively sedentary, or fully motile. Animals such as barnacles, oysters, sponges, tunicates, entoprocts, gastropods, anthozoans, mud crabs, and certain species of amphipods are common representatives of the epibenthos. Animals that live within the substrate are called infauna and include most species of annelids and bivalves, larval insects, phoronids, as well as some species of amphipods and anthozoans.

MAJOR TAXONOMIC GROUPS OF BENTHIC FAUNA IN THE YORK ESTUARY

A comprehensive checklist of benthic animals in the York River Estuary and the greater Chesapeake Bay was published by Wass (1972). It provides frequency of occurrence and habitat preferences of those animals known at the time. There is no complete benthic invertebrate species list exclusively for the York River system; however, most of the benthic fauna found in the York River Estuary are listed in the regularly updated checklist available for the Chesapeake Bay Benthic

Monitoring Program (LLANSÓ, 2005). A partial checklist of benthic organisms in the York River Estuary developed from these and other sources is provided in the Appendix.

Poriferans

Sponges are colonial macro- to megabenthic-sized organisms. They filter feed by pumping water through inhalant and exhalant pores called ostia, trapping particles along the body wall, and ingesting them by phagocytosis (BRUSCA and BRUSCA, 1990). Most sponges in the York River Estuary are limited to the meso- to polyhaline reaches. Among the most conspicuous are the red beard (*Microciona prolifera*) and brown (*Haliclona* spp.) sponges, both of which grow attached to hard substrate (Figure 1). *M. prolifera* is frequently seen on pier pilings, while *Haliclona loosanoffi* is commonly found on the blades

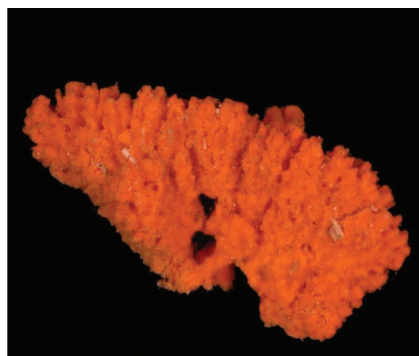


Figure 1. Unidentified red sponge. (Image courtesy of Southeastern Regional Taxonomic Center/South Carolina Department of Natural Resources)

of submerged aquatic vegetation (SAV). The boring sponges, *Cliona* spp., erode galleries of passageways through calcareous shell of molluscs, which provides protection from predators. These types of sponges are considered nuisance species by commercial shellfish harvesters because the erosion of shell material is detrimental to living molluscs. All of the sponges found in the York River Estuary are capable of both sexual and asexual reproduction. Fragments of a sponge can grow an entire new sponge, given an appropriate substrate. Sexual reproduction in sponges is through broadcast spawning with most species thought to be hermaphroditic, which means that they switch between the production of male and female gametes during different parts of their lives (BRUSCA and BRUSCA, 1990).

Cnidarians

Representatives of all three classes of cnidarians (Hydrozoa, Anthozoa, and Scyphozoa) have been observed among the macrobenthic fauna of the York River Estuary. All cnidarians possess nematocysts, responsible for the familiar stinging sensation of jellyfish, which they use for both defensive and prey capturing purposes. Hydrozoans, the most conspicuous benthic cnidarians found in the York River Estuary, settle and grow on myriad substrates along the full salinity gradient. As passive filter feeders, hydroids rely on water currents to bring food particles to their feeding tentacles. Hydromedusae are found as solitary individuals and, more commonly, as colonies of many individuals or zooids that can create substantial colonies, extending several centimeters in to the water column. Colonial hydroids are abundant in the lower York River, where the large mounds they form on the bottom support a variety of other macrobenthic organisms (Figure 2) (SCHAFFNER *et al.*, 2001). Hydrozoans have both sexual and asexual repro-

duction during different stages of their life cycle. Asexually, new hydroid zooids can be budded off an adult in an expansion of the colony, or as separate individuals in the non-colonial forms. Sexual reproduction in hydrozoans, much like the other types of cnidarians, is somewhat more complex. A free-swimming male or female medusa (jellyfish-like) stage is budded off of the benthic adult form, which in turn, releases gametes into the water column that when fertilized, form asexual, benthic individuals (BRUSCA and BRUSCA, 1990).

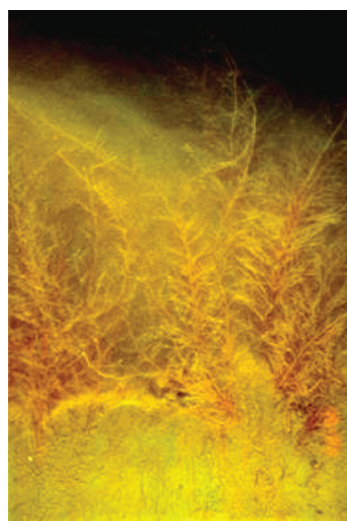


Figure 2. Colonial hydroids from the lower York River. (Image courtesy of Robert Diaz, VIMS)

Though less abundant and less diverse in the York River Estuary than hydrozoans, anthozoan (sea anemones) and scyphozoan (jelly fish) cnidarians are also found within the benthic communities. Like the hydrozoans, benthic anthozoans are passive filter feeders capable of both sexual and asexual reproduction. Anthozoans are non-simultaneous hermaphrodites that can bud off new individuals from the adult form, as well as produce male or female gametes. Anthozoans have lost the free-swimming medusa-stage of other cnidarians. The benthic adults directly release gametes to the water column, where they combine to form planular larvae that settle out of the water to form new benthic adults. Common anthozoans include epibenthic species (e.g., *Diadumene leucolena*) and infaunal species (*Cerianthopsis americanus*, *Actiniaria* sp. or *Edwardsia elegans*) (SAGASTI *et al.*, 2001; LLANSÓ 2005). Scyphozoans are only ephemeral benthic organisms, but the benthic stage is an essential part of their reproductive lifestyle that occurs at various times of the year depending upon the species (see STEINBERG and CONDON this S.I.). This benthic stage is referred to as a scyphistoma and is an asexual form that buds off the familiar, pelagic medusae seen in the estuary.

Platyhelminthes

Flatworms are a small, relatively obscure component of the benthic community that can be found all along the estuarine salinity gradient. Free-living turbellarian flatworms can be macro- or meiobenthic in size and typically live within the upper few centimeters of sandy or muddy sediments, or on hard substrate (MARTENS and SCHOCKAERT, 1986). The most common estuarine turbellarians prey or scavenge upon the smaller benthos they encounter, such as meiobenthic harpacticoid copepods or nematodes, larger protozoans like Foraminifera, as well as macrobenthic oligochaetes and chironomids (ARMITAGE and YOUNG, 1990). Although living oysters are now uncommon in the York River Estuary, the oyster flatworm *Stylochus ellipticus* remains an important component of the ecosystem's hard substrate benthic community (SAGASTI *et al.*, 2000). Parasitic flatworms (trematodes, monogenetic

flukes, and cestodes) are also found within the estuary. They live on or within a variety of estuarine fauna, including fish, gastropods, or annelids. Most of the free-living species of turbellarians are hermaphroditic and are capable of both asexual (fission) and sexual (cross-fertilization) reproduction (BRUSCA and BRUSCA, 1990).

Nemerteans

Nemerteans are highly mobile, flat, non-segmented worms, commonly referred to as “ribbon worms.” They are an ecologically important, though relatively poorly studied, taxonomic group within the benthic community of the York River. Nemerteans (Figure 3) can be quite large (often many centimeters in length) and move through the sediment by ciliary or peristaltic motion in larger species. Some of the largest nemerteans are burrowing predators (e.g., *Cerebratulus lacteus*), which move up from below to capture their prey with an ever-



Figure 3. Unidentified nemertean. (Image courtesy Southeastern Regional Taxonomic Center/South Carolina Department of Natural Resources)

sible pharynx, which may be armed with a toxin-delivering stylet (BOURQUE *et al.*, 2002). Some species have quite advanced chemosensory detection

capabilities and have been observed tracking potential prey items for some distance before striking (BRUSCA and BRUSCA, 1990). These chemosensory capabilities are also used to by nemerteans to track and locate mates for reproduction. Most nemerteans undergo sexual reproduction, with external or internal fertilization depending upon the species. Additionally, some species of the genus *Lineus*, a few species of which are observed in the York River Estuary (WASS, 1972), are also capable of asexual reproduction via fragmentation of the posterior end of the worm.

Nematodes

Meiobenthic nematodes are among the most numerically abundant benthic fauna in the York River Estuary (ALONGI *et al.*, 1982; METCALFE, 2005), though given their small size and somewhat obscure taxonomy, little species-specific research has been done on local nematode communities. These small, non-segmented round worms move through the interstitial spaces of sandy and muddy sediments. Nematodes encompass a wide variety of feeding styles, including deposit feeding, grazing, carnivory, interstitial filter-feeding, and parasitism, all of which, excluding the parasitic species, reproduce sexually with internal fertilization.

Entoprocts

Another example of a colonial filter-feeder; entoprocts (formally known as bryozoans) are epibenthos that will attach to almost any hard surface in the poly- and euhaline portions

of the York River and other estuaries. Composed of numerous individual zooids, species commonly found in the York River Estuary such as *Pedicellina cernua* (SAGASTI *et al.*, 2000), passively feed on passing plankton using ciliated tentacles (Figure 4). Entoprocts will undergo asexual budding within a given colony, but also periodically undergo sexual reproduction, broadcasting larvae into the water column to start new colonies (BRUSCA and BRUSCA, 1990). The zooids of entoprocts do not develop specialized functions like those of hydroids, but each individual is a protandric hermaphrodite, capable of both feeding and reproduction.



Figure 4. Unidentified branching, colonial entoproct. (Image courtesy of Southeastern Regional Taxonomic Center/South Carolina Department of Natural Resources)

Annelids

This group of truly segmented worms includes the polychaetes, oligochaetes, and leeches. The annelids are a numerically abundant and ecologically important component of all benthic communities, including those of the York River Estuary. Within the estuary, annelids range in size from meiobenthic juveniles to megabenthic chaetopteric polychaetes and encompass all major feeding types and living positions.

Polychaetes are the most diverse group of annelids in the saline portions of the York River Estuary, with different species dominating in different salinity zones. *Polydora cornuta* and *Sabellaria vulgaris* are tube building, epibenthos commonly found on SAV or other hard substrates throughout the York River (ORTH, 1973; SAGASTI *et al.*, 2000). There are also highly mobile carnivores (e.g., *Eteone heteropoda* and *Glycinde solitaria*) with well-developed parapodia and cirri for mobility and sensory organs for tracking prey items (Figure 5). Many species of polychaetes are sessile infauna, living with their heads and feeding appendages at the sediment-water interface (e.g., *Loimia medusa*), or head down in the sediment with their tails at the surface (e.g., *Clymenella torquata*). Deposit feeders ingest bacteria, microalgae and organic matter associated with sediment particles and are common among the polychaetes. Filter-feeding is also common in the sessile



Figure 5. A common polychaete annelid *Neanthes succinea*. (Image courtesy of Southeastern Regional Taxonomic Center/South Carolina Department of Natural Resources)

polychaetes. Some species actively pump water into their tubes/burrows with their parapodia (e.g., *Spiochaetopterus costarum*), while others are capable of switching between passive filter-feeding and surface deposit-feeding with the anterior palps (e.g., *Streblospio benedicti*) (FAUCHALD and JUMARS, 1979). Polychaetes primarily reproduce via sexual reproduction, wherein some species undergo internal fertilization and brood their larvae, while others are broadcast spawners with distinctive planktonic trochophore larvae.

Oligochaete annelids are also found throughout the York River Estuary, but are far less diverse than the polychaetes. They lack parapodia and typically have simple heads, without sensory palps or antennae, though some freshwater taxa have a proboscis for feeding (e.g., family Naidae). All of the oligochaetes found in the York River Estuary are motile, deposit feeders. Members of the genus *Tubificoides*, the naid *Paranais litoralis* and some species of the family Enchytraeidae are found in brackish and saline portions of the estuary. The tidal freshwater region contains a much more diverse assemblage of oligochaetes (e.g., *Limnodrilus hoffmeisteri*, *Aulodrilus templetoni*, *Dero digitata*). This pattern of higher diversity up-estuary reflects the radiation of oligochaetes into the estuary from freshwater systems (STEPHENSON, 1972). This contrasts with the pattern of diversity increasing with salinity in the estuary seen in many of the other estuarine invertebrates, which are descended from marine forms. All oligochaetes found in the estuary are simultaneously hermaphroditic and reproduce sexually, depositing cocoons into the mud or sand that contain a varied number of zygotes that grow and disperse after release. Some genera of oligochaetes, notably the naids, also reproduce asexually by budding offspring from their posterior regions (STEPHENSON, 1972). Asexual reproduction is a common means of reproduction during periods of favorable environmental conditions (food availability, temperature, etc.), but most species will switch to sexual reproduction when conditions become unfavorable (STEPHENSON, 1972).

