

# Intertidal meiofauna of the St Lawrence estuary (Quebec, Canada): diversity, biomass and feeding structure of nematode assemblages

G. Tita\*<sup>‡</sup>, G. Desrosiers\*, M. Vincx<sup>†</sup> and M. Clément<sup>‡</sup>

\*Université du Québec à Rimouski, Institut des Sciences de la Mer (ISMER), 310 allée des Ursulines, Rimouski (Québec), Canada, G5L 3A1. <sup>†</sup>University of Gent, Department of Biology, Marine Biology Section, Ledeganckstraat 35, Gent, B-9000, Belgium. <sup>‡</sup>Biodôme de Montréal, 4777 avenue Pierre de Coubertin, Montréal (Québec), Canada, H1V 1B3. <sup>‡</sup>Corresponding author (present address): Station Technologique Maricole des Îles-de-la-Madeleine, Ministère de l'Agriculture, des Pêcheries et de l'Alimentation du Québec, CP 658, Cap-aux-Meules (Québec), Canada, G0B 1B0. E-mail: gtita@duclos.net

The meiofauna of the St Lawrence estuary was investigated in the intertidal zone of the Parc du Bic (Quebec, Canada). Five nematode assemblages were distinguished by a cluster analysis: A1 and A2 (upper-tide level); A3 (mid-tide level); A4 and A5 (low-tide level). Discriminant function analysis showed that exposure time during low tide was the most important environmental factor in determining differences between assemblages. Chlorophyll-*a*, phaeopigments, sediment water content, and per cent of silt followed in the same order. Nematode densities (400–1500 ind 10 cm<sup>-2</sup>) were found to be lower than those generally reported for other estuarine intertidal zones of the eastern Atlantic coast. Mean nematode biomass in the five assemblages ranged between 96 ± 14 and 248 ± 86 µg C<sub>org</sub> 10 cm<sup>-2</sup>. Deposit feeders were generally the dominant nematode feeding group in terms of abundance and biomass. Correlation of epigrowth-feeders with chlorophyll-*a* and phaeopigments, respectively, suggested that in the upper-tide level, old or partially degraded phytodetritus contribute more to the diet of this nematode feeding group; and in the low-tide level epigrowth-feeders may rely more on 'fresher' phytodetritus.

## INTRODUCTION

Estuarine meiofaunal assemblages of the eastern coast of the North Atlantic have been intensively investigated (Warwick & Price, 1979; Warwick & Gee, 1984; Heip et al., 1985; Austen, 1989; Austen & Warwick, 1989; Li & Vincx, 1993; Hall & Frid, 1997). As for the western coast, investigations were limited to US estuaries (Tietjen, 1969; Coull, 1973, 1985), and no study has been carried out in the more northerly Canadian estuaries where climatic conditions are more severe. The present study has examined for the first time the intertidal meiofauna of the St Lawrence estuary (Quebec, Canada). The great length (~360 km) of this estuary is matched by widths as large as 60 km and depths of over 350 m. Two main regions are generally distinguished: the Upper estuary, with a turbidity maximum and strong salinity gradients, and the Lower estuary, which is more homogeneous and has a more oceanic character (El-Sabh & Silverberg, 1990). Another typical aspect of this sub-arctic estuary is a long winter with an ice cover lasting 4–5 months along the shore.

The study area was the Anse-à-l'Original of the Parc du Bic in the Lower estuary where the tidal regime has a semi-diurnal pattern with a mean tidal range of 3 m (El-Sabh & Murty, 1990). The surface water temperature of the Lower estuary ranges between 2° and 13°C according to the time of year (Ingram & El-Sabh, 1990). This area

offered three main advantages: an easy accessibility, a well known macrofaunal community (Vincent et al., 1987; Miron & Desrosiers, 1990; Olivier et al., 1993; Caron et al., 1993a,b; Caron, 1995; Caron et al., 1995a,b), and the presence of different sediment-bottom types. The latter was an interesting attribute for studying the different intertidal meiofaunal assemblages of this part of the estuary. Sediment type is a key factor that determines the structure of meiofaunal assemblages (Hicks & Coull, 1983; Heip et al., 1985). The macrofauna is characterized by the Atlantic boreal community of *Macoma balthica* (L.) (Desrosiers et al., 1980; Desrosiers & Brêthes, 1984; Desrosiers et al., 1984). Among the dominant macrofaunal species of this community, there are three molluscs (*Macoma balthica*, *Mya arenaria* and *Hydrobia minuta*) and two polychaetes (*Nereis virens* and *Nephtys caeca*). The polychaete *Nereis virens* has been shown to affect meiofauna either by predation or by sediment disturbance (Olivier et al., 1993; Tita et al., 2000).

The present study had two objectives: (i) describing the composition and density of the metazoan meiofauna in the selected study area; and (ii) describing the spatial distribution and the feeding structure of the nematode assemblages according to the sediment type. This study represents a complement to Tita et al. (1999) that focused attention on the size spectra and the morphological adaptations of the nematode species found in the here described assemblages.

## MATERIALS AND METHODS

### *The study area*

The study area was the intertidal zone of the Anse-à-l'Original of the Parc du Bic located in the southern shore of the St Lawrence estuary (Figure 1). Forty-eight stations, distributed over the whole area, were sampled between 4 and 7 June 1996. The stations were the same as those that were sampled by Miron & Desrosiers (1990) for a study on *Nereis virens* and *Nephtys caeca* populations. The shore profile of the Anse-à-l'Original has a complex configuration with several inlets and islets to which a correspondingly complex distribution of sediment facies is associated. As a general feature, heterogeneous muddy flats occur at the upper-tide level, and relatively homogeneous sandflats at the lower-tide level. The latter is a sandy shore, while the mid-tide level is a gravelly-sandy shore exposed to wave action in its eastern part and relatively sheltered in its western part. As for the upper-tide level, the sediment is more muddy especially in its eastern part where there is a *Spartina* salt-marsh. The whole area's surface is irregular due to the presence of large boulders and rocky substrates, especially in the eastern part, where they cover 20–40% of the bottom.

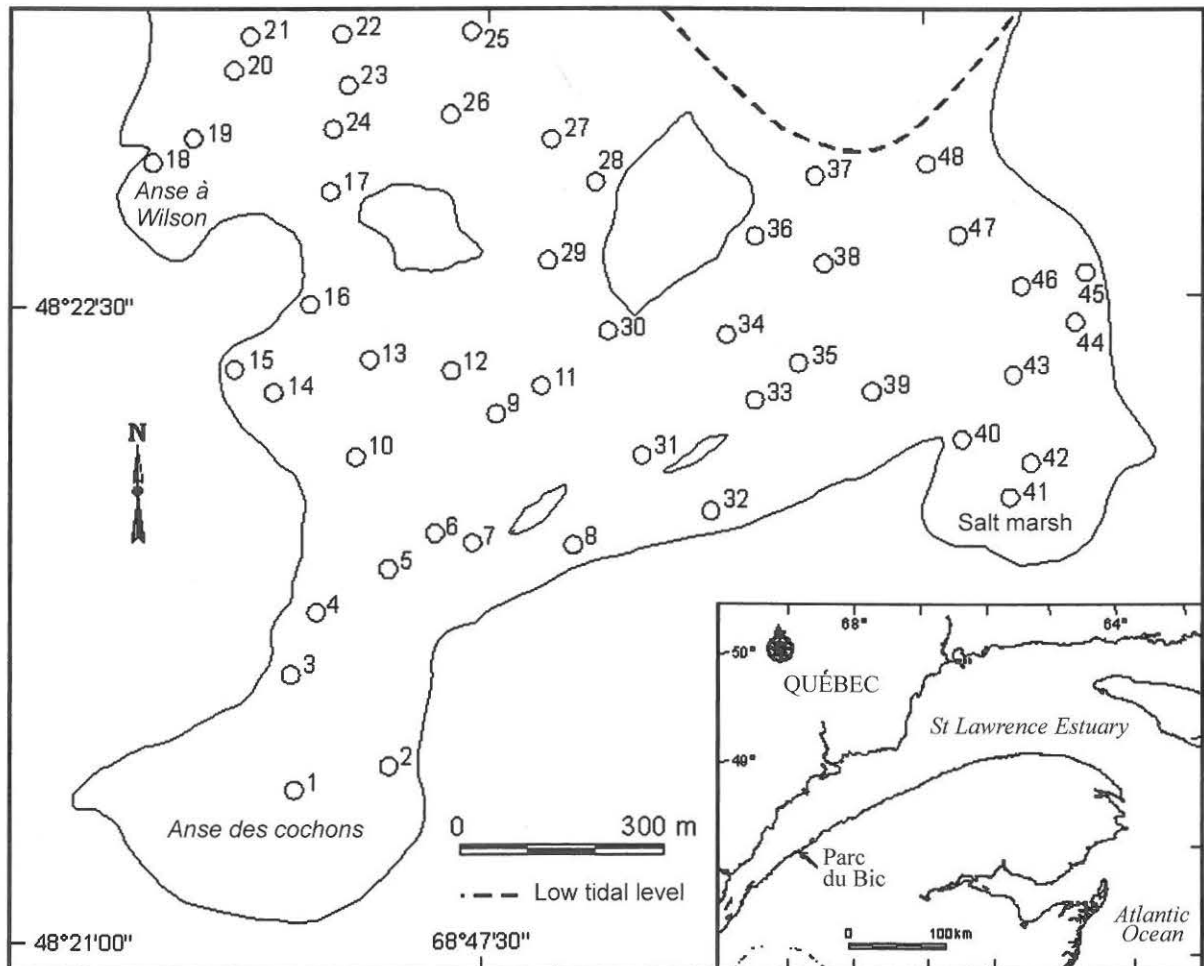
### *The environmental factors*

At each station, a sediment sample was collected from the top 2 cm of sediment using a punch with an internal

diameter of 26 mm. The sediment water content (% of total weight) was estimated by weighing the sediment before and after drying it at 50°C until a constant weight was obtained. The same sediment samples were re-dried at 90°C for total organic matter content analysis by combustion at 500°C for six hours (Luczak et al., 1997). Another sediment sample was collected for granulometric analysis. The Folk's (1974) triangular diagram was used to characterize the sediment present at all stations. Small sediment samples for the analysis of photopigment concentrations (chlorophyll-*a* and phaeopigments) were collected from the top 1 cm using short tubes as corers (internal diameter=1.2 cm). These samples were kept cool in an ice-box until arrival at the laboratory where they were preserved in a freezer at -80°C until analysis by fluorometry (Parsons et al., 1984). Exposure time ( $T_{exp}$ ) was estimated by directly timing the exposure period at neap and spring tides, and was expressed as % of tidal cycle period. Reported  $T_{exp}$  values represent the mean value from spring and neap tides. The seawater salinity and temperature were sampled with a digital portable meter (YSI, model 3050) at the low water line.

### *The meiofauna*

Meiofauna was sampled using a hand-held corer with an internal section of 7.3 cm<sup>2</sup> (Tita et al., 2000a). A preliminary study (unpublished) on the meiofauna of the sediment's top 10 cm of the same area showed that more

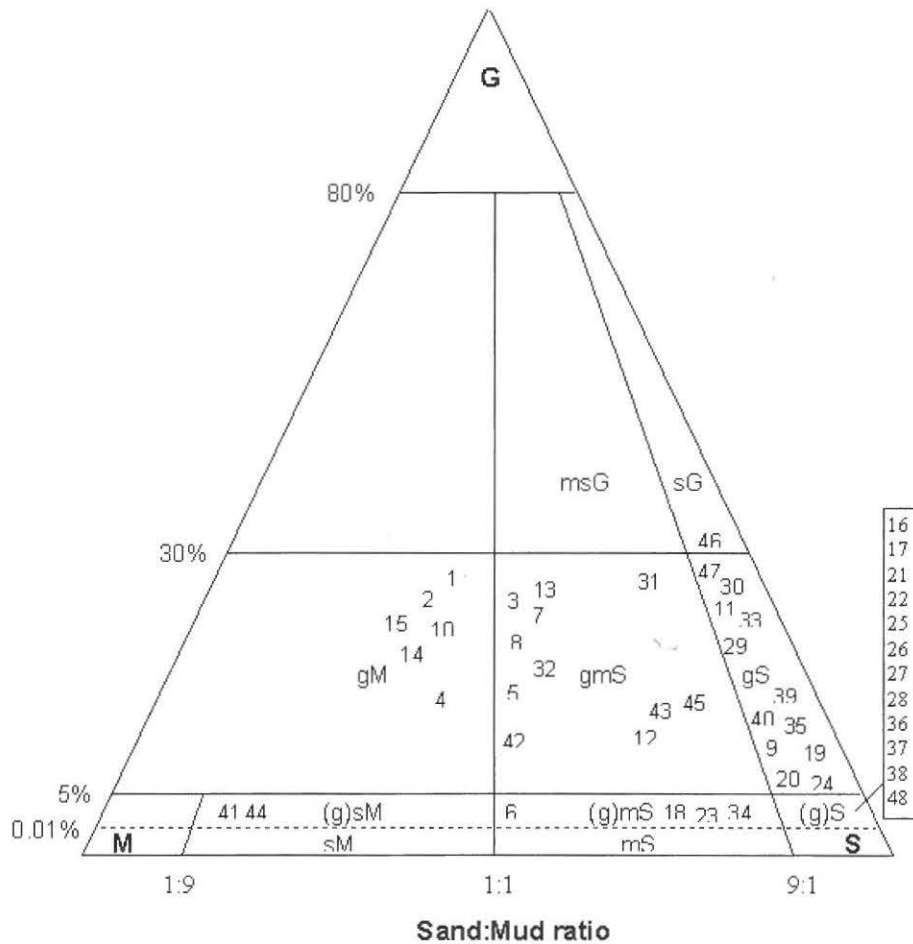


**Figure 1.** The study area (Anse-à-l'Original) and the 48 sampled stations.

**Table 1.** Minimum, maximum, average, standard deviation, and coefficient of variation for the different environmental factors in the whole study area (48 stations).

	%				$\mu\text{g } 10 \text{ cm}^{-2}$		$\mu\text{m}$	%			
	Texp	Wat	OM	P/C	Chl- <i>a</i>	Phae	Md	Mean	Fine	V. fine	Silt
Minimum	8.8	17.0	0.9	0.13	53.3	9.4	44	0.9	3.4	1.8	0.9
Maximum	63.9	41.9	7.0	1.06	115.0	57.9	390	47.2	84.8	73.5	79.5
Average	33.9	24.2	1.8	0.38	79.4	27.3	140	9.3	27.9	29.6	19.9
SD	17.5	4.5	1.2	0.25	16.3	13.7	81	10.1	22.1	19.6	21.1
CV	51.7	18.6	65.5	66.9	20.5	50.2	57.7	108.6	79.4	66.2	106.1

Texp, exposure time; Wat, water content of the sediment; OM, organic matter; P/C, Phae/Chl-*a* ratio; Chl-*a*, chlorophyll-*a*; Phae, phaeopigments; Md, granulometric median of total sediment; Mean, Fine and V. fine, mean (250–500  $\mu\text{m}$ ), fine (125–250  $\mu\text{m}$ ) and very fine (63–125  $\mu\text{m}$ ) sand; Silt, grain size <63  $\mu\text{m}$ .



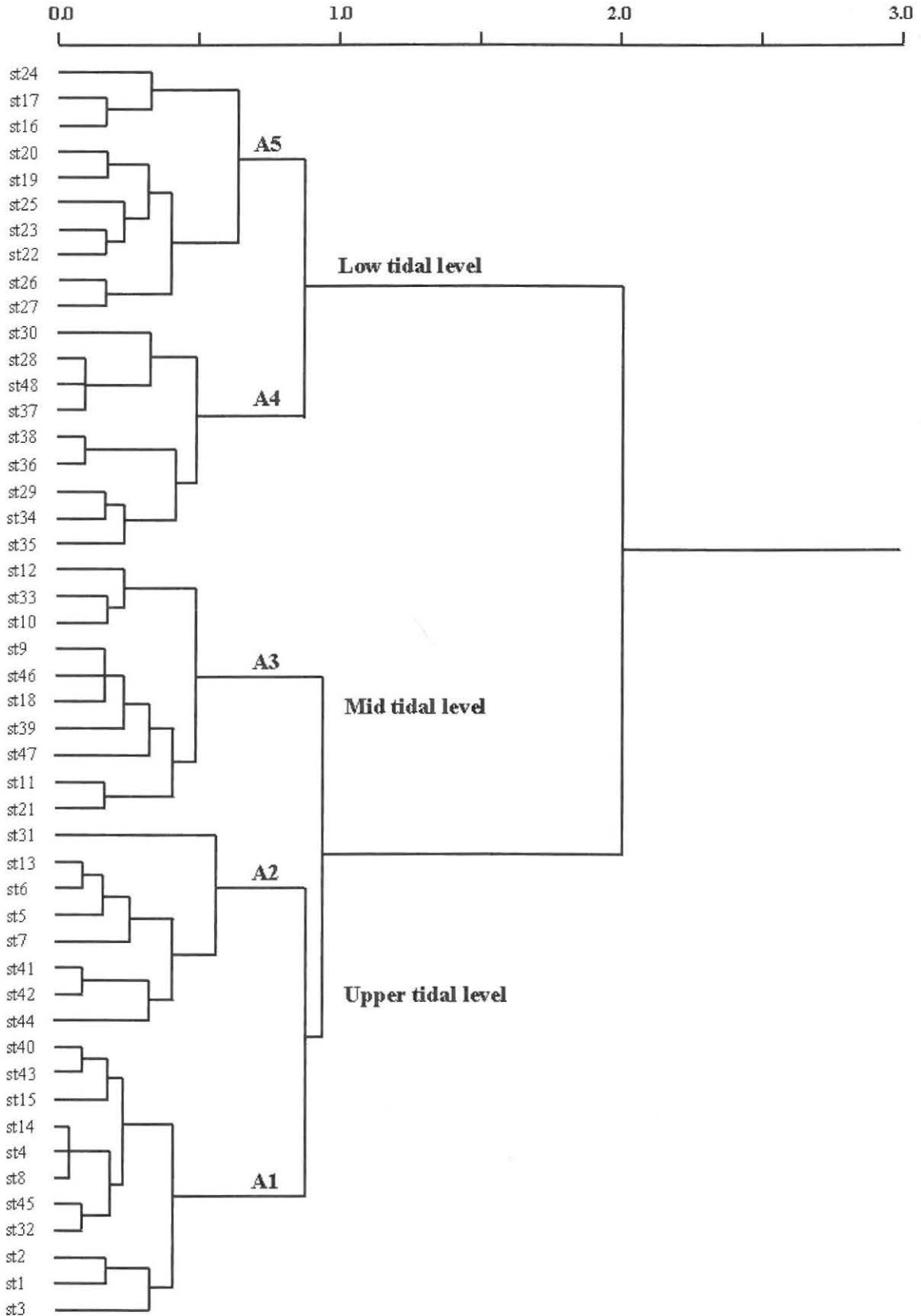
**Figure 2.** Triangular diagram (Folk, 1974) placing stations according to their sediment granulometric properties. G, gravel (>2 mm); sG, sandy gravel; msG, muddy sandy gravel; S, sand (0.063–2 mm); gS, gravelly sand; (g)S, slightly gravelly sand; gmS, gravelly muddy sand; (g)mS, slightly gravelly muddy sand; mS, muddy sand; M, mud (<0.063 mm). Per cents indicate % of gravel.