The last sub-class of annelids found in the York River Estuary is the Hirudinae (leeches). Leeches are closely related to oligochaetes and are likewise simultaneous hermaphrodites with a reduced body structure devoid of parapodia or complex setae. Unlike oligochaetes, leeches reproduce strictly through sexual reproduction, producing cocoons they deposit into the environment. Most species of Hirudinae are exoparasites (e.g., *Myzobdella lugubris*, *Calliobdella vivida*) of other animals, though a few species (e.g., *Helobdella elongata*, *H. stagnalis*) are free-living predators of smaller invertebrates such as nematodes, copepods, or oligochaetes (WASS, 1972; BRUSCA and BRUSCA, 1990). Within the York River Estuary, these free-living species are primarily limited to the tidal freshwater and oligohaline waters (J. WILLIAMS, pers. comm.).

Echiurans

Echiurans are a phylum of non-segmented, worm-like animals that live in the high mesohaline to polyhaline parts of the estuary. Wass (1972) lists *Thalassema hartmani* as the only species commonly found in the estuary. Echiurans are sessile, surface deposit feeders. They build a tube in the sediment and feed with a long a proboscis that pulls sediment below the surface to the mouth. Echiurans have separate sexes and reproduce sexually in mass spawning events where gametes are released to the water column.

Arthropods

In terms of phylogeny and body form, arthropods are possibly the most diverse group of benthic organisms in the York River. These segmented animals have hard exoskeletons and jointed appendages, but range in form from barnacles to crabs. Arthropods of the estuarine benthic community reproduce via sexual reproduction, typically with external fertilization. Most arthropods are highly motile animals capable of swimming and walking, though barnacles are a notable, sessile exception.

Pycnogonids, or sea spiders (Class Chelicerata), are epifaunal arthropods (Figure 6) most commonly observed in fouling communities; among tunicates or sponges in the polyhaline and high mesohaline portions of the York River Estuary (e.g., *Anoplodactylus pygmaeus*, *Tanystylum orbiculare*, etc.) (WASS,



Figure 6. The pycnogonid *Pallenopsis schmitti*. (Image courtesy of D. Gillett)

1972; SAGASTI *et al.*, 2000). These mobile, spider-like arthropods are mostly carnivores, which feed upon other epifauna. There are some herbivores though, which feed on the algae growing in fouling communities (BRUSCA and BRUSCA, 1990).

Though they spend only a portion of their lives as benthic fauna, larval insects, predominantly of the Orders Diptera (flies and midges) and Trichoptera (caddis flies), are an important component of the tidal freshwater and oligohaline portions of the York River Estuary. Most families of insect larvae found living within the sediments span a range of feeding modes, from carnivore/scavengers (e.g., *Tanyptus* sp.) to grazers (e.g., *Cryptochironomus* sp.). After a few weeks to months in the benthos, chironomid insect larvae metamorphose into adult dipterid and trichopterid flies and leave the system.

Crustaceans are the most taxonomically and trophically diverse group of benthic animals found in the estuary, as well the best known by the general public. Crustacean arthropods encompass the range of feeding types, including grazing, filter feeding, and deposit feeding. Macrobenthic crustaceans in the York River Estuary include sessile, filtering epifaunal organisms such as barnacles (*Balanus eburneus* and *B. improvisus*), motile, shrimp-like (peracarid) taxa like cumaceans (e.g., *Leucon americanus* or *Cyclaspis varians*) and mysids (e.g., *Neomysis americana*) that live on the sediment surface, mobile burrowing isopods (e.g., *Cyathura polita* or *Edotea triloba*), and amphipods (e.g., *Leptocheirus plumulosus*, *Protohaustorius deichmannae*, or *Caprella penantis*) (Figure 7). Decapod crustaceans include one the most famous benthic organisms of the estuary, the blue crab (*Callinectes sapidus*), as well as some smaller less well-known members, such as xanthid mud crabs (e.g., *Rhithropanopeus harrisi*). Many of the small crabs that populate the estuary are relatively cryptic, living among shells and other structured benthic habitats such as sponges. Fiddler crabs (*Uca* spp.), which live in the intertidal salt marshes that line the banks of the estuary, are a common sight to most peo-

ple. The most abundant crustaceans in the York River Estuary, meiobenthic harpacticoid copepods (e.g., *Euterpina acutifrons* or *Camuella canadensis*), reside near the sediment-water interface among sediment grains of the estuarine bottom and are important grazers of bacteria and micro-algae.



Figure 7. *Leptocheirus plumulosus*, a common amphipod in the York River Estuary. (Image courtesy of D. Gillett)

Molluscs

Benthic molluscs in the York River Estuary include the conspicuous and familiar clams and snails that can live multiple years and in some cases, e.g., oysters and mussels, are capable of creating complex, hard bottom habitats that provide living space and refugia for other benthic organisms. The most common molluscs of the York River Estuary can be divided into two groups based on the shape and number of shells they have: bivalves, with two relative concave shells, e.g., clams (*Macoma balthica* or *Mya arenaria*), oysters (*Crassostrea virginica*), and mussels (*Geukensia demissa*); or gastropod snails, which have a single, typically spiraled shell that includes whelks (*Busycon canaliculatum*) and mud snails (*Littorina littorea* or *Hydrobia* sp.).

Bivalves are found along the length of the York River Estuary in all of the salinity zones and typically comprise a significant amount of the total biomass of the infaunal benthic communities (DIAZ and SCHAFFNER, 1990; SCHAFFNER *et al.*, 2001). All of the bivalves found in the York River Estuary reproduce sexually, broadcasting their gametes into the water column, creating planktonic larvae.

Most are filter feeders (Figure 8), though one of the dominant genera in the meso- and polyhaline portions of the estuary, *Macoma*, is a functional deposit-feeder that can switch from filter feeding to deposit feeding depending upon the water currents and food availability (POHLO, 1982). Large reefs of the eastern oyster *C. virginica* were once dominant benthic features of the York River, but overfishing, habitat destruction and disease have led to their demise (Figure 9.) (HARGIS and HAVEN, 1999) and the



Figure 8. *Macoma balthica*, one of the most common infaunal bivalve molluscs in the York River Estuary. Note the incurrent and excurrent siphons protruding from the top of the shell. (Image courtesy of Heidi Mahon, Old Dominion University)

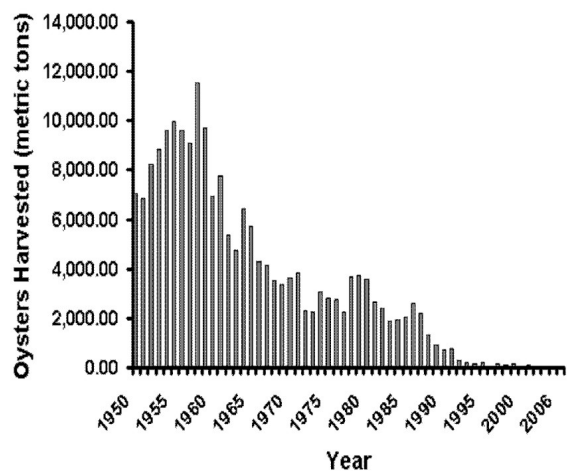


Figure 9. Commercial landings of the Eastern Oyster *Crassostrea virginica* in Virginia from 1950 – 2006. Data from the National Marine Fisheries Service (NMFS)

ecological importance of the oyster has been drastically reduced (POMEROY *et al.*, 2006; COEN *et al.*, 2007).

Gastropods are among the most voracious predators in the benthos. Large whelks, such as the channeled whelk *B. canaliculatum* and the non-native veined rapa whelk *Rapana venosa*, are a considerable problem for commercial bivalve aquaculture operations (J. HARDING, pers. comm.). Other gastropods feed on benthic microalgae in the shallow subtidal and intertidal flats of the estuary (e.g., *Hydrobia* sp. or *Narssarius obsoletus*) or on the epiphytic microbes found on the stalks of intertidal marsh grass (e.g. the marsh periwinkle *Littorina littorea*). Gastropods reproduce sexually, undergoing internal fertilization, with the females attaching their egg cases the sediment surface or some hard structure in the environment (e.g., shell material or SAV blades).

STUDIES OF BENTHIC FAUNA IN THE YORK RIVER

Because of the economic importance of the oyster fishery and the feared decline in the resource, significant effort was put into quantifying the abundance and spatial extent of eastern oyster (*C. virginica*) reefs in the York River by the state of Virginia from at least the mid 1800's, (WHEATLEY, 1959; HARGIS and HAVEN, 1999). These works, most notably the Baylor survey of 1900, represented the first surveys of benthic biota within the York River Estuary; even in light of their focus on one organism and the delineation of fishing rights. The quantitative study of the complete benthic communities of the York River Estuary began in earnest in the mid-1960's, led by scientists of the Virginia Institute of Marine Science (VIMS). Notable early studies include those of Wass (1965), Haven *et al.* (1981) [note: HAVEN *et al.* collected data in 1965-1966, but did not published until 1981], and Boesch (1971). Initial studies focused on describing benthic community composition of major York River habitats. Based on a review of the early literature for Chesapeake Bay and the major sub-estuaries, distribution and abundance patterns of dominant macrobenthic organisms of soft sediment habitats were summarized by Diaz and Schaffner (1990) (Table 1). Marsh (1970) and Orth (1973) identified the epifaunal and infaunal communities of sea grass beds in the lower York River Estuary.

Table 1. Physical and benthic community characteristics of the major benthic habitats of the York River estuary. Modified from DIAZ and SCHAFFNER, 1990.

Salinity/Habitat Type	Reserve Site(s)	Physical Characteristics	Macrobenthic Community Characteristics	Macrofauna Density / Taxa of Note	Macrofauna Biomass / Taxa of Note
Tidal Freshwater					
Shoals	Sweet Hall Marsh	Shallow depths	Stenohaline freshwater fauna	Low to moderate	Oligochaetes and bivalves high
		Mud to sand sediments	Deposit and suspension feeders	<i>Limnodrilus</i> spp.,	Others low
Channels		Wave- and tide-dominated	Infaunal predators	<i>Illydrilus templetoni</i> ,	<i>Limnodrilus</i> spp.,
		High turbidity	Many ephemeral fauna	<i>Stephensonia</i>	<i>Illydrilus templetoni</i> , and
		Low to moderate light penetration	Moderate to low diversity	<i>trivandran</i> ,	<i>Rangia cuneata</i>
				<i>Coelotanypus</i> spp.	
		Intermediate depths	Stenohaline freshwater fauna	Low, especially in areas of fluid mud	Bivalves high
		Mud to sand sediments	Deposit and Suspension feeders		Others low
		Fluid mud possible	Moderate to low diversity		
		Tide dominated			
		High turbidity			
		No light penetration			
Oligohaline					
Shoals	Taskinas Creek	Shallow depths	Euryhaline estuarine fauna	Low to high	Bivalves high
		Mud and sand sediments	Deposit and suspension feeders	<i>Tubificoides</i>	Others low
Channels		Wave- and tide-dominated	Some ephemeral fauna	<i>heterochaetus</i> ,	<i>Marenzelleria viridis</i> ,
		Region of estuarine turbidity maximum (ETM)	Low diversity	<i>Tubificoides brownae</i> ,	<i>Macoma balthica</i> ,
		High deposition		<i>Leptocheirus plumulosus</i>	<i>Cyathura polita</i>
		Low to moderate light penetration			
		Moderate depths	Euryhaline estuarine fauna	Low to high	Bivalves high
		Mud sediments	Deposit and suspension feeders	<i>Marenzelleria viridis</i> ,	Others low
		Fluid mud possible	Low diversity	<i>Leucon americanus</i>	<i>Macoma balthica</i>
		Tide-dominated			
		Region of ETM			
		High deposition			
		No light penetration			
		Occasional low oxygen			
Mesohaline					
Shoals	Cattlet Islands Timberneck Creek	Shallow depths	Euryhaline estuarine fauna	Moderate to high	Bivalves high
		Sand and mud sediments	All feeding types	<i>Streblospio benedicti</i> ,	Others moderate
Channels		Wave- and tide-dominated	Moderate diversity	<i>Mediomastus ambiseta</i> ,	<i>Macoma balthica</i> ,
		Low to moderate turbidity		<i>Leptocheirus plumulosus</i>	<i>Loimia medusa</i> ,
		Moderate light penetration			<i>Clymenella torquata</i> ,
		Occasional low oxygen			<i>Paraprionospio pinnata</i>
		Intermediate to deep depths	Euryhaline estuarine fauna	Moderate to high*	Bivalves high*
		Mud sediments	All feeding types	<i>Streblospio benedicti</i> ,	Others moderate*
		Fluid mud possible	Moderate diversity*	<i>Mediomastus ambiseta</i>	<i>Macoma balthica</i> ,
		Tide-dominated			<i>Paraprionospio pinnata</i>
		High turbidity, related to secondary ETM			
		No light penetration			
		Seasonal low oxygen			
Polyhaline					
Shoals	Goodwin Islands	Shallow depths	Stenohaline estuarine/marine fauna	Low to moderate	Low to moderate
		Sand sediments	All feeding types	<i>Streblospio benedicti</i> ,	<i>Mercenaria mercenaria</i> ,
Channels		Wave- and tide dominated	Moderate diversity	<i>Spiochaetopterus oculatus</i>	<i>Mya arenaria</i>
		High light penetration			
		Moderate to deep depths	Stenohaline estuarine/marine fauna	Moderate*	Low to high
		Mud to sand sediments	All feeding types	<i>Acteocina canaliculata</i> ,	<i>Mercenaria mercenaria</i> ,
		Tide-dominated	Moderate to high diversity*	<i>Heteromastus filiformis</i>	<i>Chaetopterus variopedatus</i>
		Moderate turbidity			
		No light penetration			
		Seasonal low oxygen			

* Except when low oxygen conditions prevail.