**Table 2.** Abundance of meiofaunal taxa (individuals  $10 \text{ cm}^{-2}$ ) in the five nematode assemblages.

	Nematodes	Copepods	Nauplii	Turbellarians	Ostracods	Rotifers	Others	Total
A1	557 $\pm$ 127	78 $\pm$ 29	171 $\pm$ 55	53 $\pm$ 16	47 $\pm$ 33	13 $\pm$ 16	10 $\pm$ 1	929 $\pm$ 217
A2	1313 $\pm$ 266	162 $\pm$ 51	398 $\pm$ 129	40 $\pm$ 9	26 $\pm$ 8	9 $\pm$ 4	12 $\pm$ 4	1960 $\pm$ 404
A3	507 $\pm$ 92	40 $\pm$ 12	62 $\pm$ 19	51 $\pm$ 10	11 $\pm$ 4	7 $\pm$ 4	3 $\pm$ 1	680 $\pm$ 101
A4	1044 $\pm$ 197	24 $\pm$ 11	85 $\pm$ 16	150 $\pm$ 18	46 $\pm$ 15		3 $\pm$ 1	1353 $\pm$ 213
A5	622 $\pm$ 104	99 $\pm$ 20	458 $\pm$ 105	46 $\pm$ 11			1 $\pm$ 1	1228 $\pm$ 117

than 95% of the nematodes in muddy sediments and 50% in sandy sediments inhabited the top 2 cm. Other meiofaunal taxa were always concentrated in the top 2 cm (>85%). For this reason and for analysis time constraints,

only the top 2 cm of the sediments were sampled and considered sufficient to describe the horizontal distribution of the nematode assemblages. Samples were fixed and preserved in 4% formalin, and stained with rose bengal.



**Figure 3.** Hierarchical analysis (1-gamma coefficient, complete linkage) based on the nematode specific composition found at the 48 stations. Clusters A1, A2, A3, A4 and A5 represent different assemblages (see text).

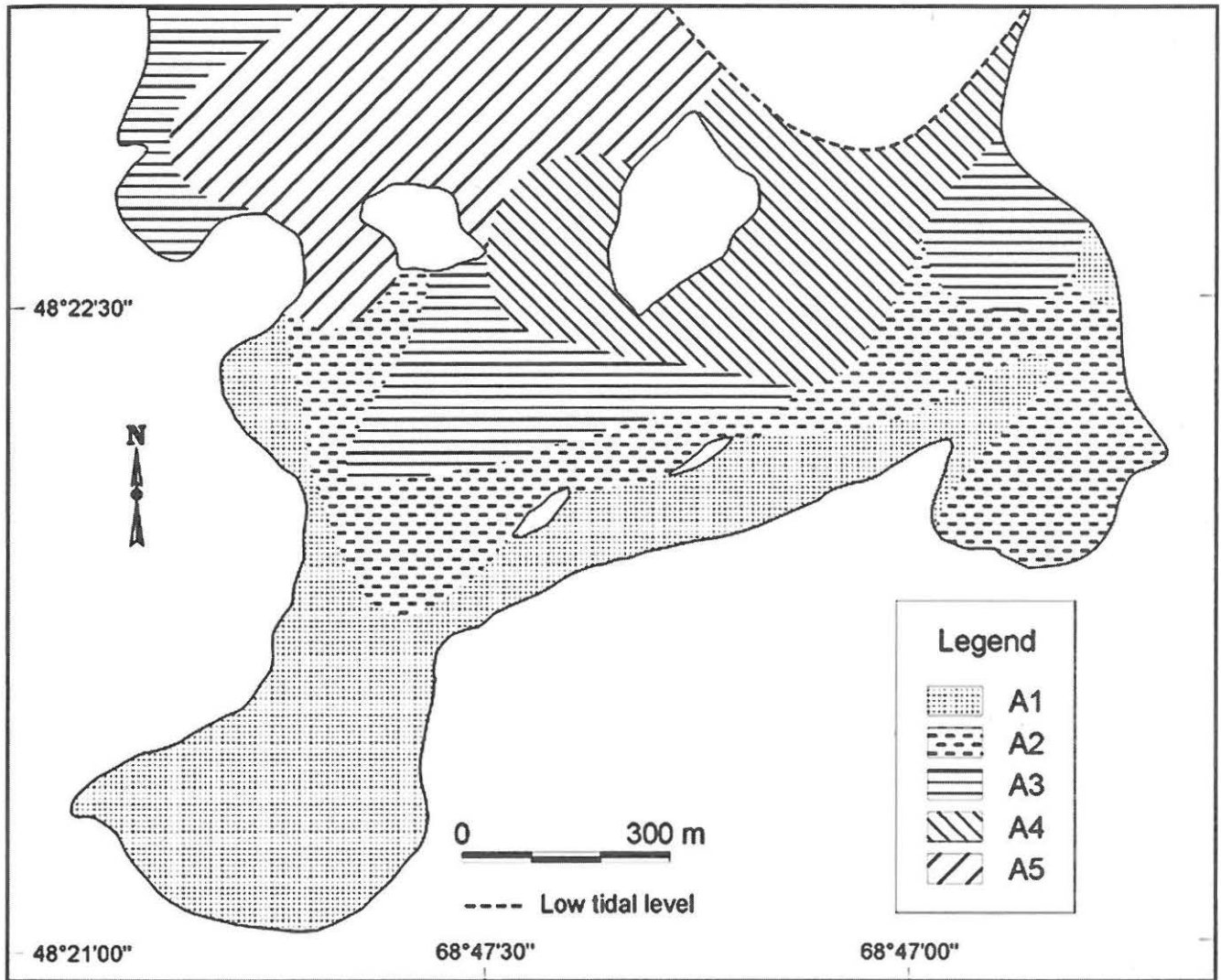


Figure 4. Spatial distribution of the five nematode assemblages (A1, A2, A3, A4, A5).

They were then washed through 1-mm and 63- $\mu\text{m}$  sieves. The sediment retained by the 63- $\mu\text{m}$  sieve was used to extract the meiofauna by centrifugation using Ludox-TM (Heip et al., 1985). Organisms were identified and counted at higher taxonomic levels, except nematodes and copepods that were identified at the species level. One hundred and ten nematodes per sample were randomly collected and mounted on glycerine slides for identification (McIntyre & Warwick, 1984). Thirty additional individuals were collected from samples of stations with the highest densities (Stations 28, 32, 36, 38, 41, 42, 44). Genus and species identifications were based on Platt & Warwick (1983, 1988) and Hopper (1969), while family systematic was based on Lorenzen (1994). Two multivariate techniques were used to describe nematode assemblages: (i) a hierarchical analysis using the 1-gamma coefficient (Goodman & Kruskal, 1954) with a complete linkage method for building the cluster graph; and (ii) a discriminant function analysis (DFA). The hierarchical analysis was used to estimate similarity between nematode species compositions in the 48 stations, and to distinguish nematode assemblages. The DFA (forward stepwise model) was used to identify the most responsible environmental factors determining the spatial distribution of nematode assemblages.

In order to investigate the influence of sediment properties in structuring nematode assemblages, the granulometric characteristics were plotted in the DFA using two variables: silt (i.e. % of the sediment fraction with grain size  $< 63 \mu\text{m}$ ), and Md63 (i.e. median of the sediment fraction with grain size  $> 63 \mu\text{m}$ ). A multivariate analysis of variance (MANOVA) was performed in order to verify significant differences of each environmental factor between nematode assemblages. Twenty individuals per nematode species were randomly sorted and used to estimate the mean individual specific biomass (m.i.b.). For the rarest species (less than 20 individuals found), the total number of recorded individuals was used for the m.i.b. estimation. Biomass was estimated with the biovolumes method (Warwick & Price, 1979). Nematode wet weight ( $\mu\text{g WW}$ ) was obtained using a specific gravity of 1.13 (Wieser, 1960). Mean total organic carbon ( $C_{\text{org}}$ ) of nematodes was estimated assuming that  $C_{\text{org}}$  was equal to 12.4% of the wet weight (Jensen, 1984). The feeding structure of nematode assemblages was described using the six feeding groups proposed by Moens & Vincx (1997): microvores (M), ciliate-feeders (CF), deposit-feeders (DF), epigrowth-feeders (EF), facultative-predators (FP), and predators (P). This classification partially derives from the Wieser



**Table 3.** Nematode species with a relative abundance greater than 1% at the five assemblages; *n*, mean abundance (ind 10 cm<sup>-2</sup>) (mean ± SE); *biom*, mean biomass (μg C<sub>org</sub> 10 cm<sup>-2</sup>) (mean ± SE).

A1		A2		A3		A4		A5	
N=563 ±129		N=1313 ±266		N=507 ±92		N=1044 ±197		N=622 ±104	
biom=177 ±41		biom=248 ±86		biom=85 ±17		biom=155 ±21		biom=96 ±14	
<i>Anoplostoma blanchardi</i>	19.1	<i>Sabatieria punctata</i>	18.8	<i>Eleutherolaimus</i> sp.	11.1	<i>Metachromadora (M.)</i> sp.	47.7	<i>Eleutherolaimus</i> sp.	26.4
<i>Theristus (D.) procerus</i>	18.8	<i>Microlaimus</i> sp. 1	12.2	<i>Microlaimus</i> sp. 1	9.0	<i>Viscosia</i> sp.	9.1	<i>Daptonema</i> sp. 2	8.6
<i>Metachromadora (M.) remanei</i>	11.7	Chromadoridae A	10.9	<i>Anoplostoma blanchardi</i>	8.5	<i>Dichromadora hyalocheile</i>	6.6	<i>Odontophora</i> sp.	7.7
<i>Daptonema tenuispiculum</i>	9.9	<i>Anoplostoma blanchardi</i>	7.5	<i>Paramonohystera</i> sp. 1	8.3	<i>Ascolaimus</i> sp.	5.1	<i>Daptonema tenuispiculum</i>	7.2
Chromadoridae A	8.6	<i>Metachromadora (M.) remanei</i>	6.2	<i>Sabatieria punctata</i>	7.3	<i>Odontophora</i> sp.	3.2	<i>Viscosia</i> sp.	6.0
<i>Sabatieria punctata</i>	6.2	<i>Desmolaimus</i> sp.	5.8	<i>Odontophora</i> sp.	5.6	<i>Pomponema sedecima</i>	2.9	<i>Paracanthochus</i> sp.	4.4
<i>Paracanthochus caecus</i>	4.6	<i>Daptonema tenuispiculum</i>	5.1	<i>Daptonema tenuispiculum</i>	5.5	<i>Chromadorita</i> sp.	2.7	Chromadoridae B	4.4
<i>Hypodontolaimus balticus</i>	2.8	<i>Paramonohystera</i> sp. 1	4.5	<i>Viscosia</i> sp.	5.0	<i>Dichromadora</i> sp.	2.5	Tripyloididae	3.3
<i>Monhystera</i> sp. 1	2.1	<i>Leptolaimus papilliger</i>	3.6	<i>Theristus (D.) procerus</i>	4.4	Chromadoridae B	2.2	<i>Microlaimus</i> sp. 1	2.7
<i>Enoplus</i> sp.	1.8	<i>Theristus (D.) procerus</i>	3.4	Chromadoridae A	3.0	<i>Daptonema tenuispiculum</i>	1.8	<i>Metachromadora (M.)</i> sp.	2.3
<i>Leptolaimus papilliger</i>	1.7	<i>Ptycholaimellus ponticus</i>	3.3	<i>Ptycholaimellus ponticus</i>	2.9	<i>Paracanthochus caecus</i>	1.5	<i>Richtersia inaequalis</i>	2.1
<i>Chromadorita</i> sp.	1.6	<i>Tripyloides</i> sp. 1	1.9	<i>Paracanthochus caecus</i>	2.4	<i>Microlaimus</i> sp. 1	1.4	<i>Pomponema sedecima</i>	1.9
<i>Sphaerolaimus</i> sp. 2	1.1	<i>Chromadorita</i> sp.	1.5	<i>Desmolaimus</i> sp.	2.3	<i>Axonolaimus</i> sp. 1	1.2	<i>Dichromadora</i> sp.	1.6
<i>Paramonohystera</i> sp. 1	1.0	<i>Halalaimus</i> sp. 1	1.4	<i>Metachromadora (M.)</i> sp.	2.0	<i>Gammanema</i> sp.	1.2	<i>Paracanthochus caecus</i>	1.4
		<i>Neochromadora poecilosoma</i>	1.1	<i>Chromadorita</i> sp.	1.8	<i>Microlaimus</i> sp. 2	1.0	<i>Southernia</i> sp.	1.3
		<i>Monhystera</i> sp. 1	1.0	<i>Daptonema</i> sp. 2	1.7			<i>Microlaimus</i> sp. 2	1.1
				<i>Sabatieria longispinosa</i>	1.5			<i>Chromadorita</i> sp.	1.1
				<i>Microlaimus</i> sp. 2	1.2			<i>Sabatieria ornata</i>	1.1
				<i>Ascolaimus</i> sp.	1.1			<i>Enoplolaimus</i> sp. 1	1.1
				<i>Dichromadora hyalocheile</i>	1.0			<i>Sabatieria longispinosa</i>	1.0
				<i>Pomponema sedecima</i>	1.0				
				<i>Hypodontolaimus balticus</i>	1.0				
Others	9.1	Others	11.5	Others	12.4	Others	10.0	Others	13.3

**Table 4.** List of functions included or excluded in the forward stepwise model of the discriminant function analysis (global Wilk's lambda=0.088;  $F=5.60$ ;  $P<0.000$ ). Only functions with  $F \geq 1.0$  were included in the model.

	Partial lambda	F	P
<i>In model</i>			
Texp	0.508	9.21	0.000
Chl- <i>a</i>	0.727	3.57	0.014
Phae	0.796	2.43	0.064
Wat	0.830	1.95	0.122
P/C	0.834	1.89	0.132
Silt	0.902	1.03	0.402
<i>Not in model</i>			
OM	0.936	0.628	0.646
Md63	0.908	0.940	0.452

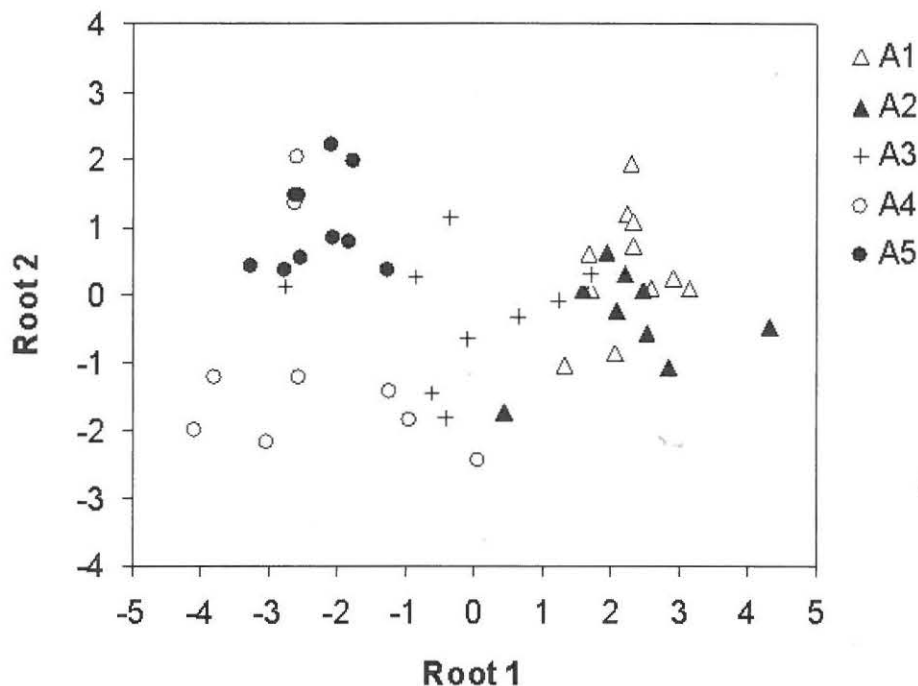
(1953) where feeding groups 1A, 1B, 2A, and 2B correspond to M, CF+DF, EF, FP+P, respectively. Moens & Vincx's feeding groups of the species recorded in our study area were established according to their buccal morphology and similarity to the 'types' described by these authors. A complete list of the species with their corresponding attributed feeding group is reported in Appendix 1.

As for copepods, all individuals of each sample were identified to species level. However, because of their absence and very low abundance in several stations, statistical analysis for describing community spatial distribution was inappropriate. Therefore, only species and corresponding relative abundance are reported.