Studies to assess the potential impact of anthropogenic disturbances in the York River Estuary were conducted by scientists at VIMS beginning in the 1970's (e.g., JORDAN *et al.*, 1975; BOESCH and ROSENBERG, 1981; ALONGI *et al.*, 1982). Monitoring of macrobenthic communities in the York River began in the 1980's as part of a larger monitoring program coordinated by the Chesapeake Bay Program (CBP), which is funded by USEPA (United States Environmental Protection Agency), NOAA (National Oceanographic and Atmospheric Association) and the states in the Chesapeake Bay watershed. Samples for infaunal macrobenthos (non-colonial forms only) of soft sediment habitats have been collected at a series of fixed and random stations throughout the Chesapeake Bay, including four fixed stations in the York and Pamunkey Rivers. The four fixed stations, all located in the main channel of the estuary, were sampled quarterly between 1984 and 1994 and subsequently reduced to the present schedule of once a year. Beginning in 1996 the sampling design was changed and 25 samples are now collected in the York-Pamunkey estuarine system each summer (July 15 – September 30) based on a probabilistic sampling design that stratifies the estuary by salinity regime and water depth (LLANSÓ *et al.*, 2006). These monitoring studies provide a wealth of information about the infauna of the York River Estuary, much of which is now available online www.chesapeakebay.net/baybio.htm. Some of the major studies describing the monitoring program and its findings are presented in Weisberg *et al.* (1997), Dauer *et al.* (2000), Alden *et al.* (2002) and Llansó *et al.* (2003).

DISTRIBUTION OF MACROBENTHIC COMMUNITIES ALONG THE ESTUARINE GRADIENT

Benthic studies of the York, James and mainstem Chesapeake Bay regions have clearly demonstrated the strong relationship between benthic community structure and salinity regime (see review by DIAZ and SCHAFFNER, 1990). For ease of comparison, the salinity regime of estuarine waters is typically referred to within the Venice salinity classification system (INTERNATIONAL ASSOCIATION OF LIMNOLOGY, 1958). Salinity in the York is relatively stable, with typical daily changes of less than 5 psu (practical salinity units) at a given location (BOESCH, 1977; SCHAFFNER *et al.*, 2001). Freshwater flow is from the Pamunkey and Mattaponi Rivers, but is relatively low overall, with the York receiving only about 6% of the freshwater entering the Chesapeake Bay from the watershed each year.

Salinity affects osmotic balance and ion regulation of most aquatic organisms. Given the variability of salinity in most estuaries, resident invertebrates must be relatively tolerant. Although some benthic organisms have a wider range of salinity tolerance than others, few species of benthic invertebrates are capable of maintaining physiological function over the full salinity range observed in an estuary, even when local populations become acclimated. Rapid changes in salinity are especially problematic and pulses of fresher water, due to major spring freshets and hurricanes, can act as disturbances to the benthic community (e.g., BOESCH and ROSENBERG, 1981; DAUER *et al.*, 2000; HOLLAND *et al.*, 2004).

A classic pattern observed for macrobenthic communities of estuarine and other brackish water environments is the relationship between salinity and species diversity (REMANE and SCHLIEPER, 1971; GAINNEY and GREENBERG, 1977; DEATON and GREENBERG, 1986). In large brackish water systems such as the

Baltic, or in estuaries that are relatively homeohaline, diversity has been shown to decrease when moving from higher salinity waters to a minimum in at 2 - 7 PSU and then increases again moving into freshwater (ATTRILL, 2002). The pattern of declining diversity with declining salinity is observed in the York River Estuary (BOESCH, 1971; BOESCH *et al.*, 1976; SCHAFFNER *et al.*, 2001), but the pattern in oligohaline to tidal freshwater is not well defined due to limited sampling. Diaz (1989, 1994) found that species diversity did not increase substantially in the tidal freshwater region of the nearby James River estuary and attributed it to the highly variable and physically stressful nature of the region.

Distribution and abundance of benthic species in soft sediment habitats of the York River Estuary is further correlated with bottom type, hydrodynamics, oxygen regime, and other variables that may covary with salinity along the estuarine gradient (see review by SCHAFFNER *et al.*, 2001). Bottom types in the estuary range from cohesive silts and clays to well-sorted sands (Figure 10) (SCHAFFNER *et al.*, 2001; SCHAFFNER unpublished). In the broad lower York, wave energy is a major factor determining sediment distribution patterns. Fine sediment is winnowed away and the bottom is floored mostly by sand and shell in shallow areas (< 10 m depth), while muds tend to ac-

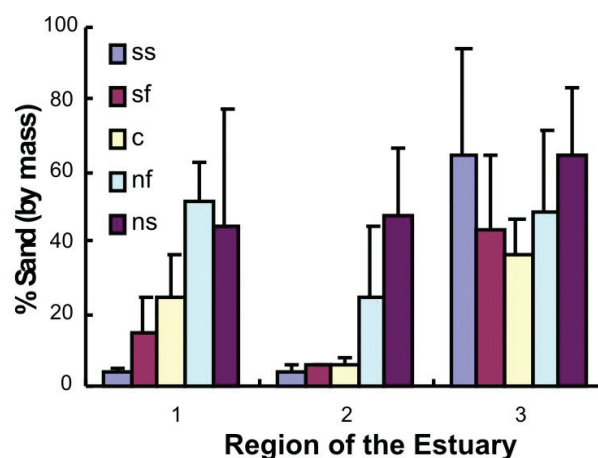


Figure 10. Mean grain size of sediment distributed throughout the York River + 1 standard error. Regions of the Estuary: 1 = upper York River, 2 = mid-York River, 3 = lower York River. ss = southwest shoal, sf = southwest flank, c = channel, nf = northeast flank, and ns = northeast shoal.

accumulate in the channel. In the middle to upper estuary, upstream of Gloucester Point, tidal energy and estuarine circulation become the more important determinants of sediment distribution. Estuarine circulation processes lead to trapping of fine particles, particularly during periods of high freshwater input. Relatively strong tidal scouring of the channel bottom, and strong wave energy on the shoals during some seasons, but not others, results in significant resuspension of sediment and physical disturbance of the bottom (DELLAPENNA *et al.*, 1998, 2003; SCHAFFNER *et al.*, 2001), which influences the structure and productivity of subtidal benthic communities in this region of the estuary (SCHAFFNER *et al.*, 2001; HINCHEY, 2002).

Benthic fauna exhibit sediment preferences that are reflected in their living positions and feeding mechanisms. As

noted above, meiobenthic fauna such as harpacticoid copepods, nematodes, and ostracods live within the spaces between individual sediment grains (the interstitial spaces), ingesting individual particles or filtering the porewater. Sediment with high clay content may become compacted and rich in sulfides, which limits habitat for meiofauna (HIGGINS and THIEL, 1988). For larger benthic organisms, feeding type may determine the suitability of a given sediment type. Highly mobile, non-selective deposit feeders (e.g. capitellid polychaetes and oligochaetes) tend to be more abundant in depositional areas where organic rich sediment particles accumulate and higher sediment water content makes burrowing easier (LOPEZ and LEVINTON, 1987; RICE and RHOADES, 1989). Sandier sediment provides favorable habitat for filter feeders, which have passive collection mechanisms (e.g., phoronids, bryozoans, or hydroids) or limited ability to sort captured particles (e.g., venerid bivalves or chaetopterid polychaetes). In turbid, soft sediment areas of the estuary, smaller silt or clay particles may clog these delicate filtering structures (LOPEZ and LEVINTON, 1987; RICE and RHOADES, 1989). Many benthic taxa of estuaries live equally well in the middle ground of muddy-sands and sandy-muds, particularly those that are capable of switching between deposit feeding and filter feeding as water flow conditions change (e.g., tellinid bivalves or spionid polychaetes) (TAGHON *et al.*, 1980; POHLO, 1982; DAUER, 1983). In the deeper waters of the York, bivalves, including both filter and surface deposit feeders, are especially abundant downstream of the estuarine turbidity maximum, which is an area high phytoplankton production (SIN *et al.*, 1999; SCHAFFNER *et al.*, 2001).

Hypoxia and anoxia are common during summer months in the deep channel of the lower York River Estuary, whereas the shallow shoals almost always remain well mixed and oxygenated. Low oxygen events, which typically last a week or less, occur primarily during periods of summer neap tides, when stratification of the water column tends to be strong and respiration is high (HAAS, 1977; DIAZ *et al.*, 1992). Oxygen is replenished to bottom waters during periods of spring tide due to physical mixing. Episodes of hypoxia or anoxia result in mortality of sensitive taxa (e.g., forams, most species of crustaceans, and some families of polychaetes) and create communities dominated by stress-resistant taxa that tolerate the events, or opportunistic taxa that are able to quickly recolonize disturbed areas (DIAZ and ROSENBERG, 1995; SAGASTI *et al.*, 2000; SAGASTI *et al.*, 2001; METCALFE, 2005).

Physical structure within estuarine habitats also influences the composition and abundance of macrobenthic communities. Oyster reefs were once a predominant feature of estuaries like the York River (HARGIS and HAVEN, 1999). Reefs provide important ecosystem services, including substrate for sessile forms, such as sponges, entoprocts, and barnacles, shelter for motile species, such as xanthid crabs, and filtration by the oyster reef community contributes to improving water clarity, which may benefit nearby sea grass meadows (COEN *et al.*, 1999; HARWELL, 2004; CERCO and NOEL, 2007). Due to over-harvest, disease, and declining water quality there are no longer large oyster reefs in the York River estuary (HARGIS and HAVEN, 1999), though shell clusters may still provide a habitat for other macrobenthos (Figure 11) (SCHAFFNER, unpublished). The proliferation of other structures in the estuary (e.g., piers, bridges, hardened shore lines, stake arrays that support fishing nets, and even ghost crab pots) have created hard sub-

strate habitat that is used by these epifaunal macroinvertebrates, though possibly not to the same degree as oyster reefs did in the past (POMEROY *et al.*, 2006).

Submerged aquatic vegetation (SAV) also increases habitat complexity, and its presence results in the formation of unique assemblages of macrobenthos in shallow estuarine waters. Orth (1973) characterized the macrobenthic infauna associated with *Zostera marina* beds in the high-mesohaline and polyhaline portions of the York River. He

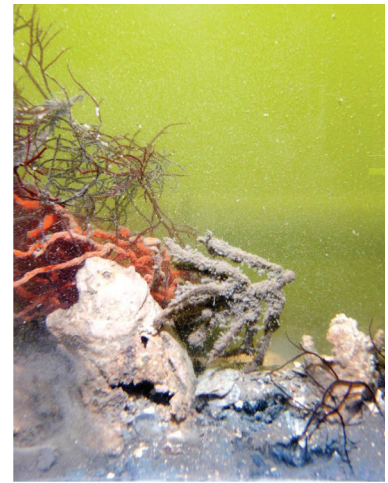


Figure 11. An epifaunal community of sponges, hydroids, entoprocts, and other fauna attached to shell rubble at Catlett Islands in the York River. (Image courtesy of Robert Diaz, VIMS)

found a community very similar in composition to that which has been found in unvegetated habitats within the same salinity zone (e.g., BOESCH, 1971; BENDER, 1972; JORDAN *et al.*, 1975). Wass (1972) provided some cataloging of the fauna attached to SAV (e.g., sponges, tunicates, etc.) and Orth and Van Montfrans (1984) and Duffy and Harvilicz (2001) have discussed the composition of the motile epifaunal grazing communities of SAV beds in the higher salinity, including amphipods, isopods and snails. Although none of the macroinvertebrates found in beds of SAV are unique to those environments, some of them may be more abundant in SAV than they are in other benthic habitats. Unfortunately, much like oyster reefs, the occurrence of SAV meadows within the York River Estuary has precipitously declined from historical levels in recent decades, due in large part to anthropogenic alterations to the estuary (MOORE *et al.*, 1996; ORTH and MOORE, 1983; MOORE, this S.I.).

Imposed upon the large-scale changes in community structure along the length of the York River Estuary, there are also changes in community structure with depth (DIAZ and SCHAFFNER, 1990; Table 1). The York River Estuary consists of a relatively deep channel (9 – 25 m) flanked by shallow (2-3 m), sometimes quite broad shoals and tidal creeks (SCHAFFNER *et al.*, 2001). In the shallow areas, light may penetrate to the sediment surface where it provides energy for the growth of microphytobenthos, an energy-rich food source for benthic fauna (MACINTYRE *et al.*, 1996; CAHOON, 1999). Phytoplankton production can also have a greater influence on the macrobenthic community in the shallow portions of the estuary, where filter feeding animals have access to the entire overlying water column and living phytoplankton, as opposed to those animals in deeper parts of the estuary that are isolated from the photic zone by stratification of the water column (GERRITSEN *et al.*, 1994). Relatively labile detrital materials may also be available due to the proximity to marshes and SAV beds. These additional food sources allow for higher productivity of the benthic community in areas where recruitment and growth are not limited by other factors (BEUKEMA and CADEE, 1997).