## RESULTS

### *Some environmental factors*

During the sampling period, the seawater salinity was  $23.4 \pm 0.2$  psu and the temperature  $10.2 \pm 0.5^\circ\text{C}$ . As a general feature, the other environmental factors showed a great heterogeneity within the study area (Table 1). The sediment



**Figure 5.** Discriminant function analysis plot applied on the five groups of stations identified with the hierarchical analysis (Wilk's lambda=0.088;  $F=5.60$ ;  $P<0.0000$ ). List of functions are reported in Table 4.

**Table 5.** Results from the discriminant function analysis: squared Mahalanobis distances ( $M$ ) between assemblages with  $F$  values, and probabilities ( $P$ ).

	A1			A2			A3			A4		
	M	F	P	M	F	P	M	F	P	M	F	P
A2	2.25	1.47	0.213									
A3	7.63	5.00	***	7.92	4.67	**						
A4	25.36	16.60	****	24.82	14.62	****	6.38	3.76	**			
A5	23.88	16.66	****	25.49	15.90	****	7.46	4.65	**	4.85	3.03	*

\*,  $P<0.05$ ; \*\*,  $P<0.01$ ; \*\*\*,  $P<0.001$ ; \*\*\*\*,  $P<0.0001$ .

**Table 6.** Results from the MANOVA applied for investigating functions' differences between the five assemblages (*Wilk's lambda*=0.077;  $P < 0.0000$ ).

	<i>F</i>	<i>P</i>
Texp	23.21	****
Phae	12.32	****
Silt	12.16	****
P/C	9.36	****
Chl- <i>a</i>	7.21	***
Wat	4.21	**
OM	3.14	*
Md63	2.43	0.062

\*,  $P < 0.05$ ; \*\*,  $P < 0.01$ ; \*\*\*,  $P < 0.001$ ; \*\*\*\*,  $P < 0.0001$ .

granulometric composition was very diverse in the different zones of the Anse-à-l'Original (Figure 2). In the low-tide level, the sediment was gravelly to slightly gravelly sand, while, in the upper-tide level, the sediment was gravelly muddy sand to gravelly mud. The sediment granulometric median (Md) was generally greater in the east side of the Anse-à-l'Original.

#### *Nematode assemblages and other meiofauna*

Nematodes were the dominant taxon in all stations accounting for 50–80% of total meiofauna (Table 2). Twenty-six families, 69 genera and 106 species of nematodes were found in the whole area (see Appendix). The hierarchical analysis clearly separated the three tide levels, i.e. lower (Ltl), mid (Mtl), and upper (Utl) (Figure 3). Utl stations clustered into two main groups (A1 and A2), as

**Table 7.** Probabilities (*P*) resulted from post hoc test (*Newman-Keuls method*) for the different functions. Average values of functions in the five assemblages are reported in italics (mean  $\pm$  SD).

	A1	A2	A3	A4	A5
<i>Texp</i> (%)	<i>49 <math>\pm</math> 7.9</i>	<i>50 <math>\pm</math> 9.2</i>	<i>32 <math>\pm</math> 11.9</i>	<i>22 <math>\pm</math> 15.0</i>	<i>15 <math>\pm</math> 5.8</i>
A2	0.836				
A3	***	**			
A4	****	****	*		
A5	****	****	**	0.134	
<i>Phae</i> ( $\mu\text{g } 10 \text{ cm}^{-2}$ )	<i>37 <math>\pm</math> 12.7</i>	<i>38 <math>\pm</math> 10.0</i>	<i>28 <math>\pm</math> 12.8</i>	<i>18 <math>\pm</math> 6.0</i>	<i>14 <math>\pm</math> 3.3</i>
A2	0.828				
A3	*	0.051			
A4	****	****	*		
A5	****	****	*	0.350	
<i>Silt</i> (%)	<i>35 <math>\pm</math> 17.9</i>	<i>40 <math>\pm</math> 25.2</i>	<i>13 <math>\pm</math> 13.2</i>	<i>3 <math>\pm</math> 2.5</i>	<i>7 <math>\pm</math> 3.3</i>
A2	0.495				
A3	**	***			
A4	****	****	0.311		
A5	***	****	0.398	0.536	
<i>P/C</i>	<i>0.58 <math>\pm</math> 0.28</i>	<i>0.55 <math>\pm</math> 0.20</i>	<i>0.36 <math>\pm</math> 0.24</i>	<i>0.18 <math>\pm</math> 0.05</i>	<i>0.19 <math>\pm</math> 0.07</i>
A2	0.806				
A3	0.052	*			
A4	***	***	0.122		
A5	***	***	0.059	0.945	
<i>Chl-a</i> ( $\mu\text{g } 10 \text{ cm}^{-2}$ )	<i>69 <math>\pm</math> 11.7</i>	<i>72.5 <math>\pm</math> 11.9</i>	<i>82 <math>\pm</math> 13.2</i>	<i>99 <math>\pm</math> 17.3</i>	<i>77 <math>\pm</math> 11.5</i>
A2	0.566				
A3	0.144	0.254			
A4	****	***	**		
A5	0.379	0.448	0.406	**	
<i>Wat</i> (%)	<i>25 <math>\pm</math> 3.6</i>	<i>28 <math>\pm</math> 7.9</i>	<i>23 <math>\pm</math> 2.2</i>	<i>23 <math>\pm</math> 1.2</i>	<i>22 <math>\pm</math> 0.9</i>
A2	*				
A3	0.420	**			
A4	0.250	**	0.918		
A5	0.498	**	0.886	0.967	
<i>OM</i> (%)	<i>2.4 <math>\pm</math> 1.62</i>	<i>2.2 <math>\pm</math> 1.06</i>	<i>1.8 <math>\pm</math> 1.26</i>	<i>1.1 <math>\pm</math> 0.09</i>	<i>1.1 <math>\pm</math> 0.19</i>
A2	0.577				
A3	0.444	0.510			
A4	*	0.099	0.154		
A5	0.069	0.155	0.310	0.976	
<i>Md63</i> ( $\mu\text{m}$ )	<i>240 <math>\pm</math> 149</i>	<i>153 <math>\pm</math> 90</i>	<i>189 <math>\pm</math> 107</i>	<i>158 <math>\pm</math> 23</i>	<i>119 <math>\pm</math> 18</i>
A2	0.201				
A3	0.243	0.691			
A4	0.152	0.913	0.480		
A5	0.056	0.432	0.380	0.641	

\*,  $P < 0.05$ ; \*\*,  $P < 0.01$ ; \*\*\*,  $P < 0.001$ ; \*\*\*\*,  $P < 0.0001$ .



**Table 8.** Significant correlations ( $r$ ) between the abundance of the dominant nematode species of the upper and lower tide levels and the environmental factors.

	OM	Chl- <i>a</i>	Phae	P/C	Texp	Wat	Silt
Upper-tide level							
<i>Anoplostoma blanchardi</i>	0.54 ****		0.53 ***	0.47 **	0.74 ****	0.53 ***	0.62 ****
<i>Theristus (D.) procerus</i>	0.40 **	-0.35 *	0.45 **	0.43 **	0.61 ****		0.56 ****
<i>Metachromadora (M.) remanei</i>		-0.30 *	0.43 **	0.43 **	0.45 **	0.60 ****	0.55 ****
<i>Sabatieria punctata</i>	0.29 *				0.40 ***	0.48 ***	0.41 **
<i>Microaimus</i> sp. 1							
Chromadoridae A	0.37 **		0.48 ***	0.45 **	0.59 ****	0.47 ***	0.59 ****
Lower-tide level							
<i>Metachromadora (M.)</i> sp.	-0.42 **	0.62 ****	-0.36 *	-0.40 **		-0.42 **	
<i>Viscosia</i> sp.	-0.33 *	0.56 ****	-0.60 ****	-0.61 ****	-0.64 ****	-0.34 *	-0.61 ****
<i>Dichromadora hyalocheile</i>		0.30 *	-0.44 **	-0.41 **	-0.53 ***		-0.39 **
<i>Eleutherolaimus</i> sp.	-0.36 **	-0.50 ***	-0.30 *		-0.32 *	-0.42 **	
<i>Daptonema</i> sp. 2	-0.32 *		-0.47 ***	-0.40 **	-0.45 **		-0.40 **
<i>Odontophora</i> sp.	-0.37 *	0.38 **	-0.70 ****	-0.66 ****	-0.61 ****	-0.42 **	-0.64 ****

\*,  $P < 0.05$ ; \*\*,  $P < 0.01$ ; \*\*\*,  $P < 0.001$ ; \*\*\*\*,  $P < 0.0001$ .

well as Ltl stations (A4 and A5). Mtl stations clustered in a single main group (A3). The five clusters A1, A2, A3, A4, and A5 were interpreted as different assemblages, each with its spatial distribution (Figure 4). Nematode assemblages were different in species composition and/or in structure (Table 3). Copepods and turbellarians had similar relative abundance (2–10%), representing alternatively the second and third dominant groups. As for copepods, 21 species were found with *Microarthridion laurenticum* dominating in the upper-tide levels (60–70%), followed by *Platychelipus littoralis* (10–25%), *Stenhelina (D.) palustris* (2–8%), *Halectinosoma curticornis* (1–3%), *Tachidius brevicornis* (0.5–2%), and *Nannopus palustris* (0–2%). In the lower-tide levels, *Rhizothrix minuta* was generally the dominant species (30–70%) followed by *Microarthridion laurenticum* (10–45%), *Thompsonula hyaenae* (0–15%), *Stenhelina (S.) divergens* (5–10%), and *Halectinosoma elongatum* (0–5%). Ostracods represented the fourth group but were not found at assemblage A5. Rotifers were exclusively found in Utl and Mtl assemblages.

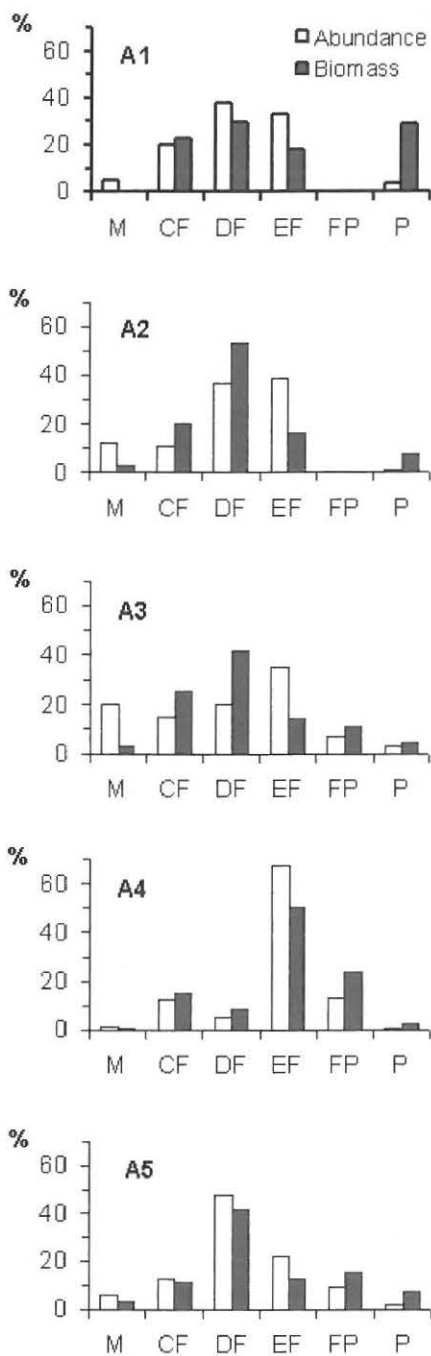
#### Environmental factors vs nematode assemblages

The DFA was performed in order to evaluate the influence of the different environmental factors in determining spatial segregation of nematode assemblages. The forward stepwise procedure was used to determine what factors were relevant ( $F \geq 1.0$ ) and therefore to be included in the model (Table 4). As the DFA plot shows, root 1 separated stations of the Ltl from stations of the Utl,

while stations of the Mtl lay in between (Figure 5). Root 2 separated assemblages of same tide levels, especially A4 and A5. Distances between assemblages resulted from the analysis are reported in Table 5. The MANOVA showed significant differences between assemblages for all environmental factors excepted for the granulometric median of the sediment fraction greater than 63  $\mu\text{m}$  (Md63) (Table 6). The Newman–Keuls multiple pairwise comparisons allowed identification of these differences (Table 7). The DFA and MANOVA results were supported by the significant correlations that the abundance of dominant nematode species showed with some environmental factors (Table 8). Species from the Utl (A1 and A2) were generally positively correlated with these factors, while species from the Ltl (A4 and A5) were negatively correlated. No significant correlation was found between species abundance and Md63.

#### Nematode feeding structure

In terms of relative abundance, DF were generally the dominant feeding group in all assemblages excepted for A3 and A4 where EF dominated (Figure 6). Facultative predators were almost absent in the Utl and were relatively abundant in the Mtl and the Ltl. In terms of relative biomass, the DF group was dominant in A1, A2, A3, and A5, while EF was dominant in A4. In A1, large P made of these feeding group the second dominant one (~30%) in terms of biomass although it represented only 5% in terms of abundance.



**Figure 6.** Relative abundance and biomass of the six nematode feeding groups (Moens & Vincx, 1997) in the five assemblages. M, microvores; CF, ciliate-feeders; DF, deposit-feeders; EF, epigrowth-feeders; FP, facultative-predators; P, predators.

## DISCUSSION

### *Comparison with other areas*

In our study area we found a meiofauna composition at higher taxonomic levels comparable to that reported for similar latitudes in the intertidal zones of the eastern Atlantic Ocean (between 45°–55° N). Nevertheless, densities were generally lower. Li & Vincx (1993) reported nematode densities between 1740 and 5328 ind 10 cm<sup>-2</sup> for the polyhaline (salinity=24.3–32.0 psu) sandy sediments of the Westerschelde (Belgium). Much higher densities

(annual average=12,400 ind 10 cm<sup>-2</sup>) were reported by Warwick & Price (1979) for the muddy sediment of the Lynher estuary (UK). However, density values in Li & Vincx refer to the sediment top 10 cm and in Warwick & Price to the top 6 cm. In our study, we sampled the top 2 cm of sediment where the meiofauna represented respectively more than 95% and 60% of the abundance integrated over a sediment depth of 10 cm, for muddy and sandy sediments, respectively. Nonetheless, by estimating densities for equivalent sampled depths we always found lower density values in our study area. Comparable densities to those we found were reported by Warwick & Gee (1984) for the muddy sediment of the Tamar estuary (UK) (nematode densities=500–1015 ind 10 cm<sup>-2</sup>). It is also worth noting that sampling for the present study was carried out in June, at the end of the spring season. No data are presently available for later months when benthic primary production (i.e. diatom blooms) significantly increases (unpublished data). This may very likely have positive effects on meiofaunal densities.

### *Nematode assemblages vs environmental factors*

Five different nematode assemblages were found in the study area: two at the upper (A1, A2), one at the mid (A3), and two at the lower (A4, A5) tide levels. The discriminant analysis showed that the Texp was the most discriminating function between nematode assemblages (i.e. lowest partial lambda). A greater Texp may be responsible for secondary effects such as wider temperature and salinity variations during low tide. In the Ltl, surface sediment temperature during the ice-cover-free season (May–October) may vary between 5–20°C, and in Utl between 1–38°C. Several authors stressed the importance of temperature as a structuring factor of meiofaunal assemblages with implications on the physiological adaptations required to inhabit specific environments (Hopper et al., 1973; Wieser et al., 1974; Wieser & Schiemer, 1977; Heip et al., 1985; Moens & Vincx, 2000a,b). Great salinity variations caused either by water evaporation (increase in salinity), especially in the warmer months, or freshwater percolation from the soil water table (decrease in salinity) may also demand particular physiological adaptations with consequences on the meiofauna species composition (e.g. Moens & Vincx, 2000a,b).

Chlorophyll-*a* and phaeopigments concentrations were found to be the second and third most important functions, respectively. The former tended to increase from Utl to Ltl, while the latter showed an opposite trend. This resulted in a gradually increasing P/C ratio (phaeopigments/chlorophyll-*a*) from Ltl to Utl. It is worth noting here that the abundance of EF and FP species dominant in the Ltl (*Metachromadora (M.)* sp., *Dichromadora hyalocheile*, and *Viscosia* sp.) were positively correlated with chlorophyll-*a* and negatively correlated with phaeopigments and P/C ratios. In contrast, EF species dominating in the Utl (*Metachromadora (M.) remanei*, *Microloaimus* sp. 1, Chromadoridae A) showed opposite correlation trends. This difference between EF of upper and lower tide levels was probably due to a different type of exploited phyto-detritus. In the upper-tide level, old or partially degraded phytodetritus may contribute more to the diet of EF, while in the low-tide level species of this feeding group

may rely more on 'fresher' phytodetritus. Vincx (1989, 1990) also reported significant correlations between the sediment concentration of chlorophyll-*a* and the meiofaunal abundance and community structure in the North Sea.

The sediment water content and silt percentage were the fourth and sixth most discriminant functions between nematode assemblages. Both of these factors give an indication of the sedimentary interstitial environment. More specifically, silt is responsible for the degree of sediment pore space filling, which determines the upper size limits of the interstitial species (Schwinghamer, 1981). In a study complementary to the present one Tita et al. (1999) investigated the nematode size spectra of assemblages A1, A3, and A5 in regard of sediment characteristics. This study showed that silt has a great importance in determining nematode lifestyle, i.e. interstitial vs burrowing, therefore significantly influencing their species composition. Tita et al. (1999) also showed that from A5 to A3 and A1 there was a gradual increase in average nematode body size (372, 424, 814 ng dw ind<sup>-1</sup>, respectively) and body width (29, 40, 70 µm, respectively). As a result of the larger body size, in assemblage A5 a lower individual respiration rate than in assemblage A3 and A1 was estimated (1.12, 1.25 and 2.26 nl O<sub>2</sub> h<sup>-1</sup>, respectively). The difference in organisms' body width can be interpreted as the dominance of burrowing lifestyle in the muddy Utl and of interstitial lifestyle in the sandy Ltl. Tita et al. (1999) found that the shift from one lifestyle to the other occurs around a body width of 32 µm. Moreover, consistently with Wieser (1959) and Coull (1988), they suggested that a median sediment grain diameter of 120 µm with a small silt fraction may virtually represent the lower granulometric limit allowing interstitial life.