While food availability may enhance the potential for high secondary productivity in shallow water areas, other factors may be limiting. Physical disturbance due to waves, strong predation, temperature extremes and other factors alter benthic community structure and may limit productivity in shallow water areas despite high food availability (EMERSON, 1989; BEUKEMA and CADEE, 1997; HARLEY *et al.*, 2006). Predation on meio- and macrobenthos is often intense in shallow water areas due to the juxtaposition of highly productive shallow water benthic habitats with marsh and SAV beds that provide smaller predators of benthic infauna, such as juvenile fish, crabs, and large infauna, refuge from larger predators (KNEIB, 1997; SEITZ *et al.*, 2005; SEITZ *et al.*, 2006). Benthic invertebrates living in shallow subtidal and intertidal zones are also subject to predation by birds (KIVIAT, 1989).

THE IMPORTANCE OF BENTHIC FAUNA

Despite their relatively small size and cryptic lifestyle, macro and meiobenthos are important components of the estuarine ecosystem, serving as critical links between the variety of organic matter sources in estuaries (e.g., phytoplankton, benthic micro- and macroalgae, detritus) and the economically, ecological, and recreationally important finfish and crustaceans that live there (CICCHETTI, 1998). Baird & Ulanowicz (1989) estimated that approximately 50% of the fish production in Chesapeake Bay is directly linked to a benthic food web. Diaz and Schaffner (1990) estimated that 194,000 metric tons of carbon is produced by benthic macrofauna in Chesapeake Bay each year (70% of which occurs in high mesohaline and polyhaline habitats) and supports a fisheries yield of 27,500 metric tons of carbon. Commercial fisheries of benthic feeding and demersal nekton (e.g., spot, croaker, blue crabs) in the Virginia portion of Chesapeake Bay yielded an annual average of 39.8 million dollars of revenue between 1998 and 2002 (NMFS, FISHERIES STATISTICS AND ECONOMICS DIVISION, 2004). Direct harvest of benthic species, especially the oysters and other bivalves, were historically important fisheries in the York River Estuary (WHEATLEY, 1959; BENDER, 1987; HARGIS and HAVEN, 1999), though now they constitute less than one million dollars in landings Bay-wide (NMFS, FISHERIES STATISTICS AND ECONOMICS DIVISION, 2004) (Figure 9). Commercial aquaculture of bivalve molluscs, particularly the hard clam *Mercenaria mercenaria*, has become an important economic force in the Chesapeake Bay as a whole (CAMARA, 2001; VA SEA GRANT, 2007), though there are no large-scale operations within the York River Estuary. Benthic communities also provide a variety of ecosystem services that affect water and sediment quality in the estuaries. In relatively shallow areas, filter feeders may effectively remove particles from the water column, which leads to deposition of organic matter from the overlying water at rates greater than natural sinking and physical mixing would allow. This can result in enhanced water clarity, which may increase the success of SAV (NEWELL and KOCH, 2004). SAV may also enhance particle deposition due to a baffling effect. Biodeposition by filter feeders also serves to shunt water column production to the sediment bed where transport, transformation and fates are then governed by benthic rather than pelagic processes (COHEN *et al.*, 1984; GERRITSEN *et al.*, 1994; NEUBAUER, 2000). Some of this organic matter will fuel the production of benthic invertebrates and their predators. Organic matter that is not assimilated by macro and meiobenthic organisms may be buried, but more

likely, it will be processed by microbes. The released nutrients and breakdown products may be retained in sediment pore waters or fluxed across the sediment-water interface.

Microbial processes generally control the rates of most important biogeochemical processes in the sediment, while meio- and macrobenthos control the mixing of constituents such as oxygen and organic matter that settles or is deposited to the estuary floor. Bioturbation and biogenic structuring of the bottom by benthic organisms has been shown to have major effects on carbon, nitrogen, phosphorus, and contaminant cycling and fate (DIAZ and SCHAFFNER, 1990). The degradation of organic matter and some contaminants is generally enhanced in the presence of infaunal organisms, due to stimulation of microbial processes, which leads to enhanced rates of mineralization (ALLER and ALLER, 1998; KRISTENSEN, 2000). Bioturbation and sediment ventilation by larger benthic organisms tend to enhance the diffusivity of dissolved constituents such as ammonium into the water column (RICE and RHOADES, 1989; MICHAUD *et al.*, 2005; MICHAUD *et al.*, 2006). Simultaneously, reduction/oxidation sensitive processes, such as nitrification-denitrification, may be enhanced in the presence of macrofauna whose tubes and burrows increase the surface area of the sediment-water interface and the depth of oxygen penetration into the sediment. The enhanced coupling of nitrification-denitrification in the presence of benthic macrofauna can lead to the production of nitrogen gas, which escapes to the atmosphere, thereby reducing the nitrogen load in the estuary (MAYER *et al.*, 1995).

THE BENTHIC FAUNA OF CBNEERVA

As noted above, the shallow waters of the York River Estuary historically contained a variety of different habitat types, with extensive SAV beds and oyster reefs interspersed with open areas of mud and sand flats. At present, the estuary is flooded mostly by unvegetated mud or sand sediments with very limited, narrow bands of SAV beds in some areas. As such, soft sediment communities have been the most well-studied, both temporally and spatially (see Studies of the Benthic Fauna of the York River, above). These habitats provide the best characterization the benthic communities throughout the whole estuary and within each of the salinity zones where the different parts of the CBNERRS VA reserve are located (Table 1). Within these generalized benthic communities though, there is almost always a considerable amount of patchiness in space for most species and in time for others, particularly those with strongly seasonal recruitment (e.g., bivalves and polychaetes) (KRAVITZ, 1983; ZOBRIST, 1988; HINCHEY, 2002).

INVASIVE/NON-NATIVE ORGANISMS IN THE YORK RIVER ESTUARY

The presence or distribution of invasive benthic fauna in the York River Estuary remains poorly studied. Invasive taxa have been found in other parts of Chesapeake Bay. The Asian clams *Corbicula manilensis* and *C. fluminea*, which are thought to have invaded other tributaries of the Chesapeake Bay around 1968 (WASS, 1972; DIAZ, 1974; PHELPS, 1994), were not historically observed in the York River Estuary (BOESCH, 1971), but have recently been collected in the Chesapeake Bay Benthic Monitoring Program (CHESAPEAKE BAY PROGRAM, 2009). There are regular observations of the veined rapa whelk *Ra-*

pana venosa (Figure 12), an invasive gastropod accidentally introduced to the high mesohaline/polyhaline York River in the mid-1990's. This species may severely impacts bivalve fisheries via predation (HARDING and MANN, 2005). Additionally, the history of colonial activity in the York River increases the likelihood that some of the species considered to be natives were introduced before scientific surveys began.



Figure 12. The invasive gastropod *Rapana venosa* collected from the York River. (Image courtesy of Juliana Harding, VIMS)

There are also examples of deliberate introduction of non-native species, most notably the non-native oysters *Crasostrea gigas* and *C. ariakensis*. These species that have been introduced to the mesohaline and polyhaline portions of the York River in the interest of supplementing/replacing the oyster fishing industry, which traditionally was based upon the native *C. virginica*. Introduced non-native species may directly compete with native fauna for resources and serve as means for unintentional introductions of parasites and other cryptic fauna associated the non-natives (DOBSON and MAY, 1986; CARLTON, 1992). In recognition of these potential problems, only sterilized, non-reproductive *C. ariakensis* have been introduced to date into the York River in experimental deployments by VIMS and the Virginia Seafood Council. In the end, the true abundance and distribution of invasive benthic taxa in estuaries like the York River and its tributaries will remain difficult to definitively quantify due to the size of the estuary, the cryptic nature of native and non-native benthic organisms, and the ephemeral and stochastic nature of most invasions (CARLTON, 1996).

HUMAN PERTURBATIONS OF BENTHIC FAUNA

The annual benthic monitoring program of the Chesapeake Bay Program assesses the quality and degree of benthic habitat degradation in the Chesapeake Bay and its tributaries using the macrobenthos and the Chesapeake Bay Benthic Index of Biotic Integrity (B-IBI) (WEISBERG *et al.*, 1997). Based upon randomly selected sites in 2005 (the most currently available data) and the B-IBI assessment approach, 73% of the area of the York River Estuary failed to meet the restoration goals set by the Chesapeake Bay Program, due in large part to low macrobenthic abundance and biomass (LLANSÓ *et al.*, 2006). The distribution of habitat quality is not uniform along the length of the estuary (Figure 13). Most of the degraded sites fall within the polyhaline and meso-

haline portions of the York River, areas known to be affected by low dissolved oxygen (LLANSÓ *et al.*, 2006). In contrast, benthic communities of sites sampled in the oligohaline and tidal freshwater parts of the York River were assessed as non-degraded (LLANSÓ *et al.*, 2006).

The hypoxic and anoxic waters observed in the York River Estuary are the end product of a complex process created by excessive nutrient inputs into Chesapeake Bay and human development and alteration of the Bay's watershed (REAY and MOORE, this S.I.; DAUER *et al.*, 2000). Hypoxic episodes in the York River are periodic in nature, lasting from hours to over a week at a time during late summer (HAAS, 1977; DIAZ *et al.*, 1992). Direct mortality of benthic fauna via suffocation will occur during persistent, multi-day episodes of hypoxia/anoxia, though the length of time an organism can survive without oxygen will vary from species to species (HOLLAND *et al.*, 1977; DIAZ and ROSENBERG, 1995; SAGASTI *et al.*, 2001; SAGASTI *et al.*, 2003). Relatively low levels of dissolved oxygen are always present in the sediment of estuaries given the abundance of organic matter and the subsequent respiration of heterotrophic bacteria. These processes result in the accumulation of reduced compounds in the sediment pore waters (e.g., sulphides, ammonia) that are toxic to benthic organisms (THEEDE *et al.*, 1973; PEARSON and ROSENBERG, 1978; SHIN *et al.*, 2006). Water column hypoxia exacerbates the sediment system, increasing the concentrations of reduced chemicals and preventing a source of oxygen to oxidize and remove these

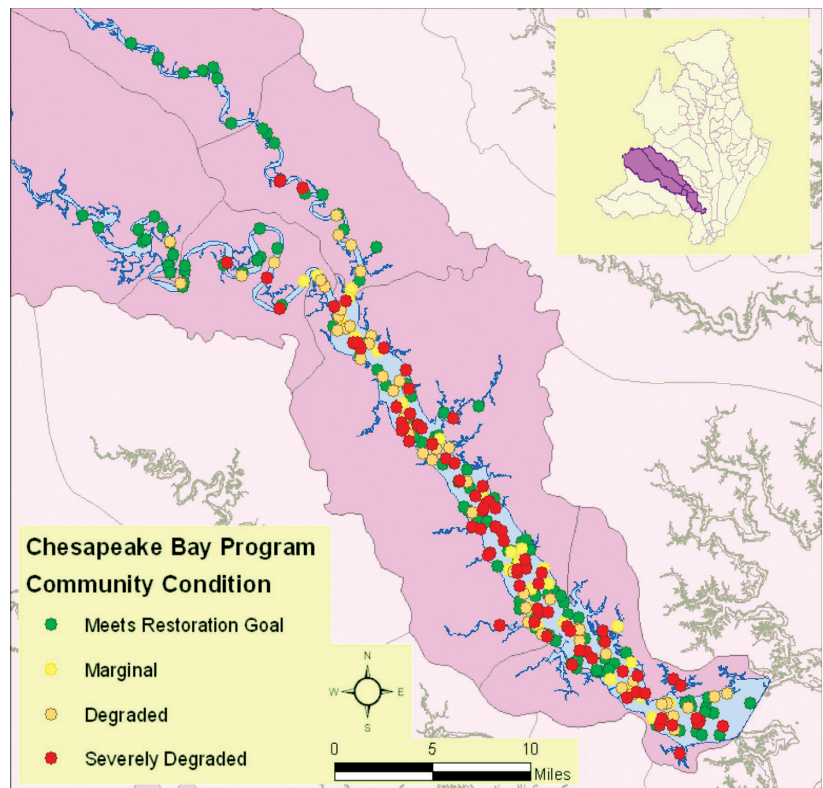


Figure 13. Benthic habitat condition at randomly selected sites within the York River Estuary from 1996 – 2006. Benthic habitat condition was assessed using the Chesapeake Bay Benthic Index of Biotic Integrity (B-IBI) and graded using the Chesapeake Bay Program's restoration guidelines: B-IBI \geq 3.0 = Meets Restoration Guidelines; 2.7 – 2.9 = Marginal; 2.1 - 2.6 = Degraded; and \leq 2.0 = Severely Degraded, as noted in the legend. (Data from CBP database and figure created by David Parrish, CBNERRSVA)

toxic chemicals (GASTON *et al.*, 1985; DIAZ and ROSENBERG, 1995; LEVIN, 2003).