#### Final considerations

This study, together with Tita et al. (1999), represents the first set of data available for the meiofaunal communities in the St Lawrence estuary. However, further studies are needed in order to estimate the local meiofaunal production, and to better understand the interactions between meiofauna and macrofauna. Moreover, a more extensive sampling of the intertidal and subtidal zones of the St Lawrence estuary should be carried out. This would provide a better general picture of meiofauna distribution and diversity in this large estuary.

This study was supported by a research grant (no. CG0003540) from the Natural Sciences and Engineering Research Council of Canada (NSERC). The first author acknowledges scholarship assistance from the Fondation de l'Université du Québec à Rimouski (FUQAR).

#### REFERENCES

Austen, M.C., 1989. Factors affecting estuarine meiobenthos assemblage structure: a multifactorial microcosm experiment. *Journal of Experimental Marine Biology and Ecology*, **130**, 167–187.  
Austen, M.C. & Warwick, R.M., 1989. Comparison of univariate and multivariate aspects of estuarine meiobenthic community structure. *Estuarine, Coastal and Shelf Science*, **29**, 23–42.

Caron, A., 1995. *Étude du partage des ressources par deux annélides polychètes, Nereis virens (Sars) et Nephtys caeca (Fabricius), dans les sédiments intertidaux de la rive sud de l'estuaire maritime du saint-Laurent*. PhD thesis, Université du Québec à Rimouski, Canada.  
Caron, A., Desrosiers, G., Miron, G. & Retière, C., 1995. Comparison of spatial overlap between the polychaetes *Nereis virens* (Sars) and *Nephtys caeca* (Fabricius) in two intertidal estuarine environments. *Marine Biology*, **124**, 537–550.  
Caron, A., Desrosiers, G., Retière, C. & Hudier, E., 1993a. Comparaison démographique des populations de deux annélides polychètes selon l'orientation des baies dans l'estuaire du Saint-Laurent. *Oceanologica Acta*, **16**, 403–412.  
Caron, A., Olivier, M., Desrosiers, G., Hudier, E., Côté, E., Koutitonsky, V., Miron, G. & Retière, C., 1993b. Distribution spatiale d'une espèce benthique épitoque en zone intertidale: Rôle de l'hydrodynamisme? *Vie et Milieu*, **43**, 85–93.  
Coull, B.C., 1985. Long-term variability of estuarine meiobenthos—an 11 year study. *Marine Ecology Progress Series*, **24**, 205–218.  
Coull, B.C., 1988. Ecology of the marine meiofauna. In *Introduction to the study of meiofauna* (ed. R.P. Higgins and H. Thiel), pp. 18–38. London, Washington DC: Smithsonian Institution Press.  
Desrosiers, G. & Brêthes, J.-C., 1984. Étude de la communauté à *Macoma balthica* de la batture de Rimouski. *Sciences et Techniques de l'Eau*, **17**, 25–30.  
Desrosiers, G., Brêthes, J.-C., Culombe, F. & Jacques, A., 1980. *Étude bionomique de l'endofaune benthique de substrat meuble de St-Fabien-sur-Mer (Québec)*. Université du Québec à Rimouski, Département d'Océanographie, Cahier d'information, no. 6, pp. 42.  
Desrosiers, G., Brêthes, J.-C. & Long, B.F., 1984. L'effet d'un glissement de terrain sur une communauté benthique médio-littorale du nord du Golfe du St-Laurent. *Oceanologica Acta*, **7**, 251–258.  
El-Sabh, M. & Murty, T.S., 1990. Mathematical modelling of tides in the St Lawrence Estuary. In *Oceanography of a large-scale estuarine system: the St Lawrence* (ed. M. El-Sabh and N. Silverberg), pp. 10–50. New York: Springer-Verlag. [Coastal and Estuarine Studies, no. 39.]  
Flint, R.W. & Kalke, R.D., 1986. Biological enhancement of estuarine benthic community structure. *Marine Ecology Progress Series*, **31**, 23–33.  
Folk, R.L., 1974. *Petrology of sedimentary rocks*, p. 182. Austin, TX: Hemphill Publishing Company.  
Goodman, L.A. & Kruskal, W.H., 1954. Measures of association for cross classifications. *Journal of the American Statistical Association*, **49**, 732–764.  
Hall, J.A. & Frid, C.L.J., 1997. Estuarine sediment remediation: effects on benthic biodiversity. *Estuarine, Coastal and Shelf Science*, **44**, Supplement, 55–61.  
Heip, C., Vincx, M. & Vranken, G., 1985. The ecology of marine nematodes. *Oceanography and Marine Biology. Annual Review*, **23**, 399–489.  
Hicks, G.R.F. & Coull, B.C., 1983. The ecology of marine meiobenthic harpacticoid copepods. *Oceanography and Marine Biology. Annual Review*, **21**, 67–175.  
Hopper, B.E., 1969. Marine nematodes of Canada. II. Marine nematodes from the Minas Basin–Scots Bay area of the Bay of Fundy, Nova Scotia. *Canadian Journal of Zoology*, **47**, 671–690.  
Hopper B.E., Fell, J.W. & Cefalu, R.C., 1973. Effect of temperature on life cycles of nematodes associated with the mangrove (*Rhizophora mangle*) detrital system. *Marine Biology*, **72**, 293–296.  
Ingram, R.G. & El-Sabh, M., 1990. Fronts and mesoscale features in the St Lawrence estuary. In *Oceanography of a large-scale estuarine system: the St Lawrence* (ed. M. El-Sabh and N. Silverberg), pp. 71–93. New York: Springer-Verlag. [Coastal and Estuarine Studies, no. 39.]  
Jensen, P., 1984. Measuring carbon content in nematodes. *Helgoländer Meeresuntersuchungen*, **38**, 83–86.

- Li, J. & Vincx, M., 1993. The temporal distribution of intertidal nematodes in the westerschelde. I. The importance of an estuarine gradient. *Netherlands Journal of Sea Research*, **27**, 319–326.
- Lorenzen, S., 1994. *The phylogenetic systematics of freeliving nematodes*. London: The Ray Society.
- Luczak, C., Janquin, M.-A. & Kupka, A., 1997. Simple standard procedure for the routine determination of organic matter in marine sediment. *Hydrobiologia*, **345**, 87–94.
- McIntyre, A.D. & Warwick, R.M., 1984. Meiofauna techniques. In *Methods for the study of marine benthos* (ed. N.A. Holme and A.D. McIntyre), pp. 217–244. Oxford: Blackwell Science. [IBP Handbook, no. 16.]
- Miron, G.Y. & Desrosiers, G., 1990. Distribution and population structures of two estuarine polychaetes in the lower St Lawrence Estuary, with special reference to environmental factors. *Marine Biology*, **105**, 297–306.
- Moens, T. & Vincx, M., 1997. Observations on the feeding ecology of estuarine nematodes. *Journal of the Marine Biological Association of the United Kingdom*, **77**, 211–227.
- Moens, T. & Vincx, M., 2000a. Temperature and salinity constraints on the life cycle of two brackish-water nematode species. *Journal of Experimental Marine Biology and Ecology*, **243**, 115–135.
- Moens, T. & Vincx, M., 2000b. Temperature, salinity and food thresholds in two brackish-water bacterivorous nematode species: assessing niches from food absorption and respiration experiments. *Marine Ecology Progress Series*, **53**, 137–154.
- Olivier, M., Desrosiers, G., Retière, C. & Brêthes, J.-C., 1993. Variations spatio-temporelles de l'alimentation chez le polychète *Nereis virens* (Sars) en zone intertidale (Estuaire maritime du Saint-Laurent, Québec). *Vie et Milieu*, **43**, 1–12.
- Parsons, T.R., Maita, Y. & Lalli, C.M., 1984. *A manual of chemical and biological methods for seawater analysis*. Oxford: Pergamon Press.
- Platt, H.M. & Warwick, R.M., 1983. *Freeliving marine nematodes*. Part 1. *British enoplids*. Pictorial key to world genera and notes for the identification of British species. Cambridge: Cambridge University Press. [Synopsis of the British Fauna, no. 28.]
- Platt, H.M. & Warwick, R.M., 1988. *Freeliving marine nematodes*. Part 2. *British chromadorids*. Pictorial key to world genera and notes for the identification of British species. Cambridge: Cambridge University Press. [Synopsis of the British Fauna, no. 38.]
- Schwinghamer P. 1981. Characteristic size distributions of integral benthic communities. *Canadian Journal of Fisheries and Aquatic Sciences*, **38**, 1255–1263.
- Tietjen, J.H., 1969. The ecology of shallow water meiobenthos in two New England estuaries. *Oecologia*, **2**, 251–291.
- Tita, G., Desrosiers G. & Vincx M., 1999. Size spectra, body width and morphotypes of intertidal nematodes: an ecological interpretation. *Journal of the Marine Biological Association of the United Kingdom*, **79**, 1007–1015.
- Tita, G., Desrosiers G. & Vincx M., 2000a. New type of hand-held corer for meiofaunal sampling and vertical profile investigation: a comparative study. *Journal of the Marine Biological Association of the United Kingdom*, **80**, 171–172.
- Tita, G., Desrosiers, G., Vincx, M. & Nozais, C., 2000b. Sediment disturbance and predation of the intertidal polychaete *Nereis virens* (Sars) on the associated meiofaunal assemblages. *Journal of Experimental Marine Biology and Ecology*, **243**, 261–282.
- Vincent, B., Brassard, C. & Harvey, M., 1987. Variations de la croissance de la coquille et de la structure d'âge du bivalve *Macoma balthica* (L.) dans une population intertidale de l'estuaire du Saint-Laurent (Québec). *Canadian Journal of Zoology*, **65**, 1906–1916.
- Vincx, M., 1989. Free-living marine nematodes. *AWLSK, Klasse der Wetenschappen, Jargang*, **51**, 37–70.
- Vincx, M., 1990. Diversity of the nematode communities in the southern bight of the North Sea. *Netherlands Journal of Sea Research*, **25**, 181–188.
- Warwick, R.M. & Gee, J.M., 1984. Community structure of estuarine meiobenthos. *Marine Ecology Progress Series*, **18**, 97–111.
- Warwick, R.M. & Price, R., 1979. Ecological and metabolic studies on free-living nematodes from an estuarine mud-flat. *Estuarine and Coastal Marine Science*, **9**, 257–271.
- Wieser, W., 1953. Die Beziehung zwischen Mundhöhlengestalt, Ernährungsweise und Vorkommen bei freilebenden marinen Nematoden. *Arkiv für Zoologie*, **2**, 439–484.
- Wieser, W., 1959. The effect of the grain size on the distribution of small invertebrates inhabiting the beaches of Puget Sound. *Limnology and Oceanography*, **4**, 181–194.
- Wieser, W., 1960. Benthic studies in Buzzards Bay. II. The meiofauna. *Limnology and Oceanography*, **5**, 121–153.
- Wieser, W., Ott, J., Schiemer, F. & Gnaiger, E., 1974. An ecophysiological study of some meiofauna species inhabiting a sandy beach at Bermuda. *Marine Biology*, **26**, 235–248.
- Wieser, W. & Schiemer, F., 1977. The ecophysiology of some marine nematodes from Bermuda: seasonal aspects. *Journal of Experimental Biology*, **26**, 97–106.