By most accounts, the York River Estuary is not systematically affected by chemically contaminated sediments, unlike more developed parts of Chesapeake Bay (e.g., Elizabeth River, Baltimore Harbor, etc.) (LLANSÓ *et al.*, 2006). That said, there are inevitably instances of local contamination in areas surrounding the various marinas along the length of the estuary, the U.S. Navy installations in the mesohaline estuary, and the coal-fired power plant and petroleum refinery in the polyhaline parts of the estuary. Fuels spills that contain toxic polycyclic aromatic hydrocarbons (PAHs) occur; older military landfills leach a variety of toxic compounds (e.g., chlorinated compounds or asbestos), and anti-fouling compounds with heavy metals leach from ships into the water column, all of which can bind to sediments and negatively impact the benthic fauna of the estuary (e.g., JORDON *et al.*, 1975; LYNCH and BULL, 2007; USEPA 2007).

AREAS OF FUTURE RESEARCH

One of the strategic goals of the National Estuarine Research Reserve System program is to characterize and monitor the biological and community conditions of the reserves, to establish reference conditions, and to quantify change (NERRS 2005). Thus, an understanding of the composition of the benthic community should be of primary concern to the CB NERRS VA program. A comprehensive baseline inventory of the benthic fauna at each of the reserve sites, from the sand and mud flats of the Goodwin Islands to the tidal creeks of Sweet Hall Marsh. Recent research projects conducted at different parts of the reserve system will provide some insight into the macro- and meiobenthic community structure (GILLETT, unpublished; SCHAFFNER and GILLETT, unpublished) and serve as a good starting point, but these studies were not designed to catalog the entire benthic community. Without knowledge of the fauna of different parts of the York River Estuary, it will be impossible to track future invasions, or to assess the role of anthropogenic factors such as development or climate change, in the alteration of benthic community structure and function. Benthic community data is most acutely lacking in the tidal freshwater and oligohaline portions of the York River Estuary. The reserve would benefit significantly by beginning a benthic community investigation at the Sweet Hall Marsh and Taskinas Creek portions of the reserve system before the further development of the watershed begins to degrade the habitat quality in those regions.

In addition to establishing the resident fauna for each portion of the reserve, habitat mapping and inter-habitat comparisons should be completed. Comparisons of the communities in the unvegetated sediment, natural and artificial hard bottom, and SAV meadows will allow the reserve managers to better assess the ecological complexity and ecosystem services rendered within the different parts of the reserve and along the salinity gradient of the York River Estuary. This is key information needed for developing restoration and mitigation plans, which will become increasingly important as human pressures on the estuary continue to grow.

Finally, very little is known concerning the spatial and temporal extent of hypoxic and anoxic conditions in the small tributaries of the York River Estuary. There is anecdotal evidence that low oxygen conditions occur in the tributaries and

creeks of the estuary that can severely impact and degrade the benthic community (Gillett personal observation), but there is little direct, quantitative evidence. Given the spatial extent of these shallow tributaries and their high primary and secondary productivity, the impact of hypoxia-induced mortality on these areas could drastically reduce the ecosystem productivity of the estuary. The CBNERRS VA program would be well equipped to investigate these areas.

FINAL OBSERVATIONS ON THE BENTHOS OF THE YORK RIVER

The York River Estuary and the component NERRS sites comprise a large, complex ecosystem. The resident benthic fauna represent a wide array of trophic and taxonomic diversity. From well-known taxa like the eastern oyster *Crassostrea virginica* or the hard clam *Mercenaria mercenaria* to the relatively obscure harpacticoid copepods or capitellid polychaetes, benthic organisms play a vital role the functioning of the estuarine system. The benthic fauna of the York, Pamunkey, and Mattaponi rivers, like all of their biological resources, are still relatively non-disturbed compared to many parts of the Chesapeake Bay. That said, the benthic communities of the estuary will change and lose their ecological and economic value as the continuing developmental pressure within the estuarine watershed continues to increase, as it has in the coastal zone around the country (BEACH, 2000; PEW OCEAN COMMISSION, 2003). The preservation and research of a diversity of benthic habitats by the Virginia CBNERRS program has been, and will continue to act as, part of the counterbalance to the forces of development in and along the York River Estuary. We have a rudimentary understanding of the functioning of the hidden and fascinating world of benthic fauna, but there is still much more for us to learn there.

ACKNOWLEDGEMENTS

The authors thank Dr. Robert Diaz for providing numerous reports, unpublished data, and benthic camera images used in the paper and Dr. Juliana Harding provided input and images on the rapa whelks in the York River. The authors also thank Mr. David Parrish for GIS assistance.

LITERATURE CITED

- ALDEN, R.W. III, D.M. DAUER, J.A. RANASINGHE, L.C. SCOTT, and R.J. LLANSÓ, 2002. Statistical verification of the Chesapeake Bay benthic index of biotic integrity. *Environmetrics*, 13, 473-498.
- ALLER, R.C. and J.Y. ALLER, 1998. The effect of biogenic irrigation intensity and solute exchange on diagenetic reaction rates in marine sediments. *Journal of Marine Research*, 56, 905-936.
- ALONGI, D.M., R.J. DIAZ, and D.F. BOESCH, 1982. Meiobenthic recolonization and succession in oiled azoic subtidal sands of the York River estuary, Virginia. *Marine Biology*, 72, 325-335.
- ARMITAGE, M.J. and J.O. YOUNG, 1990. The realized food niches of three species of stream-dwelling triclads (Turbellaria). *Freshwater Biology*, 24, 93-100.
- ATTRILL, M.J., 2002. A testable linear model for diversity trends in estuaries. *Journal of Animal Ecology*, 71, 262-269.
- BAIRD, D. and R.E. ULANOWICZ, 1989. The seasonal dynamics of the Chesapeake Bay ecosystem. *Ecological Monographs*, 59, 329-364.
- BEACH, D., 2002. Coastal Sprawl: The Effects of Urban Design on Aquatic Ecosystems in the United States. Pew Oceans Commission, Arlington, VA.

- BENDER, M.E., 1972. Benthos and Shellfish. *In*: R.A. Jordan (ed.), The Environmental and Biological Community of the Lower York River, Virginia: A Literature Review. Virginia Institute of Marine Science, Gloucester Point, VA, pp. 101-160.
- BENDER, M.E., 1987. The York River: A Brief Review of Its Physical, Chemical, and Biological Characteristics. The American Petroleum Institute, Washington, D.C.
- BEUKEMA, J.J. and G.C. CADEE, 1997. Local differences in macrozoobenthic response to enhanced food supply caused by mild eutrophication in the Wadden Sea area: food is only locally a limiting factor. *Limnology and Oceanography*, 42, 1424-1435.
- BOESCH, D.F. and R. ROSENBERG, 1981. Response to stress in marine benthic communities. *In*: G. W. Barrett and R. Rosenberg (eds.), *Stress Effects on Natural Ecosystems*. John Wiley and Sons, New York, pp. 179-200.
- BOESCH, D.F., 1971. Distribution and structure of benthic communities in a gradient estuary. Ph.D. Dissertation The College of William and Mary, Williamsburg, VA.
- BOESCH, D.F., 1977. A new look at the zonation of benthos along the estuarine gradient. *In*: B. C. Coull (ed.), *Ecology of Marine Benthos*. University of South Carolina Press, Columbia, SC, pp. 245-266.
- BOESCH, D.F., M.L. WASS, and R.W. VIRNSTEIN, 1976. The dynamics of estuarine benthic communities, p. 177-196. *In*: M. Wiley (ed.), *Estuarine Processes Volume 1 Uses, Stresses, and Adaptation to the Estuary*. Academic Press, New York.
- BOURQUE, D., G. MIRON, and T. LANDRY, 2002. Predator-prey relationship between the nemertean *Cerebratulus lacteus* and the soft-shell clam, *Mya arenaria*: Surface-exploration activity and qualitative observations on feeding behaviour. *Canadian Journal of Zoology*, 80, 1204-1211.
- BRUSCA, R.C. and G.J. BRUSCA, 1990. *Invertebrates*. Sinauer Associates, Sunderland, MA.
- CAHOON, L.B. 1999. The role of benthic microalgae in neritic ecosystems. *Oceanography and Marine Biology an Annual Review*, 37, 47-86.
- CAMARA, M., 2001. A long-term plan to domesticate wild clams. *Virginia Marine Resources Bulletin* 33, 11-14.
- CARLTON, J.T., 1992. Dispersal of living organisms into aquatic ecosystems as mediated by aquaculture and fisheries activities, p. 13-46. *In*: A. Rosenfield and R. Mann (eds.), *Dispersal of Living Organisms into Aquatic Ecosystems*. University of Maryland Press, College Park, MD.
- CARLTON, J.T., 1996. Biological invasions and cryptogenic species. *Ecology*, 77, 1653-1655.
- CERCO, C.F. and M.R. NOEL, 2007. Can oyster restoration reverse cultural eutrophication in Chesapeake Bay? *Estuaries and Coasts*, 30, 331-343.
- CHESAPEAKE BAY PROGRAM, 2009. http://www.chesapeakebay.net/data_benthic.aspx. Accessed March 10, 2009.
- CICCHETTI, G., 1998. Habitat use, secondary production, and trophic export by salt marsh nekton in shallow waters. Ph.D. Dissertation. The College of William and Mary, Williamsburg, VA.
- COEN, L.D., R.D. BRUMBAUGH, D. BUSHK, R. GRIZZLE, M. W. LUCKENBACH, M. H. POSEY, S. P. POWERS, and S. G. TOLLEY, 2007. Ecosystem services related to oyster restoration. *Marine Ecology Progress Series*, 301, 343-347.
- COEN, L.D., M.W. LUCKENBACH, and D.L. BREITBURG, 1999. The role of oyster reefs as essential fish habitat: a review of current knowledge and some new perspectives. *American Fisheries Society Symposium*, 22, 438-454.
- COHEN, R.R. H., P.V. DRESLER, E.J.P. PHILLIPS, and R.L. CORY, 1984. The effect of the Asiatic clam, *Corbicula fluminea*, on phytoplankton of the Potomac River, Maryland. *Limnology and Oceanography*, 29, 170-180.
- DAUER, D.M., 1983. Functional morphology and feeding behavior of *Scolecopsis squamata* (Polychaeta: Spionidae). *Marine Biology*, 77, 279-285.
- DAUER, D.M., J.A. RANASINGHE, and S.B. WEISBERG, 2000. Relationships between benthic community condition, water quality, nutrient loads, and land use patterns in Chesapeake Bay. *Estuaries*, 23, 80-96.
- DEATON, L.E. and M.J. GREENBERG, 1986. There is no horohalimnicum. *Estuaries*, 9, 20-30.
- DELLAPENNA, T.M., S.A. KUEHL, and L.C. SCHAFFNER, 1998. Sea-bed mixing and particle residence times in biologically and physically dominated estuarine systems: a comparison of Lower Chesapeake Bay and the York River subestuary. *Estuarine, Coastal and Shelf Science*, 46, 777-795.
- DELLAPENNA, T.M., S.A. KUEHL, and L.C. SCHAFFNER, 2003. Ephemeral deposition, seabed mixing and fine-scale strata formation in the York River estuary, Chesapeake Bay. *Estuarine, Coastal and Shelf Science*, 58, 621-643.
- DIAZ, R.J. and R. ROSENBERG, 1995. Marine benthic hypoxia: a review of its ecological effects and the behavioral responses of benthic macrofauna. *Oceanography and Marine Biology: An Annual Review*, 33, 245-303.
- DIAZ, R.J., R.J. NEUBAUER, L.C. SCHAFFNER, L. PIHL, and S.P. BADEN, 1992. Continuous monitoring of dissolved oxygen in an estuary experiencing periodic hypoxia and the effect of hypoxia on macrobenthos and fish. *Science of the Total Environment, Supplement*, 1055-1068.
- DIAZ, R.J., 1974. Asiatic Clam, *Corbicula manilensis* (Philippi), in the Tidal James River, Virginia. *Chesapeake Science*, 15, 118-120.
- DIAZ, R.J., 1989. Pollution and tidal benthic communities of the James River estuary, Virginia. *Hydrobiologia*, 180, 195-211.
- DIAZ, R.J., 1994. The response of tidal freshwater macrobenthos to sediment disturbance. *Hydrobiologia*, 278, 201-212.
- DIAZ, R.J. and L.C. SCHAFFNER, 1990. The Functional Role of Estuarine Benthos, p. 25-56. *In*: M. Haire and E.C. Krome (eds.), *Perspectives on the Chesapeake Bay, 1990. Advances in Estuarine Sciences*. United States Environmental Protection Agency, Gloucester Point, VA.
- DOBSON, A.P. and R.M. MAY, 1986. Patterns of invasions by pathogens and parasites, p. 58-76. *In*: H. A. Mooney and J. A. Drake (eds.), *Ecology of Biological Invasions of North America and Hawaii*. Springer-Verlag, New York.
- DUFFY, J.E. and A.M. HARVILICZ, 2001. Species-specific impacts of grazing amphipods in an eelgrass-bed community. *Marine Ecology Progress Series*, 223, 201-211.
- EMERSON, C.W., 1989. Wind stress limitation of benthic secondary production in shallow, soft-sediment communities. *Marine Ecology Progress Series*, 53, 65-77.
- FAUCHALD, K. and P.A. JUMARS, 1979. The diet of worms: a study of polychaete feeding guilds. *Oceanography and Marine Biology: An Annual Review*, 17, 193-284.
- GAINES, L.F., JR. and M.J. GREENBERG, 1977. Physiological basis of the species abundance-salinity relationship in molluscs: a speculation. *Marine Biology*, 40, 41-49.
- GASTON, G.R., P.A. RUTLEDGE, and M.L. WALTHER, 1985. The effects of hypoxia and brine on recolonization by macrobenthos off Cameron, Louisiana (USA). *Contributions in Marine Science*, 28, 79-93.
- GERRITSEN, J., A.F. HOLLAND, and D.E. IRVINE, 1994. Suspension-feeding bivalves and the fate of primary production: an estuarine model applied to Chesapeake Bay. *Estuaries*, 17, 403-416.
- HAAS, L.W. 1977. The effect of the spring-neap tidal cycle on the vertical salinity structure of the James, York and Rappahannock Rivers, Virginia, USA. *Estuarine and Coastal Marine Science*, 5, 485-496.
- HARDING, J.M. and R. MANN, 2005. Veined rapa whelk (*Rapana venosa*) range extensions in the Virginia waters of the Chesapeake Bay, USA. *Journal of Shellfish Research*, 24, 381-385.
- HARGIS, W.J., JR., and D.S. HAVEN, 1999. Chesapeake oyster reefs, their importance, destruction and guidelines for restoring them, p. 329-358. *In*: M. W. Luckenbach, R. Mann, and J. A. Wesson (eds.), *Oyster Reef Habitat Restoration: A Synopsis and Synthesis of Approaches*. Virginia Institute of Marine Science. Gloucester Point, VA.
- HARLEY, C.D.G., A.R. HUGHS, K.M. HULTGREN, B.G. MINER, C.J.B. SORTE, C.S. THORNER, L.F. RODRIGUEZ, L. TOMANEK, and S.L. WILLIAMS, 2006. The impacts of climate change in coastal marine systems. *Ecology Letters*, 9, 228-241.
- HARWELL, H.D., 2004. Landscape aspects of oyster reefs: Fragmentation and habitat utilization. M.S. Thesis. University of North Carolina at Wilmington, Wilmington, NC, 73p.
- HAVEN, D.S., J.N. KRAEUTER, R.C. SWARTZ, and R. MORALES-ALAMO, 1981. An Animal-Sediment Study in the Lower York River - February 1965 to February 1966. Special Scientific Report No. 108, Virginia Institute of Marine Science, Gloucester Point, VA, 70p.

- HIGGINS, R. P. and H. THIEL, 1988. *Introduction to the Study of Meiofauna*. Smithsonian Institution Press, Washington, D.C.
- HINCHEY, E.K., 2002. Organism-Sediment Interactions: The Role of Seabed Dynamics in Structuring the Mesohaline York River Macro-benthic Community. Ph. D. Dissertation. The College of William and Mary, Williamsburg, VA.
- HOLLAND, A.F., D.M. SANGER, C.P. GAWLE, S.B. LERBERG, M.S. SANTIAGO, G.H.M. RIEKERK, L.E. ZIMMERMAN, and G.I. SCOTT, 2004. Linkages between tidal creek ecosystems and the landscape and demographic attributes of their watersheds. *Journal of Experimental Marine Biology and Ecology*, 298, 151-178.
- HOLLAND, A.F., N.K. MOUNTFORD, and J.A. MIHURSKY, 1977. Temporal variation in Upper Bay mesohaline benthic communities: I. The 9-m mud habitat. *Chesapeake Science*, 18, 370-378.
- INTERNATIONAL ASSOCIATION OF LIMNOLOGY, 1958. The Venice system for the classification of marine waters according to salinity. *Limnology and Oceanography*, 3, 346-347.
- JORDAN, R.A., R.W. VIRNSTEIN, J.E. ILLOWSKY, and J. COIVOCORESSES, 1975. Yorktown Power Station Ecological Study Phase II Final Technical Report. Virginia Institute of Marine Science, Gloucester Point, VA.
- KINNE, O., 1971. Salinity. In: O. Kinne (eds.), *Marine Ecology A Comprehensive, Integrated, Treatise on Life in Oceans and Coastal Waters Volume 1 Environmental Factors Part 2*. Wiley-Interscience, New York, pp. 821-1083.
- KIVIAT, E., 1989. The role of wildlife in estuarine ecosystems. In: J. W. Day, Jr., C. A. S. Hall, W. M. Kemp, and A. Yañez-Arancibia (eds.), *Estuarine Ecology*. John Wiley and Sons, New York, pp. 438-476.
- KNEIB, R.T., 1997. The role of tidal marshes in the ecology of estuarine nekton. *Oceanography and Marine Biology: An Annual Review*, 35, 163-200.
- KRAVITZ, M.J., 1983. Niche Relations Among Polychaetes of a Meso-Polyhaline Assemblage in the York River, Virginia. Ph. D. Dissertation. The College of William and Mary, Williamsburg, VA.
- KRISTENSEN, E., 2000. Organic matter diagenesis at the oxic/anoxic interface in coastal marine sediments, with emphasis on the role of burrowing animals. *Hydrobiologia*, 426, 1-24.
- LEVIN, L.A., 2003. Oxygen minimum zone benthos: adaptations and community response to hypoxia. *Oceanography and Marine Biology: An Annual Review*, 41, 1-45.
- LLANSÓ, R.J., J. DEW, and L.C. SCOTT, 2006. Chesapeake Bay Water Quality Monitoring Program Long-Term Benthic Monitoring and Assessment Component Level 1 Comprehensive Report July 1984 - December 2005 (Volume 1). Versar, Inc, Co.
- LLANSÓ, R.J., 2005. Chesapeake Bay Soft-Bottom Benthic Macrofauna Species List, www.baybenthos.versar.com. Last accessed July 2007.
- LLANSÓ, R.J., D.M. DAUER, J.H. VÖLSTAD, and L.C. SCOTT, 2003. Application of the benthic index of biotic integrity to environmental monitoring in Chesapeake Bay. *Environmental Monitoring and Assessment*, 81: 163-174.
- LOPEZ, G.R. and J.S. LEVINTON, 1987. Ecology of deposit-feeding animals in marine sediments. *The Quarterly Review of Biology*, 62: 235-260.
- LYNCH, P. and J. M.R. BULL, 2007. Poison Left Behind. Daily Press. June 10, 2007.
- MACINTYRE, H.L., R.J. GEIDER, and D.C. MILLER, 1996. Microphytobenthos: the ecological role of the "Secret Garden" of unvegetated, shallow-water marine habitat. I. Distribution, abundance, and primary production. *Estuaries*, 19:186-201.
- MARSH, G.A., 1970. A Seasonal Study of *Zostera Epibiota* in the York River, Virginia. Ph. D. Dissertation. The College of William and Mary, Williamsburg, VA.
- MARTENS, P.M. and E.R. SCHOCKAERT, 1986. The importance of turbellarians. *Hydrobiologia*, 132: 295-303.
- MAYER, M.S., L.C. SCHAFFNER, and W.M. KEMP, 1995. Nitrification potentials of benthic macrofaunal tubes and burrow walls: effects of sediment NH₄⁺ and animal irrigation behavior. *Marine Ecology Progress Series*, 121, 157-169.
- METCALFE, W.J., 2005. Meiofauna Abundance and Distribution in Chesapeake Bay: Relationships with Environmental Stressors, Sediment Toxicity and Macrofauna. M.S. Thesis. The College of William and Mary.
- MICHAUD, E., G. DESROSIERS, F. MERMILLOD-BLONDIN, B. SUNDBY, and G. STORA, 2005. The functional group approach to bioturbation: the effects of biodifusers and gallery-diffusers of the *Macoma balthica* community on sediment oxygen uptake. *Journal of Experimental Marine Biology and Ecology*, 326, 77-88.
- MICHAUD, E., G. DESROSIERS, F. MERMILLOD-BLONDIN, B. SUNDBY, and G. STORA, 2006. The functional group approach to bioturbation: II. The effects of the *Macoma balthica* community on fluxes of nutrients and dissolved organic carbon across the sediment-water interface. *Journal of Experimental Marine Biology and Ecology*, 337, 178-189.
- MOORE, K.A., H.A. NECKLES, and R.J. ORTH, 1996. *Zostera marina* (eelgrass) growth and survival along a gradient of nutrients and turbidity in the lower Chesapeake Bay. *Marine Ecology Progress Series*, 142, 247-259.
- NATIONAL ESTUARINE RESEARCH RESERVE, 2005. Strategic Plan 2005-2010. NOAA National Ocean Service, Silver Springs, MD.
- NATIONAL MARINE FISHERIES SERVICE, 2007. 2004 Landings Data. <http://www.nmfs.noaa.gov>. Last accessed July 2007.
- NEUBAUER, M.T., 2000. Demographics, production, and benthic-pelagic coupling by the suspension feeding polychaete *Chaetopterus pergamentaceus* in the Lower Chesapeake Bay. Ph. D. Dissertation. The College of William and Mary, Williamsburg, VA.
- NEWELL, R.I.E. and E.W. KOCH, 2004. Modeling seagrass density and distribution in response to changes in turbidity stemming from bivalve filtration and seagrass sediment stabilization. *Estuaries*, 27, 793-806.
- ORTH, R.J., 1973. Benthic infauna of eelgrass, *Zostera marina*, beds. *Chesapeake Science*, 14, 258-269.
- ORTH, R.J. and J. VAN MONTFRANS, 1984. Epiphyte-seagrass relationships with an emphasis on the role of micrograzing: a review. *Aquatic Botany*, 18, 43-69.
- ORTH, R.J. and K.A. MOORE, 1983. Chesapeake Bay: an unprecedented decline in submerged aquatic vegetation in Chesapeake Bay. *Science*, 222, 51-53.
- PEW OCEANS COMMISSION, 2003. America's Living Oceans: Charting a Course for Sea Change Summary Report 2003, Pew Ocean's Commission, 35p.
- PEARSON, T.H. and R. ROSENBERG, 1978. Macrobenthic succession in relation to organic enrichment and pollution of the environment. *Oceanography and Marine Biology: An Annual Review*, 16, 229-311.
- PHELPS, H.L., 1994. The Asiatic clam (*Corbicula fluminea*) invasion and system-level ecological changes in the Potomac River estuary near Washington, D. C. *Estuaries*, 17, 614-621.
- POHLO, R., 1982. Evolution of the tellinacea (Bivalvia). *Journal of Molluscan Studies*, 48, 245-256.
- POMEROY, L.R., C.F. D'ELIA, and L.C. SCHAFFNER, 2006. Limits to top-down control of phytoplankton by oysters in Chesapeake Bay. *Marine Ecology Progress Series*, 325, 301-309.
- REMANE, A. and C. SCHLIEPER, 1971. *Biology of Brackish Water. 2nd Edition*. John Wiley and Sons, New York.
- RICE, D.L., and D.C. RHOADS, 1989. Early diagenesis of organic matter and the nutritional value of sediment, p. 59-97. In: G. Lopez, G. Taghon, and J. Levinton (eds.), *Ecology of Marine Deposit Feeders*. Springer-Verlag, New York.
- SAGASTI, A., J.E. DUFFY, and L.C. SCHAFFNER, 2003. Estuarine epifauna recruit despite periodic hypoxia. *Marine Biology*, 142, 111-122.
- SAGASTI, A., L.C. SCHAFFNER, and J.E. DUFFY, 2000. Epifaunal communities thrive in an estuary with hypoxic episodes. *Estuaries*, 23, 474-487.
- SAGASTI, A., L.C. SCHAFFNER, and J.E. DUFFY, 2001. Effects of periodic hypoxia on mortality, feeding, and predation in an estuarine epifaunal community. *Journal of Experimental Marine Biology and Ecology*, 258, 257-283.
- SCHAFFNER, L.C., T.M. DELLAPENNA, E.K. HINCHEY, C.T. FRIEDRICH, M.T. NEUBAUER, M.E. SMITH, and S.A. KUEHL, 2001. Physical energy regimes, seabed dynamics, and organism-sediment interactions along an estuarine gradient, p. 159-179. In: J. Y. Aller, S. A. Woodin, and R. C. Aller (eds), *Organism-Sediment Interactions*. University of South Carolina Press, Columbia, S.C.
- SEITZ, R.D., R.N. LIPCIUS, and M.S. SEEBO, 2005. Food availability and growth of the blue crab in seagrass and unvegetated nurseries of Chesapeake Bay. *Journal of Experimental Marine Biology and Ecology*, 319: 57-68.