Submitted 9 March 2001. Accepted 23 July 2002.



**Appendix 1.** List of nematode species recorded in the studied area with their respective feeding groups after Wieser (1953) (*W*) and Moens & Vincx (1997) (*M&V*). 1A, Selective deposit-feeders; 1B, non-selective deposit-feeders; 2A, epigrowth-feeders; 2B, omnivore-carnivores; M, microvores; DF, deposit-feeders; CF, ciliate-feeders; EF, epigrowth-feeders; FP, facultative-predators; P, predators.

Species	W	M&V	Species	W	M&V
<i>Adoncholaimus fuscus</i>	2B	FP	<i>Hypodontolaimus schuurmansstekoveni</i>	2A	EF
Aegialoalaimidae	1A	M	<i>Innocuonema</i> sp.	2A	EF
<i>Aegialoalaimus</i> sp.	1A	M	<i>Karkinochromadora</i> sp.	2A	EF
<i>Amphimonhystrella</i> sp.	1B	DF	<i>Leptolaimus elegans</i>	1A	M
<i>Anoplostoma blanchardi</i>	1B	CF	<i>Leptolaimus papilliger</i>	1A	M
<i>Antomicron</i> sp.	1A	M	Linhomoeidae	1A	M
<i>Aponema</i> sp.	2A	EF	<i>Linhomoeus</i> sp.	1A	M
<i>Ascolaimus</i> sp.	1B	CF	<i>Metachromadora (Metachromadoroides) remanei</i>	2A	EF
<i>Axonolaimus</i> sp. 1	1B	CF	<i>Metachromadora (Metachromadoroides) sp.</i>	2A	EF
<i>Axonolaimus</i> sp. 2	1B	CF	<i>Microlaimus</i> sp. 1	2A	EF
<i>Bathylaimus</i> sp.	1B	CF	<i>Microlaimus</i> sp. 2	2A	EF
<i>Bolbolaimus</i> sp. 1	2B	FP	<i>Monhystera</i> sp. 1	1B	DF
<i>Bolbolaimus</i> sp. 2	2B	FP	<i>Monhystera</i> sp. 2	1B	DF
<i>Camacolaimus</i> sp. 1	2A	EF	<i>Monoposthia costata</i>	2A	EF
<i>Camacolaimus</i> sp. 2	2A	EF	<i>Nannolaimoides effilatus</i>	2A	EF
<i>Camacolaimus tardus</i>	2A	EF	<i>Neochromadora poecilosoma</i>	2A	EF
<i>Chaetonema</i> sp.	1B	CF	<i>Neochromadora</i> sp.	2A	EF
<i>Chromadora macrolaima</i>	2A	EF	<i>Odontophora</i> sp.	1B	CF
<i>Chromadora</i> sp. 1	2A	EF	<i>Oncholaimus</i> sp.	2B	FP
Chromadoridae A	2A	EF	<i>Oxystomina</i> sp.	1A	M
Chromadoridae B	2A	EF	<i>Paracanthonchus caecus</i>	2A	EF
<i>Chromadorita</i> sp.	2A	EF	<i>Paracanthonchus</i> sp.	2A	EF
Comesomatidae	1B	DF	<i>Paradesmodora</i> sp.	2A	EF
<i>Cyartonema</i> sp.	1A	M	<i>Paralinhomoeus</i> sp.	1A	M
<i>Daptonema</i> sp. 1	1B	DF	<i>Paramonohystera</i> sp. 1	1A	M
<i>Daptonema</i> sp. 2	1B	DF	<i>Paramonohystera</i> sp. 2	1A	M
<i>Daptonema</i> sp. 3	1B	DF	<i>Paramonohystera</i> sp. 3	1A	M
<i>Daptonema tenuispiculum</i>	1B	DF	<i>Polygastrophora</i> sp.	2B	FP
<i>Desmodora</i> sp.	2A	EF	<i>Pomponema sedecima</i>	2B	FP
Desmodoridae	2A	EF	<i>Ptycholaimellus ponticus</i>	2A	EF
<i>Desmolaimus</i> sp.	1B	DF	<i>Richtersia inaequalis</i>	1B	DF
<i>Desmoscolex falcatus</i>	1A	M	<i>Sabatieria longispinosa</i>	1B	DF
<i>Dichromadora hyalocheile</i>	2A	EF	<i>Sabatieria ornata</i>	1B	DF
<i>Dichromadora</i> sp.	2A	EF	<i>Sabatieria punctata</i>	1B	DF
<i>Diplolaimella</i> sp.	1B	DF	<i>Siphonolaimus</i> sp.	2B	FP
<i>Diplopetloides</i> sp.	1A	M	<i>Southernia</i> sp.	1A	M
<i>Doliolaimus</i> sp. 1	1B	CF	Sphaerolaimidae	1B	CF
<i>Doliolaimus</i> sp. 2	1B	CF	<i>Sphaerolaimus</i> sp. 1	2B	P
<i>Eleutherolaimus</i> sp.	1B	DF	<i>Sphaerolaimus</i> sp. 2	2B	P
<i>Elzalia</i> sp.	1B	CF	<i>Spilophorella</i> sp.	2A	EF
Enchelidiidae	2B	P	<i>Spirinia</i> sp.	2A	EF
<i>Enoplolaimus</i> sp. 1	2B	P	<i>Stephanolaimus</i> sp.	1A	M
<i>Enoplolaimus</i> sp. 2	2B	P	<i>Symplocostoma</i> sp.	2B	FP
<i>Enoplus</i> sp.	2B	P	<i>Terschellingia</i> sp.	1A	M
<i>Epacanthion</i> sp.	2B	P	<i>Theristus (Daptonema) procerus</i>	1B	DF
<i>Gammanema</i> sp.	2B	FP	<i>Theristus</i> sp.	1B	DF
<i>Gnomoxyala</i> sp.	1B	CF	<i>Trefusia</i> sp.	1A	M
<i>Halalaimus</i> sp. 1	1A	M	<i>Tripyloides</i> sp. 1	1B	CF
<i>Halalaimus</i> sp. 2	1A	M	<i>Tripyloides</i> sp. 2	1B	CF
<i>Halanonchus</i> sp.	1B	CF	Tripyloididae	1B	CF
<i>Halichoanolaimus robustus</i>	2B	P	<i>Viscosia</i> sp.	2B	FP
<i>Hypodontolaimus balticus</i>	2A	EF	Xyalidae A	1B	DF
<i>Hypodontolaimus inaequalis</i>	2A	EF	Xyalidae B	1B	DF