- SEITZ, R.D., R.N. LIPCIUS, N.H. OLMSTEAD, M.S. SEEBO, and D.M. LAMBERT, 2006. Influence of shallow-water habitats and shoreline development on abundance, biomass, and diversity of benthic prey and predators in Chesapeake Bay. *Marine Ecology Progress Series*, 326, 11-27.
- SHIN, P.K.S., C.K.C. CHEUNG, and S.F. CHEUNG, 2006. Effects of nitrogen and sulphide on macroinfaunal community: A microcosm study. *Marine Pollution Bulletin*, 52, 1333-1339.
- SIN, Y., R.L. WETZEL, and I.C. ANDERSON, 1999. Spatial and temporal characteristics of nutrient and phytoplankton dynamics in the York River estuary, Virginia: analyses of long-term data. *Estuaries*, 22, 260-275.
- STEPHENSON, J., 1972. *The Oligochaeta*. Stechert-Hafner Service Agency, Inc., New York.
- TAGHON, G.L., A.R.M. NOWELL, and P.A. JUMARS, 1980. Induction of suspension feeding in Spionid polychaetes by high particle fluxes. *Science*, 210, 562-564.
- THEEDE, H., 1973. Comparative studies on the influence of oxygen deficiency and hydrogen sulphide on marine bottom invertebrates. *Netherlands Journal of Sea Research* 7:244-252.
- UNITED STATES ENVIRONMENTAL PROTECTION AGENCY. 2007. National Priorities List Site Narrative for Naval Weapons Station – Yorktown. <http://www.epa.gov/superfund/sites/npl/nar1325.htm>. Last accessed July 2007.
- VA SEAGRANT, 2007. Virginia Shellfish Aquaculture Situation and Outlook Report. Results of Virginia Shellfish Aquaculture Crop Reporting Survey 2005-2007. Virginia Institute of Marine Science, Gloucester Point, VA, 12p
- WASS, M.L., 1965. Checklist of the marine invertebrates of Virginia. Special Scientific Report no. 24. Virginia Institute of Marine Science, Gloucester Point, VA, 58p.
- WASS, M.L., 1972. A checklist of the biota of lower Chesapeake Bay: with inclusions from the upper bay and the Virginian Sea. Special Scientific Report no. 65, Virginia Institute of Marine Science, Gloucester Point, VA, 290p.
- WEISBERG, S.B., J.A. RANASINGHE, D.M. DAUER, L.C. SCHAFFNER, R.J. DIAZ, and J.B. FRITHSEN, 1997. An estuarine benthic index of biotic integrity (B-IBI) for Chesapeake Bay. *Estuaries*, 20, 149-158.
- WHEATLEY, J.J., 1959. *The Economic Implications of the York River Oyster Industry*. University of Virginia, Charlottesville, VA.
- ZOBRIST, E.C., 1988. The Influence of Post-Settlement Mortality on Recruitment Patterns in a Soft-bottom Habitat. M.S. Thesis, Virginia Institute of Marine Science, College of William and Mary, Gloucester Point, VA.

APPENDIX

A Partial Species List of Benthic Fauna Collected in the York River Estuary

Scientific name and the corresponding Integrated Taxonomic Information System Serial Code (TSN) where available.

Annelids

<i>Aglaophamus verrilli</i>	0066052	<i>Carazziella hobsonae</i>	0067003
<i>Almyracuma proximoculi</i>	0066052	<i>Caulleriella killariensis</i>	0067131
<i>Amastigos caperatus</i>		<i>Chaetopterus variopedatus</i>	0067097
<i>Ampharetidae</i>	0067718	<i>Cirratulidae</i>	0067116
<i>Amphicteis floridus</i>	0067753	<i>Clymenella torquata</i>	0067528
<i>Amphicteis gunneri</i>	0067747	<i>Cossura longocirrata</i>	0067207
<i>Ancistrosyllis commensalis</i>	0065548	<i>Demonax microphthalmus</i>	0068222
<i>Ancistrosyllis hartmanae</i>	0065543	<i>Dero digitata</i>	0068904
<i>Ancistrosyllis jonesi</i>	0065544	<i>Dero obtusa</i>	0068907
<i>Ancistrosyllis</i> sp.	0065541	<i>Dero</i> sp.	0068898
<i>Asabellides oculata</i>	0067786	<i>Diopatra cuprea</i>	0066180
<i>Aulodrilus limnobius</i>	0068682	<i>Dorvillea rudolphi</i>	0066525
<i>Aulodrilus pigueti</i>	0068680	<i>Drilonereis longa</i>	0066426
<i>Bhawania goodei</i>	0065158	<i>Drilonereis</i> sp.	0066423
<i>Bhawania heteroseta</i>	0065159	<i>Eteone heteropoda</i>	0065266
<i>Boccardiella ligERICA</i>	0067012	<i>Eteone lactea</i>	0065267
<i>Branchiura sowerbyi</i>	0068621	<i>Eumida sanguinea</i>	0065343
<i>Brania wellfleetensis</i>	0065762	<i>Glycera americana</i>	0066106
<i>Bratislavia unidentata</i>	0069023	<i>Glycera dibranchiata</i>	0066107
<i>Cabira incerta</i>	0065565	<i>Glycera</i> sp.	0066102
<i>Calliobdella virida</i>	0069351	<i>Glycinde solitaria</i>	0066132
<i>Capitella capitata</i>	0067415	<i>Gyptis</i> sp.	0065468
<i>Capitella jonesi</i>		<i>Gyptis vittata</i>	0065470
<i>Capitellidae</i>	0067413	<i>Haber speciosus</i>	0068745
<i>Capitomastus aciculatus</i>	0204558	<i>Harmothoe extenuata</i>	0064509
		<i>Harmothoe</i> sp.	0064502
		<i>Helobdella elongata</i>	0069397
		<i>Helobdella stagnalis</i>	0069398
		<i>Heteromastus filiformis</i>	0067420
		<i>Hirudinea</i>	0069290
		<i>Hobsonia florida</i>	0067755

<i>Hydroides dianthus</i>	0068282	<i>Pseudeurythoe paucibranchiata</i>	0065176
<i>Hydrilus templetoni</i>	0068662	<i>Pseudeurythoe</i> sp.	0065174
<i>Isochaetides freyi</i>	0068810	<i>Quistradrilus multisetosus</i>	0068794
<i>Laeonereis culveri</i>	0065965	<i>Sabaco elongatus</i>	BAY0341
<i>Leitoscoloplos fragilis</i>	0066656	<i>Sabella microphthalmia</i>	0068223
<i>Leitoscoloplos robustus</i>	0182728	<i>Sabellaria vulgaris</i>	0067671
<i>Leitoscoloplos</i> sp.	0066653	<i>Samythella elongata</i>	0067802
<i>Lepidametria commensalis</i>	0064703	<i>Schistomeringos rudolphi</i>	0066523
<i>Lepidonotus sublevis</i>	0064610	<i>Scolelepis bousfieldi</i>	0066944
<i>Lepidonotus variabilis</i>	0064611	<i>Scolelepis</i> sp.	0066942
<i>Levensenia gracilis</i>	0066729	<i>Scolelepis squamata</i>	0066943
<i>Limnodriloides anxius</i>	0158432	<i>Scolelepis texana</i>	0066949
<i>Limnodrilus hoffmeisteri</i>	0068639	<i>Scoloplos rubra</i>	0066603
<i>Limnodrilus profundicola</i>	0068649	<i>Sigambra bassi</i>	0065554
<i>Limnodrilus</i> sp.	0068638	<i>Sigambra tentaculata</i>	0065552
<i>Linopherus paucibranchiata</i>	0065175	<i>Spio setosa</i>	0066868
<i>Loimia medusa</i>	0068015	<i>Spiochaetopterus costarum</i>	0067107
<i>Macroclymene zonalis</i>	0067632	<i>Spiochaetopterus oculatus</i>	0067110
<i>Maldanidae</i>	0067515	<i>Spiophanes bombyx</i>	0066897
<i>Malmgreniella taylori</i>	BAY0335	<i>Spirosperma ferox</i>	0068610
<i>Manayunkia aestuarina</i>	0068171	<i>Stephensomiana trivandrana</i>	0069018
<i>Marenzelleria viridis</i>	0573739	<i>Sthenelais boa</i>	0065084
<i>Mediomastus ambiseta</i>	0067439	<i>Streblospio benedicti</i>	0066939
<i>Melinna maculata</i>	0067766	<i>Terebellidae</i>	0067899
<i>Microphthalmus szcelkowi</i>	0065477	<i>Tharyx acutus</i>	0067147
<i>Microphthalmus</i> sp.	0065476	<i>Tharyx setigera</i>	0067145
<i>Monticellina dorsobranchialis</i>	0204530	<i>Tubifex</i> sp.	0068622
<i>Mystides borealis</i>	0065307	<i>Tubificidae</i>	0068585
<i>Myzobdella lugubris</i>	0069316	<i>Tubificoides benedeni</i>	0068592
<i>Nais communis</i>	0068950	<i>Tubificoides brownae</i>	0068688
<i>Nais variabilis</i>	0068959	<i>Tubificoides diazi</i>	0068689
<i>Neanthes succinea</i>	0065918	<i>Tubificoides gabriellae</i>	0068590
<i>Nephtys incisa</i>	0066028	<i>Tubificoides heterochaetus</i>	0068595
<i>Nephtys picta</i>	0066030	<i>Tubificoides motei</i>	
<i>Nephtys</i> sp.	0066011	<i>Tubificoides</i> sp.	0068687
<i>Nereidae</i>	0065870	<i>Tubificoides wasselli</i>	0068692
<i>Nereis acuminata</i>	0065926		
<i>Notomastus</i> sp.	0067423	Ascidians	
<i>Oligochaeta</i>	0068422	<i>Ascidacea</i>	0158854
<i>Orbiniidae</i>	0066570	<i>Botryllus schlosseri</i>	0159373
<i>Palaenotus heteroseta</i>	0065152	<i>Molgula lutulenta</i>	0159581
<i>Parahesionia luteola</i>	0065493	<i>Molgula manhattensis</i>	0159557
<i>Paranais frici</i>	0068865		
<i>Paranaitis speciosa</i>	0065321	Chordates	
<i>Paraprionospio pinnata</i>	0066937	<i>Branchiostoma caribaeum</i>	0159682
<i>Pectinaria gouldi</i>	0067709	<i>Branchiostoma virginiae</i>	0206924
<i>Phyllodoce arenae</i>	0065366		
<i>Phyllodoce fragilis</i>	0065337	Cnidarians	
<i>Podarke obscura</i>	0065517	<i>Actiniaria</i> sp.	0052485
<i>Podarkeopsis brevipalpa</i>	0065532	<i>Anthozoa</i>	0051938
<i>Podarkeopsis</i> sp.	0065530	<i>Ceriantheopsis americanus</i>	0051992
<i>Pokarkeopsis levifuscina</i>	0555698	<i>Cerianthus americanus</i>	0051987
<i>Polycirrus eximius</i>	0067963	<i>Clytia cylindrica</i>	
<i>Polydora cornuta</i>	0204501	<i>Diadumene leucolena</i>	0052749
<i>Polydora ligni</i>	0066801	<i>Ectopleura dumortieri</i>	0719102
<i>Polydora socialis</i>	0066791	<i>Edwardsia elegans</i>	0052489
<i>Polydora websteri</i>	0066802	<i>Haliplanella luciae</i>	0204191
<i>Prionospio perkinsi</i>	0066854	<i>Halopteris tenella</i>	
<i>Pristina breviseta</i>	0068880	<i>Hydrozoa</i>	0048739
<i>Pristinella jenkiniae</i>	0069030	<i>Obelia bidentata</i>	0049532
<i>Pristinella osborni</i>	0069026	<i>Sertularia argentea</i>	0049914
<i>Pristinella sima</i>	0069028		

Crustaceans

<i>Aegathoa medialis</i>	0092440	<i>Melita appendiculata</i>	0093813
<i>Americamysis bigelowi</i>	0682618	<i>Melita nitida</i>	0093812
<i>Ameroculodes</i> sp.	0656551	<i>Microprotopus raneyi</i>	0094122
<i>Ampelisca abdita</i>	0093329	<i>Monocorophium tuberculatum</i>	0656762
<i>Ampelisca macrocephala</i>	0093322	<i>Monoculodes edwardsi</i>	0094539
<i>Ampelisca</i> sp.	0093321	<i>Monoculodes intermedius</i>	0094536
<i>Ampelisca vadorum</i>	0093330	<i>Neomysis americana</i>	0090062
<i>Ampelisca verrilli</i>	0093331	<i>Ogyrides alphaerostris</i>	0096737
<i>Amphiodia atra</i>	0157649	<i>Oxyurostylis smithi</i>	0090923
<i>Amphitoidae</i>	0093408	<i>Palaeomonetes pugio</i>	0096390
<i>Balanoglossus aurantiacus</i>	0158629	<i>Panopeus herbstii</i>	0098778
<i>Balanus eburneus</i>	0089621	<i>Paracaprella tenuis</i>	0095434
<i>Balanus improvisus</i>	0089622	<i>Parametopella cypris</i>	0094927
<i>Batea catharinensis</i>	0093528	<i>Paraphoxus spinosus</i>	0094756
<i>Callinectes sapidus</i>	0098696	<i>Parapleustes estuarius</i>	BAY0199
<i>Campylapsis rubicunda</i>		<i>Pinnixa chaetoptera</i>	0098998
<i>Caprella penantis</i>	0095419	<i>Pinnixa retinens</i>	0099001
<i>Cassidinidea lunifrons</i>	0092347	<i>Pinnixa sayana</i>	0099002
<i>Cerapus tubularis</i>	0093587	<i>Pinnixa</i> sp.	0098993
<i>Chiridotea almyra</i>	0092638	<i>Pleusymtes glaber</i>	0094797
<i>Chiridotea coeca</i>	0092640	<i>Polyonyx gibbesi</i>	0098083
<i>Chiridotea nigrescens</i>	0092642	<i>Ptilanthura tenuis</i>	0092155
<i>Corophium acherusicum</i>	0093590	<i>Rhithropanopeus harrisi</i>	0098790
<i>Corophium insidiosum</i>	0093600	<i>Sarsiella texana</i>	0084276
<i>Corophium lacustre</i>	0093594	<i>Sarsiella zostericola</i>	0084277
<i>Corophium simile</i>	0093595	<i>Sphaeroma quadridentatum</i>	0092339
<i>Corophium</i> sp.	0093589	<i>Squilla empusa</i>	0099143
<i>Corophium tuberculatum</i>	0093596	<i>Stenothoe minuta</i>	0094936
<i>Corophium volutator</i>	0093601	<i>Unciola irrorata</i>	0093632
<i>Cyathura burbancki</i>	0092150	<i>Unciola serrata</i>	0093633
<i>Cyathura polita</i>	0092149	<i>Unciola</i> sp.	0093629
<i>Cyclaspis varians</i>	0091033	<i>Unionicola</i>	0083073
<i>Cymadusa compta</i>	0093430	<i>Upogebia affinis</i>	0098209
<i>Decapoda</i>	0095599	<i>Xanthidae</i>	0098748
<i>Diastylis polita</i>	0090858		
<i>Dyspanopeus sayi</i>	0098901	Echinoderms	
<i>Edotea triloba</i>	0092627	<i>Holothuroidea</i>	0158140
<i>Elasmopus laevis</i>	0093761	<i>Leptosynapta tenuis</i>	0158432
<i>Erichsonella attenuata</i>	0092618	<i>Microphiopholis atra</i>	BAY0347
<i>Erichsonella filiformis</i>	0092619		
<i>Erichthoneus brasiliensis</i>	0093613	Echiurians	
<i>Eurypanopeus depressus</i>	0098759	<i>Echiura</i>	0154972
<i>Exosphaeroma</i>	0092301	<i>Thalassema hartmani</i>	0155119
<i>Gammarus daiberi</i>	0093779	<i>Thalassema</i> sp.	0155118
<i>Gammarus mucronatus</i>	0093783		
<i>Gammarus palustris</i>	0093782	Ectoprocts	
<i>Gammarus</i> sp.	0093773	<i>Anguinella palmata</i>	0155542
<i>Gammarus tigrinus</i>	0093781	<i>Bowerbankia gracilis</i>	0155559
<i>Gilvossius setimanus</i>	0552843	<i>Conopeum tenuissimum</i>	
<i>Hargeria rapax</i>	0092068	<i>Ectoprocta</i>	0155470
<i>Harpactocoida</i>		<i>Membranipora tenuis</i>	0155827
<i>Hutchinsoniella taylora</i>	0083682	<i>Pedicellina cernua</i>	0156740
<i>Hyalella azteca</i>	0094026		
<i>Idoteidae</i>	0092564	Foraminifera	
<i>Idunella smithii</i>	BAY0133	<i>Miliammina fusca</i>	0044215
<i>Lepidactylus dytiscus</i>	0093998		
<i>Leptocheirus plumulosus</i>	0093486	Hemichordates	
<i>Leucon americanus</i>	0090790	<i>Hemichordata</i>	0158616
<i>Listriella barnardi</i>	0094213	<i>Saccoglossus kowalevskii</i>	0158626
<i>Listriella clymenellae</i>	0094214		

Insects

<i>Ablabesmyia annulata</i>	0128081	<i>Xenochironomus festivus</i>	0129841
<i>Ablabesmyia parajanta</i>	0128112	<i>Xenochironomus</i> sp.	0129837
<i>Bezzia</i> sp.	0012778	<i>Zygoptera</i>	0102042
<i>Ceratopogonidae</i>	0127076	Molluscs	
<i>Chaoborus albatas</i>	0125905	<i>Acteocina canaliculata</i>	0076117
<i>Chaoborus punctipennis</i>	0125923	<i>Aligena elevata</i>	0080685
<i>Chaoborus</i> sp.	0125904	<i>Anachis obesa</i>	0073622
<i>Chironomidae</i>	0127917	<i>Anadara ovalis</i>	0079342
<i>Chironomini</i> sp.	0129229	<i>Anadara transversa</i>	0079340
<i>Chironomus</i> sp.	0129254	<i>Anomia simplex</i>	0079798
<i>Cladopelma</i> sp.		<i>Barnea truncata</i>	0081798
<i>Cladotanytarsus mancus</i>		<i>Bivalvia</i>	0079118
<i>Cladotanytarsus</i> sp.		<i>Boonea bisuturalis</i>	0075987
<i>Clinotanypus pinguis</i>	0127998	<i>Busycon canaliculatum</i>	0074097
<i>Coelotanypus</i> sp.	0128010	<i>Corbicula fluminea</i>	0081387
<i>Coleoptera</i> sp.	0109216	<i>Corbicula manilensis</i>	0081386
<i>Cricotopus</i> sp.	0128575	<i>Crassispira ostrearum</i>	0074901
<i>Cryptochironomus fulvus</i>	0129376	<i>Crassostrea virginica</i>	0079872
<i>Cryptochironomus parafulvus</i>	0129382	<i>Cratena kaoruae</i>	0078714
<i>Cryptochironomus</i> sp.	0129368	<i>Crepidula convexa</i>	0072624
<i>Cryptotendipes</i> sp.	0129394	<i>Crepidula fornicata</i>	0072623
<i>Demicrocryptochironomus</i>	0129421	<i>Cylichna alba</i>	0076148
<i>Dicrotendipes nervosus</i>	0129452	<i>Cyrtopleura costata</i>	0081796
<i>Ephemeroptera</i>	0100502	<i>Doridella obscura</i>	0078439
<i>Epoicocladus</i> sp.	0128682	<i>Ensis directus</i>	0081022
<i>Glyptotendipes</i> sp.	0129483	<i>Epitonium multistriatum</i>	0072247
<i>Gomphidae</i>	0101664	<i>Epitonium rupicola</i>	0072249
<i>Harnischia</i> sp.	0129516	<i>Epitonium</i> sp.	0072233
<i>Hexagenia limbata</i>	0101552	<i>Eupleura caudata</i>	0073300
<i>Hexagenia</i> sp.	0101537	<i>Gastropoda</i>	0069459
<i>Nanocladus</i> sp.	0128844	<i>Gemma gemma</i>	0081511
<i>Oecetis inconspicua</i>	0116613	<i>Geukensia demissa</i>	0079555
<i>Oecetis</i> sp.	0116607	<i>Haminoea solitaria</i>	0076258
<i>Palpomyia</i> sp.	0127859	<i>Hydrobia</i>	0070494
<i>Paralauterborniella</i> sp.	0129616	<i>Littoridinops tenuipes</i>	0070528
<i>Paratendipes</i> sp.	0129623	<i>Littorina littorea</i>	0070419
<i>Polypedilum convictum</i>	0129671	<i>Lucina multilineata</i>	0080389
<i>Polypedilum fallax</i>	0129676	<i>Lyonsia hyalina</i>	0081926
<i>Polypedilum flavum</i>		<i>Macoma baltica</i>	0081052
<i>Polypedilum halterale</i>	0129684	<i>Macoma mitchelli</i>	0081054
<i>Polypedilum illinoense</i>	0129686	<i>Macoma</i> sp.	0081033
<i>Polypedilum scalaenum</i>	0129708	<i>Macoma tenta</i>	0081055
<i>Polypedilum</i> sp.	0129657	<i>Mangelia plicosa</i>	0074568
<i>Polypedilum</i> sp.	0129657	<i>Mercenaria mercenaria</i>	0081496
<i>Probezzia</i> sp.	0127729	<i>Mitrella lunata</i>	0073552
<i>Procladius</i> sp.	0128277	<i>Mulinia lateralis</i>	0080959
<i>Procladius sublettei</i>	0128316	<i>Musculium</i>	0081427
<i>Pseudochironomus fulviventris</i>	0129858	<i>Mya arenaria</i>	0081692
<i>Pseudochironomus</i> sp.	0129851	<i>Nassarius obsoletus</i>	0074111
<i>Sialis</i> sp.	0115002	<i>Nassarius vibex</i>	0074107
<i>Simulium</i> sp.	0126774	<i>Nucula proxima</i>	0079132
<i>Sphaeromias</i>	0127761	<i>Nuculana messanensis</i>	0079212
<i>Stictochironomus devinctus</i>	0129790	<i>Nudibranchia</i>	0078156
<i>Stictochironomus</i> sp.	0129785	<i>Odonata</i>	0101593
<i>Tanypodinae</i>	0127994	<i>Odostomia bisuturalis</i>	0075988
<i>Tanypus neopunctipennis</i>	0128329	<i>Odostomia engonia</i>	0075504
<i>Tanypus</i> sp.	0128324	<i>Odostomia</i> sp.	0075447
<i>Tanytarsini</i> sp.	0129872	<i>Parvilucina multilineata</i>	0080388
<i>Tanytarsus</i> sp.	0129978	<i>Petricola pholadiformis</i>	0081627
<i>Trichoptera</i>	0115095	<i>Pisidium</i> sp.	0081400

<i>Polinices duplicatus</i>	0072918
<i>Polymesoda caroliniana</i>	0081383
<i>Pyramidella candida</i>	0075948
<i>Rangia cuneata</i>	0080962
<i>Rapana venosa</i>	
<i>Rictaxis punctostriatus</i>	0076083
<i>Sayella chesapeakea</i>	0070946
<i>Sphaeriidae</i>	0112737
<i>Sphaerium</i> sp.	0081391
<i>Tagelus divisus</i>	0081274
<i>Tagelus plebeius</i>	0081272
<i>Tellina agilis</i>	0081088
<i>Tellina versicolor</i>	0081100
<i>Tellinidae</i>	0081032
<i>Tenellia</i> sp.	0078547
<i>Turbonilla interrupta</i>	0075687
<i>Turbonilla</i> sp.	0053964
<i>Turridae</i>	0074555
<i>Unionidae</i>	0079913
<i>Urosalpinx cinerea</i>	0073264
<i>Yoldia limatula</i>	0079273

Nematodes

<i>Anticoma litoris</i>	0062032
<i>Axonolaimus spinosus</i>	0059512
<i>Cylindrotheristus oxyuroides</i>	0060433
<i>Desmodora</i> sp.	0060744
<i>Euchromadora</i> sp.	0061205
<i>Mesotheristus setosus</i>	0060526
<i>Metachromadora parasitifera</i>	0060715
<i>Metalinhomeus retrosetosus</i>	
<i>Metalinhomeus typicus</i>	
<i>Nematoda</i>	0059490
<i>Neotonchus punctatus</i>	0061519
<i>Oncholaimus</i> sp.	0062449
<i>Pamponema</i> sp.	
<i>Paracanthonchus</i> sp.	0061480
<i>Paramonhystera proteus</i>	
<i>Parodontophora brevamphida</i>	0059569
<i>Ptycholaimellus ponticus</i>	0061468
<i>Sabatieria pulchra</i>	0061095
<i>Sphaerolaimus balticus</i>	
<i>Steineria</i> sp.	0191219
<i>Thalassoalaimus</i> sp.	0062146

Nemerteans

<i>Carinomidae</i>	0057427
<i>Cerebratulus</i> sp.	0057446
<i>Nemertea</i>	0057411

Ostracods

<i>Ostracoda</i>	0084195
------------------	---------

Phoronids

<i>Phoronida</i>	0155456
<i>Phoronis psammophila</i>	0155467
<i>Phoronis</i> sp.	0155462

Platyhelminthes

<i>Euplana gracilis</i>	0054139
<i>Stylochus ellipticus</i>	0054089

Poriferans

<i>Cliona</i> sp.	0048523
<i>Halichondria bowerbanki</i>	0048398
<i>Haliclona loosanoffi</i>	0047774
<i>Haliclona</i> spp.	0047771
<i>Lissodendoryx carolinesis</i>	0048072
<i>Microciona prolifera</i>	0047997

Pycnogonids

<i>Anoplodactylus lentus</i>	0083644
<i>Pycnogonida</i>	0083545

Sipunculids

<i>Sipuncula</i>	0154520
------------------	---------